ᓄᓇᕗᑦ ᐅᒪᔨᓕᕆᔨᖅᑎᑦ ᑲᑎᒪᔨᖏᑦ
ᑲᑎᒪᓵᑎᕐᓂᑦ: ᑲᑎᒪᐃᓐᓇᕐᓂᑎᑐᒍᑦ ᑲᑎᒪᓂᐅᕗᑦ ᓱᓇᐅᑉᑕ RM 002-2019
ᐱᓕᕐᓂᐊᓕ, ᔪᓂ 19, 2019 (9:00 ᐅᓪᓚᑦᓂᑦ ᑎᑭᓪᓗᒍ 7:45 ᐅᓪᓄᑦᓂᑦ)
ᐊᑐᐊ, ᐊᓕᑎᐅᓚᐅᓪᓕ

ᐃᑲᖏᕐᓗ	No:	ᑲᑎᒪᓵᑎᕐᓂᖅ:	Tab:	ᐅᖃᓪᓚᕆᓂᐊᖅᑐᖅ	ᐊᑯᓂᐅᓂᓗ
9:00 AM to 9:05 AM	1	ᑲᑎᒪᒋᐊᓪᓂᖅ		ᐃᓕᓴᐅᑕᖅ	5 ᒥᓂᔪ
9:05 AM to 9:10 AM	2	ᒪᑐᐃᖅᓯᔨᑎᑦ ᐊᒻᒪ ᑭᓚᑦᑯᓂᕐᓂᓂᑦ ᓚᓗᓚᐃᖅᓯᓂᖅ		ᐃᓕᓴᐅᑕᖅ	5 ᒥᓂᔪ
9:10 AM to 9:15 AM	3	ᑲᑎᒪᓵᑎᕐᓂᑦ: ᖃᒥᖅᑐᕆᐅᓂᖏᑦ ᐊᒻᒪ ᐊᖏᖅᑕᐅᓂᖏᑦ	1	ᐃᓕᓴᐅᑕᖅ	5 ᒥᓂᔪ
		ᐱᓕᓚᐃᖓᑦ ᐊᐸᑎᓕᓚᓵᓕᑯᑦ - GN (ᓄᓇᕗᑦ ᓴᐸᒪᑯᖏᑦ - ᐊᐸᑎᓕᓚᓵᓕᑯᑦ): ᐱᓵᔪᑎᑦ/ᐃᓱᒪᓕᐅᕈᑎᑦ			
9:15 AM - 10:00 AM	4	ᐊᖅᑭᒋᐊᖅᑕᐅᓂᖓ ᑲᑎᓵᖓᓱᑦ ᐅᒪᔨᖅᑕᐅᔨᓚᖅᑐᑦ ᑭᓯᓂᖓ ᓴᓕᓂᒥᐅᑦ ᖃᑭᖅᑕᖓᓂᓄᑦ ᑐᑦᑐᒍᓴᑦ	2	ᓄᓇᕗᑦ ᓴᐸᒪᑯᖏᑦ - ᐊᐸᑎᓕᓚᓵᓕᑯᑦ	45 ᒥᓂᔪ
		ᖃᑲᑲᐃᓚᖅ			15 ᒥᓂᔪ
10:15 AM - 10:45 AM	5	ᐊᖅᑭᒋᐊᖅᑕᐅᓂᖓ ᑲᑎᓵᖓᓱᑦ ᐅᒪᔨᖅᑕᐅᔨᓚᖅᑐᑦ ᑭᓯᓂᖓ ᐅᒥᖓᓚᓵᓂᐊᖃᖏᓐᓂᒥ MX-08	3	ᓄᓇᕗᑦ ᓴᐸᒪᑯᖏᑦ - ᐊᐸᑎᓕᓚᓵᓕᑯᑦ	30 ᒥᓂᔪ
10:45 AM - 11:15 AM	6	ᐊᖅᑭᒋᐊᖅᑕᐅᓂᖓ ᑲᑎᓵᖓᓱᑦ ᐅᒪᔨᖅᑕᐅᔨᓚᖅᑐᑦ ᑭᓯᓂᖓ ᐅᒥᖓᓚᓵᓂᐊᖃᖏᓐᓂᒥ MX-09	4	ᓄᓇᕗᑦ ᓴᐸᒪᑯᖏᑦ - ᐊᐸᑎᓕᓚᓵᓕᑯᑦ	30 ᒥᓂᔪ
		ᖃᓄᓕᖅᑐᖅ ᐊᖏᔪᐊᖅᑎᑦ ᑲᑐᔾᔨᖅᑎᑑᖏᑦ: ᐱᓵᔪᑎᑦ/ᐃᓱᒪᓕᐅᕈᑎᑦ			
11:15 PM - 11:45 PM	7	Bluenose ᑐᓵᑐᖏᓪᓕᑦ ᑲᓇᖏᓇᓕᓂ ᓄᓇᓕᖏᓂ ᑐᓵᑐᓕᓚᓂᖓᔪᑦ ᐸᖏᓇᐅᑎᑦ	5	ᓄᓇᓕᖏᓂ ᐊᖏᔪᓱᓵᑐᓕᓚᓵᓕᑯᑦ	30 ᒥᓂᔪ
		ᐅᓪᓗᑭᒥᑦᓇᖅ			ᐃᑲᖏᔪᒍᑦ &15 ᒥᓂᔪ
		ᖃᑭᖅᑕᓪᓗᐣ ᐅᒪᔨᓕᓚᓵᔨᑦ ᑲᑎᒪᓵᑎᕐᓂᑦ (QWB): ᐱᓵᔪᑎᑦ/ᐃᓱᒪᓕᐅᕈᑎᑦ			
1:00 PM - 1:30 PM	8	ᐊᖅᑭᒋᐊᕈᑦ ᑲᑎᓵᖓᓱᑦ ᐅᒪᔨᖅᑕᐅᔨᓚᖅᑐᑦ ᑭᓯᓂᖓᓵᓂᑦ ᖃᑭᖅᑕᓴᓚᒥ ᑐᑦᑐᒍᓵᓂᑦ	6	QWB	45 ᒥᓂᔪ
		ᖃᑎᖅᑎᐅᑦ ᐅᒪᔨᓕᓚᓵᔨᑦ ᑲᑎᒪᓵᑎᕐᓂᑦ (KRWB)			
1:30 PM - 2:15 PM	9	ᑐᑦᓯᕆᐅᑎᑦ ᑕᓀᓴᐃᓐᓂᓵᓂᑦ ᐊᖏᔪᓀᓯᐊᖅᑎᑕᐅᔨᑦ ᑕᑦᓕᒪᓵᑦ ᐊᑭᓚᑲᐃᓐᑦ ᓂᕆᖏᑕᒃᓯᓵᓂᑦ ᐱᔨᒪᑦᓗᑎᑦ	7	KRWB	30 ᒥᓂᔪ
		ᐃᖅᑲᓕᓚᓵᔨᓀᑦ ᐊᒻᒪ ᐃᒪᖏᐅᑕᓕᓚᓵᔨᓀᑦ ᑲᓇᑕᒥ (ᐃᒪᖏᐅᑕᓕᓚᓵᔨᓀᑦ): ᐱᓵᔪᑎᑦ/ᐃᓱᒪᓕᐅᕈᑎᑦ			
2:15 PM - 3:00 PM	10	ᑲᑎᓵᖓᓱᑦ ᐃᒪᖏᐅᑕᑦ ᒥᐊᓂᓴᐅᓂᖓᓵᓂᑦ ᐸᖏᓇᐅᑎᑦ ᖃᓕᕆᓕᓵᓂᑦ-2019	8	ᐃᒪᖏᐅᑕᓕᓚᓵᔨᓀᑦ	45 ᒥᓂᔪ
		ᖃᖃᑲᑲᐃᓚᖅ			15 ᒥᓂᔪ
		ᐃᖃᓗᓕᓚᓵᔨᓀᑦ ᐊᒻᒪ ᐃᒪᖏᐅᑕᓕᓚᓵᔨᓀᑦ ᑲᓇᑕᒥ (ᐃᒪᖏᐅᑕᓕᓚᓵᔨᓀᑦ) - ᑲᔪᓯᔪᑦ			
3:15 PM - 4:00 PM	11	ᓄᓇᕗᑦ ᐊᒻᒪᓗ ᓄᓇᓖᔨ ᐃᑭᖏᑦ ᐃᖃᓗᓕᓕᐊᑲᓇᖏᑦᑕ ᑐᓂᔭᐅᓂᖓᓵᓂᑦ Davis Strait ᐱᓪᓚᓐᓇᓚᓂ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᓖᔨ ᑲᓇᓐᓇᓚᓂ--ᒥᓐᓇ ᐋᖅᑕᐅᓯᒪᔭᖅᑐᖅ ᑐᓂᔭᐅᓂᖓᓵᓂᐅᖅᓯᒪᔪᓕᑦ	9	ᐃᒪᖏᐅᑕᓕᓚᓵᔨᓀᑦ	45 ᒥᓂᔪ

4:00 PM - 4:30 PM	12	ᖃᑎᖅᑕᐅᓂᖏᑦ ᓄᓇᕗᑦ ᐊᒃᓚᒃ ᓄᓇᕕᒃ ᐃᖅᓯᓪᓚᓂᖅᑎᑦ ᐊᑕᐃᔪᑦ ᐊᒥᐊᓪᓚᖅᑕ ᓄᓇᖅᑲᑎᓐᖏᑦᑎᖅᓂᓄᑦ ᐊᐅᓪᓪᓗᖅᑎᓯᓂᖅᓯᒪᑦ ᒪᓪᓚᖏᑎᑦ ᐃᓕᓴᓪᓕᑦ ᐃᒪᖏᐅᑦ ᓴᓇᓕᕆᓂᖅᑎᑦ ᐱᖏᕐᕋᐊᖏᑎᑎᒍᑦ(MMPA)	10	ᐃᒪᖏᐅᑕᓪᓚᓱᓕᑦ	30 ᒥᓂᔪ
4:30 PM - 5:00 PM	13	ᐱᓇᓴᐅᓂᖏᑦ ᓄᓇᕗᒥ ᐃᑭᖏᑦ ᐃᖃᓗᓪᓕᐊᖅᑐᓪᓚᓂᖅᓯᒪᑦ ᒪᓪᓚᕆᓪᓕᖏᑦ	11	ᐃᒪᖏᐅᑕᓪᓚᓱᓕᑦ	30 ᒥᓂᔪ
5:00 PM - 5:45 PM	14	ᑐᙵᐃᕕᐃᑦᑐᖅ ᐃᑭᖅ ᓴᐳᒥᐅᓴᖅᑕᐅᔨᒪᓕᓪᓗᓂ ᑐᓴᓴᕆᐅᑎ	12	ᐃᒪᖏᐅᑕᓪᓚᓱᓕᑦ	45 ᒥᓂᔪ
		ᖁᑲᐃᐧᓇᖅ			15 ᒥᓂᔪ
		ᐊᙵᑎᓪᓚᓱᓕᑦ ᐊᒪᒪ ᓱᓪᓕᐅᑦ ᐊᓱᓴᖅᑕᓪᓕᐊᓂᖅᓯᒪᓂᓄᑦ ᐱᓪᓚᓇᐃᖅ ᑲᓇᑕᒥ (EC): ᐱᓴᔨᑎᑦ/ᐃᓱᒪᓪᓕᐅᕈᑎᑦ			
6:00 PM - 6:30 PM	15	*ᐊᐅᓪᓴᐃᕆᐊᖏᓂᖅᑐᑦ ᐸᖏᓇᐅᑎᑦ Porsild Bryum -ᒧᑦ, ᒪᓪᓕᓴᓪᓕᑦ ᐅᒪᓕᓱᑦ ᐃᓄᐃᔾᓗᐊᓗᒃ ᐊᐃᑦᑎᑦ ᐊᑎᑕᓗᐅᓪᓚᖅᓵᑐᑦ ᐱᖏᕐᕋᐊᖏᑎᑎᑦ (SARA)*	13	ᑲᓇᑕᒥ ᐊᙵᑎᓪᓚᓱᓕᑦ ECCC	30 ᒥᓂᔪ
6:30 PM - 7:10 PM	16	ᑲᓇᑕᒥ ᓇᓄᐃᑦ ᒥᐊᓂᓴᐅᑐᖏᑎᓄᑦ ᐸᖏᓇᐅᑎᓄᑦ ᑐᓴᓴᕆᓂ	14	ᑲᓇᑕᒥ ᐊᙵᑎᓪᓚᓱᓕᑦ ECCC	45 ᒥᓂᔪ
7:10 PM - 7:40 PM	17	ᐅᖃᐅᓴᓱᑦ ᑐᓴᖅᑎᓂᑎᓵᓴᑦ ᓄᑕᔪᓚᐊᖅᑕᐅᑐᖏᓂᓄᑦ ᑎᑭᑦᑕᖅᑐᑦ ᑎᖏᒥᐊᑦ ᒪᓪᓚᖏᑦ ᐊᓱᒪᓗ ᐅᒪᓴᓪᓚᓂᖅᓯᒪᑦ ᒪᓪᓚᖏᑦ	15	ᑲᓇᑕᒥ ᐊᙵᑎᓪᓚᓱᓕᑦ ECCC	30 ᒥᓂᔪ
7:40 PM - 7:45 PM	18	ᓄᖅᑲᖅᓂᖅ		ᐃᑦᓯᙵᐅᑕᖅ	5 ᒥᓂᔪ

ⅅⴰⵜⴰⴷⵉⵔ

ⵙⵙⵓⵖ ⴰⵍⵉⵔⵎⴰⵏⵉⵏⵉⵔ ⵃⵓⵍⵉⵣⵓⵔⵙⵙⵓⵖ

ⴰⵍⴰⵙⴰⵏⵉⵏⵉⵔ

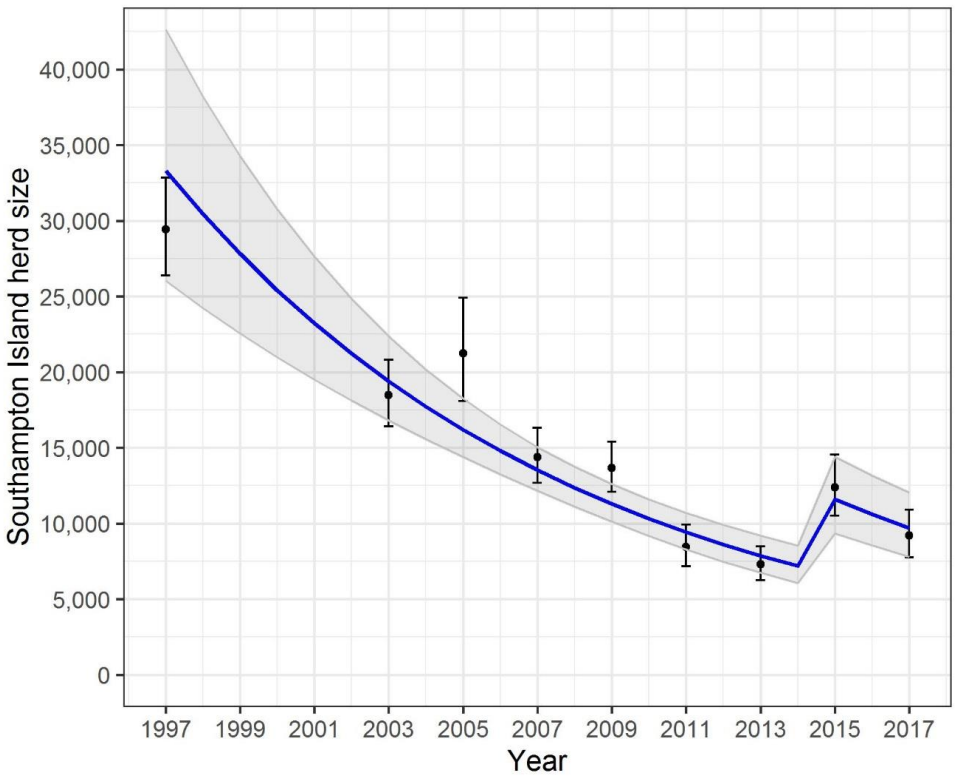
ⵜⵉⵍⵉⵃⵉⵏⵏ: ⵅⵅⵅ

ⴰⵔⵉⵎⴰⵔⵉⵜⵉⵏ ⵅⵅⵅ: X

ⵍⵉⵔⵓⵙ: ⵅⵃⵙⴰⵔⵣⵍⵣⵉⵏⵉⵔ ⵅⵉⵔⵣⵏ ⵅⵔⵔⵅⵔⵣⵏⵙ ⵇⵄⵅⵃⵜⵅⵃⵣⵔⵔⵜⵅⵔⵉ ⵜⵃⵜⵉ ⴷⵙⵏⵙⵎⵍⵣⵅⵔⵉ

ⵅⵃⵙⴰⵔⵉⴷⵅⵔⵍⵣⵍⵣⵉⵏⵉ

- ⵇⵄⵅⵃⵜⵅⵃⵣⵔⵔⵜⵅⵔⵉ ⵜⵃⵜⵣⵔⵉ ⴰⵔⵍⵃⴷⵉⴷⵅⵔⵍⵉⵄⵉ ⵅⵉⵔⵔⵜⵅⵔⵉ ⵅⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵄⵃⵄⵄⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵅⵃⵙⵔⵍⵉⵏⵜⵃⵓ ⵄⵏⵙⵙⵓⵖ 1968, ⵙⵙⵍⵉⴷⵅⵓⵏⵜⵃⵔⵔ ⵅⵉⵔⵔⵅ ⵅⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵜⵃⵜⵔⵃⵣⵔⵔ ⵅⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ 1950.
- ⵜⵃⵜⵉ ⴷⵙⵅⵔⵍⵉⵔⵣⵍⵉⴷⵅⵜⵉ ⵄⵏⵙⵙⵓⵙ 30-ⵙⵉ ⵄⵏⵍⵜ ⵙⵏⵔⵃⵅⵏⵅⵔⵍⵣⵍⵃⵃⴷⵉⵜⵃⵓ ⵜⵃⵜⵔⵍⵅⵔⵣⵍⵉⴷⵅⵜⵉ ⵜⵃⵜⵔⵍⵅⵔⵣⵏⵙⵙⵓⵖ ⵄⵏⵍⵜ ⵙⴷⵏⵔⵃⵓⵏⵅⵃⵉⴷⵜⵃⵓ 1990-ⵅⵔⵙⵙ.
- ⵅⵔⵔⵄⵏⵔⵜⵃⵜ 2003, ⵅⵉⵔⵔ ⵅⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵜⵃⵜⵉ ⵃⵙⵃⵃⵅⵉⵔⵔⵉⵔⵔⵍⵉⵔⵣⵍⵣⵉⵏⵉⵔ ⵄⵏⵍ ⵓⵔⵔⵜⵃⵜ ⵚⵙ 2011, ⴷⵙⵔⵔⵙⵙⵜⵃⵜⵍⵃⴷⵉⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵙⵏⵔⵅⵏⵅⵔⵍⵣⵍⵃⵅⵏⵅⵔⵣⵏⵙⵙⵓⵖ ⴷⵙⵙⵔⵜⵃⵜⵍⵃⵙⵙⵓⵖ.
- ⵅⵔⵔⵄⵏⵔⵜⵃⵜⵍⵃⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵃⵙⵃⵃⵅⵉⵔⵔⵉⵔⵔⵍⵉⵔⵣⵍⵣⵉⵏⵉⵔ ⵃⵓⵙⵙⵓⵖ ⵅⵔⵔⵄⵏⵔⵜⵃⵓⵔⵔ ⵅⵔⵙⵜⵃⵜⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵚⵙⵙⵓⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵄⵏⵍⵜ ⵙⵓⵔⵜⵃⵜ ⵚⵙⵙⵓⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ ⵄⵏⵍⵜ ⵙⵏⵔⵃⵅⵏⵅⵔⵍⵣⵍⵃⵃⴷⵉⵜⵃⵓ ⵅⵔⵔⵄⵏⵔⵜⵃⵜⵍⵃⵔⵔⵅⵔⵣⵏⵙⵙⵓⵖ



ᓇᓗᓇᐃᖅᓯᕐᓗ 1. – ᐊᓯᔨᖅᐊᓕᓕᐊᓂᖅᑎᑦ ᓴᓕᓕᖅ ᕿᑭᖅᑕᖓᓄᓂ ᑐᖅᑐᑦ ᐅᓄᖅᓂᓚᖏᑦ

ᐅᓪᓗᒥᐅᔪᖅ ᖃᓄᐃᓕᓂᖓᓕᖅᖢᓂ

- ᓴᓪᓕᓐᓂ ᕿᑭᖅᑕᖓᓄᓂ ᑐᓄᖅᓯᓂᓕᓕᐅᖅᑐᑦ ᒥᖅᓴᐅᓴᖅᑕᐅᓯᒪᔪᓂ 7,284-ᖓᔪᓚᖏᐅᓕᐅᖅᓯᖅᓯᒪᔪᓂ ᓚᐃ 2013-ᖓᐅᓪᓗᑐᔪ, ᐅᓄᖅᓂᖃᖅᓯᓂᖏᐅᓕᓕᐅᖅᓯᒪᔪᓂ 12,319-ᓂᒃ ᓚᐃ 2015-ᖓᐅᓪᓗᑐᔪ, ᐊᓖᓚᐅ ᐃᓄᐃᓴᓕᑎᐊᒃᑲᓂᓕᐅᖅᑐᑦ ᐃᓐᓇᐅᓂᖓᓂ 8,436-ᓄᓂ ᐊᓖᒪ ᑎᓚᑐᖃᐃᓪ ᓚᐃ 2017-ᖓᐅᓪᓗᑐᔪ.
- ᑕᒪᖅᑭᒃ ᐱᓕᓚᒪᖓᓪᓗᑐᐊ ᐊᖏᑎᓕᓚᐅᒃᑯᓂ ᐊᓖᒪᓐᓇ ᓴᓪᓕᓐᓂᓕ ᐊᖓᔪᓇᓯᑐᑦᓚᐅᒃᑯᓂ ᑎᑎᕐᑯᑎᖓᓪᓗᓂ, ᐅᒃᐱᕐᓯᒃᑐᑦ 2015-ᒥ ᐅᓄᖅᓯᓕᐊᑎᑎᓕᐅᖅᑕᖓᓪᓗᒃ ᐱᒃᔨᔪᓯᑐᖅᑐᒃ ᑐᒃᑐᑦ ᑎᑭᓚᐅᓴᓂᖓᓕᐅᖅᓯᓕᓂ ᐊᓯᐊᓂᓴᖓᓄᑦ ᐅᒃᐱᓚᓴᐅᓕᐅᖅᓯᒪᓕᒡᓕ ᓴᓪᓕᓐᓂᒥᐅᓂ ᐊᖓᔪᓇᓯᑐᑦᓚᐅᒃᑯᓂ ᐅᑭᐅᖅᓯᖓ 2013-14 ᓄᓇᒥᓴᖓᓄᑦ ᓄᓇᖅᓯᖓᓂ ᓄᓇᕐᓂ.
- ᐃᓄᐃᓴᓪᓕᓐᓂᕐᐊᓕᓕᐅᖅᑐᑦ ᐊᑯᓂᓂᖓᓂ ᓚᐃ 2015 ᐊᓖᒪᓗ 2017 ᐅᓄᖅᓂᕐᓂᓂᒃ ᐅᓴᐃᑎᓪᓗᑎᑦ, ᐃᓯᒪᔪᓂ ᓴᓪᓕᓐᓂᓕ ᐊᖓᔪᓇᓯᑐᑦᓚᐅᒃᑯᓂ ᑎᒥᖅᑎᑦ, ᐅᑭᐅᒃᑯᑦ ᐊᐳᑎᖃᓗᐊᓕᐅᖅᓯᖅᑎᑦᓚᐅᒃᑯᓂ ᓯᑯᓕᐅᖅᓯᖅᑎᑦᓚᐅᒃᑯᓂᓗ ᓄᓇ ᐱᓪᓕᖅᓕᐅᖅᓯᖅᓂᕐᓂᒃ 2015-ᖓᐅᓪᓗᑐᔪ ᓄᓇ ᓯᑯᓕᐅᖅᓯᒪᔭᕐᒪᑦ ᑭᖓᔪᖅᓚᔪᑦ 2015-ᒥ ᐅᓴᐃᓕᐅᖅᑎᓪᓗᑎᑦ.
- ᐱᓕᓚᒪᖅᑎᑦ ᐊᖏᑎᓕᓚᐅᒃᑯᓂ ᖃᐅᐅᔨᐱᖅᓂᓐᓄᖅᑐᑦ ᑐᒃᑐᑦ ᖃᓄᐃᓕᓂᒪᓕᖅᖢᓂᒋ ᐃᓕᓐᓄᖅᑎᖃᖅᑐᓂᒃ ᖃᐅᐅᔨᐱᖅᑎᓂᒃᑯᓂᒃ ᓄᓇᕐᓂᖅᑐᓂᒃ ᑐᒃᑐᑦ, ᓄᓇᖅᓯᐃᖅᑕᐅᓂᓯᐊᖏᑦᓗᓇ ᑐᒃᑐᑦ ᐊᐳᐊᓂᓴᖓᓄᑦ ᑎᑭᖃᑕᓂᖓᓂᒃ ᑐᒃᑐᑦ.
- ᐱᓕᓚᒪᖅᑎᑦ ᐊᖏᑎᓕᓚᐅᒃᑯᓂ ᐅᓄᖅᓂᕐᓂᓂᒃ ᓴᓚᐃᔨᒪᓕᖅᑐᑦ ᓴᓪᓕᓐᓂᓕ ᑐᒃᑐᑦ ᕿᑭᖅᑕᖓᓄᓂ ᐱᓕᕐᓂᖅᑐᑦ, 2019-ᒥ.

ᑐᓴᕋᓱᐊᕐᓂᖅ:

- ᐊᕐᕿᒍᐊᓂ 2017 ᑐᒃᑐᑦ ᐅᓄᕐᓂᓚᖅᓯᑦ ᖃᐅᔨᓴᖅᑕᐅᒃᑲᒃᓂᓕᕆᓂᖅᑎᑦᓯᑐᑦ ᒥᒃᓴᐅᓴᒃᑕᐅᓱᒪᓕᕐᑦ ᐃᓄᐃᓴᓕᓂᒪᐊᑲᓂᕐᓂᖅᓯᓚᐅᑦᓯᓂᖅᑐᑦᓯᓚᐅᑎᓕᐅᑦᓯᖅᑐᑦ ᐅᓱᐊᓂᓕ ᒥᐊᓂᖅᓱᕐᓴᐅᒃᓴᓂᓂᓕ ᖃᓄᓕᐅᖅᓯᑦᑐᑦ ᐱᓕᕆᐅᓱᓕᐅᖅᑐᑦ ᐅᑭᐊᒃᓴᖅᓯᓂ 2018 ᑭᖓᔪᖅᑲᑕᓕᐅᖅᓂᖅᓗᓄᑦ ᖃᑎᖅᑭᖅᐅᓂᓕᕆᓂᖅᑎᑦᓯᓚᐅᑦᓯᓂᖅᑐᑦ ᖃᐅᔨᓴᖅᑕᐅᓴᕐᔪᐊᑎᐃᓐᓇᓂᑦ ᐅᓄᕐᓂᖅᓗᓄᑦ ᐅᓴᖅᑭᑦ ᖃᐅᔨᓴᖅᑕᐅᔪᓐᓇᐃᓐᓇᓂᑦ.

