



**SUBMISSION TO THE
NUNAVUT WILDLIFE MANAGEMENT BOARD**

FOR

Information:

Decision: X

Issue: Increase in Total Allowable Harvest of muskoxen on Devon Island as a result of new survey results and increase in population assessment (Muskox Management Unit MX-04)

Background:

Devon Island has been sporadically surveyed for Peary caribou and muskoxen since 1961, usually with a focus on the productive lowland areas where muskoxen congregate. We flew a survey of Devon Island, including Philpots Island, by Twin Otter in 58 hours between March 22 and 30, 2016, to update the population estimate for caribou and muskoxen in the study area.

Survey observers were from Resolute and Grise Fiord, with 2 observers on each side of the aircraft.

During the 2016 Devon Island survey, observers recorded 14 caribou, but previous surveys in 2008 and 2001 also recorded low numbers (17 in 2008 and 37 in 2002). Hunters in Grise Fiord have not noticed an increase or decrease, and it is likely that caribou persist at low densities on the island.

Muskoxen, however, appear to have increased dramatically in abundance since the previous survey (2008), which estimated 513 muskoxen (302-864, 95%CI). The March 2016 survey found the highest reported abundance estimate for muskoxen, $1963 \pm SE343$.

This survey indicates a large increase in muskoxen on Devon Island, with more observations in all lowland areas compared to 2008, and a particular increase on Philpots Island. This population trend is mirrored on neighboring Bathurst Island to the west, surveyed in 2013, and southern Ellesmere Island to the north, surveyed in 2015.

Current Status:

There is currently a Total Allowable Harvest of 15 muskoxen for Devon Island, with tags allocated to Grise Fiord (4), Arctic Bay (4), and Resolute Bay (7). Grise Fiord generally fills Devon Island tags, although the other communities have not been harvesting muskoxen on Devon Island recently.

A survey report is being finalized from the aerial survey of Devon Island completed in March 2016 for Peary caribou and muskoxen.

Consultations:

Consultations with the Iviq HTA (Grise Fiord) took place in the community on July 18, 2016. The results of the survey were discussed and options moving forward (removal of TAH or increasing the available tags) were presented for the Board to further consider. Although there was some concern that removing the TAH altogether could result in less coordination and control of the harvest, the Board will be meeting to provide guidance on the best way forward. The options were also sent by email to the manager of the Resolute Bay HTA in early July 2016, and in-person consultation is expected in late July/early August. We will also attempt to arrange an in-person meeting with Arctic Bay to discuss survey results and management options in mid/late August, 2016.

Recommendations:

Option 1: Increase the TAH from 15 to 100, for an approximate 5% harvest rate on the estimated population. Maintaining the TAH would likely be the most reliable way to maintain harvest reporting and coordinate hunts among communities.

Option 2: Alternatively, the TAH could be removed altogether, provided the communities communicate and coordinate harvest activities on Devon Island. It is highly encouraged that harvest reporting be maintained to track the effect of changes in management actions and distribute harvest among productive lowland areas.



**DISTRIBUTION AND ABUNDANCE OF PEARY CARIBOU (*Rangifer tarandus pearyi*)
AND MUSKOXEN (*Ovibos moschatus*) ON DEVON ISLAND, MARCH 2016**

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Anderson, M. 2016. Distribution and abundance of Peary caribou (*Rangifer tarandus pearyii*) and muskoxen (*Ovibos moschatus*) on Devon Island, March 2016. Nunavut Department of Environment, Wildlife Research Section, Status Report 2016-01, Igloolik, NU.

Summary

We flew a survey of Devon Island including Philpots Island (Muskox Management Zone MX-04), by Twin Otter in 58 hours between March 22 and 30, 2016, to update the population estimate for caribou and muskoxen in the study area. The previous survey, in 2008, reported a minimum count of 17 Peary caribou and population estimate of 513 muskoxen (302-864, 95%CI). The 2016 survey found the highest reported abundance estimate for muskoxen (1,963 \pm 343 SE), and a minimum count of 14 Peary caribou suggests that they continue to persist at low densities on the island, although the low number of observations precludes calculation of a reliable population estimate.

Muskoxen were abundant in the coastal lowlands where they have been found historically, at Baring bay, Croker Bay, Dundas Harbour, and the Truelove Lowlands. They were also abundant on the north coast of the Grinnell Peninsula, and particularly abundant on Philpots Island, where we observed 310 muskoxen. Although most previous surveys covered only part of Devon Island, they did target these lowlands and their abundance estimates or minimum counts likely represent the majority of the muskox population. This survey indicates a large increase in muskoxen on Devon Island, with more observations in all lowland areas compared to 2008, and a particular increase on Philpots Island. This population trend is mirrored on neighboring Bathurst Island to the west, surveyed in 2013, and southern Ellesmere Island to the north, surveyed in 2015.

We only saw 14 Peary caribou during the survey, concentrated on the north shore of the Grinnell Peninsula, and tracks were seen south of Baring Bay. No caribou were seen in the Truelove Lowlands, although hunters from Grise Fiord have caught caribou there over the past several years. It is likely that the low density and patchy distribution of caribou in this area meant that they were not detected on the survey flights. Previous surveys also found caribou in small numbers in specific locations, including a minimum count of 17 caribou in 2008 and 37 caribou on western Devon Island in 2002. Combined with the local knowledge of residents of Grise Fiord and Resolute Bay, it is likely that this population of Peary caribou remains stable at low densities, patchily distributed on Devon Island.

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Introduction

Peary caribou (*Rangifer tarandus pearyi*) are a small, light-coloured subspecies of caribou/reindeer inhabiting the Canadian Arctic Archipelago in the Northwest Territories and Nunavut from the Boothia Peninsula in the south to Ellesmere Island in the north. They are sympatric with muskoxen (*Ovibos moschatus*) over much of their range although diet, habitat preferences, and potentially interspecific interactions separate the two species at a finer scale (Resolute Bay Hunters and Trappers Association [HTA] and Iviq HTA, pers. comm.). Arctic wolves (*Canis lupus arctos*) occur at low densities throughout Peary caribou range, but the most significant cause of population-wide mortality appears to be irregular die-offs precipitated by severe winter weather and ground-fast ice that restricts access to forage (Miller et al 1975, Miller and Gunn 2003, Miller and Barry 2009).

Peary caribou have been surveyed infrequently and irregularly on the Canadian Arctic Archipelago since Tener's 1961 survey, which provided a best guess estimate of 150 Peary caribou on Devon Island, although persistent fog prevented the Colin Archer Peninsula from being surveyed (Tener 1963). Since Tener's survey, unsystematic surveys have been conducted irregularly, usually with a focus on muskoxen in the lowland areas where they are concentrated. In 2002, the western Devon Island was surveyed as part of a program to update population estimates for Peary caribou across their range, and a minimum count of 37 was recorded (Jenkins et al. 2011). The entire island was surveyed in 2008, with a minimum count of 17 caribou (Jenkins et al. 2011). Residents of Grise Fiord and Resolute Bay have not noticed a marked increase or decline in caribou on Devon Island (Iviq HTA, pers comm.), but with higher caribou populations to the west on the Bathurst Island Complex, residents of Resolute were interested in whether caribou have moved onto northern or western Devon Island. Grise Fiord hunters regularly travel the Truelove Lowlands and catch caribou there. Community members were interested in the abundance and distribution of caribou in that area as well as in other areas where the caribou potentially move to.

Population estimates for muskoxen on Devon Island have mostly been estimated based on their abundance in discrete lowland habitat patches. In 1961, Tener surveyed the entire island (except the Colin Archer Peninsula, due to fog) at 6% coverage, and estimated that the population was about 200 muskoxen (Tener 1963). Subsequent surveys focused on the lowland areas where muskoxen could be reliably located. The overall population of muskoxen was believed to be around 300-400 through the 1970s to 1990s (Freeman 1971, Hubert 1977, Decker in Urquhart 1982, Pattie 1990, Case 1992), reaching 513 (302-864 95%CI) by 2008 (Jenkins et al 2011). This was also the first systematic survey of the entire island, although much of Devon Island is unsuitable habitat and it is unlikely that the unsystematic surveys of lowlands missed large numbers of muskoxen. Muskoxen were located consistently in the lowlands around Baring Bay, Maxwell Bay, Dundas Harbour, Philpots Island, Truelove Inlet, Sverdrup Inlet, and the northeast shores of Grinnell Peninsula.

The Peary caribou and muskoxen of western and northern Devon Island are important to the communities of Resolute Bay and Grise Fiord. Arctic Bay hunters also access the southern shores of Devon Island, and with the decline in Baffin Island caribou, Devon Island might become more important in the harvest activities of Arctic Bay. Muskoxen have been hunted in the area since the government ban on muskox hunting was lifted in 1969. As species of presumption of need, subsistence tags are currently set aside and allocated for subsistence, commercial use, and sport hunts according to the allocation of Regional Wildlife Organization (RWO) and Hunter and Trapper Organizations/Associations (HTOs/HTAs). Caribou have been regularly hunted in the region since the communities of Resolute Bay and Grise Fiord were established in the 1950s, although parts of Devon Island have been important harvest areas for centuries. This survey was conducted to update the population estimates, demographic characteristics, and distribution of Peary caribou and muskoxen on Devon Island.

Study Area

The survey area is predominantly polar desert and semi desert, with rugged topography along the mountains and fiords of the south and east coasts, which rise from sea level to 700 m, transitioning to rolling terrain dissected by deep river valleys in the interior and on the Grinnell Peninsula. The island is dominated by the 14, 590 km² Devon Ice Cap, rising to 1800 m AMSL in the center, which is also the highest point on the island. Several smaller glaciers are scattered along the south coast, Grinnell Peninsula, and Colin Archer Peninsula. Cushion forb barrens or cryptogam-herb barrens dominate the island, usually at <5% cover and <100 g/m² biomass, with isolated patches of prostrate dwarf shrub and prostrate dwarf shrub/graminoid tundra in the coastal lowlands, where vegetation cover increases to 5-50% and biomass increases to 100-500 g/m² (Gould et al. 2003, Walker et al. 2005).

Mean July temperatures are 3-5°C on the west side of the study area and 5-7°C in the east (Gould et al. 2003 and references therein). In March 2016, the average daily low and high temperatures in Resolute were -32.2°C and -26.1°C; in Grise Fiord, average daily low temperatures were -32.4°C and average daily high temperatures were -25.6°C (Environment Canada weather data, available http://climate.weather.gc.ca/index_e.html). Most of the study area was snow-covered, although some valleys, particularly along the northeast coast, were largely windswept. There was 26-29 cm snow recorded on the ground at Resolute in March 2016 and 4.3 mm of precipitation, compared to 0-5 cm of snow on the ground in Grise Fiord and 5.1 mm of precipitation (Environment Canada weather data).

The March 2016 aerial survey was flown to cover the same study area as the previous 2008 survey (Jenkins et al. 2011), excluding North Kent Island and Bailie Hamilton Island. We stratified the study area to allocate more effort to good habitat where caribou or muskoxen had previously been reported with a 5-km transect spacing and areas with moderate habitat that might have wildlife were surveyed with a 10-km spacing. We flew transects spaced 15 km apart over barren parts of the island that were unlikely to be occupied by caribou or muskoxen, but where animals could be travelling between suitable habitat patches (Figure1).

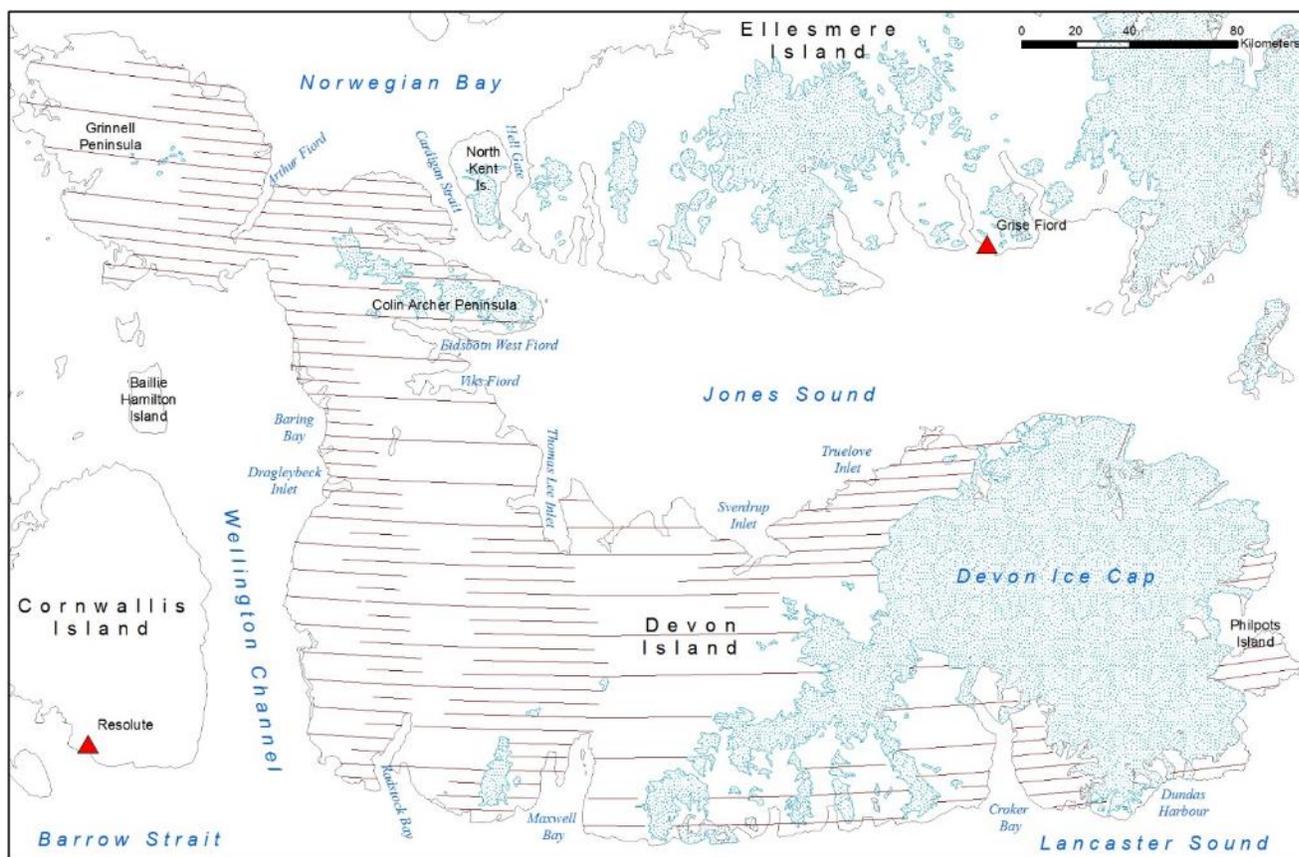


Figure 1. Major landmarks of the study area, with glaciers in stippled blue and 2016 transect lines in dark red running east-west.

Methods

Aerial Survey

Survey transects ($n=166$, Appendix 1) followed the transects established for the 2008 distance sampling helicopter survey, parallel to lines of latitude with 5, 10, or 15 km spacing and a 500 m strip on either side of the aircraft. Ice caps were excluded, and we did not detect any caribou, muskoxen, or their tracks on any ice caps during ferry flights. We stratified the study area to maximize survey effort in areas expected to have caribou or muskoxen, since much of Devon Island is barren gravel and till, unlikely to support wildlife. The high density (A) stratum was flown with transects spaced 5 km apart, the intermediate stratum (B) flown at 10 km spacing, and the low density stratum (C) was flown at 15 km spacing. Strata and transects are shown in Figure 2 and Table 1. Data used for delineation of the strata is provided in Appendix 2.

Table 1. Survey strata for Devon Island, March 22-30 2016.

Block ID	Stratum	Strata Area, Z (km ²)	Transect Spacing (km)	Transects Surveyed	Survey Area, z (km ²)	Sampling Fraction, f (%)
A	High Density	18438	5	117	3388	18.4%
B	Medium Density	6360	10	21	581	9.1%
C	Low Density	15076	15	28	1024	6.8%

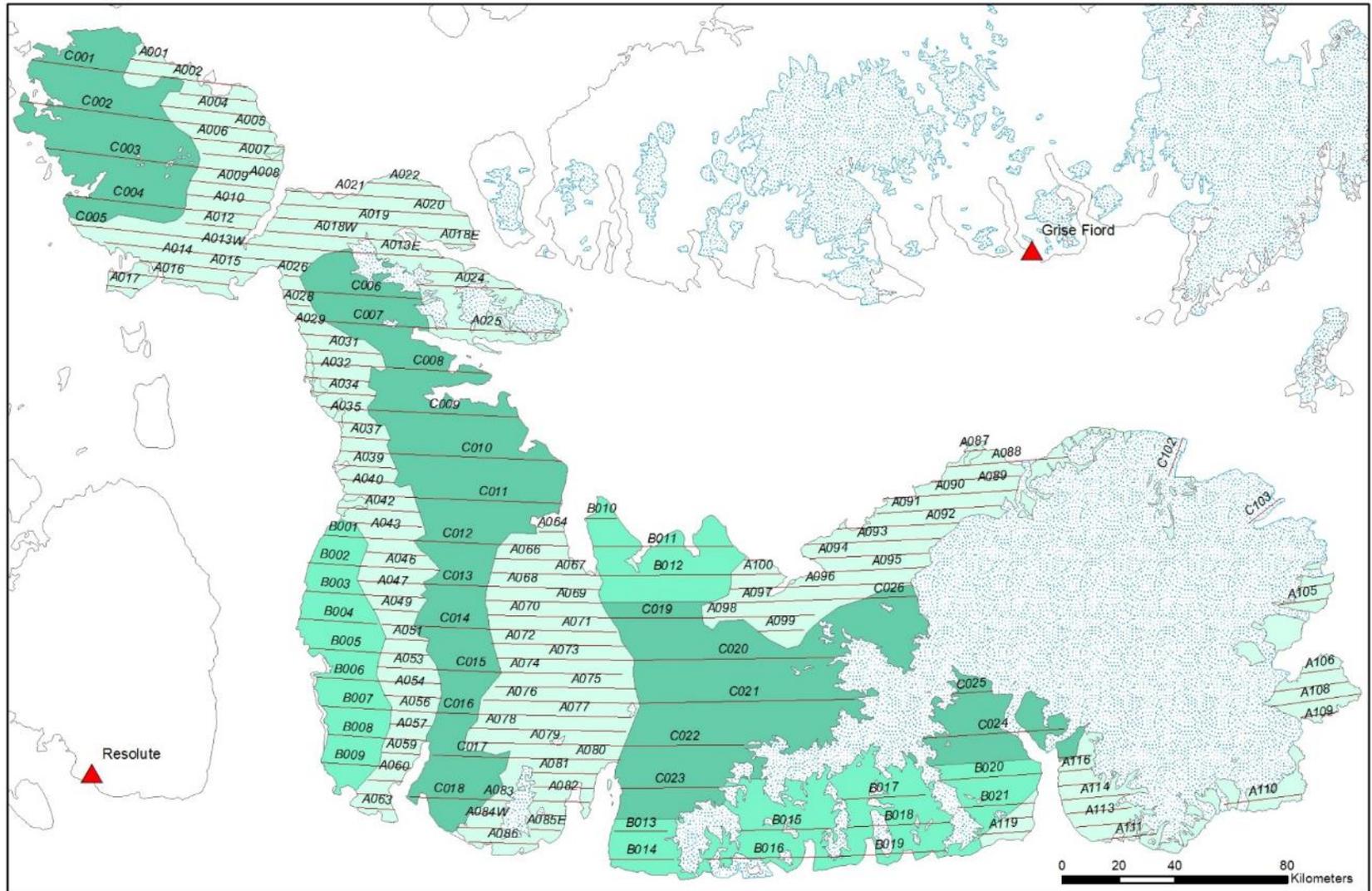


Figure 2. Transects and survey strata for Devon Island, March 22-30, 2016. A transects are the high density stratum flown with transects 5 km apart (pale green), B transects are the intermediate density stratum, flown with transects 10 km apart (bright green), and C transects are the low density stratum, flown with transects 15 km apart (dark green).

To define the transect width, we marked survey aircraft wing struts following Norton-Griffiths (1978):

$$w = W \left(\frac{h}{H} \right)$$

where W is the strip width, H is the flight height, h is the observer height when the plane is on the ground and w is calculated, measured and marked on the ground to position wing strut marks (Figure 3). For this survey we only used one mark representing 500 m marked on the wing strut. Fixed-wing strip transect sampling has been successfully used in the high arctic since 1961, and can be useful when observations are insufficient to determine the effective strip width required for distance sampling.

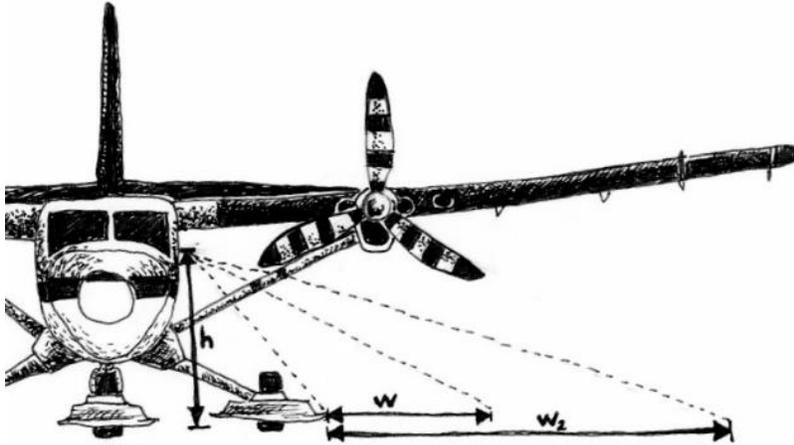


Figure 3. Derivation of wing strut marks for strip boundaries, where w and w_2 are calculated as described in the text, h is measured (2.2 m for Twin Otter on wheel-skis), and dotted lines indicate observer sightlines as modified from Norton-Griffiths (1978).

Transects were flown between 160-220 km/hr with a DeHavilland Twin Otter – higher speeds were used for uniform, snow-covered landscapes where visibility was excellent. Surveys were only conducted on good visibility days to facilitate detection of animals, tracks, and feeding craters, as well as for operational reasons to ensure crew safety. Flight height was set at 152 m (500 ft) using a radar altimeter. In rugged terrain, the flight height was adhered to as closely as possible within the constraints of crew safety and aircraft abilities.

A Twin Otter with 4-6 passengers (2 front observers, 2-4 rear observers, one of whom was also data recorder) was used to follow the double-observer methodology, which has been successful in other muskox and caribou surveys in Nunavut (see Campbell et al. 2012 for an overview of the methodology) and specifically in the High Arctic on Bathurst and Ellesmere islands (Anderson 2014, Anderson and Kingsley 2015). Front and rear observers on the same side of the plane were able to communicate and all observations by front and rear observers were combined. Estimates of group size are a potentially large source of error in calculating population estimates. However, Peary caribou and muskoxen are generally distributed in relatively small groups where observer fatigue is likely to be a more important source of error (A. Gunn, pers. comm.). We found obvious benefits of using the platform where having the added observers not only increased the accuracy of age and sex classification, but also allowed some crew members to classify with binoculars while others continued to scan for nearby groups and individuals.

All observations of wildlife and tracks were marked on a handheld Garmin Montana 650 global positioning system (GPS) unit, which also recorded the flight path every 15 seconds. Sex and age classification was limited, since the aircraft did not make multiple passes (to minimize disturbance), but adult/short yearling

(calves from the previous spring, i.e. 10-11 months old) determination was often straightforward for muskoxen and aided by binoculars. Muskoxen were frequently spotted more than a kilometer off transect due to their large aggregations and dark colour in contrast to the snowy background. Depending on distance and topography, an accurate count could not always be determined for these groups. Newborn muskoxen were obvious based on size, but their small size and close association with other animals in the herd made them difficult to count in larger groups or when muskoxen were tightly grouped. GPS tracks and waypoints were downloaded through DNR-GPS and saved in Garmin GPS eXchange Format and as ESRI shapefiles. Data was entered and manipulated in Microsoft Excel and ArcMAP (ESRI, Redlands, CA).

Analysis

Flights linking consecutive transects were removed for population analysis, although survey speed and height were maintained and all observations recorded as if on survey. Similarly, sections of transect crossing sea ice and ice fields were removed, as these areas were not included in the area used for density calculations.

Although Jolly's (1969) Method II is widely used for population estimates from surveys, it is designed for a simple random design, rather than for a systematic survey of a patchy population. For comparison, population calculations following Jolly's Method II are provided in Appendix 4, along with calculations following a systematic stratified survey design (Cochran 1977). The muskoxen and caribou detected in this survey were patchily distributed and serially correlated, not randomly distributed. For systematic samples from serially correlated populations, estimates of uncertainty based on deviations from the sample mean are expected to be upwardly biased and influenced by the degree of serial correlation; high serial correlation implies that there is less random variation in the unsurveyed sections between systematically spaced transects than if serial correlation were low (Cochran 1977). Calculating uncertainty based on nearest-neighbor differences incorporates serial correlation, and the upward bias in the uncertainty is expected to be less than if it were calculated based on deviations from the sample mean. Nearest-neighbor methods have been used previously to calculate variance around survey estimates on the unweighted ratio estimate (Kingsley et al. 1981, Stirling et al. 1982, Kingsley et al. 1985, Anderson and Kingsley 2015).

The model for observations on a transect survey following Cochran (1977) is:

$$y_i = Rz_i + \varepsilon_i\sqrt{z_i}$$

Where y_i is the number of observations on transect i of area z_i , R is the mean density and error terms ε_i are independently and identically distributed. In this model, the variance of the error term is proportional to the area surveyed. The best estimate of the mean density \hat{R} is:

$$\hat{R} = \frac{\sum_i y_i}{\sum_i z_i}$$

The error sum of squares, based on deviations from the sample mean, is given by:

$$\left(\sum_i \frac{y_i^2}{z_i} \right) - \frac{(\sum_i y_i)^2}{\sum_i z_i}$$

The finite-population corrected error variance of \hat{R} is:

$$Var(\hat{R}) = \frac{(1-f)}{(n-1)\sum_i z_i} \left(\left(\sum_i \frac{y_i^2}{z_i} \right) - \frac{(\sum_i y_i)^2}{\sum_i z_i} \right)$$

Where f is the sampling fraction and n is the number of transects. The sampling fraction also provides the scaling factor for moving from a ratio (population density) to a population estimate. It is calculated as $(\sum z_i)/Z$, where Z is the study area and $\sum z_i$ is the area surveyed. The irregular study area boundaries mean that f varies from the 20% sampling fraction expected from a 1-km survey strip and 5-km transect spacing.

If we were to apply a model $y_i = Rz_i + \varepsilon_i$ instead, then the variance of the error term would be independent of z , so the variance would depend on the number of items in the sample, but not their total size. This would lead to a least squares estimate of R of $\sum zy / \sum z^2$, rather than the more intuitive density definition and model for R presented above.

To incorporate serial correlation in the variance, we used a nearest-neighbor calculation, with the error sum of squares given by:

$$\sum_{i=1}^{n-1} \left(\frac{y_i^2}{z_i} + \frac{y_{i+1}^2}{z_{i+1}} - \frac{(y_i + y_{i+1})^2}{z_i + z_{i+1}} \right)$$

i.e. the sum of squared deviations from pairwise weighted mean densities. The nearest-neighbor error variance of \hat{R} is:

$$Var(\hat{R}) = \frac{(1-f)}{(n-1)\sum_i z_i} \sum_{i=1}^{n-1} \left(\frac{y_i^2}{z_i} + \frac{y_{i+1}^2}{z_{i+1}} - \frac{(y_i + y_{i+1})^2}{z_i + z_{i+1}} \right)$$

Both variance calculations were applied to the Devon Island survey data. In addition, calculations for these strata based on Jolly's (1969) Method II and Cochran's (1977) systematic survey models are provided in the appendices for comparison. For the final estimate, we used the nearest neighbor variance. All distance measurements used North Pole Azimuthal Equidistant projection and area-dependent work used North Pole Lambert Azimuthal Equal Area, with central meridian at 88°W and latitude of origin at 76°N (centered over the study area for high precision).

Population growth rates were calculated following the exponential growth function, which approximates growth when populations are not limited by resources or competition (Johnson 1996):

$$N_t = N_0 e^{rt} \quad \text{and} \quad \lambda = e^r$$

Where N_t is the population size at time t and N_0 is the initial population size (taken here as the previous survey in 2008). The instantaneous rate of change is r , which is also represented as a constant ratio of population sizes, λ . When $r > 0$ or $\lambda > 1$, the population is increasing; when $r < 0$ or $\lambda < 1$ the population is decreasing. Values of $r \sim 0$ or $\lambda \sim 1$ suggest a stable population.

Results

We flew surveys on March 22-30 for a total of 57.4 hours (43.2 h and 5162 km on transect). Incidental wildlife sightings are presented in Appendix 3 and daily flight summaries are presented in Appendix 4.

Visibility was excellent for all survey flights with clear skies (visual estimates of <20% cloud, except some low cloud over open water along the coasts) and high contrast. Temperatures were steady about -30°C during the survey. We saw 14 caribou and 830 muskoxen (plus 6 newborn calves) in total, including off transect sightings. This included 13 Peary caribou and 344 muskoxen on transect. Spatial data presented in Figure 4 represents waypoints taken during the survey along transects and includes on- and off-transect sightings. Except for groups observed on the transect line, waypoints have error associated with the group's distance from the plane. While observations on transect are within 500 m, some muskox groups off transect were more than 2 km away.

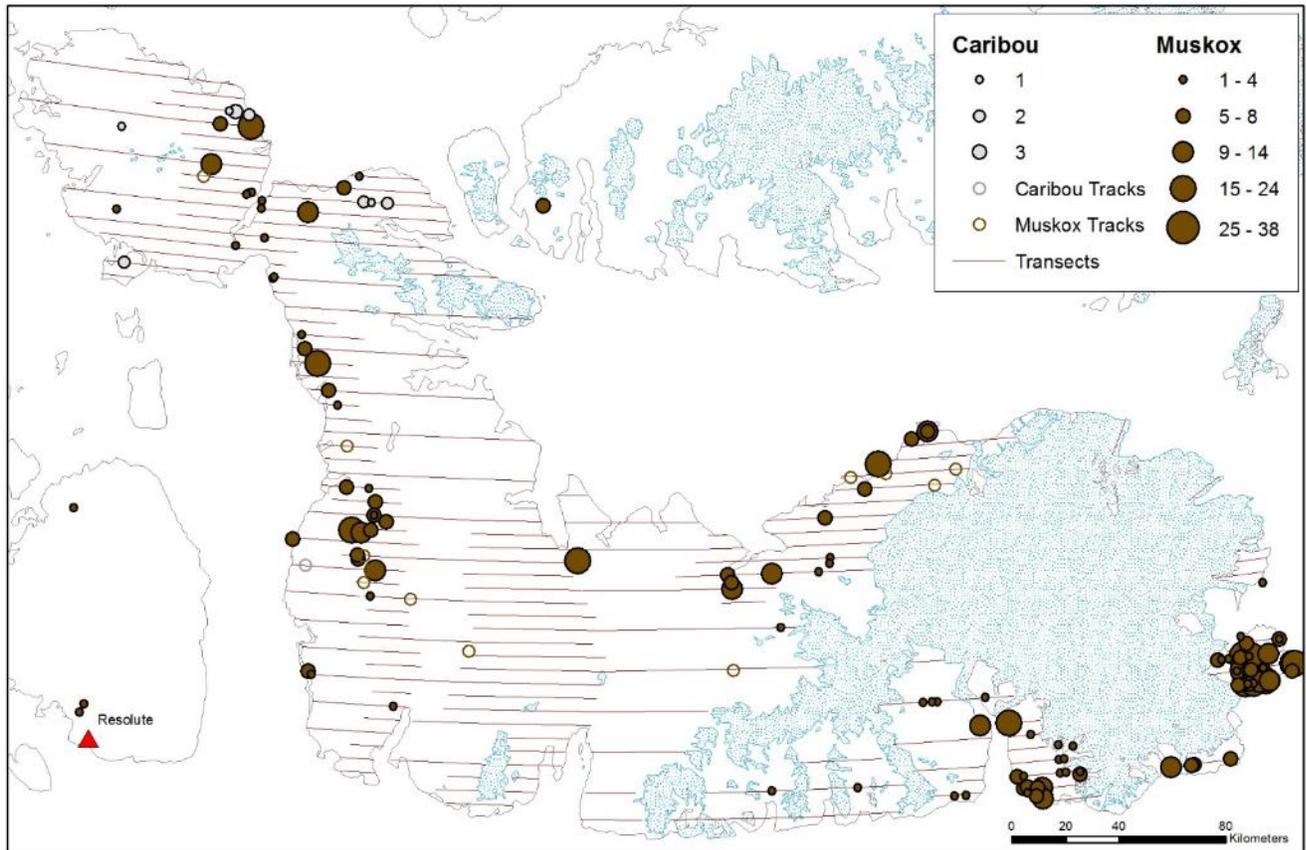


Figure 4. Observations of Peary caribou and muskoxen on Devon Island, March 2016, including observations on and off transect, and on ferry flights.

Abundance Estimates

The low number of observations in the intermediate density stratum B (9 muskoxen in 3 groups) and low density stratum C (1 group of 2 muskoxen) precluded calculation of precise population estimates for those areas, but they have been included in the overall population estimate for the island to reflect the low densities of muskoxen present in these strata. A population estimate was calculated for Peary caribou, but the few observations, which were spatially limited to the northwestern part of the study area, also prevent calculation of a precise estimate. Population estimates and variances are presented in Table 2 for muskoxen and Table 3 for caribou.

Table 2. Muskox population calculations for three strata on Devon Island with variance calculated by nearest neighbor methods and by deviations from the sample mean.

Stratum	Stratum area Z (km ²)	Surveyed area z (km ²)	Count, y	Estimate, \hat{Y}	Density, \hat{R}	Nearest Neighbor				Deviations from sample mean			
						Error Sum of Squares	Var (\hat{Y})	SE	CV	Error Sum of Squares	Var (\hat{Y})	SE	CV
High Density	18438.26	3387.77	2	1865	0.002	168.718	117524.7	342.8	0.184	246.355	171604.6	414.3	0.222
Medium Density	6359.77	580.54	9	69	0.016	1.101	2217.7	47.1	0.684	0.954	1922.6	43.8	0.637
Low Density	15076.34	1023.81	344	30	0.101	0.050	371.9	19.3	0.655	0.075	556.5	23.6	0.801
Total	39874.37	4992.12	355	1963			120114.3	346.6	0.186		174083.7	417.2	0.224

Table 3. Peary caribou population calculations for three strata on Devon Island with variance calculated by nearest neighbor methods and by deviations from the sample mean.