- ᑕᐃᒪᓕ ᑐᓴᖅᑎᑕᐅᑎᒃ ᐃᓄᐃᓴᓕᓂᔪᓕᕐᓂᖅᓇᖅᓂᖅᓂᖅᓂᖅᑐᒃ, ᓴᓪᓕᓴᓂ ᐊᖏᓚᐅᔭᑐᑕᓚᐅᑦᓯᖅ ᑎᒥᖅᓯᓂ ᐋᖅᑭᓕᓕᐅᖅᑐᑦ ᑐᐅᔪᓴᐅᑎᖅᓂᖅᓂᒃ ᐅᓄᖅᑭᓕᐅᓴᖅᓯᔪᓂᒃ ᑐᐅᑐᐃᓴᖅᓂᖅᓂᒃ ᐅᑎᖅᑭᑕᐅᔪᓐᓇᖅᑐᑦ ᐅᒪᖅᑎᑕᐅᓪᓗᑎᒃ 1,600-ᖏᓕᖅᑯᐅᑕᐅᓕᐅᖅᑐᑦ ᐅᑎᖅᑎᑕᐅᓪᓗᑎᒃ 1,000-ᓄᑦ ᑐᐅᑐᐃᓴᖅᓗᓄᑦ. ᑕᐃᒪᐃᓕᓴᐅᓕᐅᖅᑐᑦ ᐃᓕᒥᓂᖅᑎᓂᒃ ᐊᑐᓕᓚᑐᒃ ᖃᑎᖅᑭᑦᓴᐃᓐᓇᓂᑦ ᐅᓄᖅᑭᓕᐅᔪᓐᓇᖅᑐᑦ ᑭᓯᓂᖅᓂᒃ ᑭᓱᓕᕆᓂᖅ ᐊᖏᔪᒡᖅᓯᓪᓚᑐᒃ ᓯᐅᒧᐊᓕᓱᓕᓂᖅᑎᓕᐅᖅᑐᑦ.

- ᑐᐃᓴᓱᐊᓕᓂᖅᑎᓕᐅᖅᑎᓪᓗᓕᑦ ᐅᓐᓄᐃᐊᓗ 2019-ᖏᑎᓪᓗᒍ ᖃᐅᔨᓕᐅᖅᑐᑦ ᓄᓇᕗᐅᔪᖅ ᐊᓪᓚᓗ ᓄᓇᕗᓂᓂ ᐊᖏᓚᐅᔭᑐᑕᓚᐅᑦᓯᖅ ᑎᒥᖅᑎᓂ ᐃᓱᒪᓗᒍᑎᖅᑎᓂᖅᓂᒃ ᐅᓄᓕᓂᒪᓂᖅᓂᒃ ᑐᒃᑐᑦ ᐊᓪᓚᓗ ᐃᓄᐃᓴᓕᓂᒪᐊᑲᓂᕐᓂᖅᓂᒃᑲᓂᕐᓂᖅᓂᒃ ᖃᑎᖅᑭᑦᓴᐃᓐᓇᓂᑦ ᐅᓕᖅᑭᓕᐅᔪᓐᓇᖅᑐᑦ.

- ᐃᓕᓴᐅᐊᓗ 2019-ᖏᑎᓪᓗᒍ, ᓴᓪᓕᓴᓂ ᐊᖏᓚᐅᔭᑐᑕᓚᐅᑦᓯᖅ ᑎᒥᖅᑎᓂ ᑐᐃᓴᕋᓕᐅᖅᑐᑦ ᐱᓕᕆᓚᖏᖅᓗᓄᑦ ᐊᕐᓇᑎᓕᓚᐅᑦᓯᖅ ᐃᓵᖅᐅᓚᐅᓴᖅᑎᑎᖅᑎᓂᖅᓯᑦᓯᓚᐅᑦᓯᑐᑦ ᓄᓇᕗᑦ ᐅᓕᖅᑭᓕᐅᔪᓐᓇᖅᑐᑦ ᐃᓚᖅᓯᖅᑎᓗᓄᑦ ᖃᑎᖅᑭᑦᓴᐃᓐᓇᓂᑦ ᐅᓄᖅᑭᓕᐅᔪᓐᓇᖅᑐᑦ ᑭᓯᓂᖅᓂᒃ 1,000 ᐃᓄᐃᓴᓕᓂᒪᐊᑲᓂᕐᓂᖅᓯᑦᓯᓚᐅᑦᓯᖅ ᐊᓪᓚᓗ ᐅᓄᕐᓂᖅᓗᓄᑦ ᐅᓴᖅᑭᑦᓴᐃᓐᓇᓂᑦ ᑕᐃᒪᓕᐅᖅᑐᑦ ᒥᒃᓴᐅᓴᒃᑕᐅᓴᓯᓚᐅᑦᓯᖅ ᐊᓪᓚᓗ ᐅᓄᕐᓂᖅᓗᓄᑦ ᐅᓴᖅᑭᑦᓴᐃᓐᓇᓂᑦ ᑕᐃᒪᓕᐅᖅᑐᑦ ᐅᓴᖅᑭᑦᓴᐃᓐᓇᓂᑦ ᑕᐃᒪᓕᐅᖅᑐᑦ ᐃᓄᐃᓴᓕᓂᒪᐊᑲᓂᕐᓂᖅᓯᑦᓯᓚᐅᑦᓯᖅ ᐊᓪᓚᓗ ᐅᓄᕐᓂᖅᓗᓄ

ᐊᐅᓪᓚᕆᔨᓄᑦ ᓇᐃᓐᓴᖅᓯᒪᔪᑦ:

ᐊᒍᓂᐅᓂᖅᖅᑐᑦ ᓯᕕᑐᔪᒻᒪᑦ ᐊᓯᔨᔅᓴᑦᓚᓕᒫᓱᒪᓕᖅᑎᑦ ᐅᓄᔅᓂᒫᓕᖅᑎᑦ ᐊᒻᒪᓗ ᓯᐊᒻᒪᓕᖅᑎᓂᖅᓯᒪᓕᒫᓕᖅᑎᑦ

ᓴᑦᓚᑦ ᖁᑭᖅᑕᖏᓚᓂ ᑐᐅᖅᑎᑦ: 1978-2017

ᓇᓴᖅᑐᓂᖅᑑᑎᐅᑕᑦ ᑐᐅᖅᑎᑦ ᐃᓕᓴᐅᓕᐅᖅᓯᒪᔪᓄ ᓴᑦᓚᑦ ᖁᑭᖅᑕᖏᓚᓄᑦ ᐊᒃᓴᑐᑦᔪᐊᑦ ᖁᑭᖅᑕᖏᓚᓂᒃ ᐱᓴᐅᑦᓗᑎᒃ, ᑭᖓᑦᓕᖅᖅ ᓄᓇᕗᒥ, 1968-ᖃᐅᑦᓗᒍ, ᓄᖃᓕᐅᖅᑎᑦᓗᑎᑦ ᓴᑦᓚᑦ ᖁᑭᖅᑕᖏᓚᓅᓪᑐᑦ ᑐᐅᖅᑎᑦ ᐱᑎᐊᕆᓴᖅᑎᑦᓗᒍ 1950 ᐊᑦᑭᒍᖏᑦ. ᐅᔪᐊ ᖃᐅᐅᔨᓴᕐᔪᐊᑎᐃᓇᓂᐅᔪᓕᑦ ᑕᔪᐅᓴᖅᓯᓕᐅᑦᑐᖅ ᐊᓯᔨᔅᓴᑦᑕᑦᓯᓂᒪᓕᖅᑎᑦ ᐅᓄᔅᓂᒫᓕᖅᑎᑦ ᐊᒻᒪᓗ ᓯᐊᒻᒪᓕᖅᑎᓂᖅᓯᒪᓕᒫᓕᖅᑎᑦ ᑐᐅᖅᑎᑦ ᖁᑭᖅᑕᖏᓚᓂ ᑕᐃᒪᓐᓴᑦ ᐃᓕᓴᐅᓕᐅᖅᑎᓂᖅᑎᑦ. ᖁᑭᖅᑕᖏᓚᓂ ᑐᐅᖅᑎᑦ ᐅᓄᔅᓂᕆᔨᐅᔅᓴᑦ 48-ᓂᒃ ᐅᓄᔅᓂᖃᖅᑐᑦ ᐃᓕᓴᐅᓕᐅᖅᓯᒪᔪᓄ 1968-ᖃᐅᑦᓗᒍ ᐅᓄᔅᓂᖃᓕᓕᐅᖅᑐᑦ ᒥᖅᓴᕆᖅᑕᐅᓯᒪᑦᓗᑎᒃ 30,381 ᑐᖅᑐᓂᒃ (+/- 3,982, 95% CI) ᑎᑭᑦᓗᒍ ᔦᓂ 1997, ᐅᓄᔅᓯᓚᑦᓕᐊᓂᖓᓕᒫ ᓯᖃᓂᖅᖃᓕᐅᖅᑐᖅᓴᖅ ᖃᖃᓂᕐᓴᓂ 23%. ᖁᑭᖅᑕᖏᓚᓂ ᑐᐅᖅᑎᑦ ᓂᕆᐹᖅᓴᕆᔨᐊᓘᔪᐅᑦᓗᑎᒃ ᑐᐅᑐᓴᕆᖅᑎᖅᑎᓕᕆᓕᐅᖅᑐᑦ 1978-ᖃᐅᑦᓗᒍ, ᐊᒻᒪᓗ ᐅᓄᔅᓂᖅᑐᓴᓘᓂᒃ ᓂᐅᑎᑎᖅᑎᖅᑐᓂᒃᑐᖅᓯᒪᓕᖅᑎᑦ ᐱᑎᐊᖏᑦᓗᒍ 1993. ᐊᑦᑭᒍᓄᑦ 30-ᓄᑦ ᐅᓄᔅᓯᓚᑦᓕᐊᓕᐅᖅᑎᓂᓗᑎᑦ, ᑐᐅᑦ ᐃᓄᐃᔪᓕᓚᑦᓚᑦᓕᐊᓕᓕᐅᖅᑐᑦ ᒥᖅᓴᕆᖅᑕᐅᓯᒪᔪᑦ 30,381-ᖃᔨᓘᔪᐅᓕᐅᖅᓯᒪᔪᑦ ᔦᓂ 1997-ᖃᐅᑦᓗᒍ ᐅᓄᔅᓂᖅᑎᖅᓯᑎᐅᓕᓕᐅᖅᑐᑦ 20,582-ᓂᒃ (95% CI=3,065; CV=0.075) ᔦᓂ 2005-ᖃᐅᑦᓗᒍ, ᐅᓄᔅᓂᖅᑕᓕᐅᖅᑐᑦ 15,452-ᓂᒃ (95% CI=1,858; CV=0.061) ᔦᓂ 2007-ᖃᐅᑦᓗᒍ, ᐅᓄᔅᓂᖅᑕᓕᐅᖅᑐᑦ 13,953-ᓂᒃ (95% CI=1,790; CV=0.065) ᔦᓂ 2009-ᖃᐅᑦᓗᒍ, ᐅᓄᔅᓂᖅᑕᓕᐅᖅᑐᑦ 7,903-ᓂᒃ (95% CI=1,261; CV=0.081) ᔦᓂ 2011-ᖃᐅᑦᓗᒍ, ᐅᓄᔅᓂᖅᑕᓕᐅᖅᑐᑦ 7,287-ᓂᒃ (95% CI=1,045; CV=0.073) ᒪᐃ 2013-ᖃᐅᑦᓗᒍ. ᑎᑭᑦᓗᒍ ᒪᐃ 2015, ᑐᐅᑦ ᐅᓄᔅᓯᓕᐊᖅᑲᓂᓕᐅᖅᑐᑦ ᐅᓄᔅᓂᖅᑕᕐᔪᐊᑎᒃ 12,297-ᓂᒃ (95% CI = 1,844; CV = 0.076). ᑭᓯᐊᓂᓕ, ᑎᑭᑦᓗᒍ ᐊᑦᑭᒍᐊ 2017, ᑐᐅᑦ ᐅᓄᔅᓂᖏᑦ ᐃᓄᐃᔪᓕᓚᑦᓚᑦᓕᐊᖅᑲᓂᓕᐅᖅᑐᑦ ᐅᓄᔅᓂᖅᑕᕐᔪᐊᑎᒃ 9,887-ᓂᒃ (95% CI = 1,438; CV = 0.080). ᑕᐃᔨᓯᒪᓂ ᑐᐅᑦ ᐃᓄᐃᔪᓕᓚᑦᓚᑦᓕᐊᑎᒃᓗᑎᑦ ᓯᐊᒻᒪᓕᖅᑎᓂᖅᓯᒪᓕᒫᓕᖅᑎᑦ ᐃᓄᐃᔪᔭᓕᔅᓚᑦᓕᐊᓕᓕᐅᖅᑐᑦ ᓂᕆᔅᑕᓯᐊᓂ ᖁᑭᖅᑕᖏᓚᓂ ᑐᐅᖅᑎᑦ ᐊᒻᒪᓗ ᖃᐅᔨᓴᖅᑎᑦ ᑕᐃᔨᓯᒪᓂ ᐊᑦᑑᓕᓚᑦ. ᑭᖓᔪᓄᓪᓗ ᓇᔪᐊᑦᓗᑎᒃ ᐊᑦᑭᒍᑦᓗᒍ ᖃᐅᔨᓴᖅᑎᑦ ᐃᓄᐃᔪᓕᓚᑦᓚᑦᓕᐊᓕᓕᐅᖅᑐᑦ ᐃᓄᐃᔪᔭᓕᔅᓚᑦᓕᐊᓕᓕᐅᖅᑐᑦ ᓂᕆᔅᑕᓯᐊᓂ ᐃᓕᓴᐅᓕᐅᖅᑎᓂᖅᑎᑦ ᐃᓕᓴᐅᓕᐅᖅᑎᑦ ᐊᑦᑭᒍᑦᓗᒍ ᑭᓴᓕᖅᑕᐅᓕᐅᖅᑐᑦ 2011-ᒥ, ᐊᑦᑭᒍᑕᒪᓐ ᑲᑎᕐᓴᓂᒃ ᐅᒪᐃᓴᖅᓯᒪᔪᑦ ᐅᒪᓕᖅᑕᐅᓕᐅᖅᑐᑦ 1,000-ᓂᒃ ᑐᐅᑕᓄᓐᓇᓕᐅᖅᑐᑦ ᐊᑦᑭᒍᑦᓂᓐᓂ 2012, 2013, ᐊᒻᒪᓗ 2014 ᑐᐅᑕᓂᕐᓇᓂᓐᓂ ᐊᑐᖅᑕᐅᓕᓗᑎᒃ. ᐅᓄᔅᓯᓚᐊᓂᓂᑦ ᖃᐅᔨᓴᐅᔭᐅᓕᐅᖅᑎᓂᓗᑎᑦ 2015-ᖃᐅᑦᓗᒍ, ᑲᑎᕐᓴᓂᒃ ᐅᒪᓕᖅᑕᐅᓕᐅᖅᑐᑦ 1,600-ᓂᒃ ᑐᐅᑕᓄᓐᓇᓗᑎᒃ ᐊᑦᑭᒍᑕᒪᓐ ᒡᓐᓇ ᐊᑲᐊᖅᑕᐅᓕᐅᖅᑐᖅᓴᖅ ᓄᓇᕗᑦᔨᑎᖃᓂᓂ ᓴᑦᓚᑦᕆᐅᑦ 1,000-ᓂᒃ ᑐᐅᑕᓄᓐᓇᖅᓯᓗᑎᒃ 2017-ᒥ ᐃᓄᐃᔪᓚᑦᓚᑦᓕᐊᓂᓕᐅᖅᑐᑦ ᒥᖅᓴᕆᖅᑕᐅᓕᐅᖅᑎᓂᓗᑎᑦ. ᐋᓐᓂᐊᕐᓴᕆᑦᓯᓂᖅᓐᓂᓄᑦ ᐊᒻᒪ ᐅᓪᓚᕆᔨᐊᕐᓯᓄᑦ ᐋᓐᓂᐊᖅᑭᖅᑎᑦᓯᓂᖅᓐᓂᓄᑦ ᐃᖏᕐᕕᐊᓗ 2000-ᖃᐅᑦᓗᒍ. ᐊᓴᒍᑦᓴᓕᐅᑦ ᐅᓄᔅᓯᓚᑦᓕᐊᓂᕐᓂᖅᓯᒪᓕᐅᑕᓴᓕᖅᑎᓂᓗᑎᑦ 1.7%-ᒥᒃ ᐃᖏᕐᕕᐊᓗ 2000-ᖃᐅᑦᓗᒍ ᐅᓄᔅᓂᖅᑕᐅᓕᐅᖅᑐᖅᓴᖅ ᓯᖃᓂᖅᓴᒃ 58.8%-ᒥᒃ ᐃᓐᓴ 2011-ᖃᐅᑦᓗᒍ ᐊᒻᒪᓗ

ᑕᒃძᓇᓂ ᐅᓄᖅᓯᓐᓕᐊᔪᓚᐅᑎᓪᓗᑎᑦ ᓯᖏᑐᐃᔪᑦ ᐃᓄᐃᓴᓕᓐᓐᓕᐊᓕᐅᖅᑐᑦ ᑕᐃᔨᓯᒪᓂᕆᐃᐅᓇᖅ. ᓯᖏᑦᓴᐃᑦ ᐅᓄᕆᓂᖏᑦ ᐃᓄᐃᓴᓕᓂᕈᐊᓕᐅᖅᑐᑦ 93.1%-ᒥᒃ ᐃᖃᐅᑲᐊᓚ 2001-ᖏᑎᓪᓗᒍ ᑕᐅᓇ ᐊᒃᓕᓯᖏᓂᖅᓱᖏᓕᐅᖅᑐᑦᖅ 37%-ᖏᓪᓗᓂ ᐃᓴᔨ 2011-ᖏᑎᓪᓗᒍ. ᐊᓯᔨᖅᓴᓪᓕᐊᓂᓚᖏᓪᓕ ᕿᑎᕋᕆᐅᐅᓪᓗᑎᒃ ᓇᓗᓇᐃᖅᓯᓐᓕᓚᐃᔪᑦ ᓴᓪᓗᑦ ᕿᑭᖅᑕᖏᓚᓂ ᑐᒃᑐᑦ ᐅᓄᕐᓂᓚᖏᓪᓕ ᐃᓄᐃᓴᓕᓂᕈᐊᓯᒪᓂᖅᓴᓕᑕ 9%-ᒥᒃ ᐊᖏᕐᔪᒍᓴᐃᓚᵒ ᑕᐃᒪᵒᖏᓂᑦ 1997-ᒥᑦ ᐅᓴᓴᖅᑕᐅᓕᐅᖅᑎᓪᓗᑎᑦ. ᐊᐅᖏᑎᑐᒍᑦ ᖃᐅᔨᓴᖅᑕᐅᑎᓪᓗᑎᑦ ᐊᓚᒪ ᓄᓇᓕᖏᓂ ᖃᐅᔨᓚᓂᓚᐅᐅᔪᑦ ᓇᓗᓇᐃᖅᓯᓕᐅᖅᑐᑦ ᓄᑦᑕᖃᑦᑕᖏᓂᖏᓐᖏᓂᒃ ᕿᑭᖅᑕᑎᒥᑦ ᓄᓇᖅᐱᖏᒍᑦ ᐊᓚᒪᓗ ᓴᓪᓗᑦ ᕿᑭᖅᑕᖏᓚᓂ ᑐᒃᑐᑦ ᐅᑭᐅᖏᓂᓂ ᐊᖏᕐᔪᒍᓴᐃᑉᐸᓂᖏᓐᖏᓂᓂ 2014 ᐊᓚᒪᓗ 2015 ᐅᓄᖅᓯᓐᓕᐊᓂᖅᓴᖅᓴᓪᓕᐊᓕᐅᖅᑐᑦ. ᕿᓯᐊᓂᓕ ᐊᔨᐃᑲᖏᓂᓂᖏᓪᓕ ᑕᑯᓇᖏᓪᓗᑎᑦ ᐊᖏᕐᔪᒍᖏᓂᓂ 2015 ᐊᓚᒪᓗ 2017 ᐅᓴᖅᑕᐅᓯᒪᓂᖏᓂᓇᓂ ᒥᖅᐳᔨᖅᑕᐅᓯᒪᔪᑦ 9%-ᒥᒃ ᐃᓄᐃᓴᓕᓐᓐᓕᐊᓂᖏᓗ ᑲᔪᓯᓚᔪᖅ ᐊᖏᕐᔪᒍᐃᓐᓇᓂ ᒪᖏᐃᓐᓇᓂ ᐱᖏᓯᓇᓂ ᑎᑭᓪᓗᒍ ᑭᖓᔪᖅᑐᒍ. ᑕᐃᒪᐃᓚᒪᑦ ᑲᔪᓯᓴᒃᑯᑦ ᐃᓄᐃᓴᓕᓐᓐᓕᐊᓂᖅᑲᒥᐃᓕᒪᑦ ᓯᓴᑎᓕᐅᓂᖏᓐᖏᓂᓇᓂ, ᐃᓄᐃᓴᓕᓐᓕᐊᖅᑕᐅᓂᖅᑕᐅᔨᓴᒃᑯᓪᓗᑎᒃ ᑲᑎᕐᓂᐊᓚᐅᓂᖅᑐᑦ ᐅᒪᔨᓚᓕᑦ ᐅᒪᓕᐊᓂᖅᑐᑦ ᑭᑲᓐᓂᖏᑦ ᐊᖏᕐᔪᒍᓴᐃᑉᐸᓂᖏᓐᖏᓂᓂ ᑐᖅᑕᐅᓂᖅᑐᑦ.

Long-term trends in abundance and distribution of the Southampton Island caribou herd: 1978 – 2017

Government of Nunavut
Department of Environment
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Abstract:

Barren-ground caribou (*Rangifer tarandus*) were introduced onto Southampton Island (SHI) from Coats Island, in the Kivalliq Region of Nunavut, in 1968, following their extirpation from SHI in the early 1950s. This demographic study illustrates large fluctuations in abundance and distribution of caribou on SHI since its re-occupation. The SHI caribou herd grew from the 48 animals introduced in 1968 to an estimated population of 30,381 animals (+/- 3,982, 95% CI) by June 1997, with an annual rate of increase of approximately 23%. The SHI herd supported a subsistence harvest beginning in 1978, and a largescale commercial harvest beginning in 1993. After nearly 30 years of growth, herd abundance declined from the estimated 30,381 in June 1997 to 20,582 (95% CI=3,065; CV=0.075) in June 2005, to 15,452 (95% CI=1,858; CV=0.061) in June 2007, to 13,953 (95% CI=1,790; CV=0.065) in June 2009, to 7,903 (95% CI=1,261; CV=0.081) in June 2011, and then to 7,287 (95% CI=1,045; CV=0.073) in May 2013. By May 2015, the population had increased again to 12,297 (95% CI = 1,844; CV = 0.076). However, by 2017, the population had declined again to 9,887 (95% CI = 1,438; CV = 0.080). During the decline caribou distribution gradually concentrated into a core area within the south-central portion of the Island in the vicinity of the Kirchoffer River. Harvest estimates over the same periods varied widely. Following the 2011 survey, an annual Total Allowable Harvest (TAH) of 1,000 caribou was applied over the 2012, 2013, and 2014 harvesting seasons. Following the population increase detected in 2015, the TAH was increased to 1,600 caribou annually then dropped by the community of Coral Harbour to 1,000 caribou following the 2017 estimated declines. Susceptibility to disease and parasites due to low genetic heterogeneity has been a concern since the introduction of caribou to SHI, and was a likely catalyst to the wide spread infection of caribou with *Brucellosis suis* first detected in the population February 2000. Prevalence of Brucellosis climbed from 1.7% in February 2000 to 58.8% in March 2011 and this increase is thought

to have contributed to decreased pregnancy rates over the same period. Pregnancy rates dropped from a high of 93.1% in February 2001 to a low of 37% in March 2011. Trend analysis suggests that the SHI caribou population has been decreasing at a rate of 9% per year since the 1997 survey. A genetic analysis and local knowledge confirmed the occurrence of a movement event between the mainland and SHI between the winters of 2014 and 2015 which likely increased the population. However, comparison of the 2015 and 2017 survey estimates suggests the 9% decline continued over the two to three years following this event. Given the continual and steep nature of the decline, combined with the reliance of users on this population for subsistence and commercial harvesting purposes, a reduction in TAH is recommended to reduce the rate of decline and maintain the population over the long term.

Key words: Commercial harvest, barren-ground caribou, caribou, *Rangifer tarandus*, Southampton Island, Coral Harbour, Kivalliq, disease, *Brucellosis suis*, Nunavut, population survey, demographic studies.

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1.0 Introduction

Following the extirpation of caribou from Southampton Island (SHI) in the early 1950s, there was much discussion regarding their re-introduction as well as recognition of the careful husbandry that must go hand in hand with any such program (MacPherson, 1967). Discussions continued up until 1967 at which time Northwest Territories Commissioner Stuart Hodgson along with D.S. Munro, Director of the Canadian Wildlife Service (CWS), made the decision to move forward with the reintroduction of caribou to SHI. The target group for the source population was the Coats Island Herd, due to its close proximity and ecological and environmental similarities to Southampton Island. Regional Superintendent A. G. Loughery and Research Supervisor A.H. MacPherson began implementation of the program on June 7th, 1967. The very first animals on Southampton Island following their extirpation, were these 48 animals captured and transported from Coats Island.

From their start on Southampton Island, caribou were watched closely by wildlife officials. The first evidence of the success of the introduction was communicated by the game Management Officer Ed Bowden who estimated between 100 and 125 caribou ranging over the southern half of the island in the winter of 1971 (Game Management Files, 1971, 1972, 1973). The success of the re-introduction was soon realized and an aerial survey to estimate the population planned for June 1978. From 1978 to 1999 the Government of the Northwest Territories managed the progress of the 1948 reintroduction of caribou. With its formation in 1999, the Government of Nunavut (GN) took over this responsibility. The current GN management strategy follows a management plan developed in partnership with the Coral Harbour HTO and consists of a program relying upon regular aerial surveys and an extensive health monitoring program. Due to confirmed declines following the 2011 population estimate, the health monitoring component of population monitoring of the SHI herd, which included a one

hundred animal harvest for the assessment of health and condition, has been suspended to allow all available tags to go to the hunters of Coral Harbour.

The introduced caribou had steadily increased to an estimated population of 30,381 animals (95% CI=3,982; CV=0.066) in 1997, which represented the highest number of caribou ever recorded on the island. The 1997 estimate suggested a rate of increase of 23% between survey periods. However, the potential for a founder effect for this introduced population, leading to low genetic variability and increased susceptibility to disease and parasites was a concern. In February 2000, the reproductive disease *Brucellosis suis* was detected and grew to a prevalence of 58.8% by February 2011. High rates of *Brucellosis* in the population are thought to have been the main catalyst behind later declines observed following the June 1997 survey (Campbell, 2015). An aerial population survey conducted in 2003 detected the first decline of caribou since their introduction, showing a population estimate of 17,981 +/- 2,127 (95% CI=2,127; CV = 0.06) animals. The population remained relatively stable, or increased slightly, between June 2003 and a follow-up survey flown in June 2005. The June 2005 abundance survey estimated 20,582 +/- 3,056 (95% CI=3,056; CV=0.075), but the observed increase from 2003 results was not found to be statistically significant. The first evidence of a significant drop in abundance was recorded between June 2005 and June 2007 when survey results estimated 15,452 (95% CI=1,858; CV=0.06) caribou. This suggested a 14% decline from the June 1997 estimate (Campbell, 2015). The SHI caribou population continued its decline to 13,953 (95% CI=1,790; CV=0.07) in 2009, and to 7,902 (95% CI=1,261; CV=0.08) in June 2011. Following the introduction of a Total Allowable Harvest (TAH) by the 2012 harvesting season, the decline slowed and by May 2013, abundance was statistically stable at an estimated 7,287 (95% CI=1,045; CV=0.07) caribou (Campbell, 2015).

Over this same period, body condition did not appear to change, with the exception of a condition study in February and March of 2011 which showed that

the Southampton Island Herd was in the poorest condition reported since the initiation of the health monitoring program in March 1993 (Campbell, 2015). During the winters of 2010 and 2011, hunters reported numerous freezing rain events and extensive icing across the island. These icing events likely made winter forage less accessible to caribou (Tyler, 2010). Icing events that reduced accessibility to food could also have been associated with observed declines in condition, which further reduced reproductive success (Cameron et al. 1993, Gerhart et al. 1997). Support for this hypothesis stemmed from numerous local reports of starving and dead caribou during mid to late winter 2011 (Campbell, 2015).