Stratum	Stratum area Z (km ²)	Surveyed area z (km ²)	Count, y	Estimate, \hat{Y}	Density, \hat{R}	Nearest Neighbor				Deviations from sample mean			
						Error Sum of Squares	Var (\hat{Y})	SE	CV	Error Sum of Squares	Var (\hat{Y})	SE	CV
High Density	18438.26	3387.77	13	69	0.004	1.314	2658.0	51.6	0.751	1.380	930.7	30.5	0.445
Medium Density	6359.77	580.54	0	0	0								
Low Density	15076.34	1023.81	0	0	0								
Total	39874.37	4992.12	13	69			2658.0	51.6	0.751		930.7	30.5	0.445

Population Trends

Muskoxen have increased since the last survey in 2008. Based on a population estimate of $1963 \pm SE343$ in 2016 and 513 in 2008 (302-864, 95%CI; Jenkins et al. 2011), the instantaneous growth rate r is 0.08, and lambda λ is 1.09. More sophisticated analyses incorporating uncertainty in the estimates have not been undertaken.

A population estimate for caribou was not calculated in 2008 due to the small number of observations. If the groups observed in 2008 had been observed in 2016 with a fixed-width strip transect survey instead, then 3 of the 4 groups (13 of 17 individuals) would have been on transect in the high density stratum. The 2008 population estimate would have been $69 \pm SE47$, compared to the 2016 estimate of $69 \pm SE52$. The wide confidence interval and few observations in both years make these estimates questionable. Furthermore, neither survey detected caribou in the Truelove Lowlands, where they are known to occur. The 2016 survey also did not detect caribou around Baring Bay, another area where they are known to exist. Lack of observations could be due to movement of animals out of these areas, but it is also possible that they were present but not detected.

Calf Recruitment

Although we observed 119 groups of muskoxen, many of these were too far away or individuals were grouped too closely for sex/age identification, and 59 of these groups had at least some individuals with an unknown age. It is also likely that newborn calves were missed in tightly grouped herds, since they are still small and would be inconspicuous or deliberately hidden behind the adults. Newborns were identified in herds with 5, 7, 7, 8, and 15 1+-year-old muskoxen – larger or more tightly clumped groups could easily have concealed others. The distinct size difference between yearlings and adults would also be less obvious under these circumstances. Eleven yearlings were conclusively identified in groups without any unknown age class animals, making them 4.8% of the population. This is based on a biased sample of groups, however, since the larger groups which had animals of unknown age and sex class likely had more yearlings.

Group Size

We observed 119 groups of muskoxen, with group sizes ranging from single animals to a herd of 38, with an average of 7.0 muskoxen per group (SD=6.0). Caribou were seen in smaller groups of 1 to 4.

Discussion

Population Trends

Previous surveys of Devon Island have used different survey platforms (Piper Super Cub and deHavilland Beaver, Tener 1963; ground surveys, Freeman 1971; Bell 206 helicopter, Case 1992, Jenkins et al. 2011; Twin Otter, this survey). They have also concentrated on different parts of the island, usually with the goal of estimating muskox populations and therefore focusing on the lowland areas of the north, west, and southeast coasts. The largely unsuitable habitat for caribou or muskoxen on the rest of the island minimizes the bias in estimates derived from these surveys however, especially compared to other island groups that have historically been partially surveyed. Case (1992) did note that muskoxen on the 1990 survey may have been missed inland from Baring Bay and a search of that area would have improved the survey results.

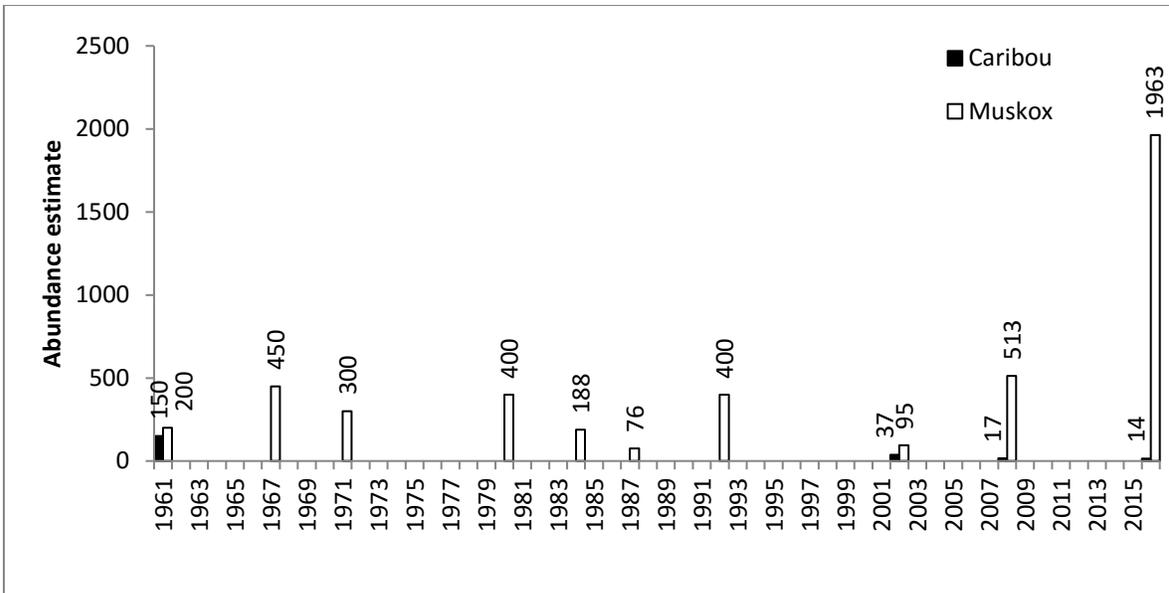


Figure 5. Population estimates for muskoxen and caribou on Devon Island. Muskox estimates prior to 1980 were extrapolations from minimum counts (Tener 1963, Freeman 1971, Hubert 1977, Decker in Urquhart 1982, Case 1992), followed by minimum counts (Pattie 1990, GN data unpublished for 2002) and then systematic surveys covering part (GN data unpublished for 2002) or all (Jenkins et al. 2011 and this survey) of Devon Island. Caribou estimates are guesses (Tener 1963) or minimum counts (Jenkins et al. 2011, this survey).

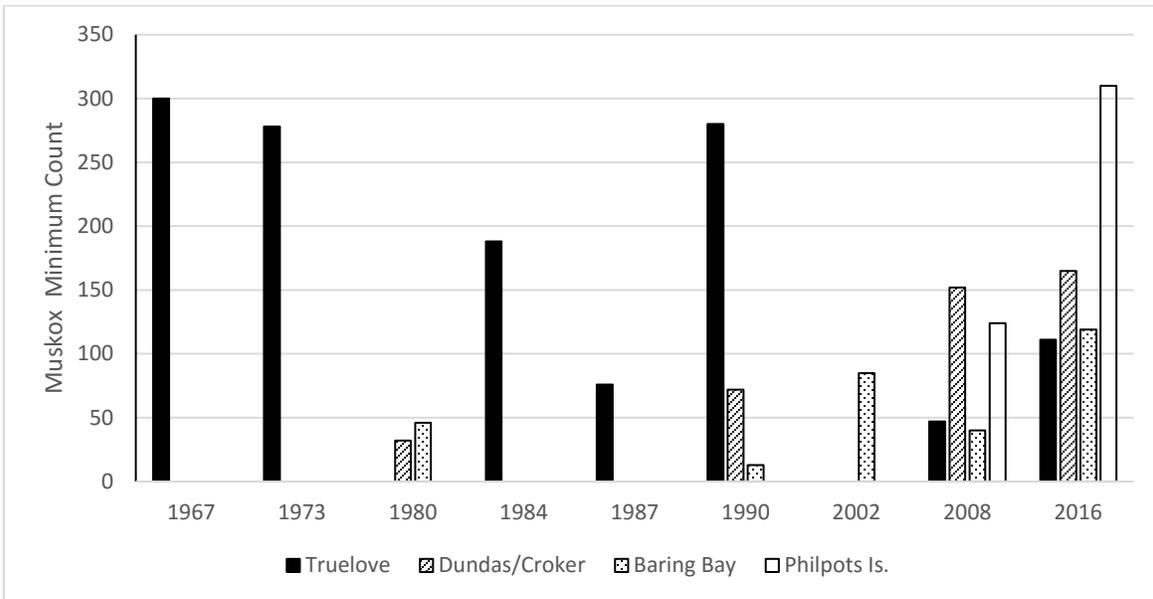


Figure 6. Minimum counts of muskoxen recorded on surveys of lowland areas where muskoxen congregate (Freeman 1971, Hubert 1977, Decker in Urquhart 1982, Pattie 1990, Case 1992, GN data unpublished for 2002 and 2008, Jenkins et al. 2011, and this survey). Not all areas were surveyed in all years.

Muskox and Caribou Distribution

Muskox concentrations have been reported consistently in the lowlands around Baring Bay, Truelove/Sverdrup Inlet, Dundas Harbour, and Philpots Island, and these continued to be places with high muskox densities. The area around Arthur Fiord on the Grinnell Peninsula also supported

relatively high densities of muskoxen. Although the distribution has not changed dramatically, each of the lowland areas, and particularly Philpots Island, has experienced an increase in muskox population since the last survey in 2008. The Truelove Lowlands have historically supported larger muskox populations than the number observed during this survey, although more survey effort in these areas in the past compared to a systematic survey makes it difficult to directly compare this years' observations with historic counts. The increasing muskox population is still largely confined to discrete areas of suitable habitat, however, and the unsuitable habitat in the barren interior of the island remains largely unoccupied. Increasing populations on the Bathurst Island Complex and on southern Ellesmere Island indicate that muskox populations are increasing across the region. The increase on Devon Island may be due to recruitment within the population rather than large-scale movement of muskoxen from other neighboring island groups. High calf recruitment of 15-20% starting with a population of 531 muskoxen over the last 8 years could account for an increase to a 2016 population of 1600-2300 muskoxen, but this would be contingent on other factors like adult survival. Relatively little is known about muskox movements in the area.

Caribou distribution has apparently also remained similar to previous surveys and reports. We were unable to locate caribou in the Truelove Lowlands, despite local knowledge of their presence. This may not be surprising if the caribou persist at low densities in small isolated habitat patches. We were also unlikely to have found tracks across this part of the study area, since much of the lowlands were either windswept or had hard-packed snow, which was not conducive to track detection.

We also checked for tracks and animals along the sea ice and shorelines during short ferry flights between transects, allowing us to cover 50% of the shoreline. We did not see any caribou or muskox tracks on the sea ice that would suggest recent movement among islands, and no major movement to or from Devon Island was evident during the survey.

Calf Recruitment

The recorded proportion of muskox yearlings in the population (5%) was much lower than recorded for southern Ellesmere Island in summer 2014 (24%, Anderson and Kingsley 2015), and lower than the 10.5% calf production which Freeman (1971) estimated would be required to offset natural mortality based on observations in 1965 and 1967. Since no unusual mortality or calf crop losses have been noticed by harvesters, it is likely that the recorded proportion of yearlings represents biased sampling of small, dispersed, and often adult-dominated, muskox groups, without taking into account the proportion of yearlings in larger or tightly grouped herds. The proportion of newborn calves will be biased low due to detectability, and because the survey was at the beginning of calving season.

Lack of observations prevents any conclusions on calf recruitment for Peary caribou.

Group Sizes

Muskox groups are largest early in the spring and smaller as summer progresses (Freeman 1971, Gray 1973), with winter (including April and May) groups about 1.7 times larger than summer groups (Heard 1992). Muskoxen were encountered in herds of 2-38, with some lone adults seen as well, and averaged 7.0 muskoxen per herd. This is slightly smaller than the 10.0 muskoxen per herd encountered by Freeman (1971) and slightly smaller than herd sizes encountered in March 2015 on southern Ellesmere Island (8.9-12.1 muskoxen/group, 95%CI, Anderson and Kingsley 2015), although the degree to which muskoxen move among the two islands is not clear and group size could be different for different populations.

Ferguson (1991) suggested that caribou groups are largest in August and smaller in late winter, and Fischer and Duncan (1976) noted that groups across the Arctic islands averaged 4.0 caribou in late winter, 2.8 caribou in early summer, and 8.8 caribou in mid-summer. Peary caribou were seen singly or in small groups of 2-4, but not enough groups were observed to make any meaningful conclusions on group sizes.

Management Recommendations

Peary caribou and muskoxen on Devon Island are an important source of country food and cultural persistence for Inuit. Consistent with the Nunavut Land Claim Agreement, and the Management Plan for High Arctic Muskoxen of the Qikiqtaaluk Region, 2012-2017 (DOE 2014), these management recommendations emphasize the importance of maintaining healthy populations of caribou and muskox that support sustainable harvest.

Under the Management Plan for the High Arctic Muskoxen of the Qikiqtaaluk Region, 2013-2018 (DOE 2014), Devon Island is considered a single management unit, MX-04, with a Total Allowable Harvest (TAH) of 15. The high numbers of muskox suggest that the TAH could be increased or removed, although with 3 communities harvesting from the island, maintaining a TAH might facilitate harvest management and co-ordination by the 3 HTAs (i.e. maintaining tags to track harvest, but setting the TAH high enough to ensure any interested hunter could receive a tag). The current TAH reflects a conservative harvest rate of 4% on a population of about 400 muskoxen, which is close to the population estimates from the 1970s until 2008. The 2016 population estimate, however, is close to four times the 2008 estimate. At the same harvest rate of 4%, 79 muskox tags could be issued. At a 5% harvest rate, 98 tags could be issued. Muskoxen do move across the barren interior of the island and among habitat patches (based on unpublished GN telemetry data, and local knowledge in Grise Fiord and Resolute), but dispersing harvest among several lowlands would prevent having to wait for muskoxen to re-establish themselves in areas that might be more isolated.

It is highly recommended that a harvest reporting system be maintained even if the TAH is removed. This would allow biologists, community members, and decision makers to track harvest patterns over time and to determine whether changes to management zones or harvest restrictions have the desired effect. With muskoxen concentrated in discrete lowland habitats that can be reliably accessed for harvesting, it may be particularly useful to distribute harvest pressure among these areas or to target under-utilized areas for larger community hunts. A large, coordinated community hunt with distributed harvest pressure in the more concentrated areas could be considered for 2016-17 with harvest monitoring recommended without TAH limitation. Following this, a TAH of 100 could be considered for future years, with ongoing population monitoring at regular intervals. As local knowledge and previous surveys have demonstrated, population changes can be rapid and unexpected if severe weather causes localized or widespread starvation or movement, so continuous monitoring and adaptive management is necessary even when populations are at high levels.

Harvest trends for muskoxen over the last decade suggest that Grise Fiord and Resolute Bay harvest fewer muskoxen than in the 1990s (Anderson 2016), but changing the configuration of management zones may encourage more harvesting in areas that were previously accessible but not included in a management unit. The major decline in caribou on Baffin Island, and subsequent harvest restrictions, has also reduced the availability of country food for Baffin communities,

including Arctic Bay, which has harvested muskoxen on Devon Island in the past. The community of Arctic Bay has been in discussions with Grise Fiord to determine whether they would be able to harvest several muskoxen to offset the lack of Baffin caribou, and this should be further considered given the healthy populations of muskoxen on southern Ellesmere and Devon islands.

Although we saw only 14 caribou during the survey, the results of previous surveys over the same areas suggest that caribou have persisted at relatively low densities on Devon Island. There may or may not have been a decline from the 2008 survey, the few observations recorded from both surveys make it difficult to tell. Most caribou harvest activity from Resolute Bay has been focused on Bathurst Island, reducing the available recent knowledge of caribou on Devon Island, although residents of Resolute still visit Devon Island for other harvesting activities and during travel. Hunters from Grise Fiord report seeing caribou fairly regularly in the Truelove Lowlands, and a few are caught there each year. It is unlikely that harvest restrictions on Peary caribou will result in any marked increase in the population, as harvest is restricted to a small human population with limited access to the caribou range, and lack of suitable habitat on Devon Island is likely a more important factor limiting caribou population growth in the area. Monitoring sightings and harvest will continue to provide a more complete picture of where caribou are on the landscape.

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Appendix 1. Devon Island survey transects, 2016.

Table 4. Transect end points and strata on Devon Island for a fixed-wing survey, March 2016.