Along with severe weather events, reproductive disease is thought to be a major contributor to overall population declines. Pregnancy rates declined from approximately 80% in 1997, to 60% in 2003, reaching a low of 36.3% in 2008, then climbing to 55.6% in 2010 only to decline again to 37.0% in 2011 (Campbell, 2015). The reproductive disease *Brucellosis suis* (Brucella) was first detected in February 2000 at a rate of 1.7% and by March 2011, rates of infection had risen to 58.8% by March 2011 (Campbell, 2015). High Brucella infection rates raised concerns regarding human health, as well as the ability of the SHI caribou herd to sustain and recover from substantial commercial harvesting and subsistence harvesting pressures.

Brucellosis and icing events are not the only issues threatening the SHI caribou population. Over-harvest has become a dominant threat to the long term sustainability of this population. In particular, a growing export market within Nunavut territory, driven in large part through caribou meat sales via social media, has been driving harvest levels beyond sustainable limits since 2011. Elements of the unrestricted sale of caribou meat are also driving increased harvest pressure on breeding females: customers on social media offer higher payment for fat caribou, which during the winter and spring seasons are predominantly pregnant females.

In this report we summarize the findings of over 20 years of monitoring on the SHI caribou herd and discuss trends in abundance, disease, harvest, and other long-term threats to the herd and their implications for management of this population.

2.0 Study Area

At 43,000 km^2, Southampton Island is the largest island in Hudson Bay. The island is divided into the Northern and Southern Arctic ecozones. The Northern Arctic ecozone covers White Island, and the northeastern third of Southampton Island including northern Bell Peninsula and can be further divided into the Boothia-Foxe Shield eco-province and then the Wager Bay Plateau ecoregion (**Figure 1**).

The Wager Bay Plateau ecoregion covers the northeastern Kivalliq Region, extending westward from the northern portion of Southampton Island on Hudson Strait to Chesterfield Inlet in the south, and as far west as the Back River (Wiken, 1986; Natural Resources Canada, 2001). The mean annual temperature is approximately -11°C with a summer mean of 4.5°C and a winter mean of -26.5°C. The mean annual precipitation ranges from 200 to 300 mm. This ecoregion is classified as having a low Arctic ecoclimate and is characterized by a discontinuous cover of tundra vegetation, consisting mainly of dwarf birch (*Betulaglandulosa*), willow (*Salix spp*.), northern Labrador tea (*Ledumdecumbens*), mountain avens (*Dryas integrafolia)*, and *Vaccinium* spp. Taller dwarf birch, willow, and alder (*Alnusspp.*) occur on warm sites, while wet sites are dominated by willow and sedge (*Carex spp*.). Lichen-covered rock outcroppings are prominent throughout the ecoregion. This ecoregion is composed of massive Archean rocks of the Canadian Shield that form broad, sloping uplands, plains, and valleys. It rises gradually westward from Chesterfield Inlet to 600 m ASL (above sea level) elevation, where it is deeply dissected. Turbic and static cryosols developed on discontinuous, thin, sandy moraine and alluvial deposits are the dominant soils in the ecoregion, while large areas of regosolic static cryosols are associated with marine deposits along the coast. Permafrost is continuous with low ice content. Naujaat and Baker Lake are the

main settlements within the ecoregion (Wiken, 1986; Natural Resources Canada, 2001).

The Southampton Island Plain ecoregion includes the remainder of Southampton Island and all of Coats and Mansel Islands (**Figure 1**). The mean annual temperature is approximately -11°C with a summer mean of 3°C and a winter mean of -24.5°C. The mean annual precipitation ranges from 200 to 300 mm (Wiken, 1986; Natural Resources Canada, 2001). This ecoregion is classified as having a low Arctic ecoclimate and is characterized by a nearly continuous cover of low Arctic shrub tundra vegetation, consisting of dwarf birch, willow, northern Labrador tea, mountain avens, and *Vaccinium* spp. Wet sites are dominated by willow, sedge, and moss. The region is composed of the partly submerged blanket of flat-lying Paleozoic carbonate rocks and is generally less than 90 m ASL in elevation. Bedrock outcrops are common. Static and turbic cryosols developed on level to undulating morainal and marine deposits are the dominant soils. The maritime influence is limited to the late summer and early fall. Coastal ice and fog persist for long periods in the summer when the sea ice is absent. The ecoregion is underlain by continuous permafrost with medium ice content composed of ice wedges. Coral Harbour is the largest settlement within this ecoregion (Wiken, 1986; Natural Resources Canada, 2001).

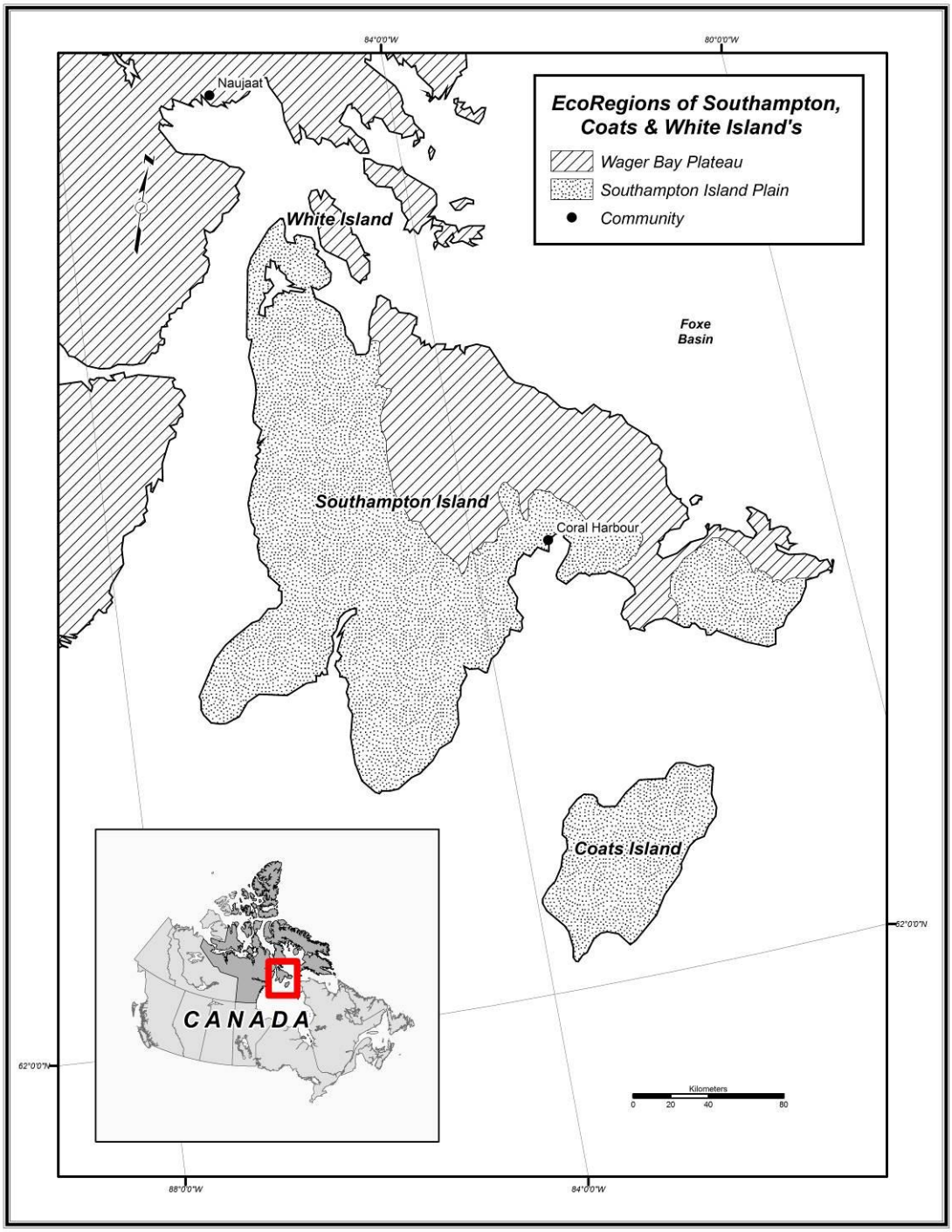

Figure 1 Ecoregions of the Southampton Island, Coats Island and White Island study areas (Wiken, 1986; Natural Resources Canada, 2001).

3.0 Methods

3.1 Caribou Introduction (1967) – An Historical Account:

Caribou reintroduced to Southampton Island from the Coats Island Herd were initially immobilized from a G2 helicopter using a CO_2 gas-operated Palmer 'Cap-chur' gun and both 2 cc and 5 cc darts. The darts used during the initial capture contained a pre-measured dose of crystalline succinylcholine ('Anectine') dissolved in isotonic water at a concentration of 5mg/cc and administered at a rate of 5 mg per 100 pounds (Eskimo 1968). The tranquillizer 'Largactil' at a concentration of 25 mg/cc and a dosage of 125mg per 100 pounds was used to maintain immobility. Up to seven animals were captured in this way, per day. Captured animals were taken to a base camp with an enclosure on Coats Island, where they were weighed, medicated with Vitamin E and Selenium, as well as an antihistaminic and anti-biotic, injected into the shoulder. Animals were held for up to one week. From the enclosure, animals were re-captured for transport by roping or tackling, tied up in slings, tranquillized and placed in single and twin Otter fixed wing aircraft for their final transport and release onto Southampton Island, in the vicinity of the airport. In total, 66 caribou, comprising 12 bulls, 26 cows (one pregnant), and 10 calves (8 male and 2 female), were captured and released onto Southampton Island. Of the original 66, 18 animals died, two from dart wounds, two from broken legs, and six from what appeared to be capture myopathy (CWS correspondence, 1969; Eskimo 1968). Reasons for the remaining eight deaths do not appear in the records examined. In total, 48 animals survived the reintroduction and make up the founding group of Southampton Island's reintroduced caribou herd.

3.2 Aerial Surveys (1978-2012):

Following their reintroduction onto Southampton Island, caribou were monitored periodically by both local and Government wildlife officials, primarily using ground-based methods. Kraft's aerial survey flown in November 1978 was the first scientific population estimate made since their re-introduction (Kraft, 1981). Kraft used a stratified transect survey method to cover 3 strata that were believed to represent the full extent of the Southampton Island Herd's distribution. The survey was flown between November 22nd and 25th, 1978 and utilized one observer on each of the left and right side of the single engine high wing DE Havilland Otter aircraft. Transects were placed 6.44 km apart for a total of 12.5 % coverage of the entire survey area. Effective strip width was a total of 800 meters, 400 meters out each side of the aircraft, while survey elevation was 122 meters AGL (above ground level) with a mean survey speed of 140 Kph (Kraft 1981). Population estimates were derived by calculating the density of caribou observed for all transect strips, and multiplying density by the total stratum area.

3.3.2 Random Block Survey Method:

A second survey method was employed in June 1986 and consisted of a stratified random block survey design (Heard and Grey, 1987). The census zone was divided into 5 strata which received differential coverage ranging from 11% to 54%. The stratification into census zones was based on a pre-survey reconnaissance, habitat and range preference (Parker, 1975), and recent observational data from both local hunters, wildlife service personnel, and previous survey observations of caribou. A Bell 206B helicopter was used as the survey vehicle at variable speeds and altitudes. The survey personnel consisted of two rear seat observers, a front left seat navigator, and a pilot. Sightings from all personnel were recorded. Each animal observed was approached and circled so that its sex and age could be determined. Heard and Grey also attempted to determine sightability through the re-surveying of portions of three blocks at three times the initial survey intensity and determining the differences recorded

between these surveys. This method of determining sightability was, however, unsuccessful do to the movement of animals between surveys. A third survey flown June 1991 aimed at improving the 1987 survey effort. The June 1991 survey followed, for the most part, the same methodology employed in 1987 by Heard and Grey (Ouellet, 1992). The main modifications made to the 1987 methods were made to ensure complete coverage of the island and involved the delineation of two strata defined as low density which were surveyed using an aerial strip transect survey flown with a Cessna 337 fixed-wing aircraft. Sampling intensity varied from 11% to 51% over 48 transects and/or blocks flown.

3.3.3 Single Observer Pair Method:

The November 1978, March 1990, June 1995, 1997, 2003, 2005, 2007, and 2009, surveys were flown using a single observer pair stratified systematic aerial strip transect method while the June 2011, May 2015, and May 2017 surveys were flown using a dependent double observer pair stratified aerial strip transect method. Little of the method has been documented for the March 1990 survey that was undertaken to estimate the SHI caribou population and distribution (Ouellet, 1992). The March 1990 survey was flown using a Cessna 337 fixed wing aircraft at 120 meters above ground level (AGL) at various speeds between 100 and 120 knots. The survey crew included two rear seat observers, a front right seat navigator, and the pilot. The strip width on each side of the aircraft was 400 meters. The survey covered the entire Island using 18 transects, which yielded 4% coverage, leading to low survey intensity and precision, resulting in an estimate of 9,319 (95% CI=6,341) caribou (Ouellet, 1992). Because of the low precision of the 1990 estimate, the survey was repeated in 1991 utilizing a different quadrat method. The 1991 survey estimated 13,676 (95% CI=3,105; CV=0.12), and being of greater precision, should be the relied upon estimate. A single observer pair stratified systematic aerial strip transect survey was flown in late June and early July of 1995, but there were serious problems of sightability as caribou were extremely hard to see due to their darker summer coats. Surveyors consider their population estimate result from the 1995 survey of

18,275 +/- 1,390 (95% CI) to be a major underestimate, and these results are therefore excluded from this report. Due to the sightability issues with the 1995 survey, a specific recommendation was made to conduct surveys earlier in June or before, prior to moulting (Mulders, pers. comm.). The survey to re-estimate the population was later flown in June 1997 and resulted in a population estimate of 30,381 (95% CI=3,982; CV=0.066).

Survey efforts in June 1997, 2003, 2005, 2007, 2009, 2011, 2012, May 2013, 2015 and 2017 utilized a stratified systematic aerial strip transect method flown with a high wing single engine turbine or gas, fixed wing aircraft. In 2013 surveys were moved from early June to mid-May as weather modelling indicated more "flyable" days during May. Additionally, the month of May provided more continuous snow cover for improved sightability, while maintaining distributions within June based strata. These findings lead to a permanent change in survey scheduling to May. Reconnaissance surveys used to delineate strata extents were flown in June of 1997, 2003, 2005, and 2007 (**Figure 2**). Although densities of caribou declined between 1997 and 2013, strata remained consistent with an even drop in relative densities across all survey strata, with the exception of White Island (**Figure 3**), where caribou abundance declined disproportionately more than on Southampton Island. Though strata remained similar between surveys, transect spacing did increase with decreasing relative densities within the Bell Peninsula and White Island strata. The largest single modification to strata occurred within the Low South strata in 2005 as a result of extensive flooding along the Boas River, which travels through the strata. In this case transects over the Boas River area were shortened to avoid flooded areas where caribou would not be found. Strip width (w) for all surveys were established using streamers or dowels attached to the wing struts, based on calculations described in Norton-Griffiths (1978) (**Figure 4**), and as follows:

$w = W * h/H$

where:

W = the required strip width**;**
h = the height of the observer's eye from the tarmac; and
H = the required flying height

Strip width calculations were confirmed by flying perpendicularly over runway distance markers or other fixed distance markers periodically throughout the survey. The strip width area for all abundance surveys was 400 meters per side.

Standardized reconnaissance transects with a total observation strip of 800 meters (400 meters per side) were flown during the June 1997, 2003, and 2005 surveys and used to stratify caribou into areas of similar relative densities, used later to allocate effort for the abundance phase (Heard 1987). A stratified random transect method was then used during the abundance phase of all surveys (**Figure 5 to 8**). Attempts were made to maintain a constant altitude of 400 ft. during the 1978, 1990, 1995 and 1997 surveys. A radar altimeter was employed during the 2003, 2005, 2007, 2009, 2011, 2012, 2013, 2015 and 2017 surveys to increase altitude precision between transects and survey years. The first transect within each of three strata (Low, Medium and High) was randomly placed along a line of latitude or otherwise randomly selected, with each sequential line being evenly spaced.

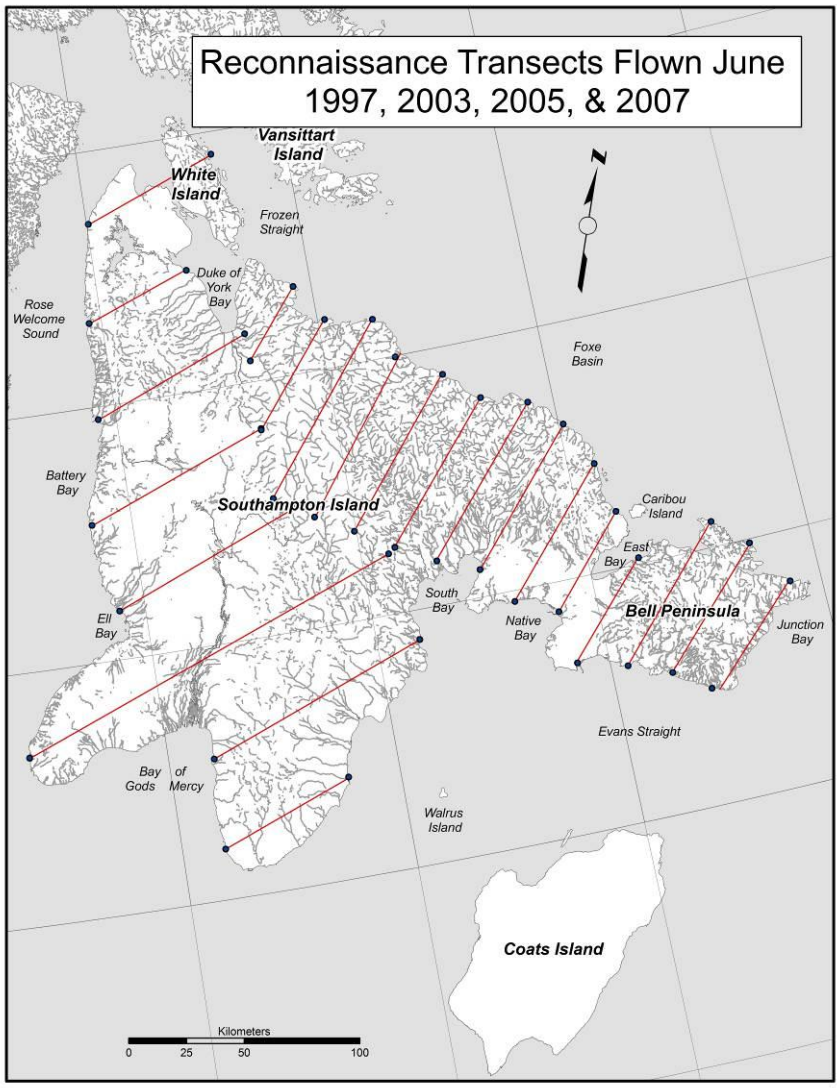

Figure 2. Reconnaissance transects flown in June of 1997, 2003, 2005, and 2007, to delineate abundance strata used to estimate Southampton Islands (including White Island) caribou population.

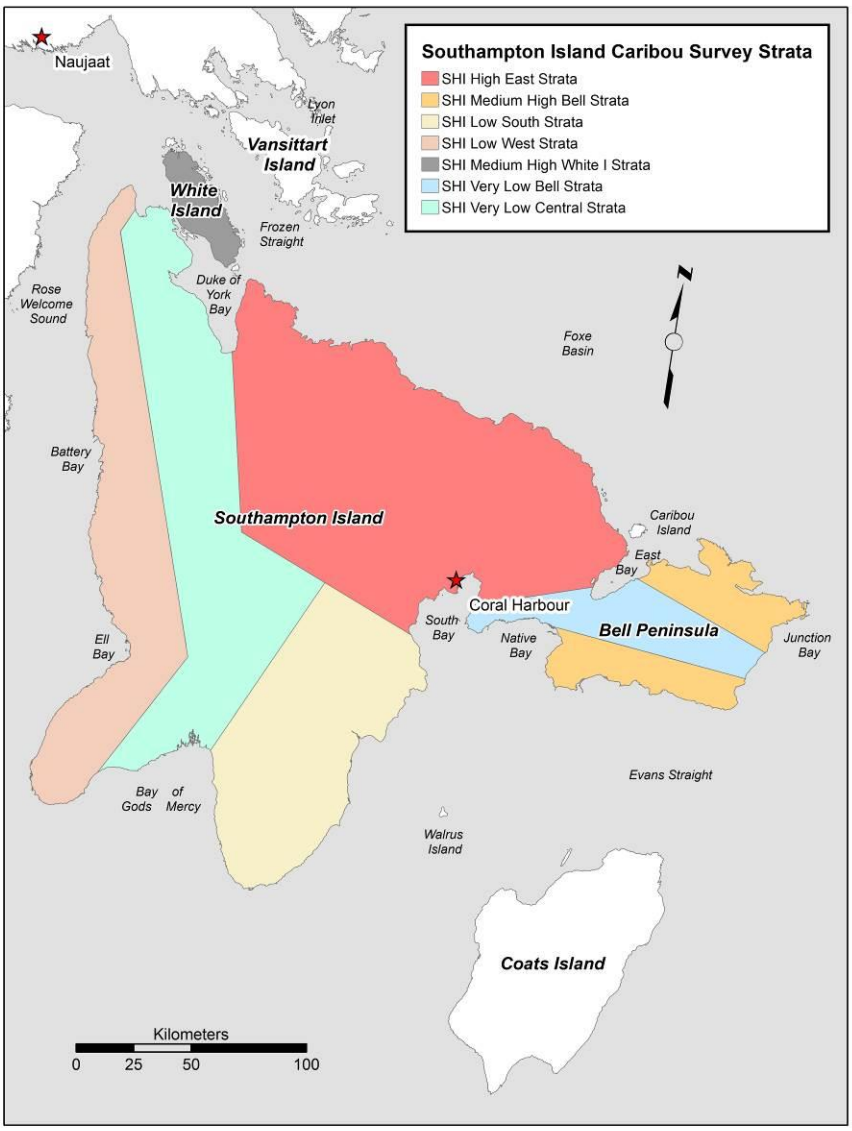

Figure 3. Abundance strata initially delineated using reconnaissance flights to map relative densities of caribou. As caribou distribution changed little across all survey years, these strata were utilized for all surveys post-2007.

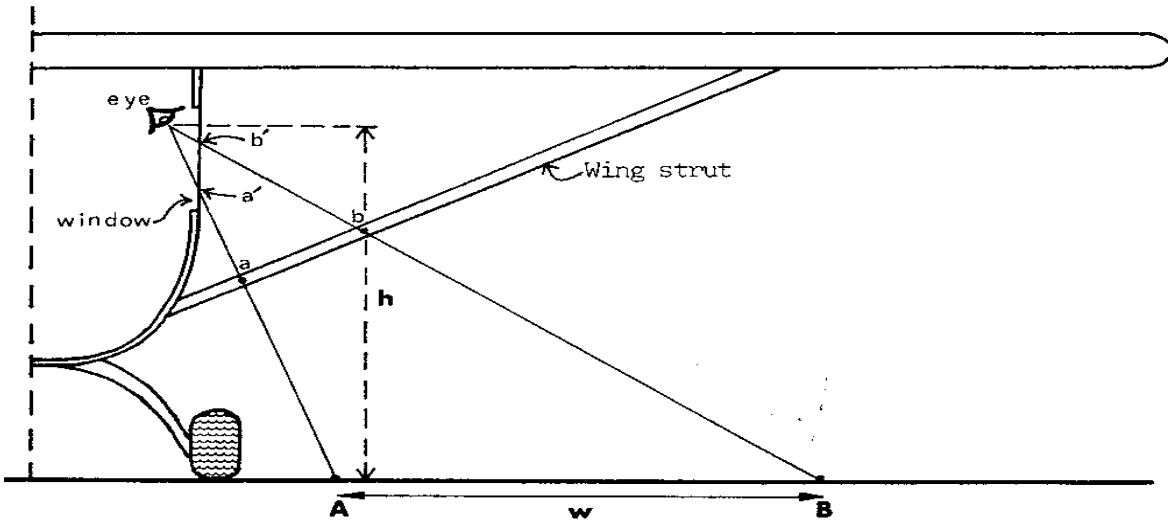

Figure 4. Schematic diagram of aircraft configuration for strip width sampling (Norton-Griffiths, 1978). W is marked out on the tarmac, and the two lines of sight a' – a – A and b' – b – B established. The dowels/streamers are attached to the struts at *a* and *b*. a' and b' are the window marks.

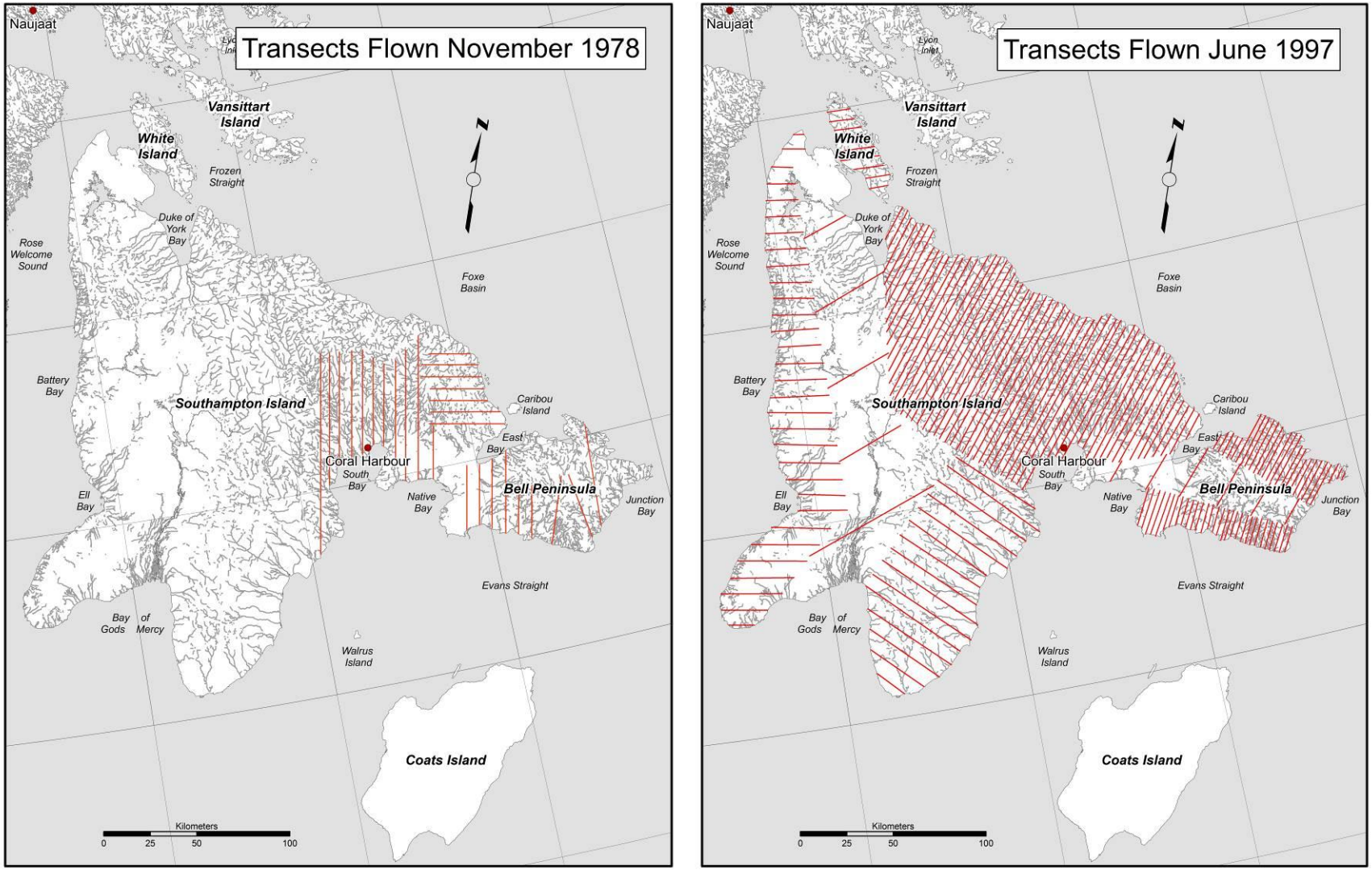

Figure 5. Stratified random transect surveys flown in November 1978 and June 1997.

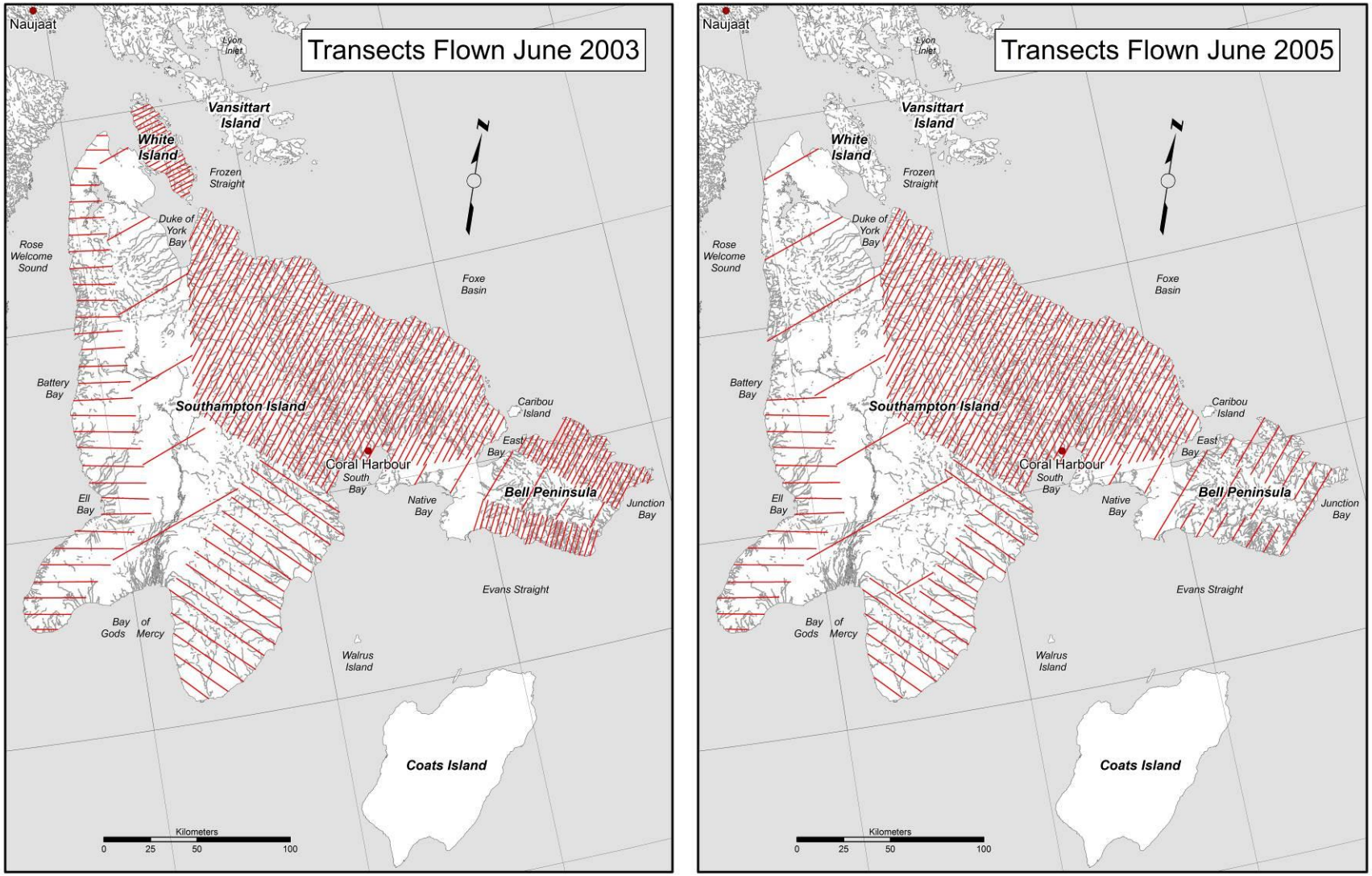

Figure 6. Stratified random transect surveys flown in June 2003 and 2005.

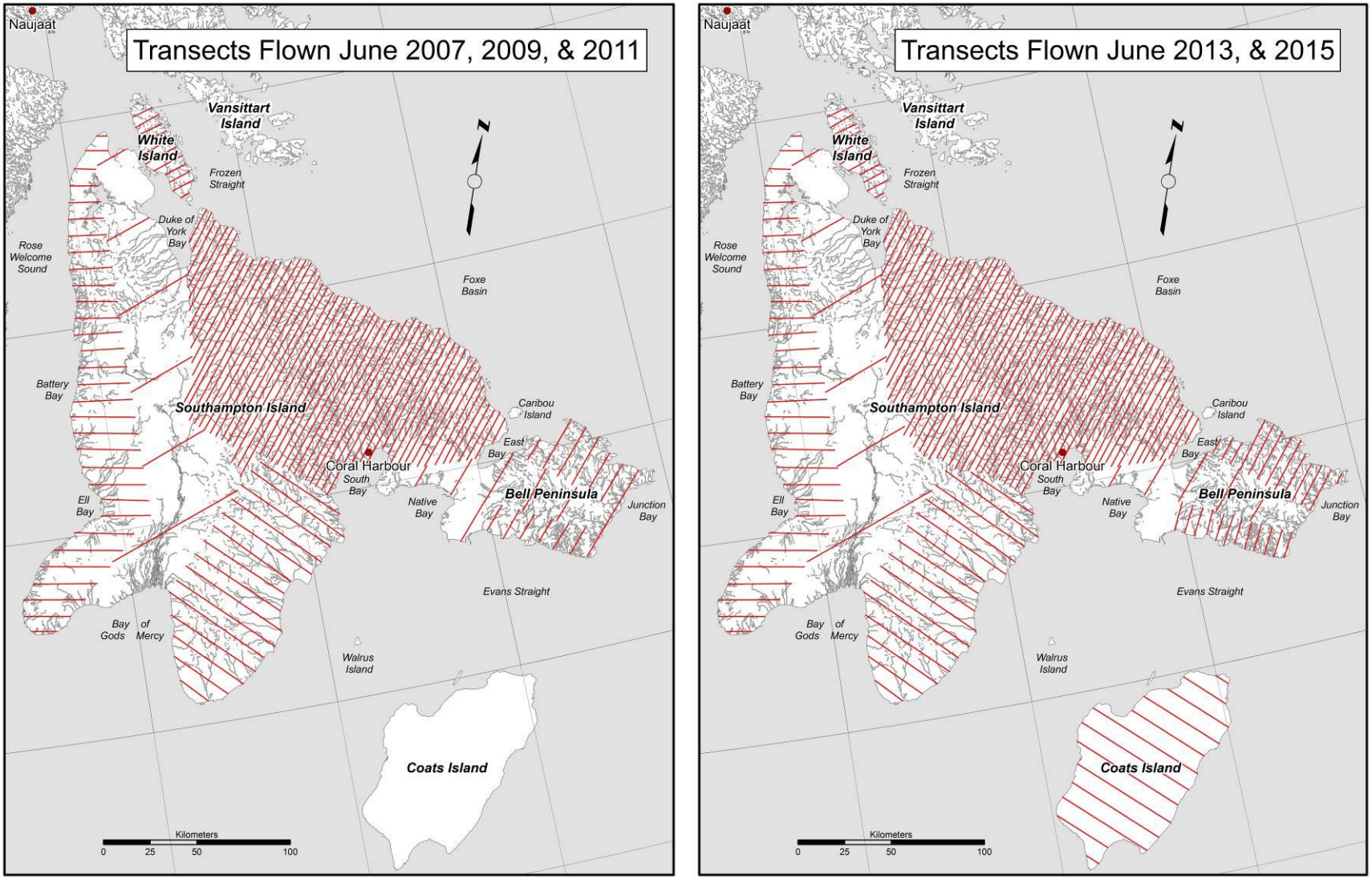

Figure 7. Stratified random transect surveys flown in June 2007, 2009, 2011 and June 2013 and 2015.

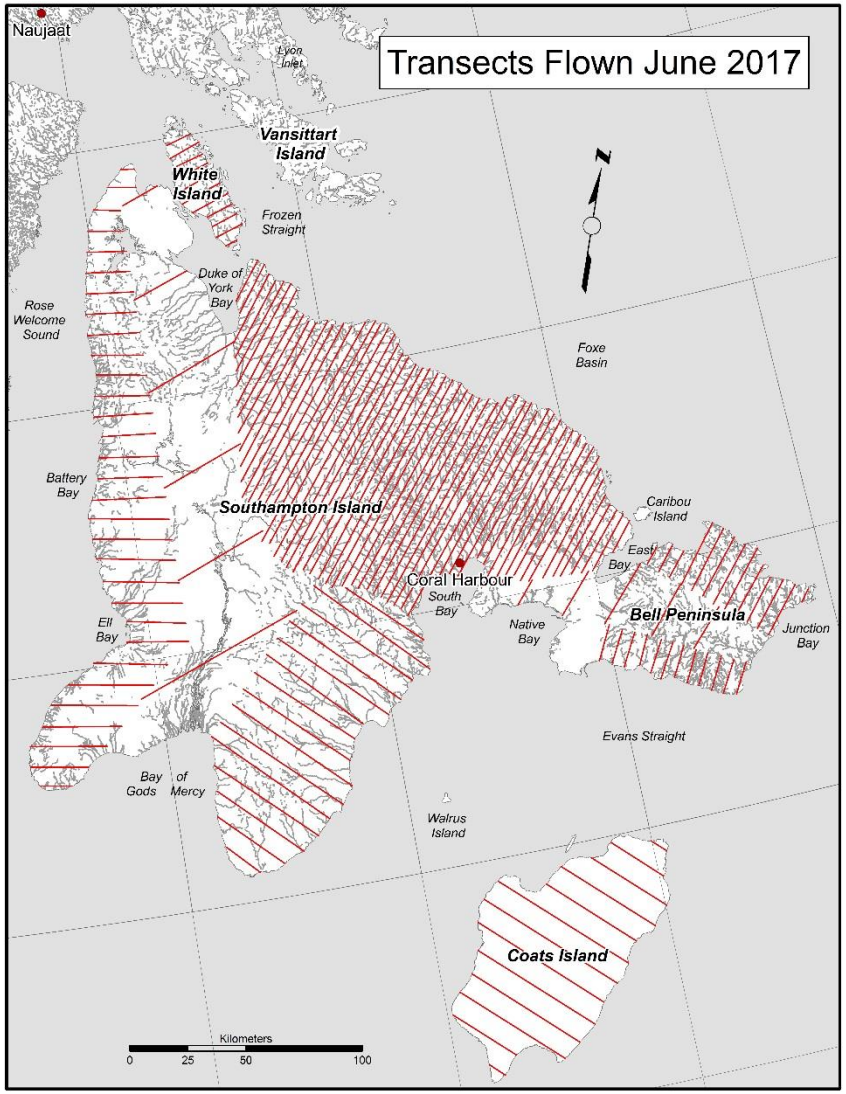

Figure 8. Stratified random transect survey flown in June 2017.

During the 1978, 1997, 2003, 2005, 2007, and 2009 population estimates the survey crew included a pilot (front left seat), a data recorder/navigator (front right seat), a left rear seat observer and a right rear seat observer. The pilot monitored air speed and altitude while following transects pre-drawn on 1:250,000 topographic maps (November 1978) or shooting waypoints on a Trimble GPS (June 1995, 1997). During the 2003, 2005, 2007, 2009, 2011, 2012, 2013, 2015 and 2017 surveys, transects were navigated using cloned pre-programmed routes on two Garmin C-176 (203 through 2013) and Montana 650T (2015 and 2017) geographic positioning system (GPS) units set to WGS 1984 datum and true north. The data recorder/navigator was responsible for assisting in the navigation of transects (1978 and 1997), and monitoring a second identically programmed GPS unit for the purposes of double-checking the position, altitude, distance from transect, and ground speed (2003, 2005, 2007, 2009, 2011, 2012, 2013, 2015, and 2017). Geographic coordinates (waypoints) and numbers of adult and calf caribou were either recorded on compact tape recorders with associated positions marked on a map (1997), or recorded on both maps and data sheets (1978), or recorded on data sheets (2003, 2005, 2007, 2009, 2011, 2012, 2013, 2015, and 2017). The responsibilities of the left side and right side observers was to monitor their 400 meter strips and call out numbers of caribou separated by adults and calves, both on and off transect as indicated by wing strut markers. The 2003, 2007, 2009 and 2017 air crews remained the same throughout the survey, while during the 2005, 2011 and 2015 surveys, one observer was changed part way through the survey. Information on the 1978, and 1997 surveys concerning consistency in air crews is lacking. All observational data including position were archived in a digital database and are included in Appendix 1.

Survey data from all surveys were initially analyzed using Jolly's Method 2 for unequal sample sizes (Jolly 1969 *In* Norton-Griffiths 1978). Only counts of adults and yearlings were used for the final population estimates as calves are not considered fully recruited into the population until they have survived their first

winter. Lake areas were not subtracted from the total area calculations used in density calculations.

3.3.4 Dependent Double Observer Pair Method:

The June 2011, May 2013, 2015, and 2017 surveys were marked by a change in visual survey method. An additional 2 observers and one data recorder were added to the survey crew increasing the crew to 7 individuals including the pilot. The method has been adopted to all Kivalliq regional ungulate surveys. Pilot studies conducted on Muskox abundance in 2010 and barren-ground caribou abundance in 2011, confirmed fewer animals were being missed while using this new configuration. Additionally more HTO representatives could be involved in the survey while maintaining two experienced observers covering each side of the survey aircraft. The new method is termed a dependent double observer pair visual method, and is set up with two left side observers and two right side observers with a data recorder for each.

The dependent double-observer pair method involves one “primary” (front) observer who sits in the front seat of the plane and a “secondary observer” (rear) observer who sits behind the primary observer on the same side of the plane (**Figure 9**). One data recorder sitting to the right of the pilot was assigned the right primary and secondary observers while the second data recorder, sitting on the rear left side was assigned the left primary and secondary observers. The method adhered to five basic steps; ***1*** **-** The primary observer called out all groups of caribou (number of caribou and location) he/she saw within the 400 meter wide strip transect before they passed halfway between the primary and secondary observer (approximately at the wing strut). This counting included caribou groups that were between approximately 12 and 3 o’clock for right side observers and 9 and 12 o’clock for left side observers. The main requirement was that the primary observer be given time to call out all caribou seen before the secondary observer called them out; ***2*** **-** The secondary observer called out whether he/she saw the caribou that the first observer saw and observations of

any additional caribou groups. Both the primary and secondary observers waited to call out caribou until the group observed passed half way between observers (between 3 and 6 o'clock for right side observers and 6 and 9 o'clock for left side observer); ***3 -*** The observers discussed any differences in group counts to ensure that they are calling out the same groups or different groups and to ensure accurate counts of larger groups; ***4 -*** The data recorder categorized and recorded counts of caribou groups into "primary only", "secondary only", and "both", entered as separate records; ***5 -*** The observers switched places approximately half way through each survey day (i.e. during re-fueling) to monitor observer ability. The recorder noted the names of the primary and secondary observers.

The sample unit for the survey was "*groups of caribou*" not individual caribou. This created problems for the data recorder trying to determine when a group of caribou differed from individual caribou that were separated by short distances. To resolve this issue, recorders and observers were instructed to consider individuals to be those caribou that were observed independent of other individual caribou and/or groups of caribou through an estimated separation of 100 meters.

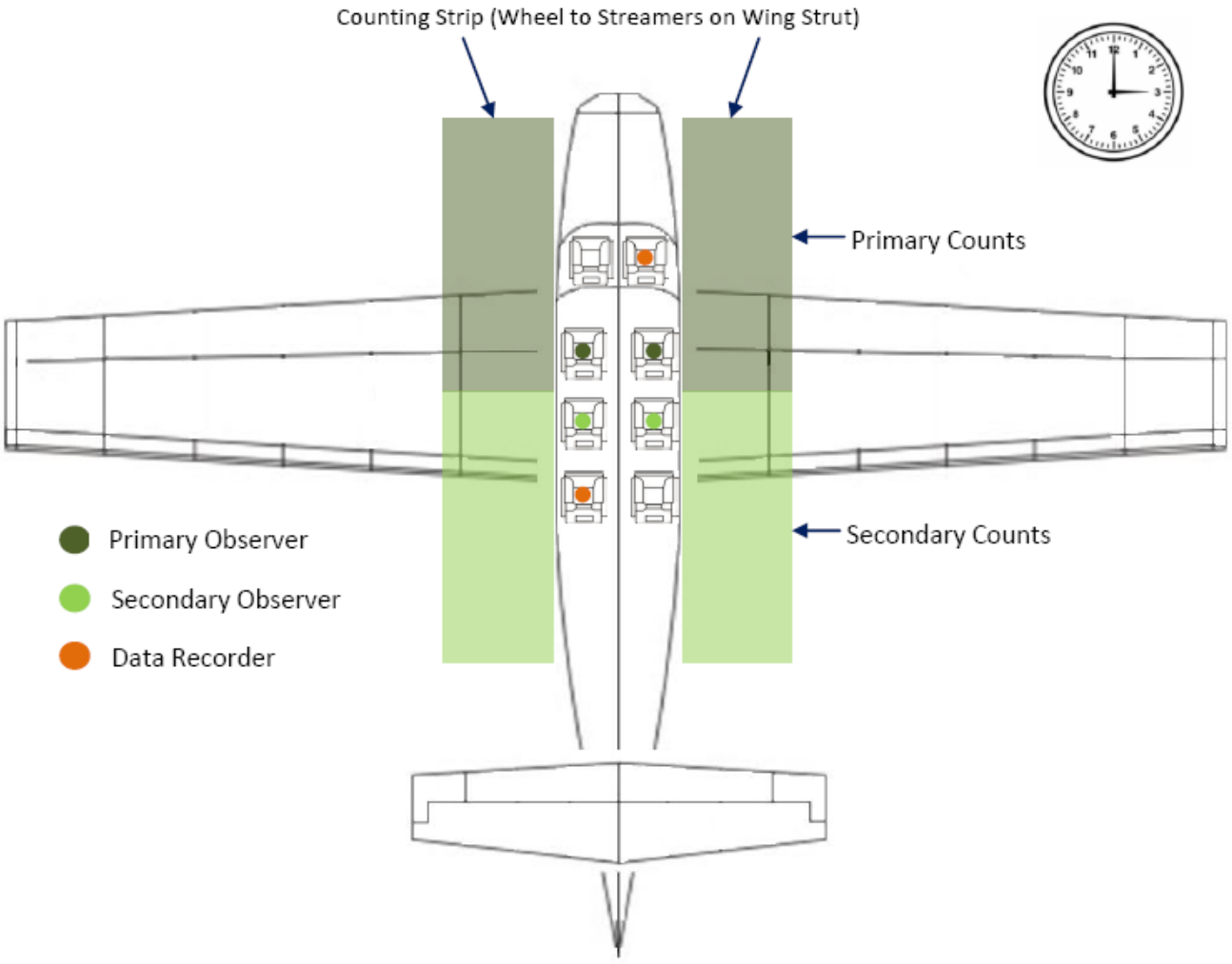

Figure 9. Observer position for the dependent double observer pair method employed on the 2011, 2013, 2015, and 2017 Southampton Island caribou abundance surveys. The secondary observer calls caribou not seen by the primary observer after the caribou have passed through the main field of vision of the primary observer. The small hand on a clock is used to reference relative locations of caribou groups (e.g. “Caribou group at 3 o’clock” would suggest a caribou group 90^{o} to the right of the aircrafts longitudinal axis.).

3.3 Distribution:

Distribution maps were developed to graphically summarize survey data for survey observations up to and including June 2007. The distribution maps were generated through an interpolation which provided an estimate of the number of animals present at un-sampled locations based on the known values gathered at surrounding locations. This type of analysis generates a surface consisting of cells, each with an attribute (in this case, population density), used to interpret the spatial distribution of geographic points and then convert them into a continuous distribution reflecting estimates of point densities. In this study an Inverse Distance Weighted (IDW) interpolation technique was used within ArcMap's Spatial Analyst extension. IDW is an effective means of interpolating scattered data points. It assumes that the resulting interpolated surface should be influenced most by the nearby points and less by more distant points. It estimates values by calculating a weighted average. The farther a sampled point is from the cell being evaluated, the less weight it has in the calculation of the cell's value (Watson and Phillip, 1985).

To account for null data, all survey observations were first buffered to ten kilometers. To account for nil records those portions of the transect not covered by the observation buffers were then divided into 5 kilometer segments with the first starting 5 km from the edge of the nearest buffered observation or transect starting point. At each division between 5-km segments, a point with an observed value of zero was inserted. The analysis was then run using the survey observation values as well as the newly populated zero values. The analysis requires that a series of parameters be defined. The parameters, along with a description and the settings used are summarized in **Table 1**.

The resulting surfaces were themed by the population density attribute and overlaid on a base map of Southampton and White Islands to develop the

figures. Density class or “bins” are developed to reveal the most visual information and highlight and estimate distributional changes between surveys. As area estimates of relative densities on Southampton Island are mostly aggregated in the 0 to 5 caribou/km^2 class, the bins were developed to accentuate these lower relative densities.

Table 1. Inverse distance weighting (IDW) analysis parameters employed in the analysis of Southampton Island caribou densities from 1978 through 2007.

Parameter	Description	Settings
Z Value Field	The Z value is the attribute being used to derive the interpolated surface.	The population attribute stored in the field Number.
Power	The higher the Power value, the greater the influence of values closest to the interpolated point. The most common value for the Power parameter is 2.	A value of 2 was selected.
Search radius type	The search radius can be either variable or a fixed distance.	As the sample points were not evenly distributed, a variable search radius was used that assessed the data points nearest to the particular cell of interest.
Search radius setting	The search radius setting specifies either the maximum distance of a fixed radius search or the number of points for a variable type.	The number of points considered in each of the analyses was 12.
Output cell size	The resolution (or cell size) of the grid (the surface) resulting from the analysis.	An out cell size of 100 m2 was specified resulting in a population density of animals per hectare.

3.4 Statistical Analysis of Abundance and Trend:

3.4.1 Strip transect surveys (1997-2017):

The standard Jolly estimator (Jolly 1969, Krebs 1998) was applied to the strip transect data for all years with an assumed strip with of 800 meters for all years except 2015 where the strip width was 918 meters. Strip transect data for 2017 was created using the first 2 bins of the distance sampling data which amounted to an 800 meter strip. Strata were estimated separately and then combined for a total estimate for the Island. Coats Island was also surveyed in 2013, 2015, and 2017 and was excluded from the South Hampton Island estimate. Log-normal confidence limits were generated on the estimates (Thompson 1992).

3.4.2 Double observer/strip transect analyses (2011, 2013, 2015, and 2017):

Given that this method assume equal sightability between observers it is essential that the observers switch seats over the course of the survey (Cook and Jacobsen 1979). Estimates of herd size and associated variance were measured using the mark-recapture distance sampling (MRDS) package (Laake et al. 2012) in the program R (R Development Core Team 2009). In MRDS, a full independence removal estimator which models sightability using only dependent double observer pair information (Laake et al. 2008a, Laake et al. 2008b) was used making it possible to derive dependent double observer pair strip transect estimates. Strata-specific variance estimates were calculated using the formulas of Innes et al. (2002). Data were explored graphically using the ggplot2 (Wickham 2009) package in R.

3.4.3 Modelling of sighting probability variation:

One assumption of the dependent double observer pair method is that each caribou group observed had an equal probability of being sighted. To account for differences in sightability we also considered the following sightability covariates in the MRDS analysis (**Table 2**). Each observer pair was assigned a binary individual covariate and models were introduced that tested whether each pair had a unique sighting probability. Previous

analyses (Campbell et al. 2012; Boulanger et al. 2014) suggested that the size of the group of caribou had strong influence on sighting probabilities and therefore we considered linear and log-linear relationships between group size and sightability (**Table 2**). Cloud and snow cover were recorded by data recorders as they changed and were included in the analysis as ordinal rankings. We suspected that sightability was most likely lowest in mixed snow cover conditions and therefore we considered both categorical and linear models to describe variation in sightability caused by snow cover. Cloud cover could also influence sightability by causing glare, flat light, or variable lighting. We used the same basic strategy to model cloud cover variation as we did for snow cover variation.

The fit of models was evaluated using the Akaike Information Criterion corrected for small sample size (AIC_c). The model with the lowest AIC_c score was considered the most parsimonious, thus minimizing estimate bias and optimizing precision (Burnham and Anderson 1998). The difference in AIC_c values between the most supported model and other models (ΔAIC_c) was also used to evaluate the fit of models when their AIC_c scores were close. In general, any models with a ΔAIC_c score of less than 2 between them were considered to have equivalent statistical support.

3.4.4 Distance sampling analyses (2017):

For the 2017 survey, distances of caribou groups from the survey planes were binned into intervals (0-200m, 201-400m, 401-600m, 601-1000m, and 1001m-1500m), based upon markers on wing struts of the survey plane, as was done in the 2014 Baffin Island caribou survey (Campbell et al. 2015). In addition, the dependent double observer pair also assessed sightability of caribou in the 0-200 meter strip closest to the aircraft.

A combined distance sampling and mark-recapture approach was used to estimate abundance for the 2017 data set. The basic approach involved using mark-recapture analytical methods to estimate the probability of detection of caribou at 0 distance from the survey plane and distance sampling methods to estimate the decrease in probability of detection at greater distances from the plane. This approach ensured a more robust

estimate than using distance sampling methods alone which assume that the probability of detection of caribou groups at 0 distance from the plane is 1 (Borchers et al. 1998, Buckland et al. 2004, Laake et al. 2008a, Laake et al. 2008b, Buckland et al. 2010, Laake et al. 2012).

As with the dependent double observer pair analysis, the MRDS R package (Laake et al. 2012) was used to build mark-recapture and distance sampling models. The general approach used was to build distance sampling models with the mark-recapture model parameters held constant. Once a parsimonious distance sampling model was identified, the mark-recapture model was built to further assess sightability of caribou in immediate proximity to the aircraft. The same general set of covariates used in the dependent double observer pair analysis (**Table 2**) were used for both the dependent double observer pair and distance sampling models. As with the dependent double observer pair analysis, AIC methods were used to assess model fit. Overall model fit was also assessed using goodness of fit tests as well as graphical comparison of detection functions with histograms of frequencies of observations from the survey.

3.4.5 Trend analyses:

We used log-linear models to analyze trends for the increase and decrease phase of the caribou abundance dataset (McCullough and Nelder 1989, Thompson et al. 1998, Williams et al. 2002). Our models assumed an underlying quasi-Poisson distribution of estimates with population change occurring on the exponential scale. Abundance survey estimates were weighted by the inverse of their variance therefore giving more weight to the more precise estimates. A log-link was used for the analysis allowing direct estimates of yearly rate of change as one of the regression β terms. Additive terms were used to estimate phase-specific trends and the effect of a possible immigration event, likely occurring between May 2013 and May 2017, on SHI herd trend.

Table 2. Covariates used to model variation in sightability of caribou for the dependent double observer pair analysis conducted on the 2017 abundance survey of the Southampton Island caribou herd.

Covariate	Acronym	Description
Observer pair	observers	each unique observer pair
Group size	size	size of caribou group observed
	Log(size)	Natural log of group size
Snow cover	snow	snow cover (0,25,75,100)
	snowc	continuous
Cloud cover	cloud	cloud cover (0,25,75,100)
	cloudc	continuous

3.5 Condition and Disease Sampling:

The health and condition of SHI caribou condition was monitored through the collection of harvest samples, beginning in 1995 (Campbell, 2015). Variables measured included a ratio of bare kidney to kidney fat index, the recording and sampling of any apparent disease and/or diseased tissue, the recording and sampling of parasitic infections, the measurement of back fat, bone marrow condition (in some years), pregnancy rates, fetal sex (in some years), and age through the analysis of cementum-annuli from the sampling of I-1 (the first incisor) from the lower jaw. In the case of the GN health studies, all anatomical components of an individual caribou being sampled and/or measured were recorded along with a common tag number and the associated harvest year. This common tag number allowed for the pooling of analysis results to provide a comprehensive description of the health, age and sex of the individual being sampled. From 1995 through 1999, approximately 400 animals per year were sampled in this way. Sampling across February and March 2000 through 2009 was reduced to approximately 200 to 300 animals (excluding 2001, 2002 and 2003). Prior to 2009, all sampling was carried out in conjunction with the commercial harvest, which ran from mid-February through to early April in most years. Following the cessation of the commercial harvest in 2009, harvest numbers in 2010 and 2011 were reduced to 100 animals. Following the 2011 survey results and subsequent application of a total allowable harvest (TAH) in 2012, the community of Coral Harbour requested that the 100 animals harvested for body condition be suspended so that all TAH allocations could be provided to the community. The suspension of the condition sampling harvest has remained in effect to-date.