Transect	Stratum	Lon (West)	Lat (West)	Lon (East)	Lat (East)
A001	High	-95.4515	76.9729	-94.5496	76.9736
A002	High	-95.5004	76.9283	-93.7822	76.9278
A003	High	-94.7150	76.8833	-93.6314	76.8824
A004	High	-94.9700	76.8372	-93.5000	76.8371
A005	High	-94.7862	76.7916	-93.3761	76.7913
A006	High	-94.5015	76.7461	-93.1818	76.7466
A007	High	-94.3147	76.7014	-93.2004	76.7013
A008	High	-94.2895	76.6557	-93.2195	76.6559
A009	High	-94.4366	76.6110	-93.2781	76.6106
A010	High	-94.4592	76.5652	-93.3219	76.5653
A011	High	-94.4104	76.5201	-93.4145	76.5200
A012	High	-94.4379	76.4753	-93.5371	76.4743
A013	High	-95.5015	76.4292	-90.8734	76.4297
A014	High	-95.5037	76.3837	-92.6704	76.3843
A015	High	-95.0002	76.3382	-93.4020	76.3383
A016	High	-95.4086	76.2931	-93.7747	76.2934
A017	High	-95.3984	76.2480	-94.9366	76.2486
A018	High	-93.3600	76.4744	-90.4714	76.4742
A019	High	-93.2573	76.5203	-90.5103	76.5198
A020	High	-93.1695	76.5650	-90.5891	76.5652
A021	High	-92.2238	76.6103	-90.8334	76.6108
A022	High	-91.9925	76.6557	-90.9958	76.6558
A023	High	-91.1194	76.3840	-90.2572	76.3837
A024	High	-91.2429	76.3394	-89.8187	76.3386
A025	High	-91.0414	76.2040	-89.3047	76.2023
A026	High	-93.0451	76.3390	-92.8219	76.3387
A027	High	-93.0268	76.2946	-92.7023	76.2936
A028	High	-92.9776	76.2478	-92.6527	76.2479
A029	High	-92.7997	76.2024	-92.4764	76.2025
A030	High	-92.7452	76.1573	-92.0528	76.1574
A031	High	-92.6659	76.1118	-91.6568	76.1119
A032	High	-92.6472	76.0663	-91.8596	76.0690
A033	High	-92.6542	76.0211	-91.7933	76.0209
A034	High	-92.5567	75.9763	-91.6839	75.9766
A035	High	-92.4049	75.9306	-91.7767	75.9314
A036	High	-92.1608	75.8853	-91.6562	75.8857
A037	High	-92.1191	75.8399	-91.4810	75.8389
A038	High	-92.1076	75.7946	-91.4616	75.7956
A039	High	-92.1276	75.7492	-91.3693	75.7499
A040	High	-92.0838	75.7040	-91.3943	75.7037

Transect	Stratum	Lon (West)	Lat (West)	Lon (East)	Lat (East)
A041	High	-92.0019	75.6590	-91.0036	75.6591
A042	High	-92.0969	75.6130	-91.0329	75.6135
A043	High	-91.9416	75.5678	-91.0005	75.5677
A044	High	-91.7431	75.5229	-91.0916	75.5224
A045	High	-91.6967	75.4771	-90.9195	75.4770
A046	High	-91.7627	75.4319	-90.7419	75.4321
A047	High	-91.7767	75.3852	-90.9011	75.3865
A048	High	-91.6819	75.3410	-90.9186	75.3418
A049	High	-91.5624	75.2962	-90.9901	75.2959
A050	High	-91.5011	75.2503	-91.4406	75.2504
A051	High	-91.3900	75.2043	-90.8364	75.2054
A052	High	-91.4369	75.1599	-90.8372	75.1600
A053	High	-91.4721	75.1145	-90.6935	75.1145
A054	High	-91.4349	75.0696	-90.6713	75.0696
A055	High	-91.3613	75.0243	-90.6430	75.0249
A056	High	-91.2629	74.9785	-90.7010	74.9785
A057	High	-91.2693	74.9338	-90.7616	74.9338
A058	High	-91.3129	74.8880	-90.8166	74.8878
A059	High	-91.3528	74.8429	-90.8916	74.8427
A060	High	-91.4164	74.7973	-90.9834	74.7973
A061	High	-91.5014	74.7520	-91.0738	74.7524
A062	High	-91.6261	74.7065	-91.1911	74.7067
A063	High	-91.6055	74.6614	-91.1491	74.6611
A064	High	-89.4999	75.5675	-89.1716	75.5679
A065	High	-89.9996	75.5219	-89.1295	75.5227
A066	High	-90.0587	75.4768	-88.9798	75.4771
A067	High	-90.0836	75.4316	-88.6913	75.4319
A068	High	-90.1396	75.3866	-88.7039	75.3865
A069	High	-90.1529	75.3415	-88.6963	75.3418
A070	High	-90.1137	75.2960	-88.5720	75.2960
A071	High	-90.0533	75.2507	-88.4995	75.2504
A072	High	-90.0618	75.2053	-88.3821	75.2051
A073	High	-89.9997	75.1599	-88.3181	75.1599
A074	High	-89.9242	75.1146	-88.4192	75.1148
A075	High	-89.9997	75.0694	-88.2573	75.0698
A076	High	-90.1350	75.0240	-88.2871	75.0240
A077	High	-90.1881	74.9784	-88.3126	74.9785
A078	High	-90.2849	74.9335	-88.3626	74.9326
A079	High	-90.3433	74.8877	-88.3949	74.8878
A080	High	-89.7865	74.8426	-88.4362	74.8430
A081	High	-89.5025	74.7966	-88.7032	74.7974
A082	High	-89.5010	74.7520	-88.7710	74.7525

Transect	Stratum	Lon (West)	Lat (West)	Lon (East)	Lat (East)
A083	High	-90.0525	74.7068	-88.8018	74.7066
A084	High	-90.2588	74.6614	-89.1049	74.6617
A085	High	-90.3994	74.6157	-89.1419	74.6158
A086	High	-90.2637	74.5705	-89.4294	74.5707
A087	High	-84.0040	75.7944	-83.7115	75.7945
A088	High	-84.2406	75.7490	-82.6540	75.7493
A089	High	-84.4573	75.7043	-83.2935	75.7036
A090	High	-85.0695	75.6586	-83.3435	75.6585
A091	High	-85.1656	75.6136	-83.5736	75.6132
A092	High	-85.6929	75.5679	-84.1179	75.5682
A093	High	-86.1124	75.5229	-84.3768	75.5222
A094	High	-86.0504	75.4775	-84.5217	75.4767
A095	High	-85.8495	75.4322	-84.5947	75.4319
A096	High	-87.0686	75.3870	-84.7760	75.3866
A097	High	-87.3422	75.3412	-85.4043	75.3412
A098	High	-87.4193	75.2959	-86.0030	75.2958
A099	High	-86.7360	75.2507	-86.1493	75.2508
A100	High	-86.9984	75.4318	-86.4147	75.4318
A101	High	-86.8880	75.4776	-86.6370	75.4777
A104	High	-79.9315	75.2511	-79.5140	75.2505
A105	High	-80.2150	75.2056	-79.5756	75.2049
A106	High	-80.0445	74.9786	-79.5506	74.9782
A107	High	-80.4098	74.9334	-79.4804	74.9333
A108	High	-80.4593	74.8879	-79.3482	74.8880
A109	High	-80.1173	74.8423	-79.6645	74.8426
A110	High	-81.1654	74.5974	-80.2187	74.5974
A111	High	-82.6603	74.5250	-81.9998	74.5251
A112	High	-82.9629	74.5704	-82.2931	74.5706
A113	High	-83.0611	74.6157	-82.2674	74.6165
A114	High	-83.1139	74.6612	-82.2818	74.6615
A115	High	-83.1294	74.7063	-82.6106	74.7063
A116	High	-83.1117	74.7522	-82.6943	74.7522
A117	High	-83.1035	74.7973	-82.6953	74.7969
A118	High	-83.8110	74.6163	-83.4697	74.6147
A119	High	-84.1586	74.5706	-83.4989	74.5710
B001	Medium	-92.2611	75.5225	-91.7431	75.5229
B002	Medium	-92.4253	75.4319	-91.7627	75.4319
B003	Medium	-92.4319	75.3413	-91.6819	75.3410
B004	Medium	-92.4867	75.2507	-91.5011	75.2503
B005	Medium	-92.3308	75.1599	-91.4369	75.1599
B006	Medium	-92.2119	75.0691	-91.4349	75.0696
B007	Medium	-92.1187	74.9786	-91.2629	74.9785

Transect	Stratum	Lon (West)	Lat (West)	Lon (East)	Lat (East)
B008	Medium	-92.0224	74.8882	-91.3129	74.8880
B009	Medium	-92.0776	74.7972	-91.4164	74.7973
B010	Medium	-88.9181	75.6133	-88.4986	75.6130
B011	Medium	-88.8185	75.5222	-87.0187	75.5222
B012	Medium	-88.6913	75.4319	-86.9984	75.4318
B013	Medium	-88.5004	74.6154	-87.8639	74.6159
B014	Medium	-88.5684	74.5253	-87.7900	74.5256
B015	Medium	-86.9721	74.6158	-85.8925	74.6148
B016	Medium	-87.4396	74.5258	-85.8803	74.5241
B017	Medium	-85.7538	74.7063	-84.7336	74.7067
B018	Medium	-85.6999	74.6155	-84.4595	74.6160
B019	Medium	-85.8803	74.5241	-84.5302	74.5254
B020	Medium	-84.5960	74.7519	-83.3208	74.7524
B021	Medium	-84.3724	74.6606	-83.4318	74.6597
C001	Low	-96.8561	76.9279	-95.5004	76.9283
C002	Low	-96.9199	76.7922	-94.7862	76.7916
C003	Low	-96.4657	76.6559	-94.2895	76.6557
C004	Low	-96.1059	76.5168	-94.4104	76.5201
C005	Low	-95.9554	76.4259	-95.5015	76.4292
C006	Low	-92.7023	76.2936	-91.1545	76.2941
C007	Low	-92.4764	76.2025	-91.2682	76.2033
C008	Low	-91.8596	76.0690	-90.2112	76.0666
C009	Low	-91.7767	75.9314	-89.8083	75.9305
C010	Low	-91.4616	75.7956	-89.2222	75.7949
C011	Low	-91.0036	75.6591	-89.2157	75.6586
C012	Low	-91.0916	75.5224	-89.9996	75.5219
C013	Low	-90.9011	75.3865	-90.1396	75.3866
C014	Low	-91.0024	75.2506	-90.0533	75.2507
C015	Low	-90.6935	75.1145	-89.9242	75.1146
C016	Low	-90.7010	74.9785	-90.1881	74.9784
C017	Low	-90.7393	74.8424	-89.7865	74.8426
C018	Low	-90.9777	74.7063	-90.0525	74.7068
C019	Low	-88.5720	75.2960	-87.4193	75.2959
C020	Low	-88.3181	75.1599	-85.7670	75.1600
C021	Low	-88.2871	75.0240	-85.5456	75.0245
C022	Low	-88.3949	74.8878	-86.9173	74.8876
C023	Low	-88.4291	74.7518	-87.2422	74.7531
C024	Low	-84.7369	74.8883	-83.0024	74.8876
C025	Low	-84.3593	75.0242	-83.8459	75.0245
C026	Low	-85.4042	75.3443	-84.7132	75.3416
C102	Low	-81.3948	75.6551	-81.1461	75.7744
C103	Low	-80.4653	75.4746	-80.0000	75.5419

Appendix 2. Delineation of survey strata for Devon Island.

The following figures show the boundaries for high, intermediate, and low density strata for caribou and muskoxen. Both species were considered together, since much of the information indicated overlapping ranges and both species were targeted for the survey. In addition to the maps provided below, we used maps provided in Case (1992) of high muskox density areas and locations indicated by community members (summarized in Taylor 2005 and Johnson et al. 2016, but also indicated by elders and hunters prior to and during the survey).

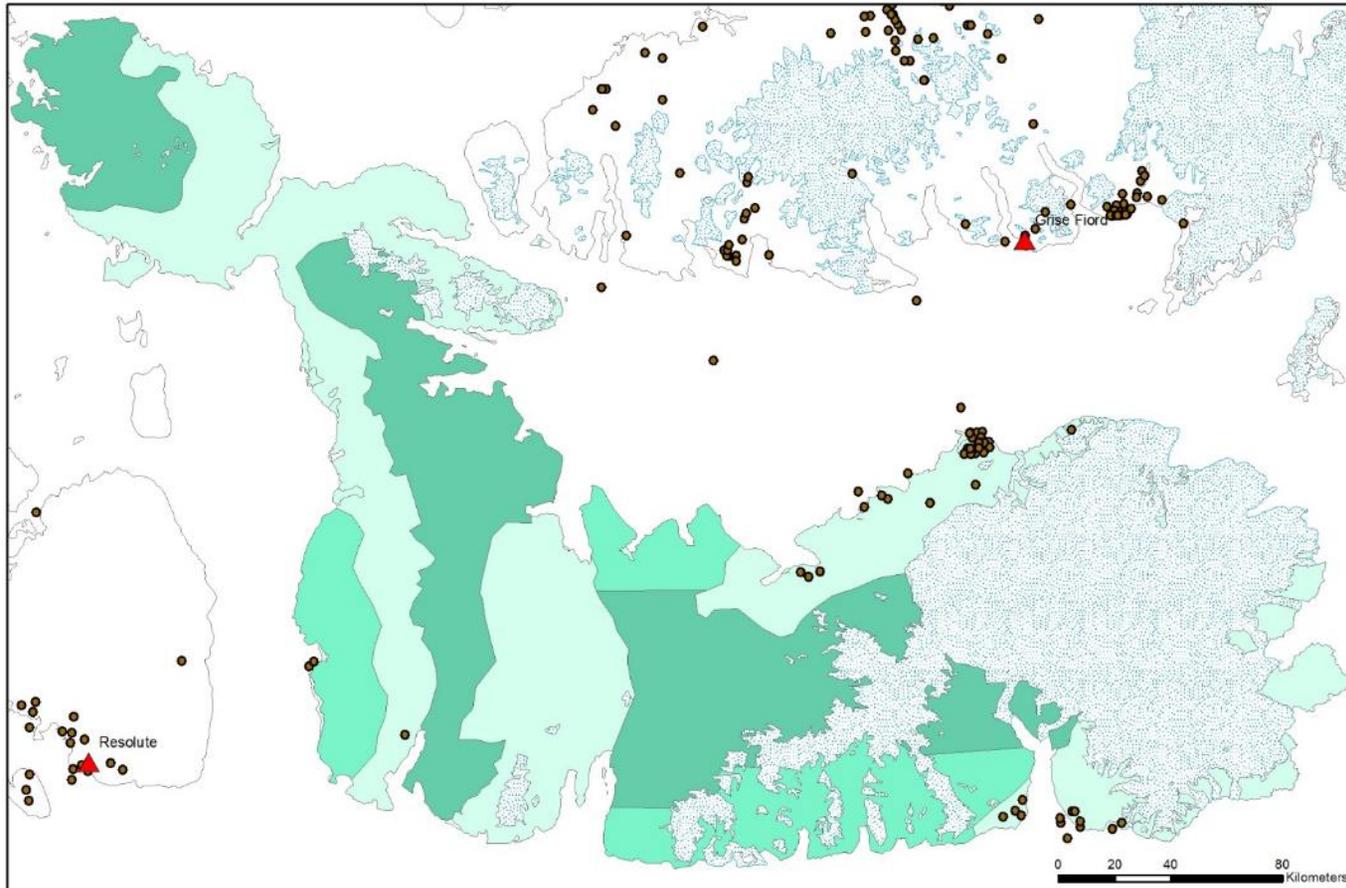


Figure 7. Locations of muskox harvest from Grise Fiord, Resolute Bay, and Arctic Bay, 1990-2015. Survey strata are indicated by shaded green – high density (pale green), intermediate density (bright green), low density (dark green).

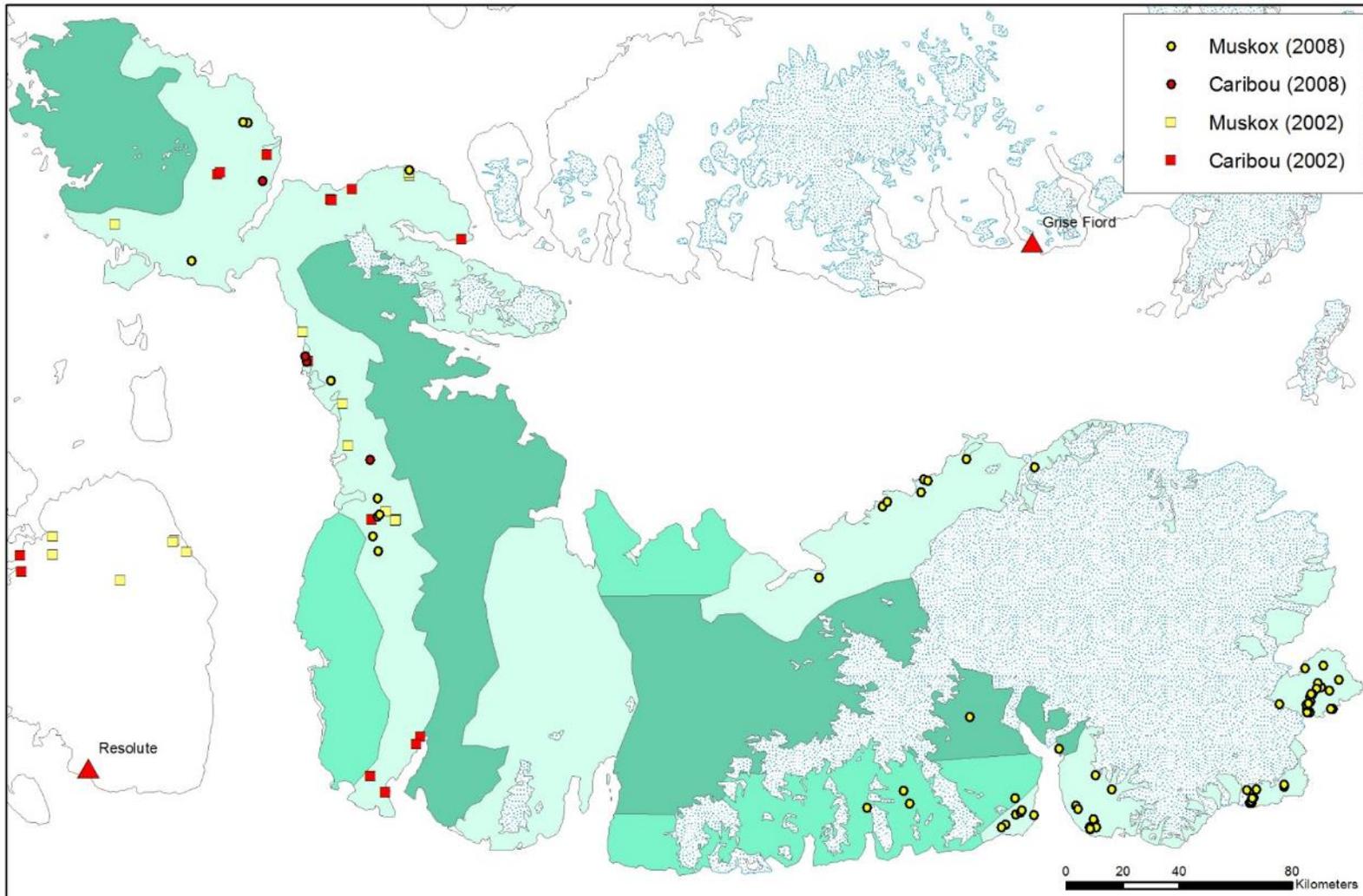


Figure 8. Locations of caribou and muskoxen seen on aerial surveys in 2002 and 2008. Survey strata are indicated by shaded green – high density (pale green), intermediate density (bright green), low density (dark green).