The kidneys sampled for the kidney fat index (either left or right) were selected based on the amount of fat surrounding the kidney. In all cases the fatter of the two was chosen. The thickness of back fat was measured along a line 5 to 10 centimeters from the base of the tail, perpendicular to the spine. Measurements were taken from the thickest fat deposits on the rump (one to two inches to the left and right of the base of the tail) on

mainly the left side but also on the right side when left-side fat was obviously removed during the skinning process.

As a standard protocol, the Canadian Food Inspection Agency (CFIA) randomly collected between 300 and 400 blood samples from commercially harvested animals from 1993 through 2007. From 2007 to 2011, blood samples were collected by GN biologists from animals collected for health and condition harvests, from remaining ventricular and/or arterial blood. Sampled blood was drained into red topped vacutainers, left to stand approximately two hours at between five and ten degrees Celsius, then spun down in a centrifuge for approximately ten to fifteen minutes to separate the serum from cellular material. Individual serums were poured off into new sterile red-topped vacutainers, carefully packed and allowed to freeze at approximately -20^{o} to -30^{o} degrees Celsius. Frozen blood serums were then transported first to labs in Lethbridge, Alberta for Brucellosis, and Tuberculosis screening then to the CFIA lab in Ontario for further disease testing. Adult female reproductive tracts were also collected in 2005 for the purposes of identifying reproductive stress and/or disease. All sampling pre-2009 was carried out in conjunction with the commercial harvest which ran from mid-February to early April. The GN did not have access to all CFIA test results.

3.6 Genetic Analysis – Movement:

Over the winter of 2014, Coral Harbour hunters reported caribou tracks crossing the ice from the mainland across to the northwestern extents of SHI. Though no estimates of the total number of caribou involved in this crossing were communicated beyond “hundreds”, local hunters had observed more calves in June 2014 and increased densities of caribou in the following harvesting year, compared to preceding hunting seasons. Results from the May 2015 abundance survey estimated a significant increase in SHI caribou abundance of both adults and calves, compared with 2013 results. This population increase was theorized by both the community of Coral Harbour and Wildlife officials’ to

be related to the immigration of mainland caribou onto SHI. We set out to further investigate this hypothesis using population genetics.

We engaged Wildlife Genetics International (WGI) to pursue this question using the clustering programs Structure and Genetix, which produce accessible visual summaries of the results (Paetkau, 2015). Caribou tissue collected on SHI in 2004, and 2014, as well as tissue samples collected from hunters in the Naujaat (Repulse Bay) area, were compared for assessments of ancestry. Additionally, WGI used archived samples from the Qamanirjuaq caribou herd for added comparative analysis with the SHI herd. WGI used GeneClass2 to assess ancestry hypotheses, explicitly. Initial explorations included data from South Baffin Island, Melville Peninsula, and Ahiak/Beverly, but these explorations did not identify any associations of relevance to SHI.

Genotyping was performed by Wildlife Genetics International (WGI) using a standard set of 18 highly variable microsatellites that they had consistently employed for other caribou genetic analyses in Nunavut, Northwest Territories, British Columbia, and Alberta. The analysis proceeded in two rounds of 9 markers (including gender markers), as all 18 markers cannot be loaded into a single sequencer lane. After completing a first pass with all 18 markers, WGI did a round of reanalysis ('cleanup') of individual data points that were scored with low confidence (1) during the first pass (Paetkau, 2015). This reanalysis used 5 µL of DNA per reaction, up from the 3 µL used for first pass. In some cases multiple attempts were made to confirm problematic data points. At the end of the cleanup phase, 6 samples from SHI still had low-confidence scores in their genotypes (Paetkau, 2015). In total, WGI was able to successfully genotype complete 18-locus genotypes for 37 samples from Naujaat, and 131 from SHI. With genotyping completed, WGI defined an individual for each unique multilocus genotype, taking identifiers from the first sample to be assigned to each individual, of which 37 samples from Naujaat were assigned to 34 individuals (10M:24F), and the 131 samples from SHI in 2014 were assigned to 127 individuals (76M:51F). None of these animals had previous detections in the greater Nunavut dataset including samples from the 2004 harvesting season.

Paetkau (2015) then used resampling in the software GeneClass2 to generate 10,000 simulated mainland and island genotypes, and plotted the distribution of the island/mainland likelihood ratio to produce critical values for statistical testing (Paetkau *et al.* 2004 *Mol. Ecol.*). By way of example, 99% of simulated island genotypes had a log likelihood ratio in excess of 9.

4.0 Results and Discussion:

4.1 Population Distribution:

In discussing changes in distribution of caribou on Southampton Island, it is important to note that although an island population, some exchange with the mainland likely occurred on a very small scale during winters when an ice bridge had formed across Roes Welcome Sound (Local Knowledge). According to island residents, during most winters, Roes Welcome sound does not freeze over completely creating an effective barrier to caribou movement. If such a change was to occur however, this exchange would most likely have been with the Wager Bay population of caribou occupying the Lyon Inlet area due to its closer proximity to SHI. Though tracks have been observed of caribou on the ice of Roes Welcome Sound in late winter (going both east to the island and west from the island) there has been no documented evidence of a successful crossing.

Between 1968 and 1978, the first ten years of caribou occupancy on SHI following re-introduction, monitoring was mainly conducted using ground observations. During this period observations of caribou taken during patrols whether by ground or by air suggested caribou had spread considerably across the Island (Figure 10). Kraft's aerial survey in November 1978 was the first to estimate the population since its re-introduction. Both transects and points of observation were digitized off report figures and used with IDW to produce an estimate of abundance of the newly introduced Southampton Island Caribou herd (**Figure 11**). At this time, caribou were largely aggregated in the shoulder area east and northeast of Coral Harbour, the south shore of Bell Peninsula, and along the coast just south of the town of Coral Harbour. No animals were observed by either hunters nor during aerial reconnaissance conducted in previous years, anywhere further north nor west of the areas indicated.

After the examination of all available observational data up to 1987, Heard and Gray (1987) concluded that there always appeared to be caribou in the core areas of Bell Peninsula and the Kirchoffer River uplands, northeast of Coral Harbour (Heard and Grey, 1987). These observations made during their 1987 aerial population estimate showed an expansion of the herds distribution further north and west, and throughout the coastal strip encompassing Bell Peninsula. Unfortunately, point data are not available from this survey. Caribou distribution estimated on SHI in June 1991 was similar to that recorded in 1987. Oullett (1992) suggested that caribou range did not appear to expand between 1987 and 1991 even though their numbers increased substantially. Oullett (1992) found that to accommodate growth, densities simply increased within the existing range. Once again, point data is not available for either of Oullett's 1990 or 1991 surveys.

All surveys conducted from 1997 to present have point data, from which to base the analysis. The 1997 results also showed little in the way of distributional change since Oullett's observations in 1992, although densities had continued to increase significantly (**Table 3 & 4**) (**Figure 12**).

An examination of distributional change following introduction was made using IDW with ground and aerial survey point data (**Figures 12, 13, 14, and 15**). These analyses suggests that caribou distribution increased across the Island from the point of introduction, up to the 1987 survey year, at which time the now rapidly increasing population stopped expanding its range, suggesting that the herd had reached full occupancy of usable caribou habitat on the island. Caribou had occupied the southern portions of the island including Bell Peninsula and inland toward the central portions of the island along the Kirchoffer River watershed by as early as 1983. As observed by Oullett (1992) densities increased over the same geographic areas from 1987 up until a period between 1998 and 2002, at which time densities over the same areas decreased significantly. The most dramatic decrease in densities was in Bell Peninsula

between the 2005 and 2007 surveys (Figure 14 and 15). Caribou had all but abandoned the area, likely as a result of overgrazing from years of higher relative densities. Today the central portion of the island remains the most highly used habitat by SHI caribou. Ecologically, this area is where the western flats meets the eastern highland, creating an ecotone between the Wager Bay Plateau and Southampton Island Plain Ecoregions.

Observed distributions from the 2003 survey indicated little change from distributions observed during previous surveys though localized densities had decreased. The first distributional change was recorded in June 2005, at which time, there was a noticeable decrease in the numbers of caribou occupying Bell Peninsula (**Figure 14, and 15**). According to survey density estimates, this declining trend in Bell Peninsula continued through June 2007, and was also reported by local hunters.

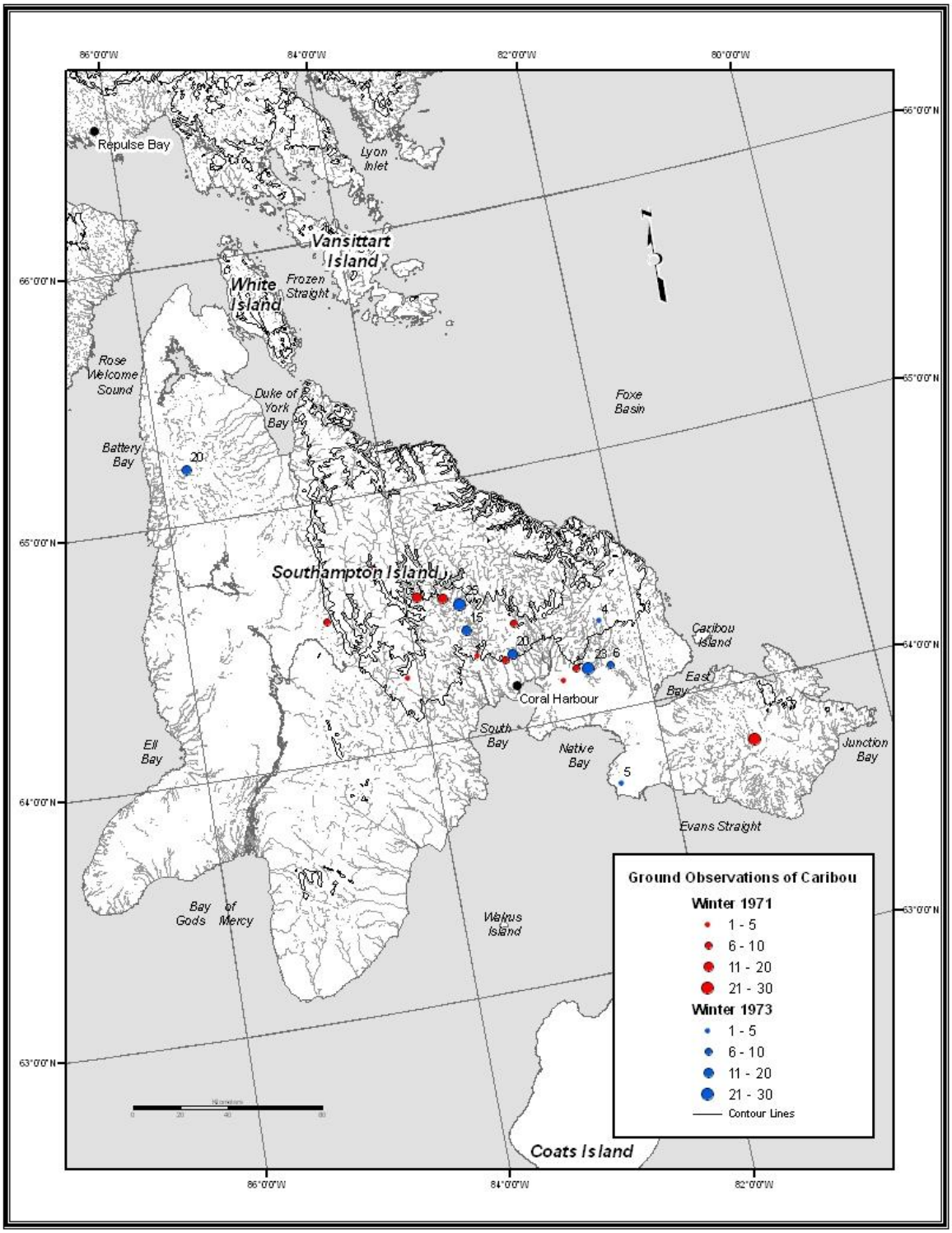

Figure 10. Observations of caribou on Southampton Island (1970 to 1973).

Table 3. Inverse distance weighting (IDW) values for the entire Southampton Island study area including White Island and Bell Peninsula showing changes in adult caribou density through time. Results from June 2005 were removed as White Island was not surveyed in that year.

Caribou/km²	1978		1997		2003		2007		% Change			
	km²	Percent	km²	Percent	km²	Percent	km²	Percent	1978 to 1997	1997 to 2003	2003 to 2007	1997 to 2007
0-1	43,471	98.5%	31,591	71.6%	32,674	74.0%	36,514	82.7%	-26.9%	2.4%	8.7%	15.6%
1-2	345	0.8%	4,535	10.3%	4,741	10.7%	3,862	8.8%	9.5%	0.4%	-1.9%	-1.5%
2-5	230	0.5%	4,836	11.0%	5,190	11.8%	3,248	7.4%	10.5%	0.8%	-4.4%	-3.6%
5-8	58	0.1%	1,680	3.8%	1,202	2.7%	434	1.0%	3.7%	-1.1%	-1.7%	-2.8%
>8	22	0.0%	1,484	3.4%	319	0.7%	68	0.2%	3.4%	-2.7%	-0.5%	-3.2%
Total	**44,126**	**100.0%**	**44,126**	**100.0%**	**44,126**	**100.0%**	**44,126**	**100.0%**				

Table 4. Inverse distance weighting (IDW) values for the entire Southampton Island study area including White Island and Bell Peninsula showing changes in the density of observed calves through time.

Caribou/km²	2003		2005		2007		% Change		
	km²	Percent	km²	Percent	km²	Percent	2003 to 2005	2005 to 2007	2003 to 2007
0-1	43,276	98.1%	44,062	99.9%	44,042	99.8%	1.8%	-0.1%	1.7%
1-3	801	1.8%	60	0.1%	82	0.2%	-1.7%	0.1%	-1.6%
>3	49	0.1%	4	0.0%	2	0.0%	-0.1%	0.0%	-0.1%
Total	**44,126**	**100.0%**	**44,126**	**100.0%**	**44,126**	**100.0%**			

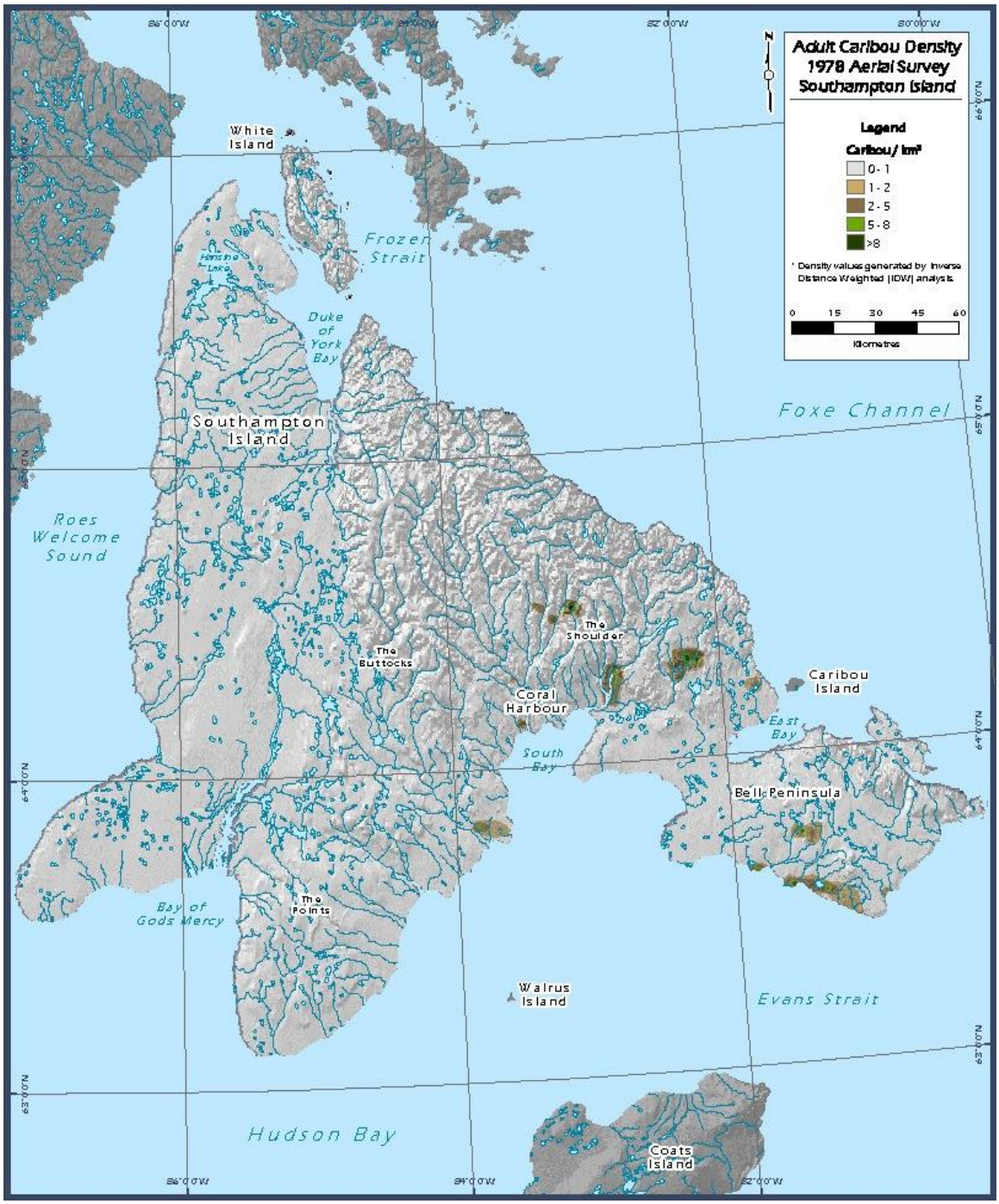

Figure 11. Results of the inverse distance weighted (IDW) interpolation technique applied to November 1978 abundance survey observations showing relative density of barren-ground caribou (*Rangifer tarandus*) on Southampton Island..

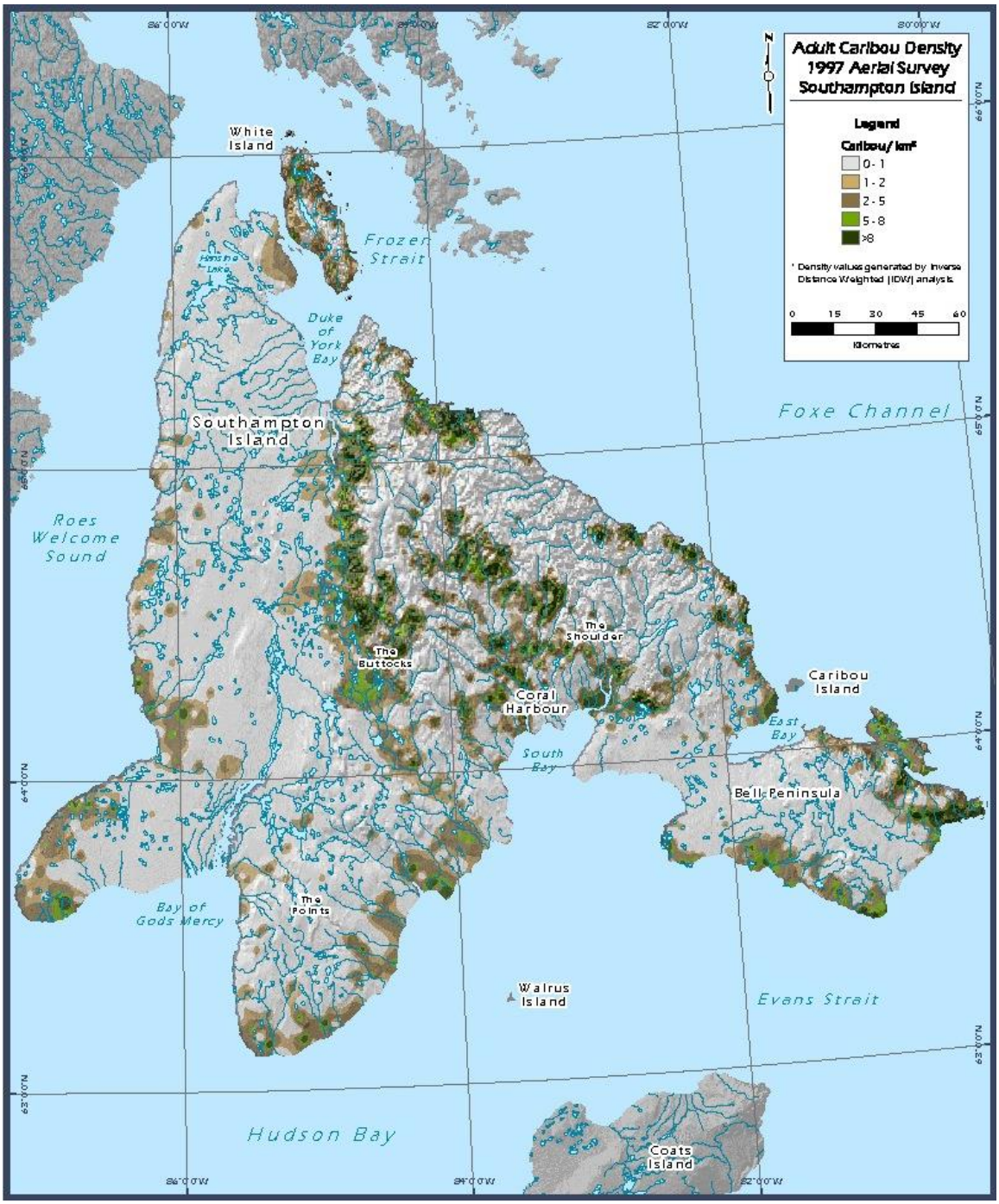

Figure 12. Results of the inverse distance weighted (IDW) interpolation technique applied to June 1997 abundance survey observations showing relative density of barren-ground caribou (*Rangifer tarandus*) on Southampton Island.

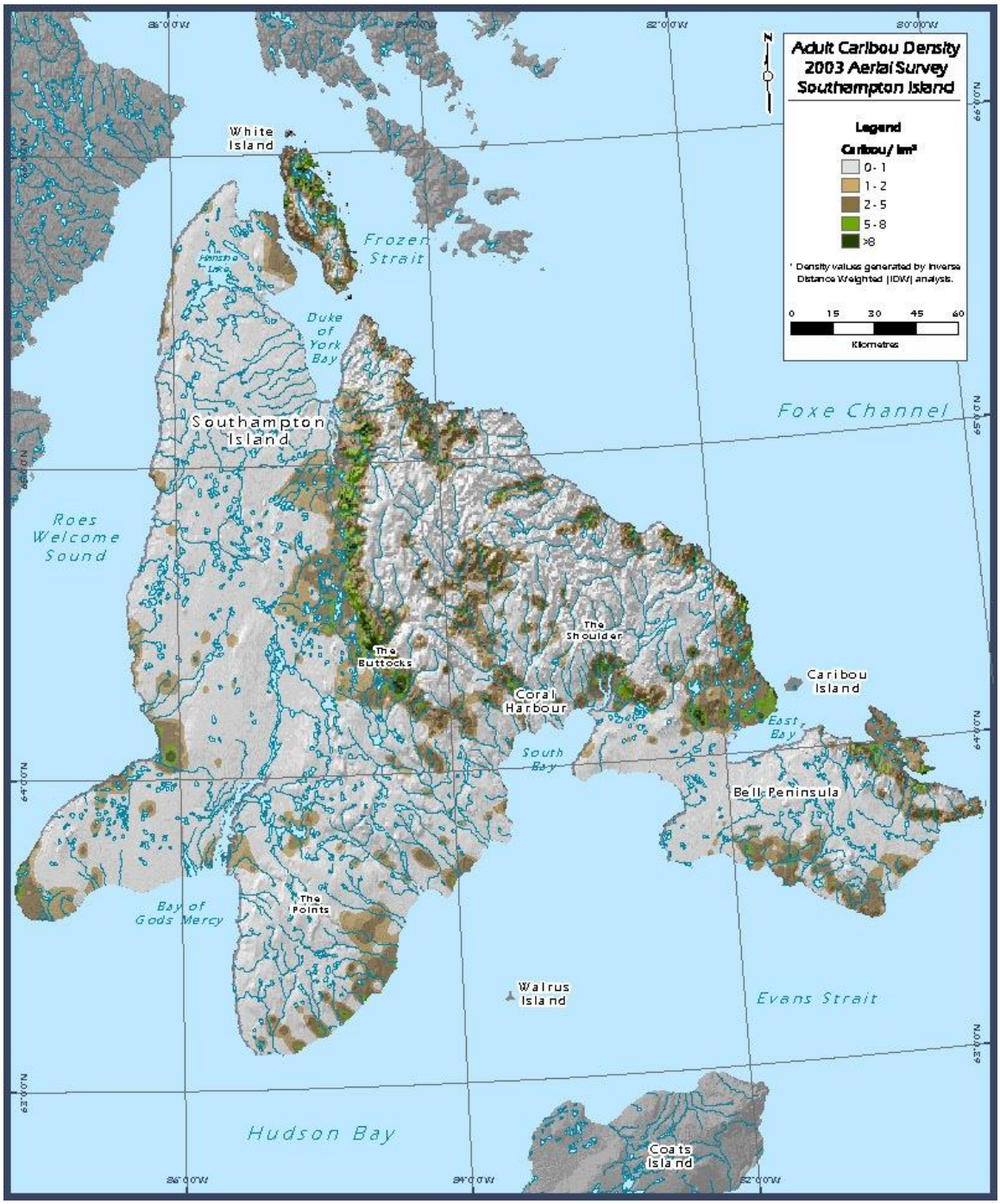

Figure 13. Results of the inverse distance weighted (IDW) interpolation technique applied to June 2003 abundance survey observations showing relative density of barren-ground caribou (*Rangifer tarandus*) on Southampton Island.