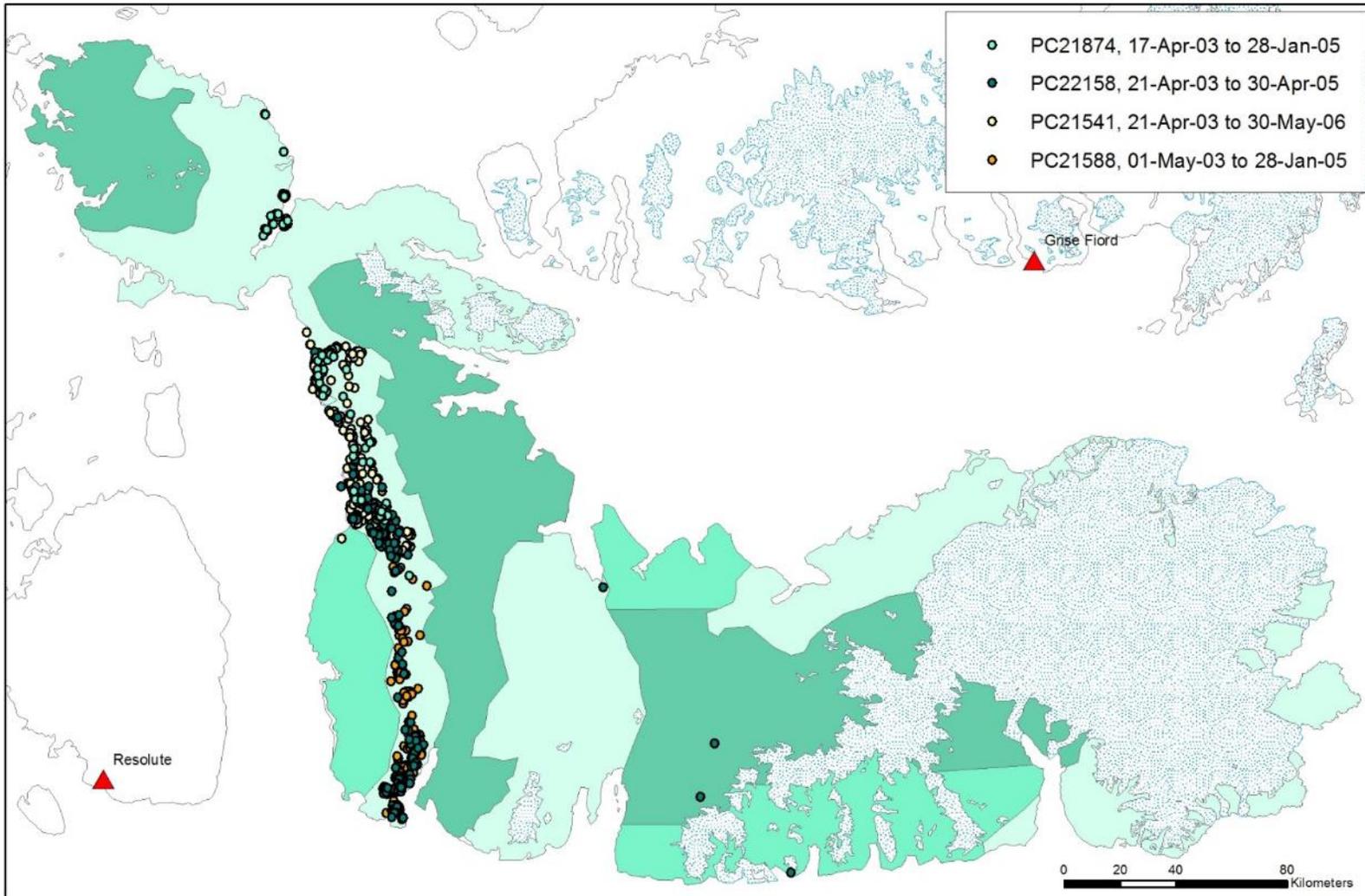


Figure 9. Telemetry locations of 4 collared female caribou, 2003-2006, on Devon Island. Survey strata are indicated by shaded green – high density (pale green), intermediate density (bright green), low density (dark green).

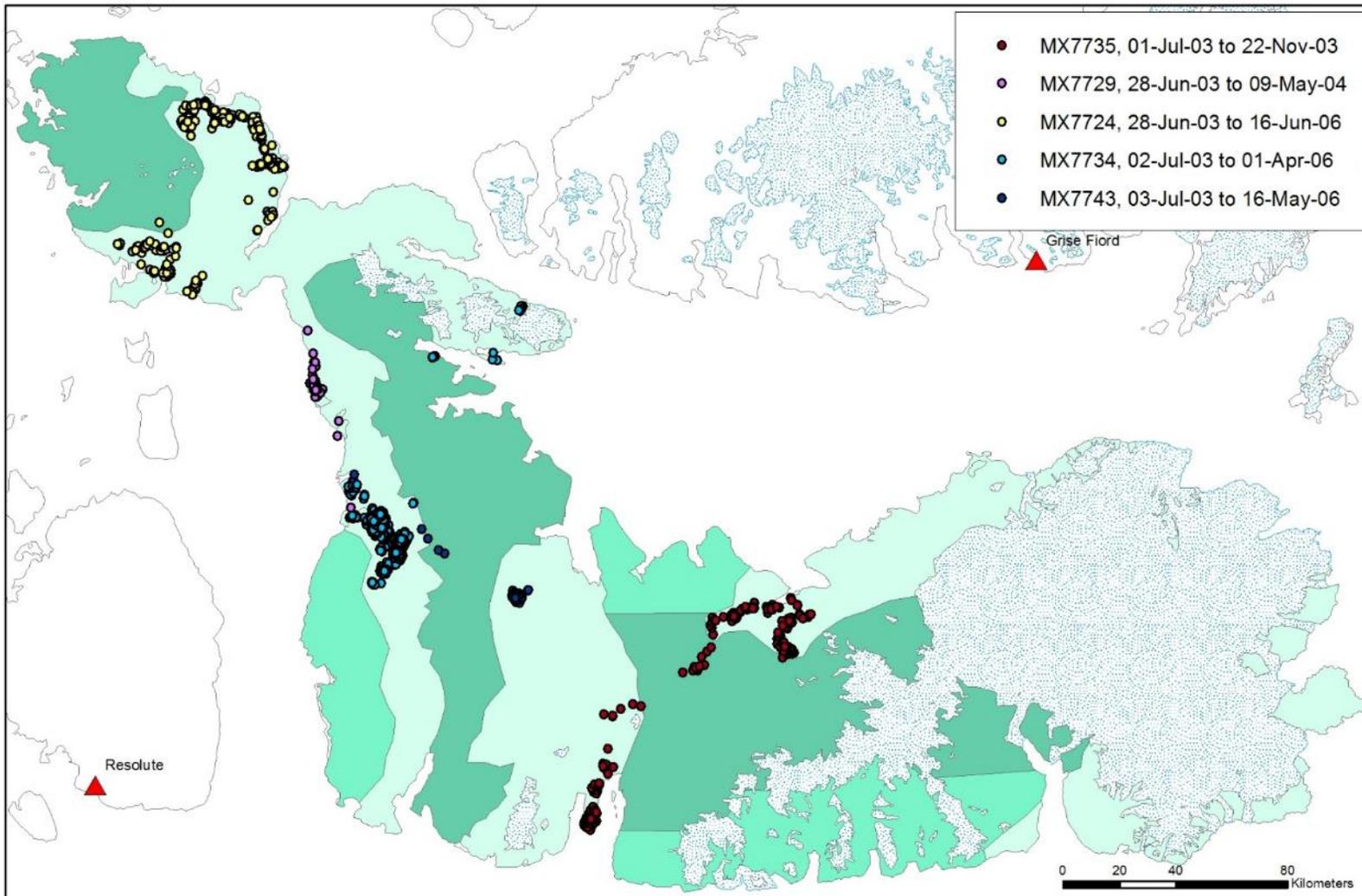


Figure 10. Telemetry locations of 5 collared female muskoxen, 2003-2006, on Devon Island. Survey strata are indicated by shaded green – high density (pale green), intermediate density (bright green), low density (dark green).

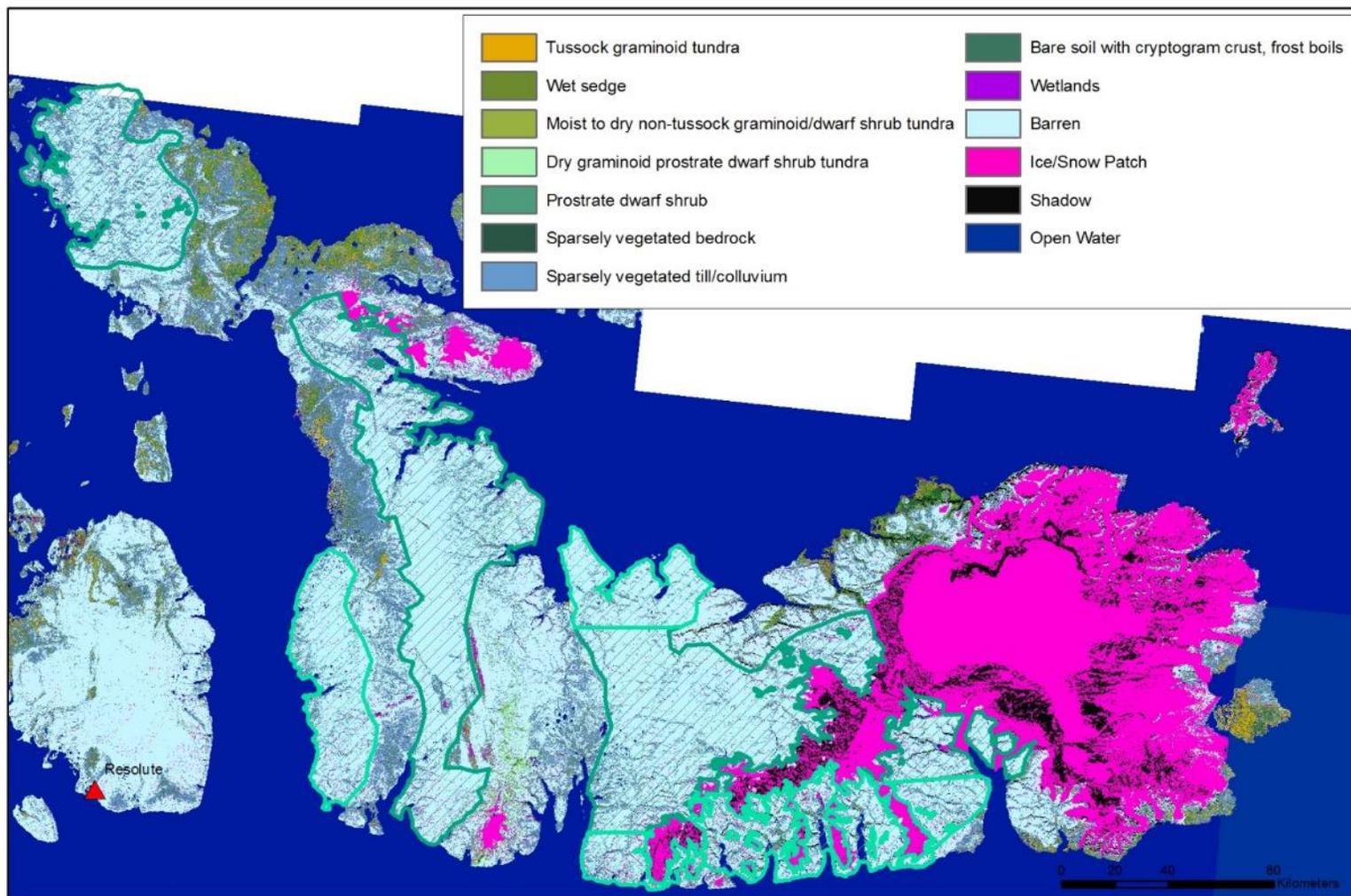


Figure 11. Land cover classification developed from Landsat imagery 1999-2002 (Olthof et al. 2008; available online through Natural Resources Canada). Survey strata are outlined and hatched by light green (intermediate density) or dark green (low density), with remaining non-icecap areas as high density strata.

Appendix 3. Alternate population calculations.

Jolly Method II Calculations

In this report, we used a systematic sampling approach to analysis, since we were estimating abundance of a patch population rather than estimating density in a habitat (which varied across the study area). Other systematic aerial surveys have frequently used Jolly's Method II, and estimates derived from both analyses were similar. Population estimates for fixed-width strip sampling using Jolly's Method 2 for uneven sample sizes (Jolly 1969; summarized in Caughley 1977) are derived as follows:

$$\hat{Y} = RZ = Z \frac{\sum_i y_i}{\sum_i z_i}$$

Where \hat{Y} is the estimated number of animals in the population, R is the observed density of animals (sum of animals seen on all transects $\sum_i y_i$ divided by the total area surveyed $\sum_i z_i$), and Z is the total study area. The variance is given by:

$$Var(\hat{Y}) = \frac{N(N-n)}{n} (s_y^2 - 2Rs_{zy} + R^2s_z^2)$$

Where N is the total number of transects required to completely cover study area Z , and n is the number of transects sampled in the survey. s_y^2 is the variance in counts, s_z^2 is the variance in areas surveyed on transects, and s_{zy} is the covariance. The estimate \hat{Y} and variance $Var(\hat{Y})$ are calculated for each stratum and summed. The Coefficient of Variation ($CV = \sigma/\hat{Y}$) was calculated as a measure of precision.

Table 5. Abundance estimates (Jolly 1969 Method II) for muskoxen on Devon Island, March 2016. N is the total number of transects required to completely cover study area Z , n is the number of transects sampled in the survey covering area z , y is the observed muskoxen, Y is the estimated muskoxen with variance $Var(Y)$. The coefficient of variation (CV) is also included.

Stratum	Y	Var(Y)	n	Z (km ²)	z (km ²)	N	y	Density (per km ²)	CV
A	1815.81	39767.06	117	18438.26	3479.02	288	344	0.098	0.110
B	95.85	847.56	21	6359.77	597.17	138	9	0.015	0.400
C	27.77	undefined	28	15076.34	1085.68	288	2	0.002	
Total	1939.43		166	39874.37	5161.87	288	355		

Table 6. Abundance estimates (Jolly 1969 Method II) for Peary caribou on Devon Island, March 2016. N is the total number of transects required to completely cover study area Z , n is the number of transects sampled in the survey covering area z , y is the observed caribou, Y is the estimated caribou with variance $Var(Y)$. The coefficient of variation (CV) is also included.

Stratum	Y	Var(Y)	n	Z (km ²)	z (km ²)	N	y	Density (per km ²)	CV
A	70.46	1806.83	117	18438.26	3479.02	288	13	0.004	0.603
B	0		21	6359.77	597.17	138	0	0	
C	0		28	15076.34	1085.68	288	0	0	
Total	70.46		166	39874.37	5161.87	288	13		

Stratified Systematic Survey Calculations

Following Cochran (1977), the abundance estimate for a systematic survey is given by:

$$\hat{Y} = \frac{S}{w} \times \sum n_i$$

Where \hat{Y} is the population estimate, S is the transect spacing (5 km), w is the transect width (1 km), and n_i is the total number of animals observed on transect i , the sum of which is all animals observed on I transects in the survey. The configuration of the study area may mean that the actual sampling fraction (proportion of the study area that is surveyed) varies, which was partly why Cochran's ratio estimator was used instead, and why the estimate varied between methods and stratification regimes. The variance is based on the sum of squared differences in counts between consecutive transects:

$$Var(\hat{Y}) = \frac{S}{w} \times \frac{\left(\frac{S}{w} - 1\right) \times I}{2 \times (I - 1)} \times \sum (n_i - n_{i-1})^2$$

Table 7. Abundance estimates for a stratified systematic survey (Cochran 1977) of muskoxen on Devon Island, March 2016. I is the number of transects sampled.