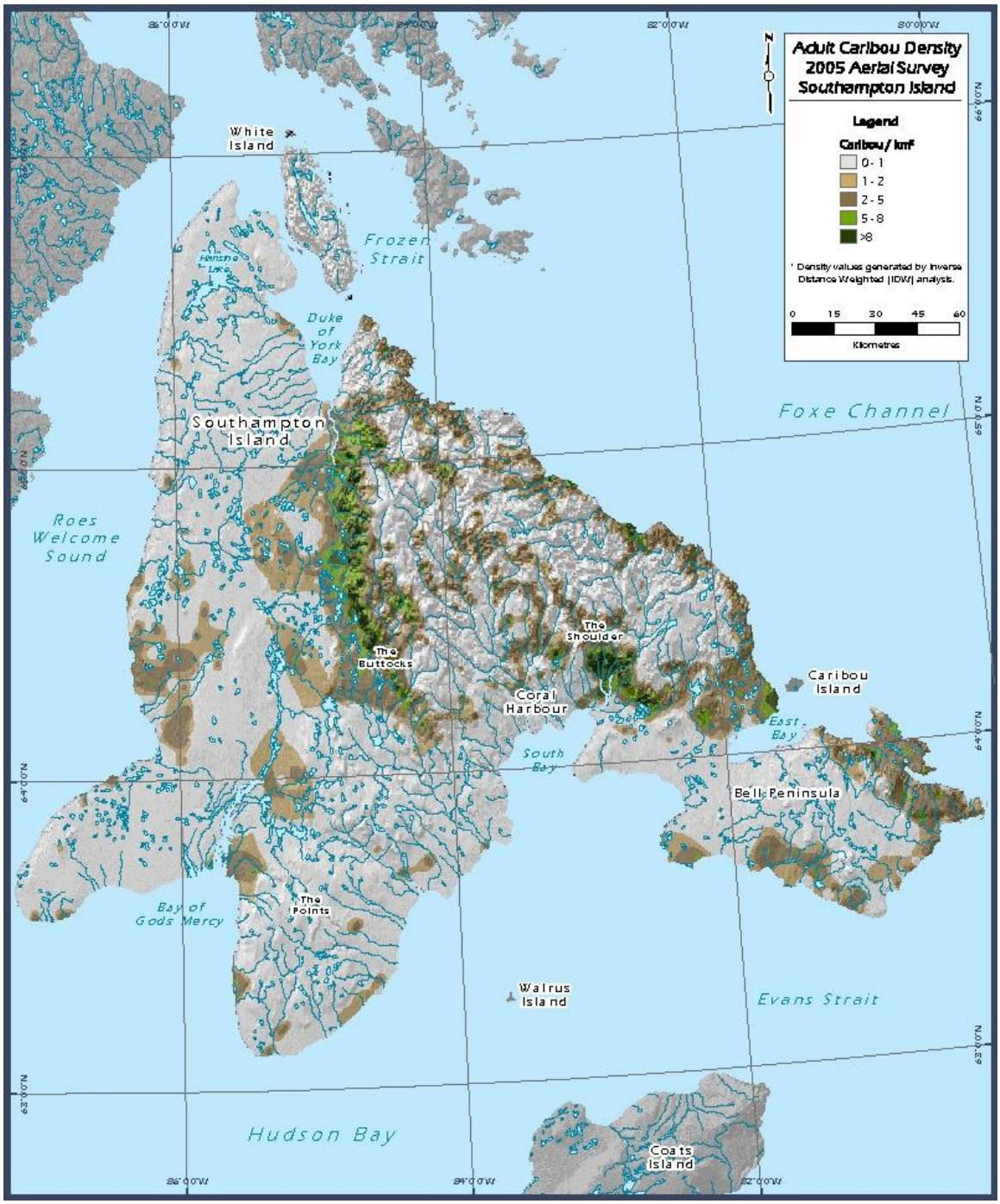

Figure 14. Results of the inverse distance weighted (IDW) interpolation technique applied to June 2005 abundance survey observations showing relative density of barren-ground caribou (*Rangifer tarandus*) on Southampton Island.

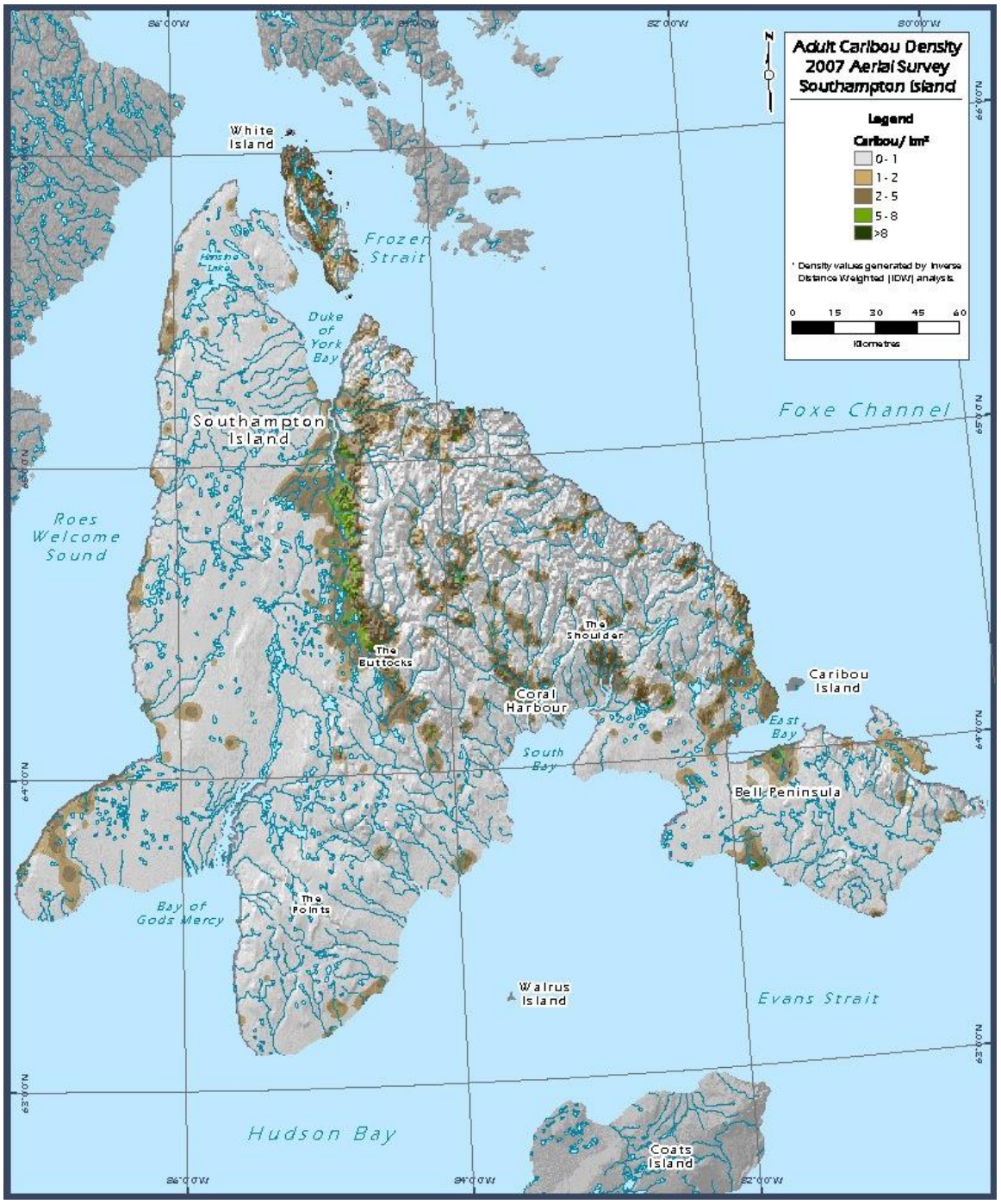

Figure 15. Results of the inverse distance weighted (IDW) interpolation technique applied to June 2007 abundance survey observations showing relative density of barren-ground caribou (*Rangifer tarandus*) on Southampton Island.

4.2 Strip Transect Surveys:

Overall, population abundance estimates from 1997 to 2017 were reasonably precise, with Coefficients of Variation (CVs) of less than 10% in all years (**Table 5, Figure 16**). **Figure 16** displays the estimates from **Table 5** and strata-specific estimates are shown in **Figure 17**. A tabular listing of estimates is provided in excel worksheets with this report. Coats Island is included in **Figure 17**, however, it was not included in overall Southampton Island estimates. Population declines occurred in all strata from 1997 to 2013, and again from 2015 to 2017. The use of different scales on the graph in **Figure 17** aids in interpretation of stratum-specific trends but it is also misleading in terms of the relative abundance of caribou in each strata. For this reason, the same estimates of caribou numbers are plotted on the same scale (**Figure 18**), clearly indicating that the majority of caribou on SHI occurred on the High Eastern SHI strata, in all years.

Table 5. Strip transect estimates of caribou on Southampton Island, showing the number of strata sampled each year, the number of caribou counted on transect, and population estimates with descriptive statistics (SE = standard error, CV = coefficient of variation) are given for each year of surveys from 1997 through 2017.

Year	Strata	Caribou	Strip transect estimates				
	sampled	Counted	N	SE	Confidence Limits		CV
1997	7	5777	29,425	1622.5	26,375	32,827	5.5%
2003	7	3833	18,479	1099.8	16,420	20,797	6.0%
2005	6	4079	21,227	1701.8	18,098	24,896	8.0%
2007	7	2689	14,389	914.6	12,684	16,325	6.4%
2009	6	2521	13,651	833.1	12,091	15,412	6.1%
2011	7	1667	7,937	580.4	6,861	9,182	7.3%
2013	7	1597	7,284	525.3	6,307	8,413	7.2%
2015	7	3068	12,319	931.6	10,591	14,328	7.6%
2017	7	1685	8,436	680.8	7,184	9,906	8.1%

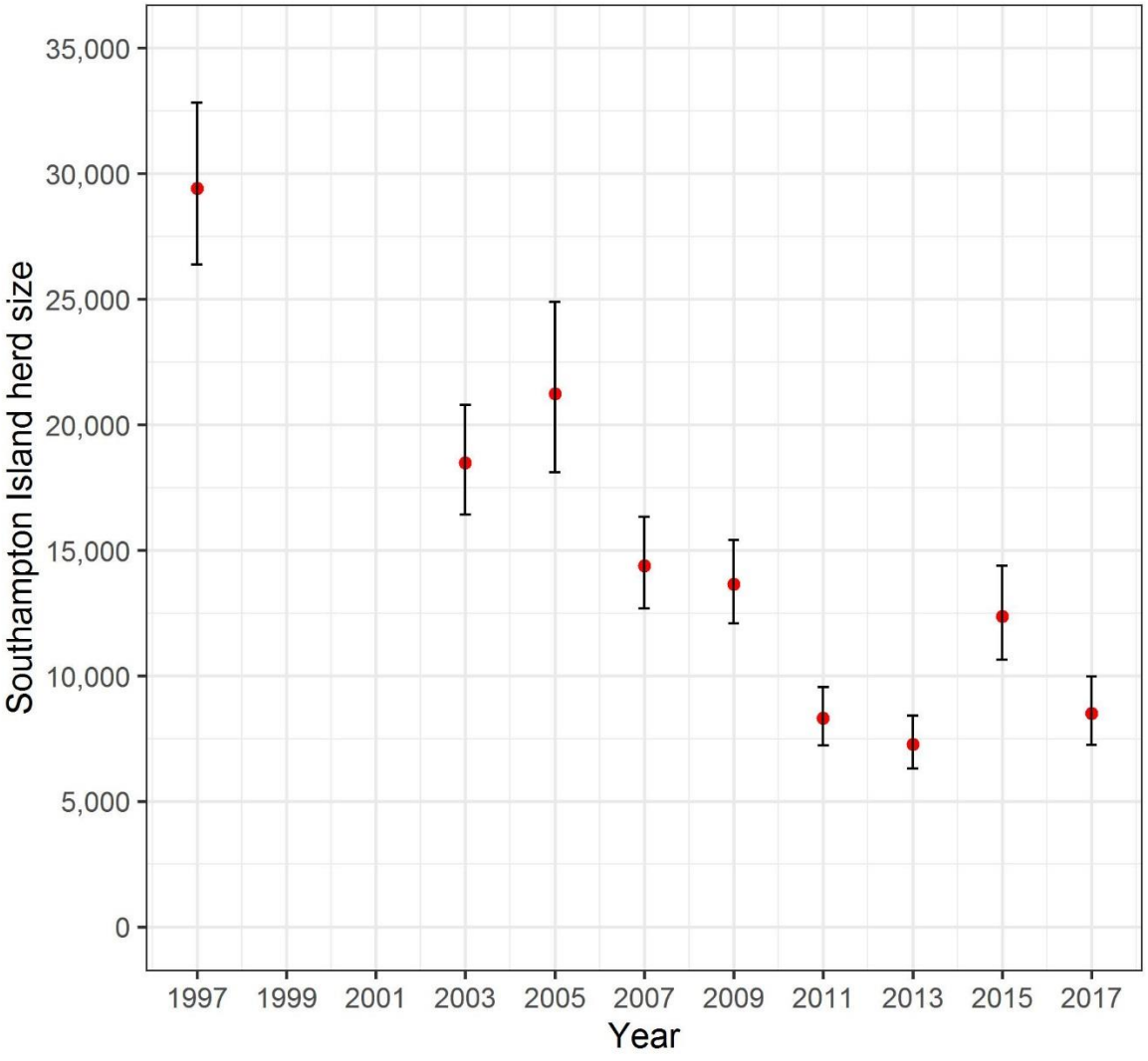

Figure 16. Population abundance estimates of the Southampton Island caribou herd using a strip transect estimator, according to strata listed in Table 5.

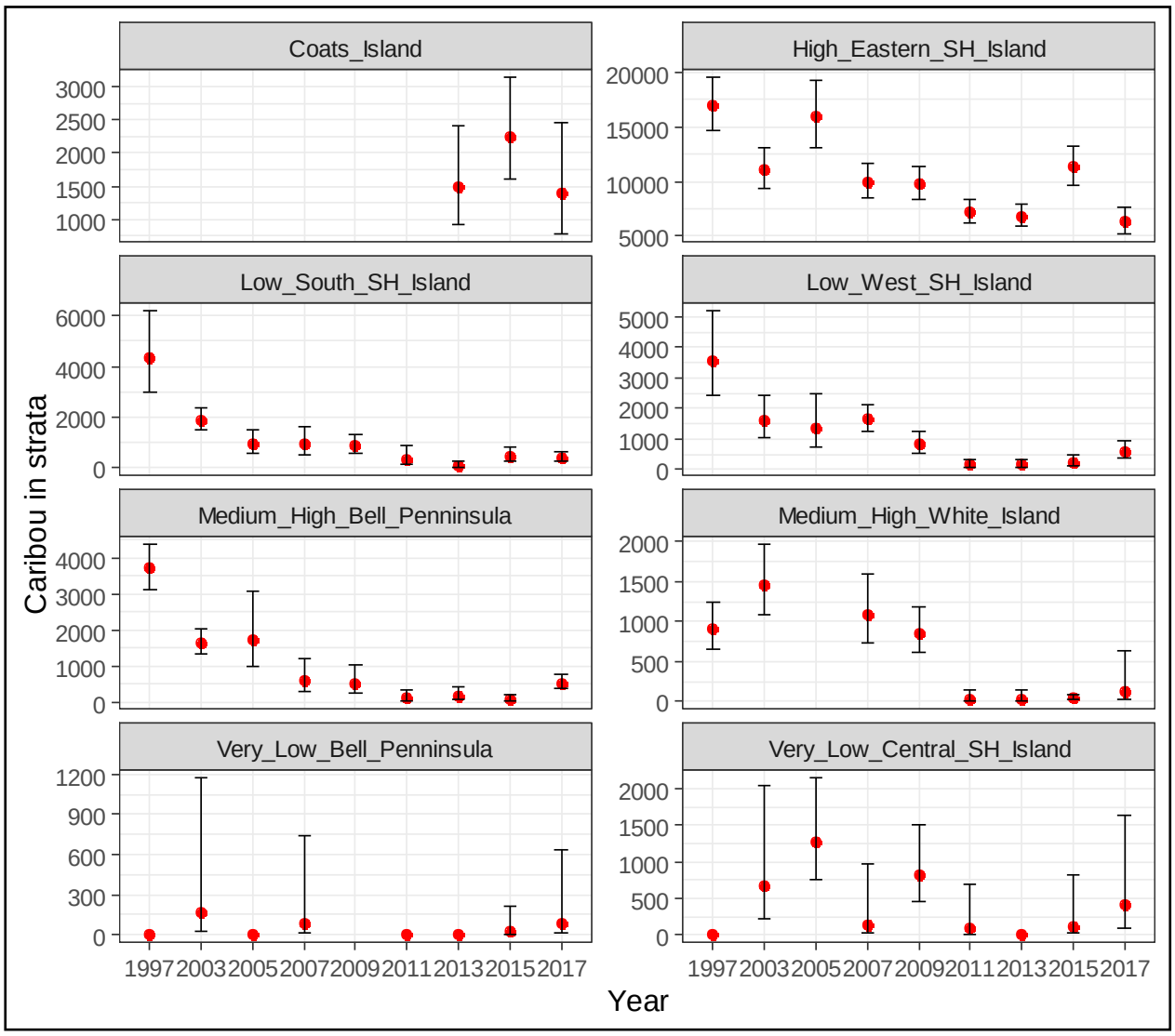

Figure 17. Strata-specific estimates of strata sampled using a strip transect estimator. Note that the y-scales are different for each graph.

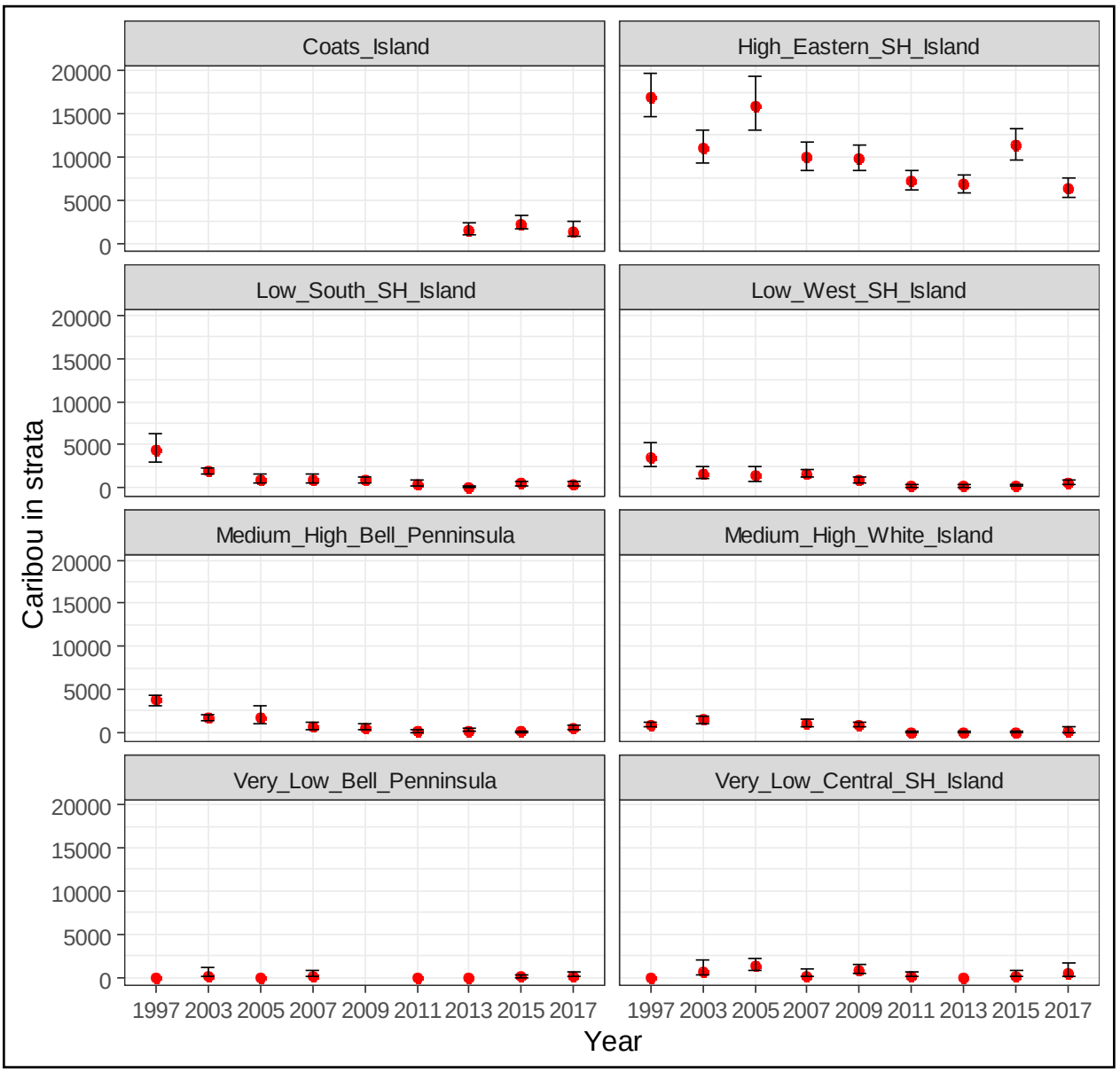

Figure 18. Strata-specific estimates of strata sampled using a strip transect estimator with the same scale used on each graph.

4.2.1 Dependent double observer analyses (2011-2015):

Dependent double observer pair data were collected using fixed-wing aerial surveys in 2011, 2013, 2015, and 2017. In 2017, we used binned distance markers on wing struts to allow for distance sampling methods, as described in the methods section of this report. Survey conditions, group sizes, and observer efficiency varied between each survey year. These data were explored graphically to help assess dominant forms of variation prior to identifying a statistical model for population estimates derived from the dependant double observer pair method. The distribution of group sizes was relatively similar during each survey year with larger groups observed in 2015 (**Figure 19**).

In general, smaller group sizes were more likely to be seen only by a single observer. Observers were placed into 12 pair combinations, of which 5 pairs switched between primary and secondary roles, and 7 did not. The assumption of the dependent double observer method is that the two observers have similar sighting probabilities and therefore, estimates may be biased when observers do not switch places during the survey. The sighting probability of pairs varied between observers for some pairs (i.e. in particular for pair 7) showing a higher relative frequency of only one observer seeing a group of caribou (**Figure 20**). A detailed listing of observer pairs is given in Appendix 1.

Between 1997 and 2011, all surveys were flown in early June close to or during the onset of spring melt. From 2013 to present, survey deployment was changed to early to mid-May (see methods) though no detectable variation in relative densities and their related strata were found. Regardless, snow cover varied each survey year with 2011 having a full range of snow cover and other years showing primarily high snow cover particularly 2013 and on following the change in survey timing to May. Sighting probabilities were lower in 2011 as shown by higher frequencies of single observer sightings (**Figure 21**). Cloud cover also varied for each year (**Figure 22**), with no discernable patterns.

Dependent double observer pair model selection, performed by sequentially calculating differences in Akaike's information criterion (AICc), suggested that sighting probabilities varied according to a combination of observer, year, size of caribou groups observed, and cloud and snow cover categorized in 25% intervals (**Table 6**). A total of three models were supported by differences in AICc values of less than 2. However, the difference (>1.51) suggested the most support for model 1, which also had the least number of parameters of the three models and would therefore produce the most precise estimates. Therefore, model 1 was used to infer covariates and population estimates. The support for year as a sightability term suggested that there were year-specific factors affecting sightability that were not accounted for by other covariates. Observer pairs in the analyses were reduced to the main pairs that exhibited lower sighting probabilities, given that a model with all observer pairs parameterized did not converge. Using this strategy, the main observer pairs that displayed lower or higher probabilities were accounted for with other observer pairs set to a mean value. The predictions of the most supported model (model 1 in **Table 6**) are shown graphically, demonstrating that sightability was lower in 2011 for both observer pairings and as a function of cloud and snow cover (**Figure 23 and 24**). Estimates from dependent double observer methods are compared to those from strip transects in a later section of the report.

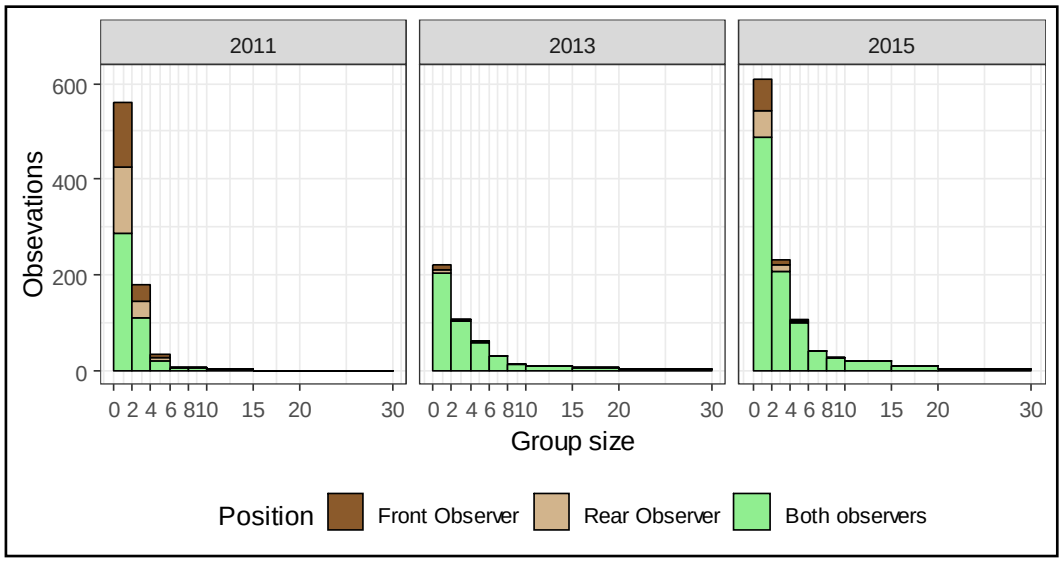

Figure 19. Group sizes of caribou observed each year for surveys conducted in 2011, 2013, and 2015 with frequency of sightings made by front, rear, and both observers.s.

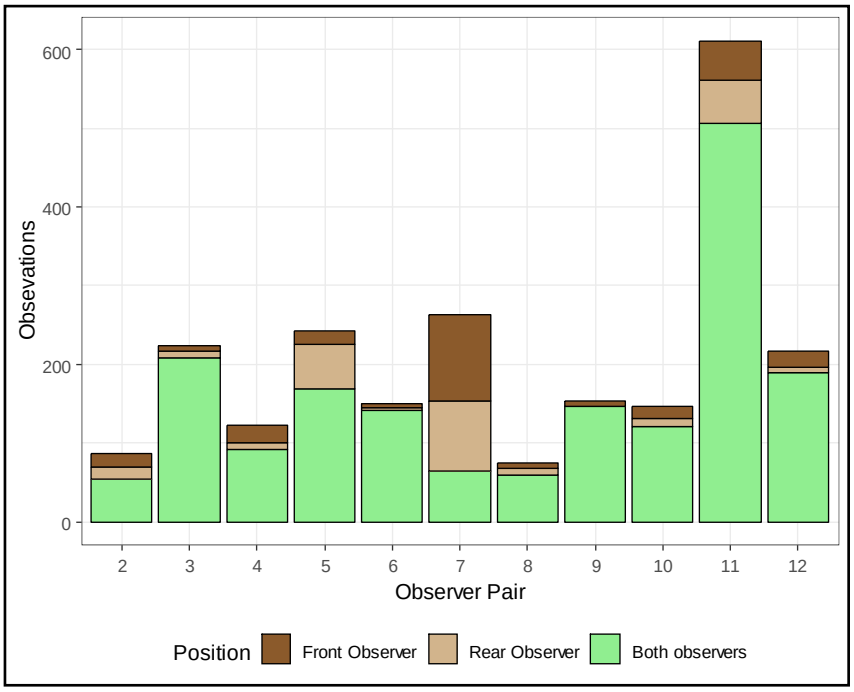

Figure 20. Observer pairings with frequencies of sighting by front, rear, and both observers.

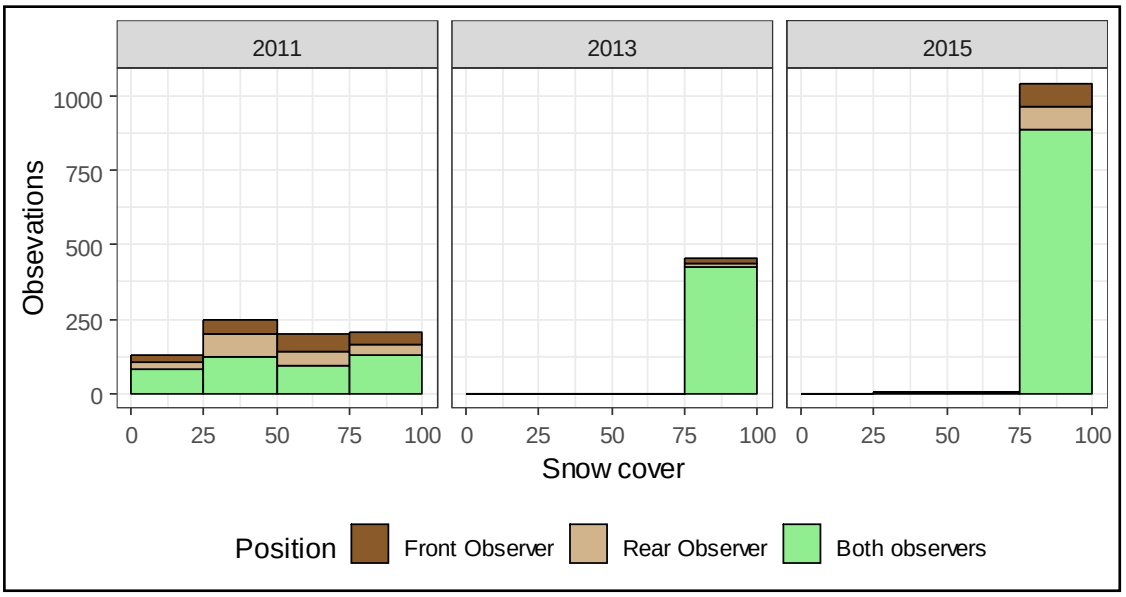

Figure 21. Snow cover during each year of the survey with frequencies of sightings by front, rear, and both observers.

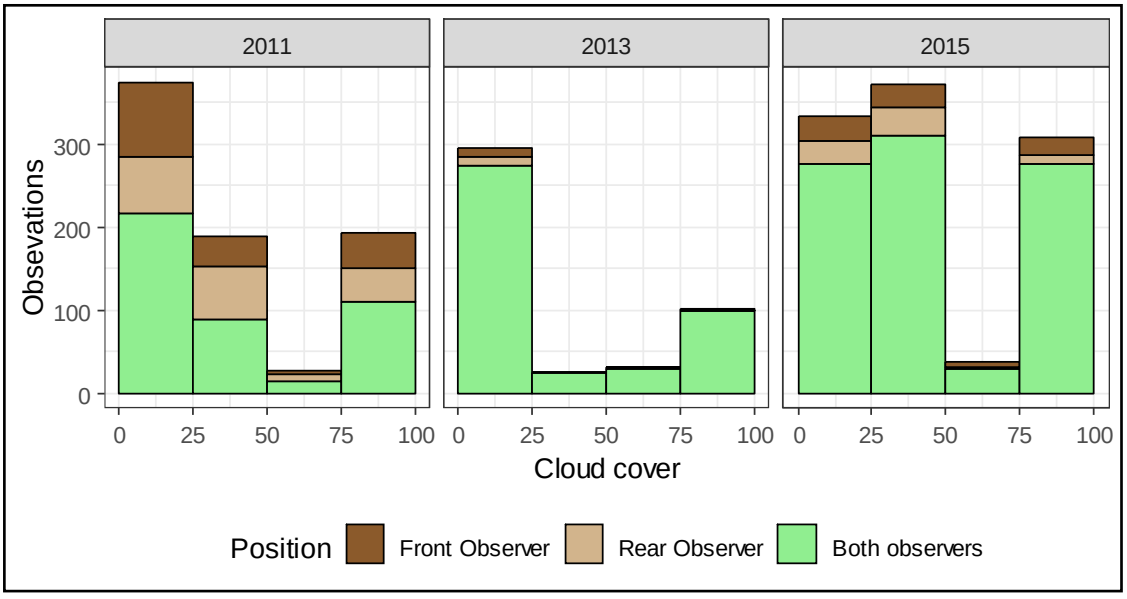

Figure 22. Cloud cover during each year of the survey with frequencies of sightings by front, rear, and both observers.

Table 6. Dependent double observer model selection results. Sample size adjusted Akaike Information Criterion (AICc), the difference in AICc between the most supported model. For each model (ΔAIC_c), AIC_c weight (w_i), number of model parameters (K), and deviance is given. See Table 1 for covariate definitions

No.	Model	AIC_c	ΔAIC_c	w_i	K	LL
1	observers (reduced) +Year + size + cloud+snow	1424.16	0.00	0.53	14	-698.0
2	observers (reduced) +Year + size + cloud+snowc*size	1425.66	1.51	0.25	15	-697.7
3	observers (reduced) +Year + size + cloud+Year*size	1425.91	1.75	0.22	16	-696.8
4	Year + size + cloud + snow	1456.59	32.44	0.00	10	-718.2
5	observers (all)	1463.83	39.67	0.00	11	-720.9
6	size + snow + cloud	1487.26	63.10	0.00	8	-735.6
7	Year	1493.91	69.75	0.00	3	-743.9
8	snow + cloud	1517.70	93.55	0.00	7	-751.8
9	snow	1538.96	114.80	0.00	4	-765.5
10	snowc + cloudc + snowc * cloudc	1565.15	140.99	0.00	4	-778.6
11	size	1597.33	173.17	0.00	2	-796.7
12	Log(size)	1608.79	184.63	0.00	2	-802.4
13	Cloud	1628.97	204.81	0.00	4	-810.5
14	constant	2666.43	1242.27	0.00	1	-1332.2

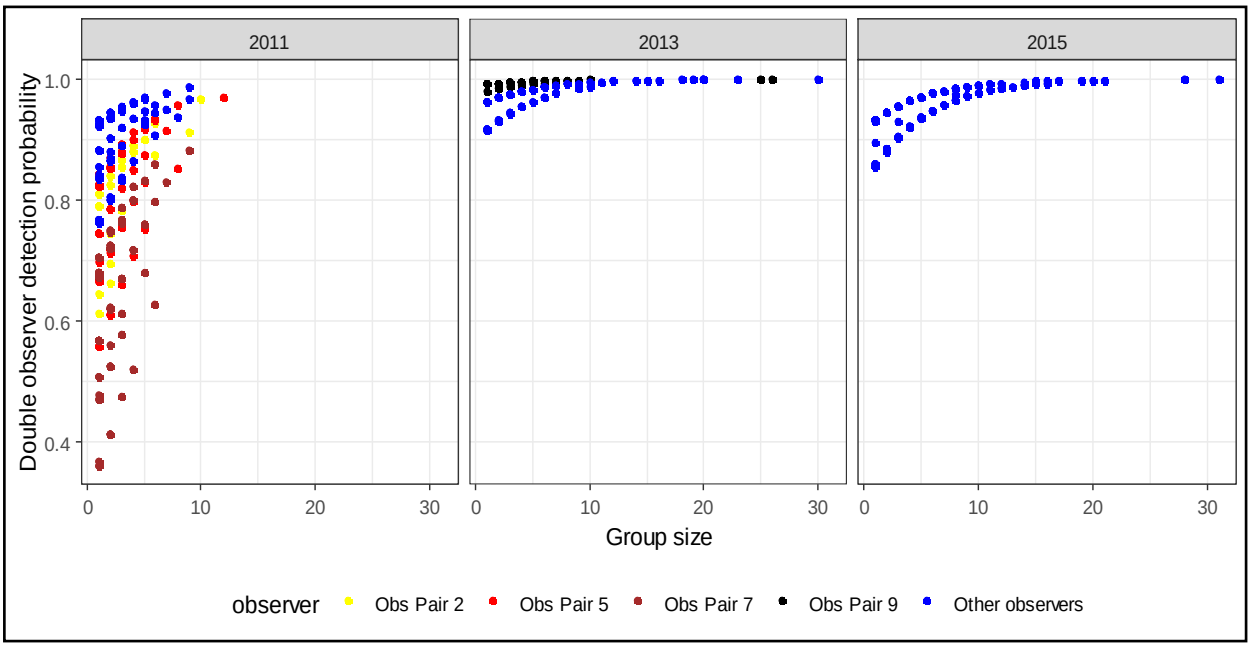

Figure 23. Dependent double observer detection probabilities as a function of year, group size, for selected observer pairs.

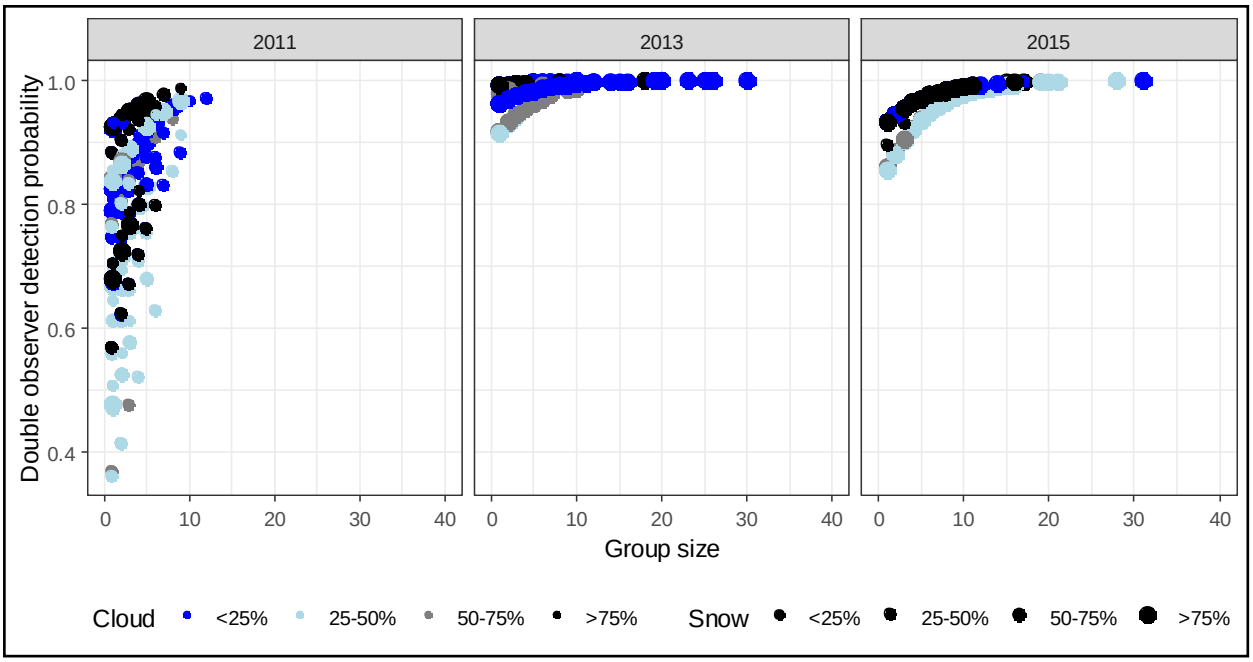

Figure 24. Dependent double observer probabilities as a function of year, group size, snow and cloud cover.

4.2.2 Distance sampling/double observer pair sampling in 2017:

During the 2017 survey, frequencies of observations by distance bins revealed different detection probability curves between observer pairs 1 and 2. Observer pair 1 had a higher frequency of observations near the aircraft whereas observer pair 2 had a higher frequency away from the plane. Compared to previous years of dependent double observer pair sampling, there was a higher frequency of observations from data recorders in 2017, suggesting a higher level of observation experience by the data recorders. To utilize these recorder observations, we categorized them as single observer observations and assumed that the data recorder had similar sighting probabilities to the other observers (**Figure 25**). Snow cover was greater than 50% in the area of most observations, and the results of the 2017 observations suggested that sightability was lower when snow cover conditions were below 50% (**Figure 26**).

Model selection proceeded by building distance sampling models with the mark-recapture model parameters held constant, and by initially comparing half normal and hazard rate models. Of these, the hazard rate model was the most supported, with observer pair and snow (continuous cover) as covariates. Once this model was selected, dependent double observer pair mark-recapture models were compared with observer pair and snow, as well as the most supported covariates. Group size (log transformed) was also supported as a distance sampling covariate (Model 1, **Table 7**). Goodness-of-fit for model 1 was marginal (chi-square=19.8, df=7,p=0.006), however most of the lack of fit came from the 600-1,000 meter distance bin which would have less influence on estimates given low observation frequency rates in this bin (**Figure 27**). An additional analysis was conducted, which used the first 2 distance bins of data to fit dependent double observer pair only models to the data, without the distance component (**Table 8**). The same suite of dependent double observer pair models was applied to the data set as used in previous years analysis and as listed in **Table 7**. According to this subsequent analysis, a model with observer, snow

(continuous) and the log of group size was most supported. Population abundance estimates from this model were thus compared to the distance sampling and strip transect estimates.

4.2.3 Comparison of estimates from strip transect, dependent double observer pair, and distance sampling:

Comparison of strip transect, dependent double observer pair, and distance sampling estimates suggests reasonable agreement between estimates, with the confidence intervals from each method all overlapping. Estimates from the dependent double observer method when compared with the single observer jolly estimator were 6% higher in 2011, similar in 2013 and 2015, and 4% higher in 2017. In 2017, distance sampling estimates were 9.1% higher than strip transect estimates and 5% higher than dependent double observer pair estimates suggesting that distance sampling was better able to accommodate observations where one observer may be over sampling further distance bins (**Table 9**, **Figure 28**).

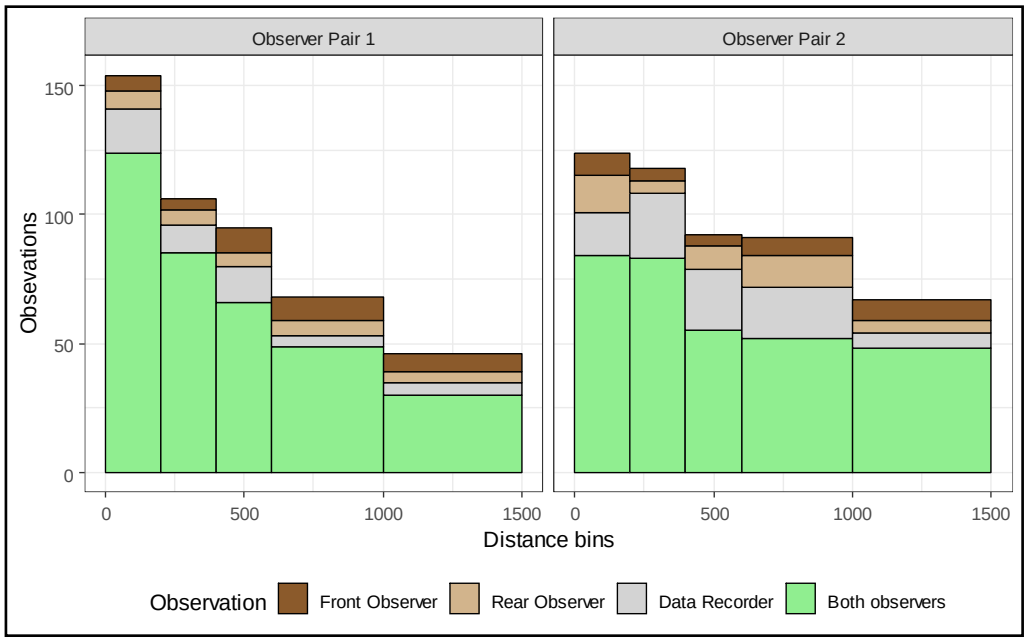

Figure 25. Frequencies of observations by distance bin for the 2 observer pairs in May 2017.

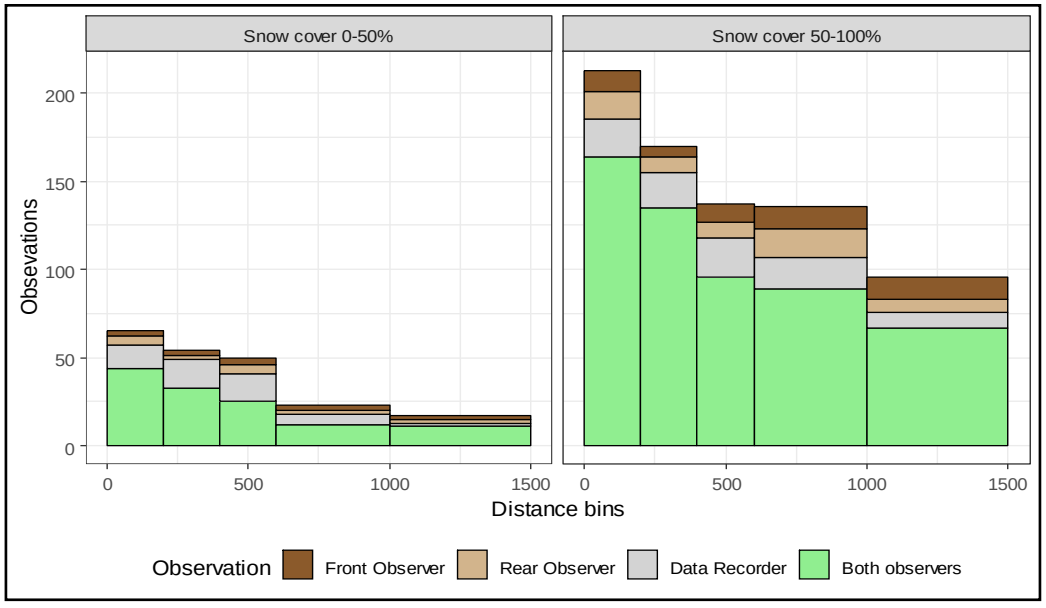

Figure 26. Frequencies of observations by distance bin for 2 levels of snow cover.

Table 7. Dependent double observer model selection results. Sample size adjusted Akaike Information Criterion (AICc), the difference in AICc between the most supported model for each model (ΔAICc), AICc weight (wi), number of model parameters (K) and deviance is given. See Table 1 for covariate definitions.

		Distance sampling	2x observer	Model fit				
No.	**DF**	**Distance covariates**	**covariates**	AIC_c	ΔAIC_c	w_i	K	LL
		Distance /Double observer models						
1	HR	obs+snowc+log(size)	obs+snowc	4000.0	0.00	0.95	8	-1992
2	HR	obs+snowc+log(size)	obs+snow	4006.6	6.59	0.04	10	-1993
3	HR	obss+snowc+size	obs+snow	4009.3	9.24	0.01	10	-1994
4	HR	obs+snowc	obs+snow	4010.8	10.75	0.00	9	-1996
5	HR	obs+log(size)	obs+snowc	4012.7	12.66	0.00	7	-1999
6	HN	obss+snowc+logsize	obs+snow	4019.2	19.16	0.00	9	-2000
7	HN	obs+snowc	obs+snow	4019.8	19.71	0.00	8	-2001
8	HR	obss+snowc	obs	4020.1	20.04	0.00	6	-2004
9	HR	obss+snowc	size	4028.0	27.94	0.00	6	-2008
		Distance sampling models						
10	HR	obss+snowc	constant	4031.2	31.14	0.00	5	-2010
11	HN	obss+snowc	constant	4040.2	40.11	0.00	4	-2016
12	HN	obs+snowc+size	constant	4040.2	40.15	0.00	5	-2015
13	HR	obs	constant	4041.7	41.66	0.00	4	-2016
14	HN	snowc	constant	4045.3	45.28	0.00	3	-2019
15	HR	size	constant	4047.0	46.96	0.00	4	-2019
16	HR	constant	constant	4047.3	47.27	0.00	3	-2020
17	HN	obs	constant	4061.1	61.09	0.00	3	-2027
18	HN	constant	constant	4067.3	67.22	0.00	2	-2031
19	HN	size	constant	4067.9	67.89	0.00	3	-2031

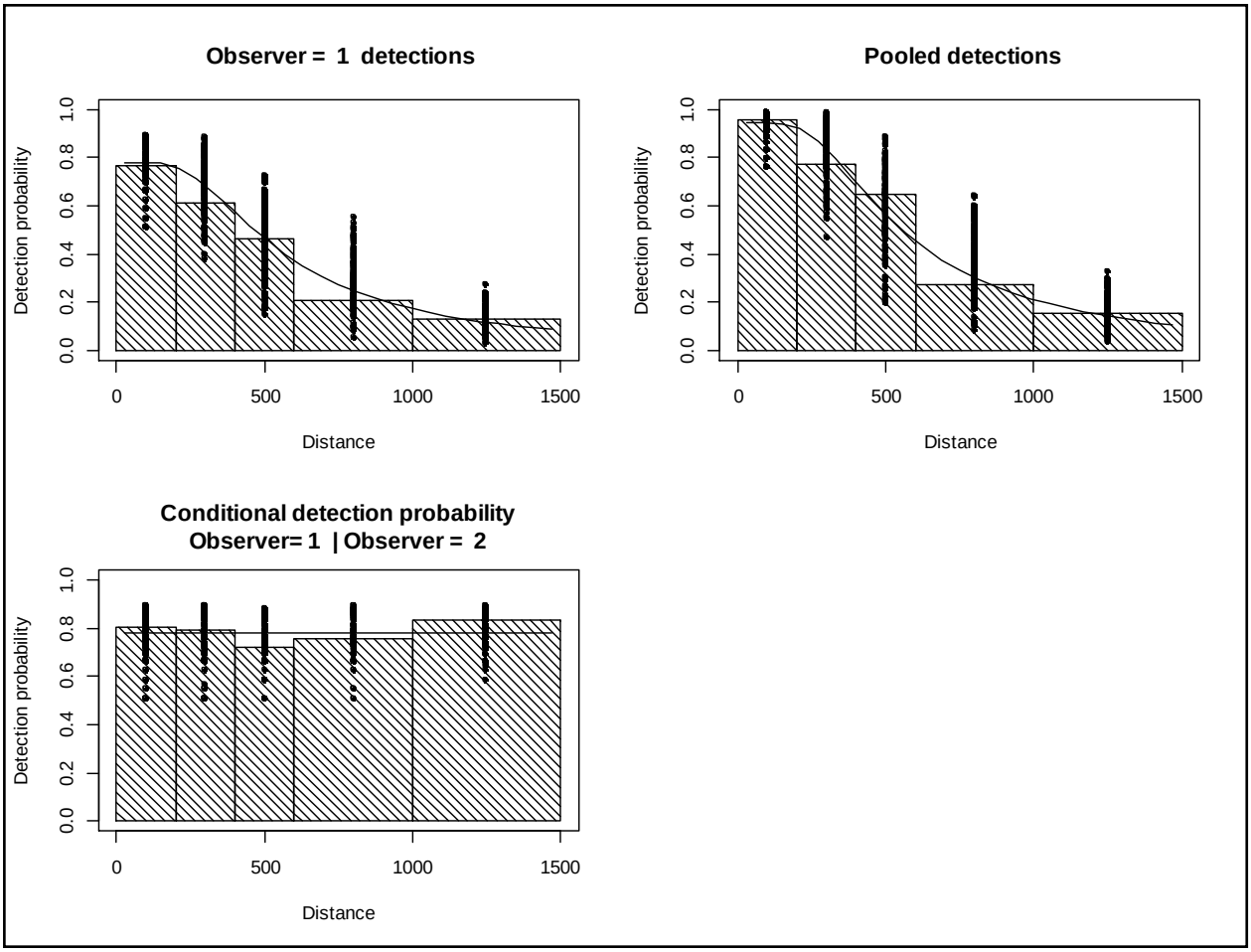

Figure 27. Graphical representation of goodness-of-fit of the most supported double observer model (Model 1, Table 7).

Table 8. Dependent double observer model selection results. Sample size adjusted Akaike Information Criterion (AICc), the difference in AICc between the most supported models for each model (ΔAIC_c), AIC_c weight (w_i), number of model parameters (K), and deviance is given. See Table 1 for covariate definitions

No	Model	AIC_c	ΔAIC_c	w_i	K	LL
1	obs+snowc+log(size)	1,193.0	0.00	0.40	4	-592.5
2	obs+snowc+size	1,193.1	0.07	0.38	4	-592.5
3	obs+snowc	1,194.7	1.70	0.17	3	-594.3
4	obs+snow_factor	1,197.8	4.75	0.04	5	-593.8
5	obs+snow_factor+cloud_factor	1,200.3	7.26	0.01	8	-592.0
6	constant	1,208.7	15.73	0.00	1	-603.4

Table 9. Comparison of estimates of Southampton Island caribou using strip transect, double observer, and distance sampling/double observer (2017 only)

Year	Method	Caribou counted	N	SE	Confidence Limit		CV
2011	Strip transect	1667	7,937	580.4	6,861	9,182	7.3%
2011	2x Observer strip transect	1667	8,442	691.9	7,171	9,937	8.2%
2013	Strip transect	1597	7,284	525.3	6,307	8,413	7.2%
2013	2x Observer strip transect	1597	7,287	557.2	6,255	8,490	7.6%
2015	Strip transect	3068	12,319	931.6	10,591	14,328	7.6%
2015	2x Observer strip transect	3068	12,368	1002.6	10,518	14,542	8.1%
2017	Strip transect	1685	8,436	680.8	7,184	9,906	8.1%
2017	Distance 2x observer	1653	9,200	796.4	7,755	10,915	8.7%
2017	2x Observer strip transect	1665	8,752	759.5	7,365	10,399	8.7%

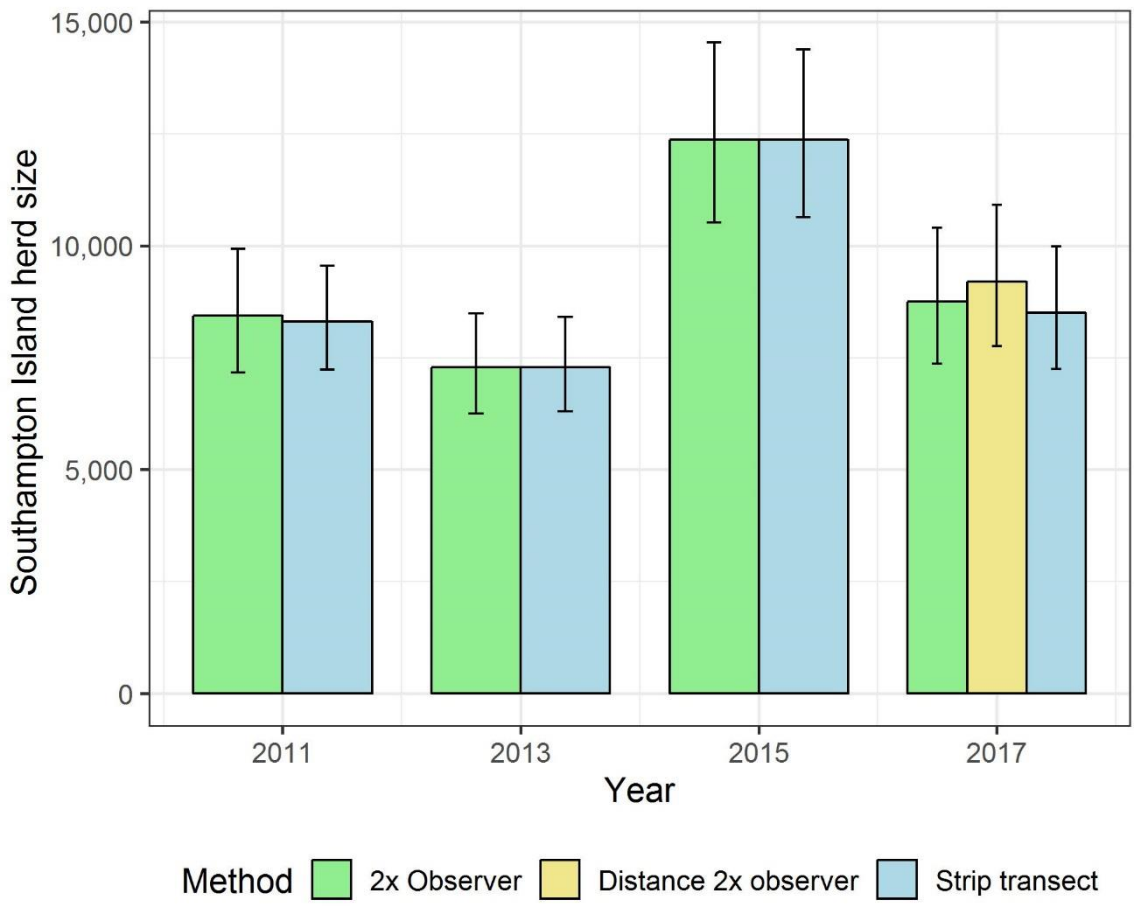

Figure 28. Comparison of Southampton Island caribou herd abundance estimates from strip transect, dependent double observer, and distance sampling/double observer analyses (2017).

4.3 Trend estimates:

Our trend analyses covers two separate phases of Southampton Island caribou abundance: prior to 1997 when herd abundance was increasing, and from 1997 to 2017 when the herd was declining. Prior to 1997, and since their reintroduction, the abundance of the SHI caribou herd was increasing. Despite a statistically significant increase between the May 2013 and May 2015 survey estimates, the SHI herd has exhibited an overall decline since 1997.

4.3.1 Trend from 1997 to 2017 (the decline phase):

Data from 1997 to 2017 included strip transect, dependent double observer pair, and distance sampling surveys. The use of different methods had minimal effects on the overall abundance trends identified. However, the best estimates for 2011, 2013, and 2015, based on model fit and lowest CV's, were dependent double observer pair estimates which accounted for sightability, especially in 2011. For 2017, the distance sampling estimate was least biased because of observer error. For this reason we used strip transect data for estimates from surveys up to 2009, followed by dependent double observer pair estimates for 2011 to 2015, and distance sampling estimates for 2017.