Stratum	Estimated Abundance \hat{Y}	Var(\hat{Y})	I	Transect Spacing S (km)	Transect Width w (km)	Observed Individuals y	Density (per km ²)	CV
A	1720	67436.38	117	5	1	344	0.098	0.151
B	90	2740.50	21	10	1	9	0.015	0.582
C	30	871.11	28	15	1	2	0.002	0.984
Total	1840	71047.99	166			355		

Table 8. Abundance estimates for a stratified systematic survey (Cochran 1977) of Peary caribou on Devon Island, March 2016. I is the number of transects sampled.

Stratum	Estimated Abundance \hat{Y}	Var(\hat{Y})	I	Transect Spacing S (km)	Transect Width w (km)	Observed Individuals y	Density (per km ²)	CV
A	65	67436.38	117	5	1	13	0.004	0.557
B	0	2740.50	21	10	1	0	0	
C	0	871.11	28	15	1	0	0	
Total	65	71047.99	166			13		

Appendix 4. Daily flight summaries for south Ellesmere survey flown by Twin Otter, March 2015.

Table 9. Summary by day of survey flights and weather conditions for March 2015 Peary caribou and muskox survey, southern Ellesmere Island.

Date	Time Up	Time Down	Time Up 2	Time Down 2	Time Up 3	Time Down 3	Flying Time	Transect Time	Area	Comment
22-Mar-16	9:35	13:15	13:54	17:15			7:01	4:18	Grinnell Peninsula	Clear, calm, -31°C, light wind ~20 kph at Arthur Fiord for fuel; right engine 'hiccup' but likely just water/ice in fuel line and fixed itself
23-Mar-16	10:00	13:45					3:45	1:03	Grinnell Peninsula	Sunny clear calm -32°C except severe/moderate turbulence in hills of Arthur Fiord; left generator not working so only one flight
24-Mar-16	9:05	13:20	14:25	17:35			7:25	4:59	Colin Archer Peninsula; west coast	Clear -32°C slight wind N/NW ice crystals
25-Mar-16	8:45	13:00	13:41	17:34			8:08	5:17	West coast	Clear -32°C with ice crystals/fog along south shore (unable to fly below 3000' so moved north); burning off in pm
26-Mar-16	9:08	13:35	14:15	18:11			8:23	5:28	West central	-29°C clear some cloud/ice crystals/foggy cover at south end but burned off in pm. Late start/one flight since autofeather not engaging.
27-Mar-16	10:07	12:41					2:34	0:51	YRB-YGF, some lines in between	-29°C clear, some low cloud west of transects
28-Mar-16	8:34	12:46	13:26	13:56	14:41	17:30	7:31	3:27	Truelove and east coast	-30°C calm clear, landed at Truelove cache and scraped teflon off the left ski, so no more offstrip until its back to YRB for repair
29-Mar-16	7:50	12:00	12:46	16:25			7:49	16:13	Dundas Harbor and south coast	-30°C clear calm, some cloud south over Lancaster Sound
30-Mar-16	9:54	13:05					3:11	1:35	YGF-YRB, some lines in between	-30°C clear calm

Pilots – Phil Amos, Reagan Schroeder; Navigator - Morgan Anderson

Observers:

Mar 22 – Morgan Anderson, Saroomie Manik, PJ Attagootak, James Iqaluk, Oolat Iqaluk

Mar 23 – Morgan Anderson, Saroomie Manik, PJ Attagootak, James Iqaluk, Oolat Iqaluk

Mar 24 – Morgan Anderson, Saroomie Manik, PJ Attagootak, James Iqaluk, Oolat Iqaluk

Mar 25 – Morgan Anderson, PJ Attagootak, Debbie Iqaluk, Oolat Iqaluk

Mar 26 – Morgan Anderson, PJ Attagootak, Debbie Iqaluk, Oolat Iqaluk

Mar 27 – Morgan Anderson, PJ Attagootak

Mar 28 – Morgan Anderson, Jopee Kiguktak, Aksakjuk Ningiuk, Frankie Noah, Simon Singoorie, Olaph Christianson

Mar 29 – Morgan Anderson, Jopee Kiguktak, Aksakjuk Ningiuk, Frankie Noah, Simon Singoorie, Junior Kakkee

Mar 30 – Morgan Anderson, PJ Attagootak

Appendix 5. Incidental wildlife observations.

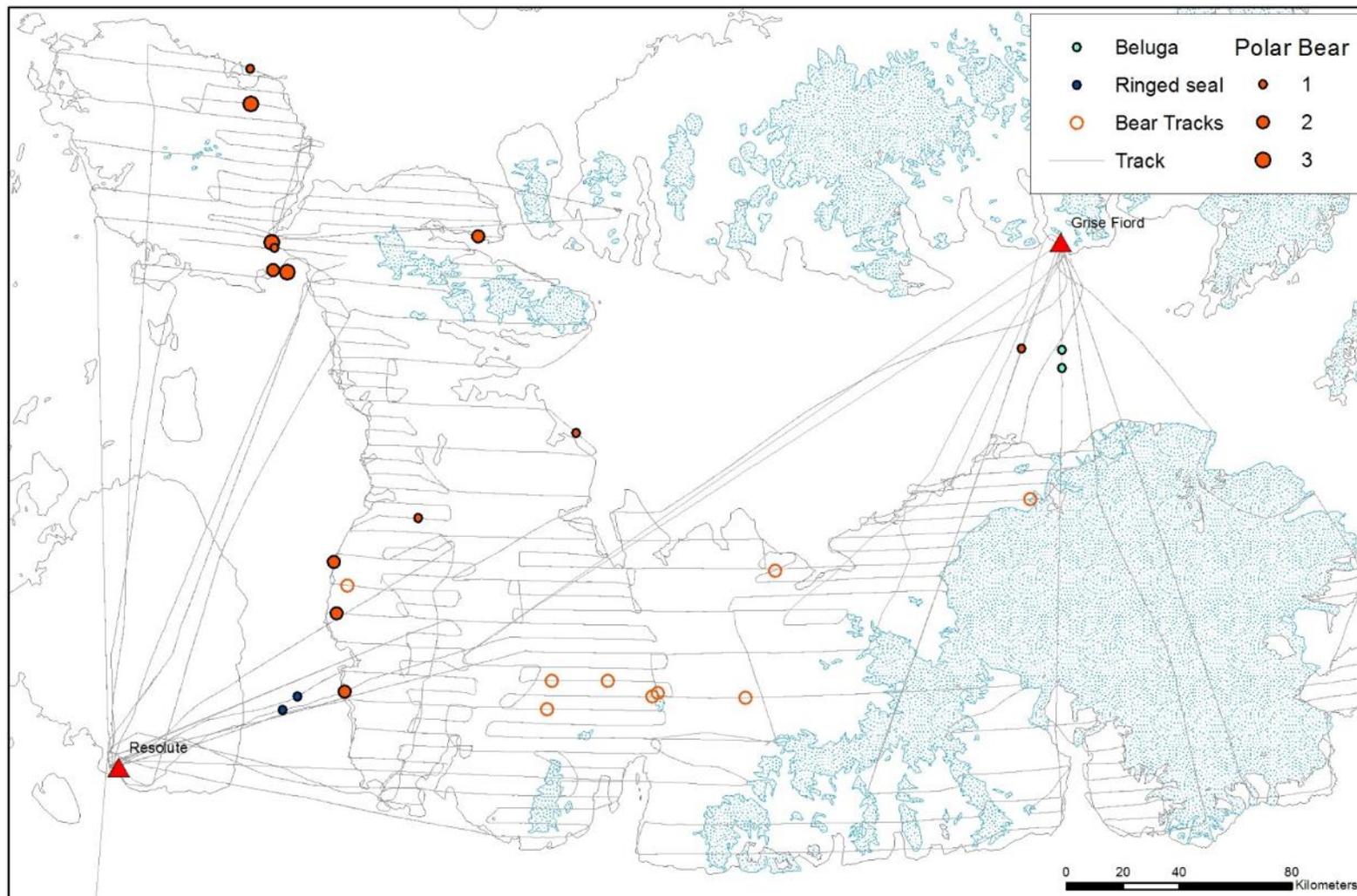


Figure 12. Incidental observations, Mar 22-30 2016, and flight lines for an aerial survey of Devon Island. Some track lines are incomplete due to loss of satellite coverage. A total of 37 polar bears were observed, as well as 5 ringed seals basking on the sea ice in Wellington Channel, and 2 groups of beluga (6 and 7 individuals) along the floe edge south of Grise Fiord. Polar bear family groups included very small cubs recently emerged from dens, and one den was seen with tracks, 40 km northwest of Maxwell Bay.



**DISTRIBUTION AND ABUNDANCE OF PEARY CARIBOU (*Rangifer tarandus pearyi*)
AND MUSKOXEN (*Ovibos moschatus*) ON DEVON ISLAND, MARCH 2016**

MORGAN ANDERSON¹

Version: 13 July 2016

¹Wildlife Biologist High Arctic, Department of Environment
Wildlife Research Section, Government of Nunavut Box 209 Igloolik NU X0A 0L0

STATUS REPORT 2016-01
NUNAVUT DEPARTMENT OF ENVIRONMENT
WILDLIFE RESEARCH SECTION
IGLOOLIK, NU

Anderson, M. 2016. Distribution and abundance of Peary caribou (*Rangifer tarandus pearyii*) and muskoxen (*Ovibos moschatus*) on Devon Island, March 2016. Nunavut Department of Environment, Wildlife Research Section, Status Report 2016-01, Igloolik, NU.

Summary

We flew a survey of Devon Island including Philpots Island (Muskox Management Zone MX-04), by Twin Otter in 58 hours between March 22 and 30, 2016, to update the population estimate for caribou and muskoxen in the study area. The previous survey, in 2008, reported a minimum count of 17 Peary caribou and population estimate of 513 muskoxen (302-864, 95%CI). The 2016 survey found the highest reported abundance estimate for muskoxen (1,963 \pm 343 SE), and a minimum count of 14 Peary caribou suggests that they continue to persist at low densities on the island, although the low number of observations precludes calculation of a reliable population estimate.

Muskoxen were abundant in the coastal lowlands where they have been found historically, at Baring bay, Croker Bay, Dundas Harbour, and the Truelove Lowlands. They were also abundant on the north coast of the Grinnell Peninsula, and particularly abundant on Philpots Island, where we observed 310 muskoxen. Although most previous surveys covered only part of Devon Island, they did target these lowlands and their abundance estimates or minimum counts likely represent the majority of the muskox population. This survey indicates a large increase in muskoxen on Devon Island, with more observations in all lowland areas compared to 2008, and a particular increase on Philpots Island. This population trend is mirrored on neighboring Bathurst Island to the west, surveyed in 2013, and southern Ellesmere Island to the north, surveyed in 2015.

We only saw 14 Peary caribou during the survey, concentrated on the north shore of the Grinnell Peninsula, and tracks were seen south of Baring Bay. No caribou were seen in the Truelove Lowlands, although hunters from Grise Fiord have caught caribou there over the past several years. It is likely that the low density and patchy distribution of caribou in this area meant that they were not detected on the survey flights. Previous surveys also found caribou in small numbers in specific locations, including a minimum count of 17 caribou in 2008 and 37 caribou on western Devon Island in 2002. Combined with the local knowledge of residents of Grise Fiord and Resolute Bay, it is likely that this population of Peary caribou remains stable at low densities, patchily distributed on Devon Island.

Management Recommendations

Peary caribou and muskoxen on Devon Island are an important source of country food and cultural persistence for Inuit. Consistent with the Nunavut Land Claim Agreement, and the Management Plan for High Arctic Muskoxen of the Qikiqtaaluk Region, 2012-2017 (DOE 2014), these management recommendations emphasize the importance of maintaining healthy populations of caribou and muskox that support sustainable harvest.

Under the Management Plan for the High Arctic Muskoxen of the Qikiqtaaluk Region, 2013-2018 (DOE 2014), Devon Island is considered a single management unit, MX-04, with a Total Allowable Harvest (TAH) of 15. The high numbers of muskox suggest that the TAH could be increased or removed, although with 3 communities harvesting from the island, maintaining a TAH might facilitate harvest management and coordination by the 3 HTAs (i.e. maintaining tags to track harvest, but setting the TAH high enough to ensure any interested hunter could receive a tag). The current TAH reflects a conservative harvest rate of 4% on a population of about 400 muskoxen, which is close to the population estimates from the 1970s until 2008. The 2016 population estimate, however, is close to four times the 2008 estimate. At the same harvest rate of 4%, 79 muskox tags could be issued. At a 5% harvest rate, 98 tags could be issued. Muskoxen do move across the barren interior of the island and among habitat patches (based on unpublished GN telemetry data, and local knowledge in Grise Fiord and Resolute), but dispersing harvest among several lowlands would prevent having to wait for muskoxen to re-establish themselves in areas that might be more isolated.

It is highly recommended that a harvest reporting system be maintained even if the TAH is removed. This would allow biologists, community members, and decision makers to track harvest patterns over time and to determine whether changes to management zones or harvest restrictions have the desired effect. With muskoxen concentrated in discrete lowland habitats that can be reliably accessed for harvesting, it may be particularly useful to distribute harvest pressure among these areas or to target under-utilized areas for larger community hunts. A large, coordinated community hunt with distributed harvest pressure in the more concentrated areas could be considered for 2016-17 with harvest monitoring recommended without TAH limitation. Following this, a TAH of 100 could be considered for future years, with ongoing population monitoring at regular intervals. As local knowledge and previous surveys have demonstrated, population changes can be rapid and unexpected if severe weather causes localized or widespread starvation or movement, so continuous monitoring and adaptive management is necessary even when populations are at high levels.

Harvest trends for muskoxen over the last decade suggest that Grise Fiord and Resolute Bay harvest fewer muskoxen than in the 1990s (Anderson 2016), but changing the configuration of management zones may encourage more harvesting in areas that were previously accessible but not included in a management unit. The major decline in caribou on Baffin Island, and subsequent harvest restrictions, has also reduced the availability of country food for Baffin communities, including Arctic Bay, which has harvested muskoxen on Devon Island in the past. The community of Arctic Bay has been in discussions with Grise Fiord to determine whether they would be able to harvest several muskoxen to offset the lack of Baffin caribou, and this should be further considered given the healthy populations of muskoxen on southern Ellesmere and Devon islands.

Although we saw only 14 caribou during the survey, the results of previous surveys over the same areas suggest that caribou have persisted at relatively low densities on Devon Island. There may or may not have been a decline from the 2008 survey, the few observations recorded from both surveys make it difficult to tell. Most caribou harvest activity from Resolute Bay has been focused on Bathurst Island, reducing the

available recent knowledge of caribou on Devon Island, although residents of Resolute still visit Devon Island for other harvesting activities and during travel. Hunters from Grise Fiord report seeing caribou fairly regularly in the Truelove Lowlands, and a few are caught there each year. It is unlikely that harvest restrictions on Peary caribou will result in any marked increase in the population, as harvest is restricted to a small human population with limited access to the caribou range, and lack of suitable habitat on Devon Island is likely a more important factor limiting caribou population growth in the area. Monitoring sightings and harvest will continue to provide a more complete picture of where caribou are on the landscape.

Research Project Updates and Proposals, July 18 2016
Grise Fiord Hamlet Building 19:30-21:30
Meeting with Iviq Hunters and Trappers Association

In attendance: Jaypetee Akeeagok (chair), Jopee Kiguktak, Amon Akeeagok, Imooshie Nutaraqjuk, Aksakjuk Ningiuk, Etuangat Akeeagok, Charlie Noah, Monasie (secretary-manager filling in for Terry Noah), Morgan Anderson.

Jaypetee introduced Morgan and the purpose of the meeting; Morgan provided an overview of research results to date and upcoming projects for comment; Monasie provided translation throughout the meeting.

Devon Island survey – Morgan showed maps of the transects and survey strata and rationale, followed by observations of muskox and caribou groups and tracks on the island and total estimates (minimum count of 14 caribou – not an estimate – and $1963 \pm SE343$ muskoxen). Concentration areas for both caribou and muskoxen were in areas where they had previously been observed, although we did not see any caribou around Truelove – they may have been missed between transects if they are at such low densities, and the report acknowledges this. Caribou are believed to be stable at low density on the island, but muskoxen have almost quadrupled from historic estimates, so we can look at changing management for muskoxen on Devon Island. Morgan proposed that the TAH could be increased from the 15 tags currently available (a conservative harvest of 5% of the population would be about 100 tags), and maintaining tags might allow multiple communities to better coordinate harvest. Alternatively, the TAH could be removed entirely, but coordination would still be important. Morgan showed the difference she found between the voluntary reporting of the Nunavut Wildlife Harvest Study and the mandatory reporting of muskox tags. More muskoxen were reported when it was a requirement, and this is important for establishing basic needs level (if it ever needed to be determined for muskoxen) and provides a good dataset for making management changes and supporting decision-making. Morgan pointed out that prior to any official changes for the Devon Island TAH through NWMB in September, if people are interested in doing a hunt, we can put through an exemption to increase the number of tags available for it.

Comments – Jaypetee suggested that the Board further discuss options for Devon Island. His personal opinion was that opening up the harvest completely could be problematic, especially if communities that are not used to hunting muskoxen might not know the best ways to harvest them responsibly. Maintaining tags but increasing the number might be a good approach. Jaypetee and Aksakjuk both reminded everyone that the muskox might be in a 'boom' right now, but that population booms are followed by busts, and we still need to be careful. Aksakjuk pointed out that increasing muskox harvest now, while their numbers are high, could be beneficial for caribou, since Peary caribou tend to be at low numbers when muskoxen are abundant. The Board will be meeting on July 21 and will further discuss.

From: Orman, Lynda
Sent: August 11, 2016 2:59 PM
To: 'hto_ab@qiniq.com'
Cc: Anderson, Morgan
Subject: Devon Island Muskox
Importance: High

Hello Grace and Jobie,

Biologist Morgan Anderson is currently out on the Prince of Wales and Sommerset Island muskox and Peary caribou survey or she would be contacting you today.

Further to Biologist Morgan Anderson's recent letter to you and then following our telephone call last week, I was wondering if the Arctic Bay HTO Board has some recommendations on muskox harvest management for Devon Island? This would be in terms of the following options which I believe Morgan discussed with you:

- 1) remove the TAH in light of good abundance; however, harvest monitoring and community coordination would not be easily facilitated without the use of tags
- 2) increase the TAH from 15 to 100 muskox with tags, which can help facilitate harvest coordination and help maintain the harvest data base for co-management
- 3) Or a combination whereby the TAH could be removed the first year (2016-17) with unlimited Inuit harvest, communities could agree to having harvest monitoring in place, and a coordinated community hunt could proceed among the three communities (Arctic Bay, Resolute Bay, Grise Fiord); following this, a TAH of 100 could be put in place for subsequent years.

The Department of Environment recognizes that it would be beneficial for Inuit to harvest Muskox on Devon Island not only for the immediate families of these communities, but perhaps also to help facilitate the supply of meat to other communities on Baffin Island where caribou country food is currently scarce. Along with option 3) for your consideration, DOE would be receptive to a request from the communities for assistance in helping to coordinate such a large community hunt in 2016-17, with an intention to try to facilitate some of the logistics in terms of getting the meat distributed among communities if the hunt was successful. While DOE is open to such a request, this would be subject to the Department finding the resources to assist in this way, if asked.

Please let me know the Board's thoughts on the above 3 options, and I would appreciate it if you could get back to me by reply email tomorrow, and/or with your Board's Resolution on this.

Many Thanks,

Qujannamiik,

Lynda Orman
Manager Wildlife Research
Department of Environment

From: Orman, Lynda
Sent: August 3, 2016 12:24 PM
To: 'hto_ab@qiniq.com'
Cc: Anderson, Morgan
Subject: HTO AB resolution on Devon Island MX

Hi Grace,

Nice to talk to you and Jobie on the phone this morning.

Here is my email contact and that of GN Biologist, Morgan Anderson, if you could please forward the resolution by the Arctic Bay HTO on Devon Island MX management action in light of population increase (fourfold) this Friday, if possible.

This would be in response to the options proposed by Biologist, Morgan Anderson, in her information letter to the Board, which Jobie said was read out at yesterday's meeting.

Qujannamiik,

Consultation Record with Grise Fiord

From: Orman, Lynda
Sent: August 11, 2016 3:16 PM
To: 'gfiviq_hta@qiniq.com'
Cc: Anderson, Morgan
Subject: FW: Devon Island Muskox
Importance: High

Hello Jaypetee and Terry,

Biologist Morgan Anderson is currently out on the Prince of Wales and Sommerset Island muskox and Peary caribou survey or she would be contacting you today.

Further to Biologist Morgan Anderson's recent meeting with you, I was wondering if the Grise Fiord HTA Board has some recommendations on muskox harvest management for Devon Island? This would be in terms of the following options which I believe Morgan discussed with you:

- 1) remove the TAH in light of good abundance; however, harvest monitoring and community coordination would not be easily facilitated without the use of tags
- 2) increase the TAH from 15 to 100 muskox with tags, which can help facilitate harvest coordination and help maintain the harvest data base for co-management
- 3) Or a combination whereby the TAH could be removed the first year (2016-17) with unlimited Inuit harvest, communities could agree to having harvest monitoring in place, and a coordinated community hunt could proceed among the three communities (Arctic Bay, Resolute Bay, Grise Fiord); following this, a TAH of 100 could be put in place for subsequent years.

The Department of Environment recognizes that it would be beneficial for Inuit to harvest Muskox on Devon Island not only for the immediate families of these communities, but perhaps also to help facilitate the supply of meat to other communities on Baffin Island where caribou country food is

currently scarce. Along with option 3) for your consideration, DOE would be receptive to a request from the communities for assistance in helping to coordinate such a large community hunt in 2016-17, with an intention to try to facilitate some of the logistics in terms of getting the meat distributed among communities if the hunt was successful. While DOE is open to such a request, this would be subject to the Department finding the resources to assist in this way, if asked.

Please let me know the Board's thoughts on the above 3 options, and I would appreciate it if you could get back to me by reply email tomorrow, and/or with your Board's Resolution on this.

Many Thanks,

Qujannamiik,

Lynda Orman
Manager Wildlife Research
Department of Environment



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Lynda Orman
Manager Wildlife Research
Wildlife Management Division
Department of Environment
Government of Nunavut

Lynda Orman
Atannguyaq Uumayunik Qauyihaqti
Uumayunik Aulapkainikkut Havakvik
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867-934-2183

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lorman@gov.nu.ca

Consultation Record with Resolute

From: Orman, Lynda
Sent: August 11, 2016 3:20 PM
To: 'rbhta@qiniq.com'
Cc: Anderson, Morgan
Subject: FW: Devon Island Muskox
Importance: High

Hello Philip and Nancy,

Biologist Morgan Anderson is currently out on the Prince of Wales and Sommerset Island muskox and Peary caribou survey or she would be contacting you today.

-----Original Message-----

From: Anderson, Morgan
Sent: August 6, 2016 11:23 AM
To: Orman, Lynda
Subject: RE: HTO AB resolution on Devon Island MX

....

I heard back from Resolute that they support raising the TAH to 100 and they'll do a resolution to that effect, so that's good, and I'll follow up with Arctic Bay.

From: Orman, Lynda
Sent: Friday, August 05, 2016 10:25 AM
To: Anderson, Morgan
Subject: RE: HTO AB resolution on Devon Island MX

Hi Morgan,
Jobie said they read out your letter at the meeting and discussed it (supportive) and will make their resolution to send today. So I don't actually know what it is yet and in conversation I just referred to the options you had laid out in your letter to them, not wanting to influence them one way or the other.

Please follow-up with Grace/ Jobie and see if you can receive their resolution today?

Thank you,
Lynda

....

-----Original Message-----
From: Anderson, Morgan
Sent: August 5, 2016 10:15 AM
To: Orman, Lynda
Subject: RE: HTO AB resolution on Devon Island MX

Hi Lynda - what was the resolution that AB passed? I've requested Grise and Resolute also submit board resolutions and Grise was intending to but I haven't heard if they've submitted anything.

Morgan

From: Orman, Lynda
Sent: Wednesday, August 03, 2016 12:26 PM
To: Anderson, Morgan
Subject: RE: HTO AB resolution on Devon Island MX

Hi Morgan,

Similarly, if you're also able to obtain HTO resolutions from Grise and Resolute, we can send those resolutions with your Devon MX BN NWMB and Report abstract/discussion to NWMB (deadline next week).

Jobie called this morning and said that their HTO would be able to send us the resolution by this Friday.

Thanks,



SUBMISSION TO THE
NUNAVUT WILDLIFE MANAGEMENT BOARD

FOR

Information: X

Decision:

Issue: 2016 field season up-date for the M'Clintock Channel polar bear genetic biopsy study

Background:

M'Clintock Channel (MC) is a smaller polar bear subpopulation managed by Nunavut (Figure 1). This subpopulation is currently hunted by residents of Gjoa Haven and Cambridge Bay with a Total Allowable Harvest (TAH) of 12 bears per year. An initial mark-recapture study was done from 1973-78 (Furnell and Schweinsburg, 1984) for MC and Gulf of Boothia (GB), but it did not identify them as individual demographic units. However, a summed population estimate for both areas of 1081 was derived. The estimate was known to be biased by non-representative sampling, and was subsequently increased to 900 for GB and 900 for MC based on the belief that the current harvests were sustainable, and the estimated number was the one required to sustain the harvest.

In the mid-1990s, the MC estimate was revised downwards to 700 based on hunter reports of reduced densities of polar bears. Both populations were later delineated based on movements and recoveries of tagged bears, movements of satellite radio-collared adult female bears in adjacent areas (Taylor and Lee, 1995; Taylor et al., 2001), and local knowledge of Inuit about how local conditions may influence the movements of polar bears. Past harvests of 34 bears/year from 1979-1999 were unsustainable, and a moratorium from 2001/2002 – 2003/2004 was implemented, followed by a reduction in TAH down to 3 bears per year (2004 – 2014), and subsequent increase to 12 bears per year for the 2015/16 harvest season. The subpopulation has been managed to achieve recovery, and in fact local traditional knowledge confirms that there are more bears being seen in recent years. The past abundance estimate for MC, based on a physical mark-recapture study (1998-2000) was 284 bears (Taylor et al. 2006). At such low abundance levels, the population remains at risk.

In accordance with commitments under the 2005 Polar Bear Memorandum of Understanding (MOU) for MC and in response to concerns regarding the subpopulation status, harvest opportunities and the temporary loss of traditional and cultural opportunities, DOE proposed to undertake a 3-year study (2014-2016) involving genetic mark-recapture to reassess the size and status of the MC polar bear subpopulation.

Current Status:

DOE began a new 3-year study in 2014 which utilized a less-invasive technique given the expressed concerns by hunters and HTOs over handling of wildlife. DOE staff used genetic biopsy sampling, similar to the studies conducted in Baffin Bay and Kane Basin, which requires the collection of a small skin tissue sample through a biopsy dart. This genetic tissue then becomes the mark and can be used to identify the animal in subsequent recaptures through DNA analysis. If insufficient marks are recaptured in years 2 and 3 of the study, a fourth year may be required in order to increase the robustness of the population estimate. A population estimate is derived over multiple years of genetic mark-recapture analysis.

The 2016 field season lasted from 1 April until 8 June with field operations occurring between 19 April and 7 June. During the study, a total of 98 polar bears in 65 groups of both genders and varying age groups were encountered, of which 85 biopsy samples were obtained for analyses.

All tissue samples are currently being analysed by a special wildlife lab where results are expected later this year. A 2016 field report will be prepared and made available to all co-management and funding partners in due time. Upon receiving the genetic results, analyses of the 3-year data set will commence so that a final report can be produced and finalized by winter of 2017.

Consultations:

Community consultations were held with HTO representatives from Taloyoak, Gjoa Haven, and Cambridge Bay in 2013 about the planned study. Over the years, results have been presented and reported back, either written, orally or both to the Regional Wildlife Organization, the communities, and the co-management partners.

References:

- Furnell, D.J., Schweinsburg, R.E. 1984. Population dynamics of central Arctic polar bears. *Journal of Wildlife Management* 48: 722-728.
- Taylor, M.K., Lee, L.J. 1995. Distribution and abundance of Canadian polar bear populations: a management perspective. *Arctic* 48: 147–154.
- Taylor, M.K., Akeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* 79: 690–709.
- Taylor, M.K., Laake, J.L., McLoughlin, P.D., Cluff, H.D., Messier, F. 2006. Demographic parameters and harvest-explicit population viability analysis for polar bears in M'Clintock Channel, Nunavut. *Journal of Wildlife Management* 70: 1667–1673.

Date: 12 July 2016

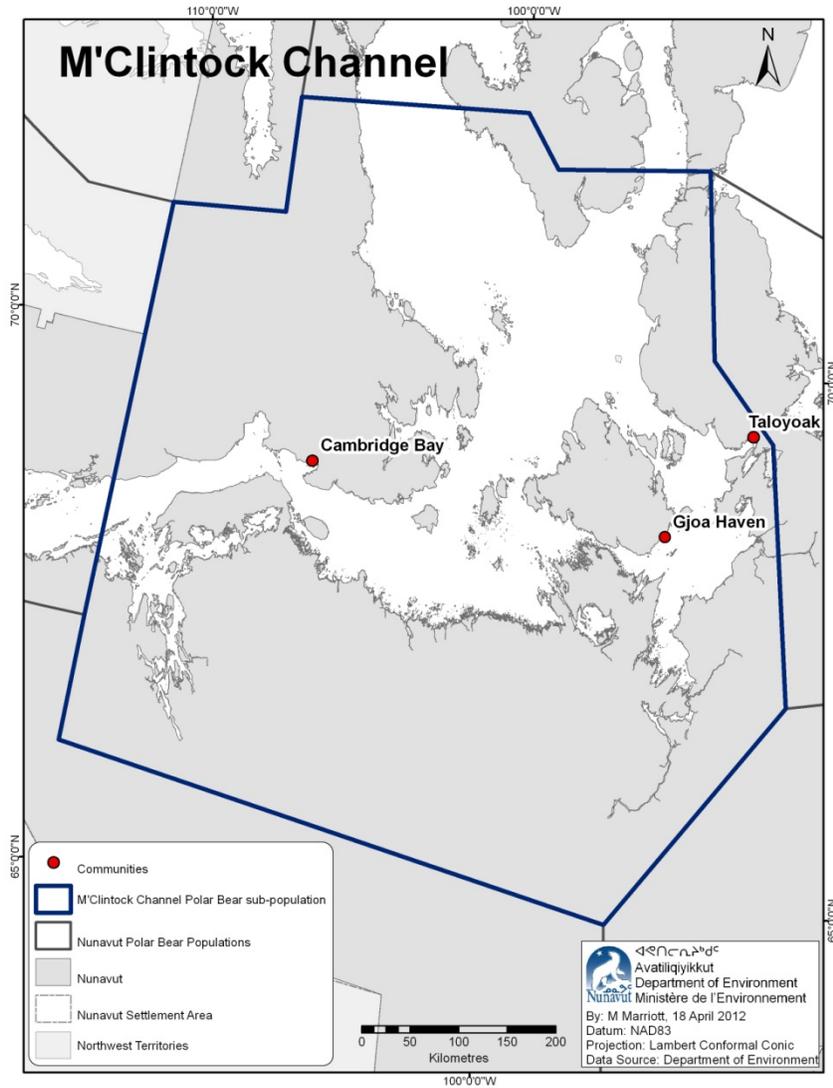


Figure 1. Map of the M'Clintock Channel polar bear sub-population boundary and location of communities that harvest from within.



SUBMISSION TO THE

NUNAVUT WILDLIFE MANAGEMENT BOARD

FOR

Information: X

Decision:

Issue: 2016 field season up-date for the Gulf of Boothia Polar Bear subpopulation

Background:

An initial mark-recapture study was done from 1973-78 (Furnell and Schweinsburg, 1984), where M'Clintock Channel (MC) and Gulf of Boothia (GB) were not identified as demographic units. However, a summed population estimate for both areas of 1081 was derived. The estimate was known to be biased by non-representative sampling, and was subsequently increased to 900 for GB and 900 for MC based on the belief that the current harvests were sustainable, and the estimated number was the one required to sustain the harvest. In the mid-1990s, the MC estimate was revised downwards to 700 based on hunter reports of reduced densities of polar bears. Both populations were later delineated based on movements of satellite radio-collared adult female bears in adjacent areas, tag returns of harvested bears (Taylor and Lee, 1995; Taylor et al., 2001), and local knowledge of Inuit about how local conditions may influence the movements of polar bears.

The last population inventory work for GB was completed in 2000, where abundance for GB (mean \pm SE) was estimated to be 1592 ± 361 polar bears (Taylor et al., 2009). Based on the latest GB inventory and results, the population is considered to be stable or very likely increasing due to high recruitment and survival rates (COSEWIC, 2008; Taylor et al., 2009); however, caution is warranted regarding long-term trends especially when one considers observed environmental changes in other polar bear populations (e.g., Foxe Basin [Sanahatian and Derocher, 2012]; Baffin Bay and Davis Strait [Rode et al., 2012]; Western Hudson Bay [Regehr et al., 2007a]; Beaufort Sea [Regehr et al., 2007b]). The GB sub-population is currently harvested at an annual TAH of 74 bears (mean harvest of 61 bears between 2005/06 – 2010/11). Although no genetic similarities between these two populations were detected in a past genetic study (Campagna et al. 2013; Paetkau et al., 1999), in recent years, the distinction between the MC/GB subpopulations has been put into question by Inuit hunters (Keith et al., 2005), and new genetic analyses (Saunders, 2005) suggesting some considerable interchange between both subpopulations.

Current Status:

The current data for GB are dated, and in accordance with commitments under the 2005 Polar Bear Memorandum of Understanding (MOU) for GB, DOE proposed to undertake a 3-year study (2015-2017) involving genetic mark-recapture to reassess the size and status of the GB polar bear subpopulation.

Between 19 April and 14 May 2016, a total of 161 bears in various age and sex groups were encountered on the sea ice, of which 120 were biopsied. HTO representatives from Taloyoak and Kugaaruk participated in the field operations. Tissue samples have been submitted to a specialized wildlife lab where genetic results are expected to be available in October, 2016. A field report is in preparation and will be distributed to all co-management and funding partners.

Consultations:

Community consultations were held with HTO representatives from Taloyoak, Gjoa Haven, Igloolik, Repulse Bay, Kugaaruk, and Hall Beach in 2013 about the planned study. The results of last year's field season have been presented and reported back, either written, orally or both to the Regional Wildlife Organization, the communities, and the co-management partners.

References:

- Campagna, L., Van Coeverden de Groot, P.J., Saunders, B., Atkinson, S., Weber, D., Dyck, M.G., Boag, P.T., Lougheed, S.C. 2013. Extensive sampling of Polar Bears (*Ursus maritimus*) in the Northwest Passage (Canadian Arctic Archipelago) reveals population differentiation across multiple spatial and temporal scales. *Ecology and Evolution* 3: 3152-3165.
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- Keith, D.J. Arqvig, L. Kamookak, J. Ameralik and the Gjoa Haven Hunters' and Trappers' Organization. 2005. Inuit Qaujimaningit Nanurnut, Inuit Knowledge of Polar Bears. Gjoa Haven Hunters' and Trappers' Organization and CCI Press, Edmonton, Alberta.
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- Regehr, E.V.S., Lunn, N.J., Stirling, I., and Amstrup, S.C. 2007a. Effects of earlier sea ice breakup on survival and population size of polar bears in western Hudson Bay. *Journal of Wildlife Management* 71:2673-2683.

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- Rode, K.D., Peacock, E., Taylor, M., Stirling, I., Born, E.W., Laidre, K.L., and Wiig, Ø. 2012. A tale of two polar bear populations: ice habitat, harvest, and body condition. *Population Ecology* 54:3-18.
- Sanahatien, V., and Derocher, A. E. 2012. Monitoring sea ice habitat fragmentation for polar bear conservation. *Animal Conservation*. doi:10.1111/j.1469-1795.2012.00529.x.
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- Taylor, M.K., J.L. Laake, P.D. McLoughlin, H.D. Cluff, and F. Messier. 2009. Demography and population viability of polar bears in the Gulf of Boothia, Nunavut. *Marine Mammal Science* 25:778-796.
- Taylor, M.K., Lee, L.J. 1995. Distribution and abundance of Canadian polar bear populations: a management perspective. *Arctic* 48: 147–154.
- Taylor, M.K., Akeeagok, S., Andriashek, D., Barbour, W., Born, E.W., Calvert, W., Cluff, H.D., Ferguson, S., Laake, J., Rosing-Asvid, A., Stirling, I., Messier, F. 2001. Delineating Canadian and Greenland polar bear (*Ursus maritimus*) populations by cluster analysis of movements. *Canadian Journal of Zoology* 79: 690–709.

Date: 12 July 2016

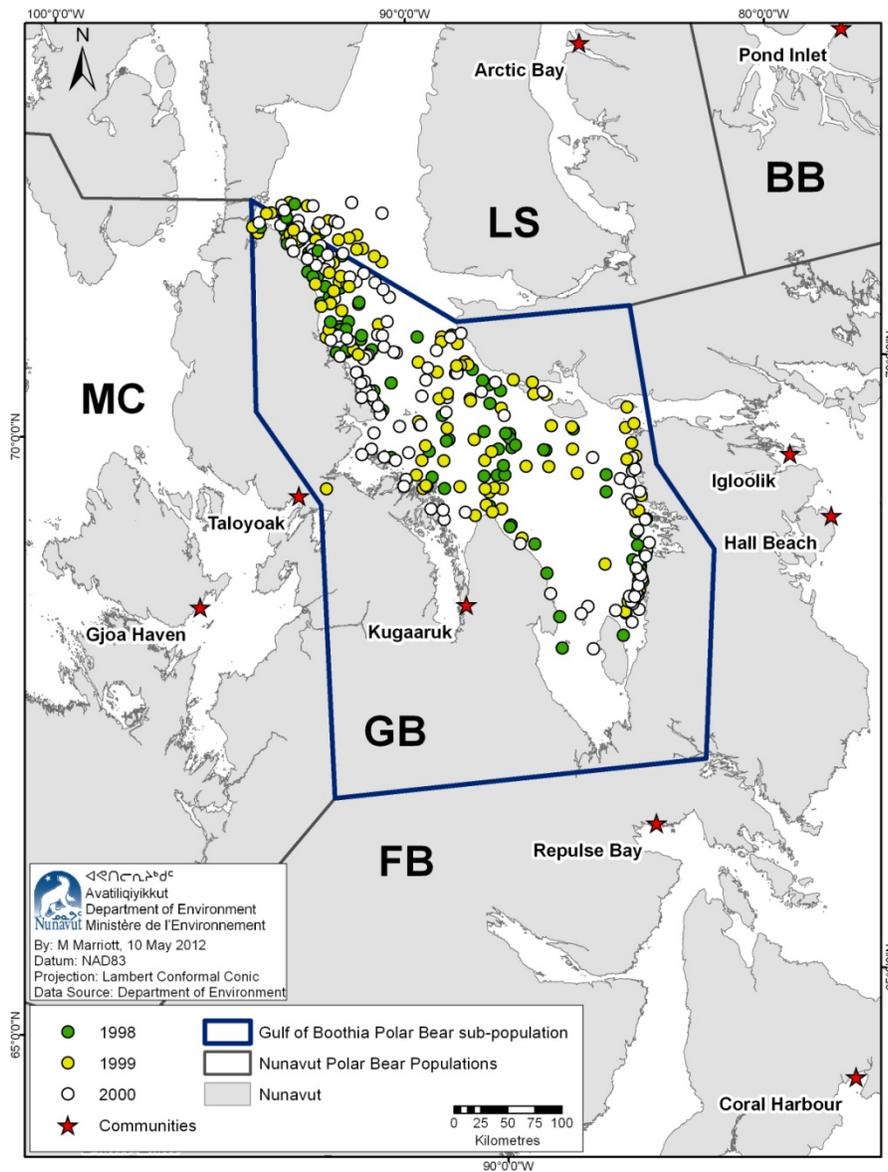


Figure 1. Map of the and Gulf of Boothia (GB) polar bear subpopulation boundaries and location of Nunavut communities (NB: previous capture locations from the past inventory study are shown as well).



SUBMISSION TO THE

NUNAVUT WILDLIFE MANAGEMENT BOARD

FOR

Information: X

Decision:

Issue: 2016 field season up-date for the western Hudson Bay polar bear aerial survey

Background:

The Western Hudson Bay (WH) polar bear subpopulation (Figure 1) is one of the most intensively studied large carnivore populations in the world, with research activities dating back to the late 1960s (Jonkel et al., 1972; Stirling et al., 1977; Derocher and Stirling, 1995a, b; Regehr et al., 2007). Although WH is shared among several jurisdictions (i.e., Nunavut, Manitoba, and Parks Canada), research activities to obtain abundance estimates were primarily conducted by Environment Canada. Through long-term mark-recapture and tagging studies and the collection of traditional local knowledge, it appears that the subpopulation increased throughout the 1960s and 1970s, remained stable for part of the 1980s, and declined between 1984 and 2004 by approximately 22% (Derocher and Stirling, 1995a; Lunn et al., 1997; Regehr et al., 2007, Tyrrell 2006; Nirlungayuk and Lee 2009). Population viability analyses using the 2004 demographic data predicted that the WH subpopulation would further decline in abundance. However, the aerial survey conducted in 2011 and the most recent EC mark recapture analysis has shown the population to be stable (Lunn et al. 2016).

The decline in abundance of WH bears has been debated in recent years, fueled in part by apparent differences between scientific and local knowledge. Local knowledge, including reports of more widespread encounters with polar bears, suggested that bear abundance had not decreased (Tyrell 2006), despite the scientific findings. Scientific information has also identified declines in survival rates, reproductive output and body condition, correlating these changes with reductions in sea ice extent and duration (Stirling et al. 1999, Regehr et al. 2007). In response to this debate, the uncertainty about the status of this population, and also to satisfy the need to advance non-invasive monitoring tools for polar bears in contrast to the traditional handling and capture activities, the Government of Nunavut, Department of Environment, in collaboration with Manitoba Conservation, Parks Canada, Environment and Climate Change Canada, and the University of Minnesota tested and implemented an aerial survey during August of 2011. Results from this comprehensive aerial survey indicated that with an abundance estimate of 1030 bears (95% CI: 754-1406) the subpopulation had not declined to levels that were previously predicted (Stapleton et al. 2014). A more recent study by

Environment Canada that analysed data between 1984 and 2011 also suggests that the WH subpopulation remained stable between 2000 and 2011 (Lunn et al. 2016). Their abundance estimate of 806 bears (95% CI: 715-1398) is consistent with the results of the aerial survey.

The collection of a full complement of demographic data through traditional mark-recapture studies provides opportunities to predict how a population will likely perform, given a set of met assumptions. Aerial surveys as non-invasive methods do not provide these options since they are a “snap-shot” in time. As a consequence, only several consecutive aerial surveys can aid in determining a population trend. As the Government of Nunavut has moved toward non-invasive research for some of Nunavut’s polar bear subpopulations, a more frequent monitoring of these subpopulations is required to provide meaningful input into adaptive decision-making, especially in light of environmental change and the uncertainty of how populations will be affected by it.

Current Status:

DOE is conducting a new aerial survey for western Hudson Bay during the summer (August – September) of 2016 in collaboration with Parks Canada, Manitoba Department of Sustainable Development, Environment and Climate Change Canada (ECCC). The preparations have been ongoing since earlier this year.

This aerial survey will provide another abundance estimate for polar bears of western Hudson Bay, and together with the 2011 study and the recent ECCC results, it will be possible to re-assess the status of this population. All the information collected throughout the survey (e.g., body condition, number of offspring, etc.) will be used by decision makers when assessing the Total Allowable Harvest levels.

Consultations:

Consultations for this planned survey occurred through meetings at the Kivalliq Wildlife Board Meeting during the fall of 2015, and in-community consultation meetings during the spring (Feb - Mar 2016).

References:

- Derocher, A.E., Stirling, I., 1995a. Estimation of polar bear population size and survival in western Hudson Bay. *J. Wildl. Manage.* 59, 215–221.
- Derocher, A.E., Stirling, I., 1995b. Temporal variation in reproduction and body mass of polar bears in western Hudson Bay. *Can. J. Zool.* 73, 1657–1665.
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- Jonkel, C.J., Kolenosky, G.B., Robertson, R.J., Russell, R.H., 1972. Further notes on polar bear denning habits. *Int. Conf. Bear Res. Manage.* 2, 142–158.
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Date: 12 July 2016

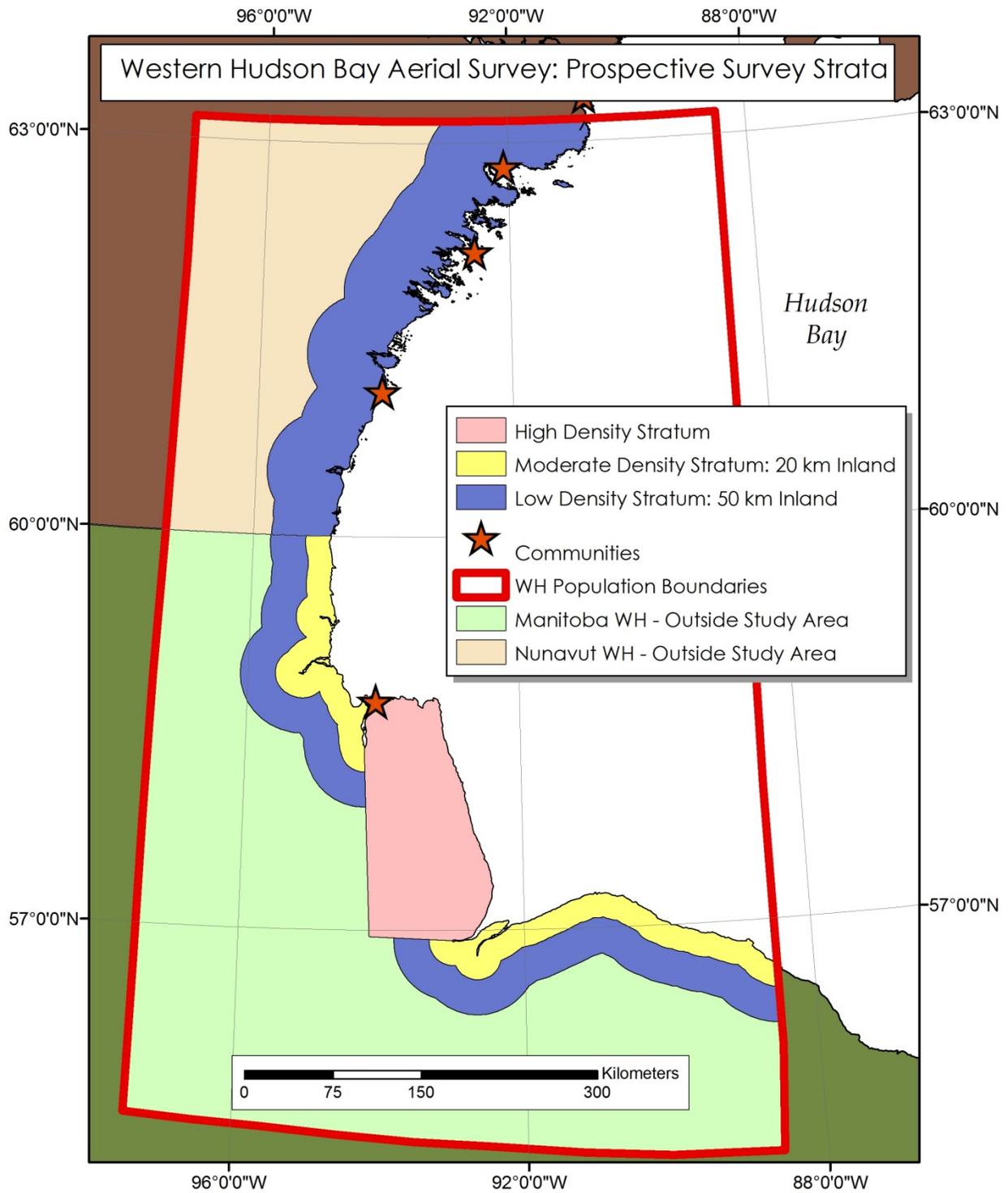


Figure 1. Proposed survey strata for the 2016 Western Hudson Bay aerial survey (as per Stapleton et al. 2014).



**SUBMISSION TO THE
NUNAVUT WILDLIFE MANAGEMENT BOARD**

FOR

Information: X

Decision:

Issue: Preliminary reports and investigation of a die-off of caribou on Prince Charles Island

Background:

Prince Charles Island is a large, mostly flat island in Foxe Basin where a large proportion of Baffin Island caribou have been seen on recent surveys. In 2014, an aerial survey of the Baffin Island caribou range estimated 4,652 caribou on Baffin Island (3,462-6,250 95% CI). Of this, 1,603 (95%CI=1158-2220) were estimated on Prince Charles Island, about a third of the total population of Baffin Island caribou.

In the first week of July 2016, a Canadian Wildlife Service crew looking at sites for bird research on Prince Charles Island, encountered several caribou carcasses on the central eastern part of the island. A local observer from Cape Dorset was also present with the crew. At the time, they estimated over 30 carcasses, but had not immediately recognized the dark shapes surrounded by white rings of hair as dead caribou and had not been counting. They also reported seeing relatively few live caribou, fewer than 20, during their time on the island. While not a survey to estimate population, they were surprised to have seen so few on an island that had previously had high densities of caribou. They reported their sightings to the Department of Environment in Igloolik, and two biologists visited the esker ridge where the carcasses had been found on July 8 by helicopter. Limited helicopter time resulted in a short visit to the area, where 47 carcasses were found and 18 sampled. Biologists collected hind legs, skin, blood, tissue, fecal pellets, photographs and noted location, sex, age, and condition of the carcasses. There was minimal scavenging and many carcasses were in a position as if bedded, while others were on their sides as if they had fallen over from a standing position. Males and females and adults, old animals, and young animals were all represented in the dead animals sampled. Carcasses were in an advanced state of decay, so samples of internal organs could not be obtained. Carcasses were on both sides of a 15-30 m high esker, 1.5-2 km long, with some further northeast. The biologists did not see any live caribou in their limited time on the island.

The most recent survey work, spring composition surveys in April 2016, did not indicate large numbers of dead caribou, and biologists did not have difficulty locating groups of caribou. Combined with the advanced decomposition of the carcasses, the event that triggered the die-off is therefore believed to have been

in late spring, possibly linked to heavy snow and icing conditions, although further investigations of weather conditions on Prince Charles Island have not been completed to date. Similar weather-related die-offs have been recorded for island-living caribou when forage becomes inaccessible under hard packed snow or ground-fast ice, leading to long-distance movements and/or starvation.

Current Status:

Investigations into the cause and extent of the die-off and health of the deceased caribou is on-going.

Consultations:

No consultations are currently planned. Co-management partners and community members will be provided with results of analyses from sampled carcasses and with any additional updates as the situation is further investigated.