T-tests were used to compare the significance of the difference between sequential estimates (**Table 10**). Of the 8 survey estimate comparisons the 1997 to 2003, 2005 to 2007, 2009 to 2011, 2013 to 2015, and the 2015 to 2017 periods showed statistically significant change. Of these comparisons, only the 2013 to 2015 estimates showed a statistically significant increase in the SHI caribou population, all others displaying significant declines. Annual change in population size, based on a year to year comparison of estimates (expressed as ratios), varied between 0.79 and 1.30. SHI caribou abundance estimates from 1997 to 2017 are shown graphically in **Figure 29**. To estimate the effect of a potential immigration event on the overall trend, prior to the 2015 survey, an

additive term was applied to model use to generate the 2015-2017 survey estimates. This term basically assumed that the SHI population was increased by a constant amount during this time due to immigration. These terms were both found to have a significant effect on the trend in caribou abundance (**Table 11)**. The year term provided an estimate of long term annual rate of change for the SHI population (0.91 CI=0.89-0.93) which was not, overall, affected by the immigration event. This translates to a 9% (CI=7-11%) decline in caribou abundance each year, from 1997 through 2017. The immigration term estimated the gross rate of increase (1.76, CI=1.3-2.4) in the SHI population between the 2013-2015 surveys, additive to the year term.

A plot of model predictions reveals good fit of the model to estimates with predictions intersecting the confidence limits of 7 of the 9 estimates (**Figure 30**). Namely, the model suggests that the herd declined at a constant rate from 1997-2014, followed by an immigration event sometime between May 2013 and May 2015 (Patkeau, 2015), and then continued to decline at a similar rate as it had previously, from 2015-2017 (**Figure 30**). Using this model, and assuming a constant rate of decline (9%) over the period, we estimated that approximately 5,024 caribou would have had to immigrate to SHI between May 2013 and May 2015 to account for the increased number of animals observed in May 2015. If the immigration event had not occurred, and the population continued to decrease at the 9% rate, then there would be approximately 4,200 caribou remaining on the island as opposed to the 9,200 estimated in the 2017 survey.

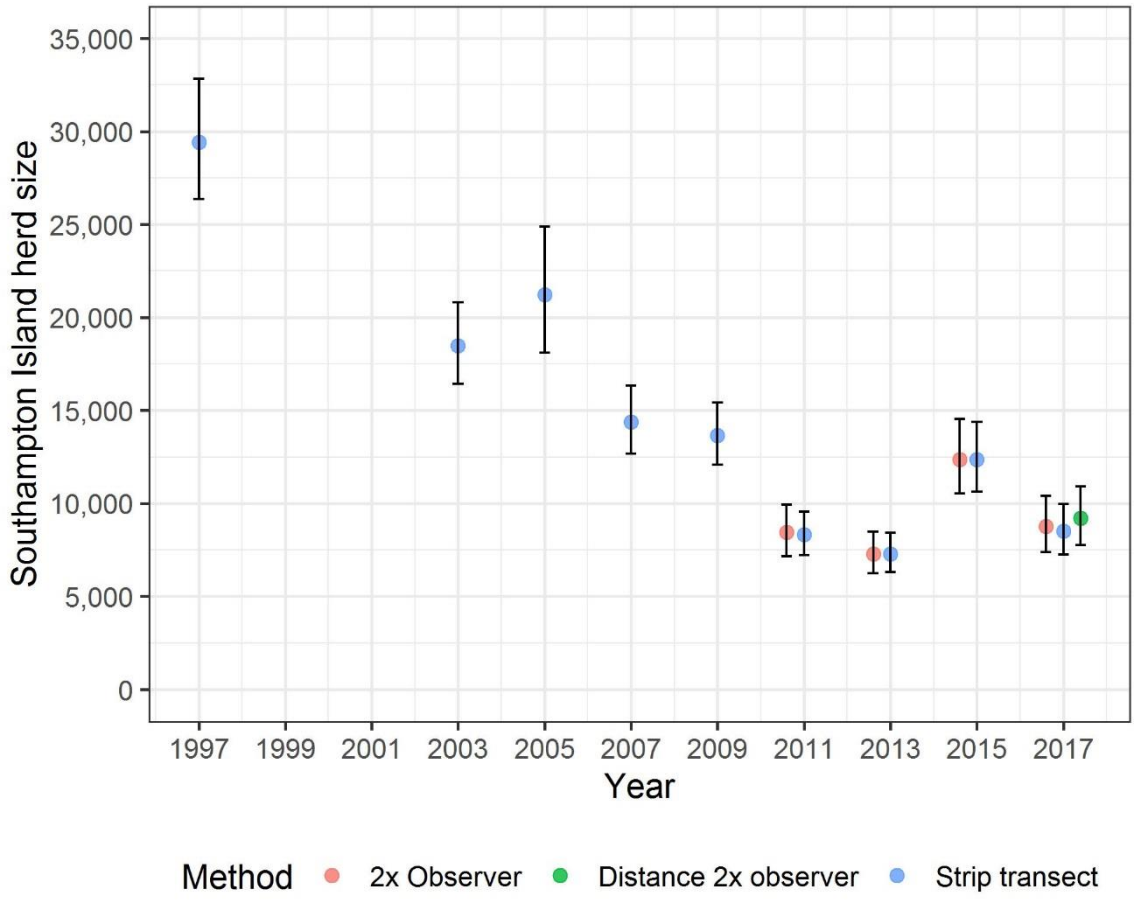

Figure 29. Comparison of strip transect, dependent double observer pair, and distance sampling/double observer pair estimates (2017).

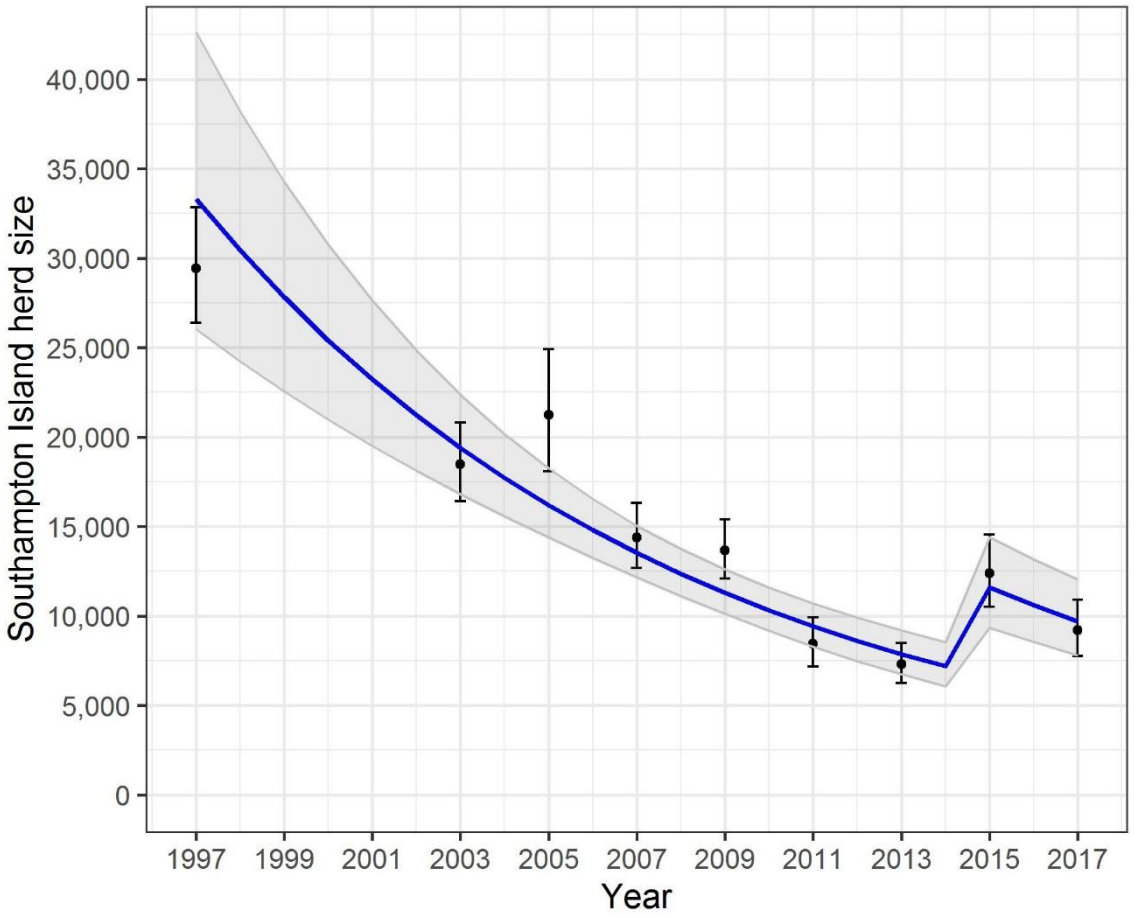

Figure 30. Predictions of herd size of the Southampton Island caribou population from the log-linear model (Table 11), which assumes a constant decline in population size with an immigration event that occurred before the 2015 survey. Confidence limits are provided as shaded regions on the plots.

Table 10. Estimates used for the 1997 to 2017 trend analysis of Southampton Island caribou abundance, with the results of t-tests comparing the estimates of successive surveys. Also shown are estimates of gross and annual change based on the ratios of successive estimates.

Year	method	N	SE	CV	df	ttest	df	p-value	Gross change	Annual change
1997	Strip	29,425	1622.5	5.5%	93					
2003	Strip	18,479	1099.8	6.0%	90	-5.58	163	0.000	0.63	0.93
2005	Strip	21,227	1701.8	8.0%	76	1.36	132	0.177	1.15	1.07
2007	Strip	14,389	914.6	6.4%	88	-3.54	117	0.001	0.68	0.82
2009	Strip	13,651	833.1	6.1%	80	-0.60	168	0.551	0.95	0.97
2011	2x Obs	8,442	691.9	8.2%	73	-4.81	151	0.000	0.62	0.79
2013	2x Obs	7,287	557.2	7.6%	59	-1.30	131	0.196	0.86	0.93
2015	2x Obs	12,368	1002.6	8.1%	59	4.43	92	0.000	1.70	1.30
2017	Distance	9,200	796.4	8.7%	134	-2.47	133	0.015	0.74	0.86

Table 11. Log-linear model parameter estimates for trend analysis (1997-2017).

Term	β	SE (β)	t	p-value	Confidence limit	
Intercept	36448.28	0.14	77.39	0.0000	27,708.8	47,188.1
Year (λ)	0.91	0.01	-8.37	0.0002	0.89	0.93
Immigration	1.76	0.15	3.68	0.0103	1.30	2.38

4.3.2 Trend from 1978 to 1997(the increase phase):

The historic data set (1978-1991) was added to the analysis to obtain an estimate of trends in the SHI caribou population during the phase of increase that occurred from 1978 to 1997. This was accomplished by adding terms to account for the decrease phase, which allowed us to estimate an annual rate of increase of 1.19 (CI=1.16-1.22), or, 19% (CI=16-22%), from 1978 to 1997 (**Table 12**). A plot model for these predictions is shown in **Figure 31**.

4.3.3 Sampling effort and error

A comparison of strip transect, dependent double observer pair, and distance sampling estimates suggest that the assumption of perfect sightability on the 400 meter survey strip was met in 2013 and 2015 with estimates being close for dependent double observer pair and strip transect estimates (**Figure 28**). In 2011, variability in observers and snow cover reduced the strip transect estimates compared to the dependent double observer pair estimates. In this context, the dependent double observer pair method provided a test of assumptions of the strip transect method and corrected estimates when the assumption of perfect sightability was violated. In 2017, distance sampling estimates were higher than dependent double observer pair and strip transect estimates. This may have been due to one of the observer pairs not putting enough survey effort to the distance bins closer to the aircraft (**Figure 25**), as indicated by different shapes of the detection histograms for the two observer pairs. This would have caused a negative bias in both strip transect and dependent double observer pair estimates and illustrates a potential issue with distance sampling; observers spending too much time looking out at further bins which are often easier to view than the closer bins. In the case of conditions of excellent sightability, this can lead to a significant over estimate. The dependent double observer pair method partially accounted for this by also estimating the sighting probabilities of observers near the survey line. The dependent double observer pair method assumes that the two observers in a pair have equal sighting probabilities. It is therefore essential that observers switch places half

way through the day to ensure robust estimates from this method. Of the 14 observer pairings on surveys, 7 switched places which may have affected the overall quality of the dependent double observer estimates. If observers cannot switch places then an independent observer method should be considered especially when caribou density is not high.

4.3.4 Overview of Abundance and Trend Analysis:

Overall, trend analysis suggests that the SHI population has been decreasing at a rate of 9% per year since the 1997 survey. An immigration event in 2015 increased the population, however, comparison of the 2015 and 2017 survey estimates suggests the 9% decline continued, even after this immigration event (**Figure 30**).

Table 12. Log-linear model parameter estimates for trend analysis (1978-2017).

term	β	SE (β)	t	p-.value	Confidence Limit	
Intercept	907.80	0.18	38.8	0.0000	633.80	1262.22
Trend (1978-1997)	1.19	0.01	13.1	0.0000	1.16	1.22
Decrease-Intercept	350.46	0.54	10.9	0.0000	121.56	991.77
Immigration (2015)	1.92	0.17	3.8	0.0049	1.38	2.68
Year*Decrease	0.76	0.02	-13.4	0.0000	0.73	0.79

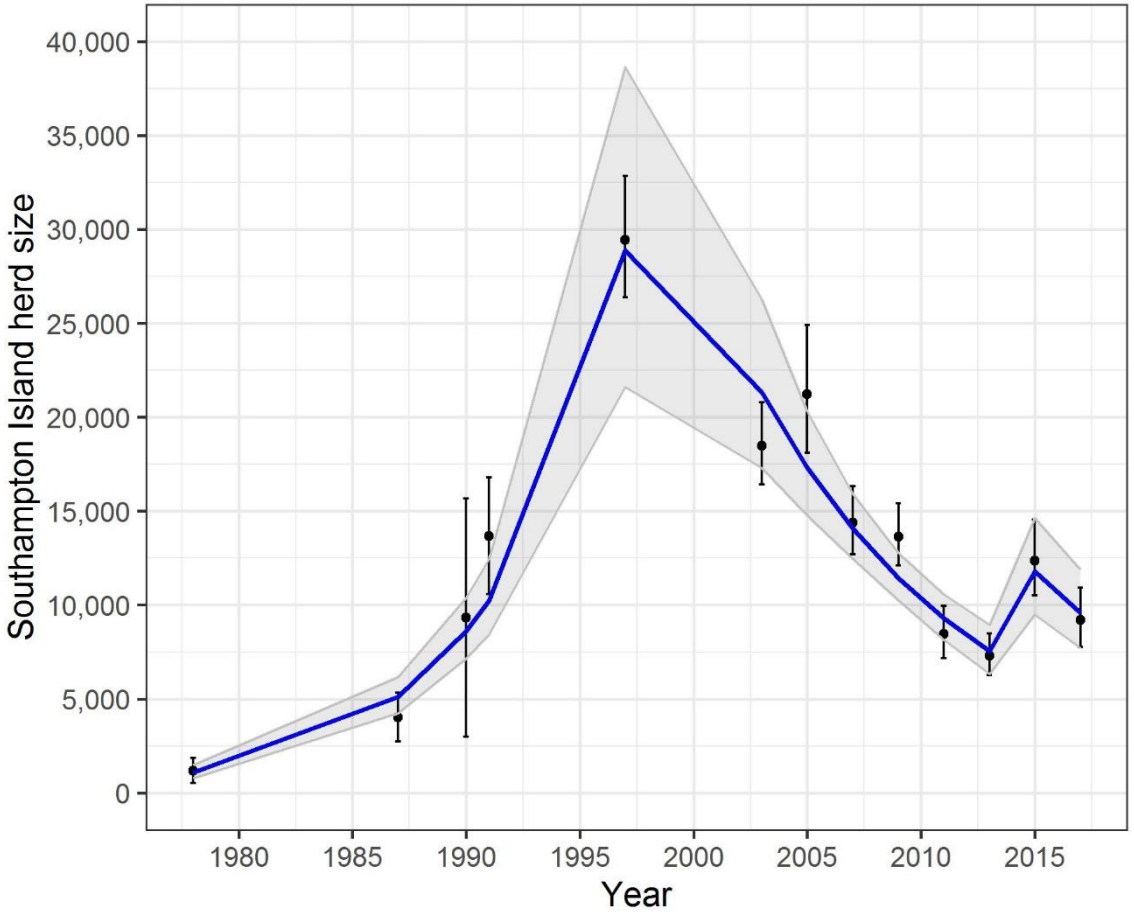

Figure 31. Predictions of herd size of the Southampton Island from the log-linear model (Table 12) which assumes a constant decline in population size after 1997 with an immigration event that occurred before the 2015 survey .

4.4 Effect of Disease on Abundance:

Brucellosis is an infectious disease caused by the Bacteria *Brucella*. Many different animal species including humans can become infected. The form of Brucellosis that occurs in wild caribou is *Brucella suis* Type IV. In caribou this bacteria occurs primarily within tissues of the reproductive system but also commonly occurs within leg joints (Williams et al. 2001; CDC 2016; Corbel, 2006). The bacteria can also be found in the milk, blood, urine and semen of infected animals (CDC 2016; Corbel, 2006). Animals can get the bacteria by either oral ingestion, direct contact with the mucus membranes of the eyes, nose, or mouth, or through breaks in the skin. Brucella can also be transmitted by contaminated objects (fomites) (Corbel, 2006). Some animals are carriers and can have the bacteria without showing signs of the illness. Animals in these cases can shed the bacteria into the environment for long periods, infecting other animals in the herd. Brucellosis can cause reproductive problems such as abortions, still birth and infertility. Other signs can include arthritis, swelling of the joints and testicles, and udder infections (mastitis) (Williams et al. 2001; CDC 2016; Corbel, 2006). Tissues and fluids associated with abortions, drainage of fluid from swollen joints, vaginal discharge, fetal fluids, and semen can be highly infective and can spread the bacterium into the immediate environment where uninfected animals can become infected through the ingestion of infected tissues and objects such as plants. The potential for environmental concentration of this disease makes Brucellosis a density-dependent disease. Areas of concentration such as migratory corridors, rutting areas and, particularly, calving grounds would represent some of the higher risk areas for the spreading of this disease (Williams et al. 2001; CDC 2016; Corbel, 2006). Predation and scavenging of diseased tissue can also contribute to the bacterium's spread throughout the environment.

Health monitoring of the SHI barren-ground caribou had its beginnings in 1988 when Heard (departmental correspondence) sampled 20 cows in March to

determine their reproductive status and general condition. These small condition studies continued through 1991 (Adamczewski and Heard data) at which time the condition studies ceased. The analysis of condition was started up again in February 1996 in association with the initiation of the large scale commercial harvest in March 1993. Due to the small sample sizes in the early condition data, for the most part, these were not included in this analysis. The first samples did, however, give results that were consistent with hunter reports of caribou on SHI in excellent health and condition at this time. By 1995, the condition and productivity of the herd had changed little, an assessment that would remain up until the 2000 harvesting season when CFIA random blood testing identified the beginning of what would become a rapid induction of the bacterial disease *Brucella suis* serovar 4 in the SHI caribou herd (**Figure 32**). There is no evidence of this disease within this population prior to the 2000 harvesting season.

Concurrent with the decline from the 1997 to 2005 survey estimate, there were at first subtle, then more dramatic shifts in range use by 2005. Range use changed significantly as densities dropped in most areas, with the exception of the north central portions of the islands where use remained consistent between years although densities slowly dropped up to present (Figure 12 and 13). In addition, the first cases of *Brucella suis* were reported during the 2000 harvest year (1.7% of 400 animals tested) and had reached a prevalence of 19.5% in 2003, 28.6% in 2005, 48.8% by 2007, 39.1 % by 2009 and 58.8 % by 2011. Pregnancy rates, which are affected by Brucellosis, initially dropped from 93.1% in 2001 to 37.9% in 2005, and then increased to 64.4% in 2007. The hopes that the disease was declining in the population were dashed when a 2009 screening showed pregnancy rates dropping further to 44.3%. The last major condition study conducted in March 2011, prior to the application of a TAH, recorded pregnancy rates of 37% (**Figure 32**).

In 1992 the Canadian Polar Commission released a status report on Brucellosis in the Circumpolar Arctic (O'Reilly, 1992). In the report, O'Reilly summarized the incidence of Brucellosis across the Circumpolar arctic (**Table 13**). Brucellosis prevalence within the Southampton Island population reached a high of 58.9% in 2011 which represents the highest prevalence amongst any caribou and/or reindeer populations' worldwide (O'Reilly, 1992). Currently levels are unknown due to a cessation of the annual caribou condition harvest. With the human health issues associated with Brucellosis through either the consumption or handling of infective tissues, Coral Harbour residents are concerned over the future of their caribou herd.

4.41 Brucellosis and heard trend:

Concurrent with the rising prevalence of the reproductive disease *Brucella suis* was the reported declines in abundance from 1997 through 2013 (**Figure 32**). It appears clear that Brucellosis was a contributing factor to the steady declines observed in this population of caribou. However, with high commercial harvest rates of the SHI herd up to 2009, it is likely that both commercial hunting pressure and disease together, contributed significantly to a declining trend in caribou abundance. By 2003, three years following the first confirmed cases of Brucellosis in SHI caribou, pregnancy rates were still over 85% and the population was still over the hypothesized carrying capacity of the island of an estimated 15,000 animals (Oulett et al. 1996). With Brucellosis being a density dependent disease, it was decided by all co-managers that a further reduction in caribou abundance would be beneficial to the long term viability of the SHI population. In the meantime, continual monitoring and population assessments every 2 years would provide an early warning system, should the decline steepen.

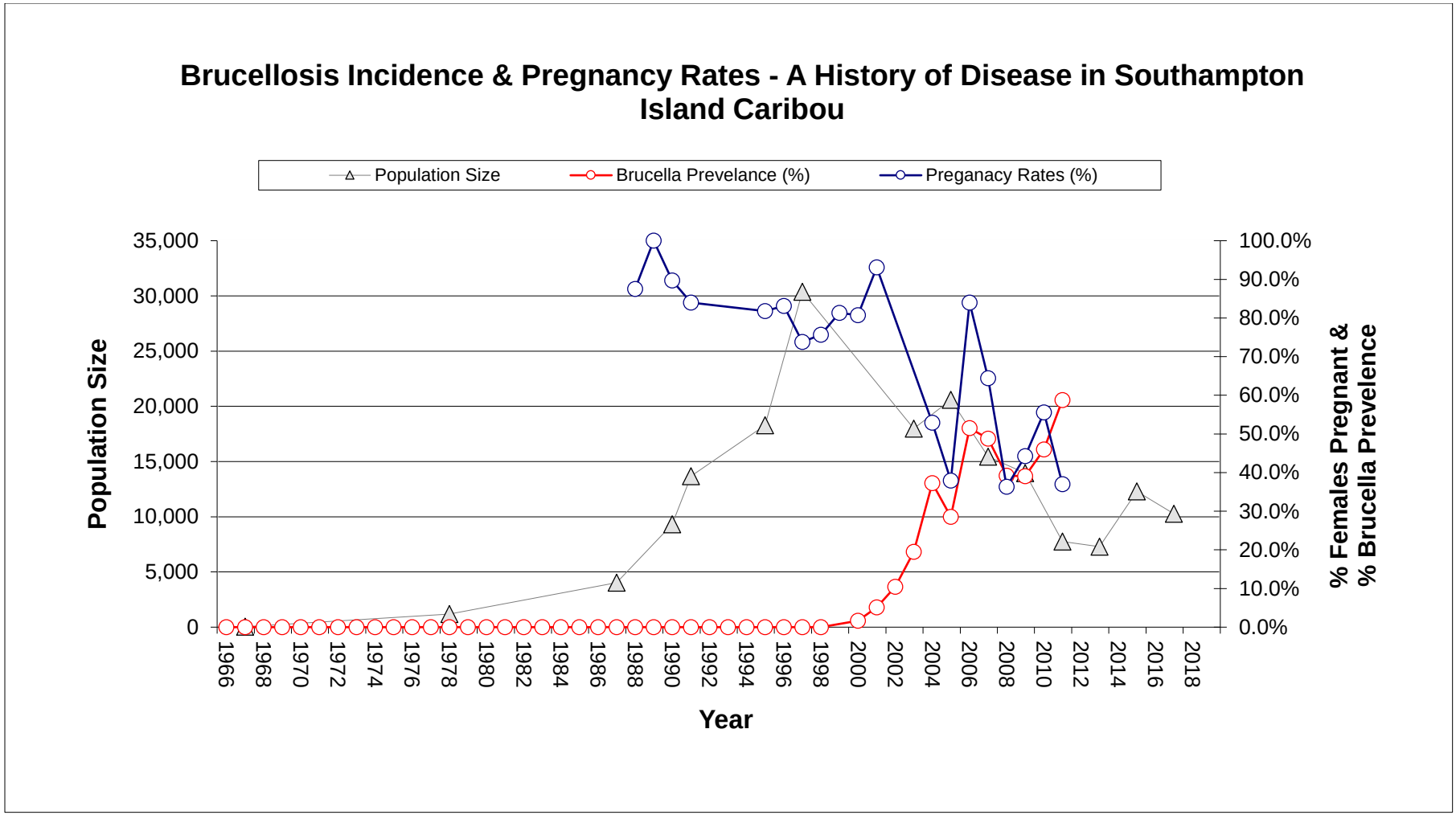

Figure 32. A history of abundance, pregnancy rates and *Brucellosis suis* prevalence for the Southampton Island caribou herd originally introduced onto the island from Coats Island in 1968.

Table 13. Circumpolar Incidence of Brucellosis in barren-ground Caribou and Reindeer across arctic North America (after O'Reilly, 1992).

Herd	Incidence (%)	Date	Remarks	Source
Southampton	Not Present	1990	75 samples	(NWT Wildlife notes)
Qamanirjuaq	4%	1966-68		(NWT Wildlife notes, 1983)
Beverly	< 2%	1983	118 samples	(Goldfarb, 1990)
Bathurst	Present	1981-1983	3 samples	(NWT Wildlife notes, 1983)
Baffin Island	14-35%	Mid-1980s	N Baffin highest	(O'Reilly, 1992)
Melville/Boothia	20-35%	1980s	17 samples	(O'Reilly, 1992; Gunn et al. 1991)
Ahiak	?			
Porcupine	15-20%	1980s	?	(O'Reilly, 1992)
Central Arctic	15-20%	1980s	?	(O'Reilly, 1992)
Western Arctic	</= 30%	1960-1980	?	(O'Reilly, 1992; Neiland et al. 1968)
Nechina	1-6.5%	1962-65	?	(Neiland et al. 1968)
George River	Not Present	1987-88	?	(Forbes 1991; Greenberg et al. 1958)
QEI Peary	Present	1980s	1 sample (P. of W. Island)	(Forbes, 1991)

4.5 Harvest:

Throughout the reintroduction of barren-ground caribou to SHI, wildlife managers of the time were vigilant in their on-going management of the herd. Management recommendations were, in all cases, based on research results, and particularly quantitative population estimates. In February 1978, the first caribou hunt since the 1968 introduction, was carried out on SHI. The quota was set at 25 bulls and was based on observations from a reconnaissance survey flown in 1977 that sighted a total of 172 caribou, 79 of which were adult males, 54 adult females, and 39 yearlings, suggesting a sex ratio skewed towards males (Kraft, 1978) (**Table 14**) (**Figure 33**). In August 1979, the TAH (quota) for bulls was increased to 50 largely based largely on the findings of the November 1978 population survey. Early in 1983 the first cow harvest was approved with a TAH set at 20. Regulations were developed along with this new TAH stipulating that 10 cows be harvested in the spring and the remaining 10 in the fall. The TAH was then raised from 50 to 250 bulls, and from 20 to 50 cows, based on recommendations generated following the 1987 population estimate (Heard and Grey, 1987).

During the 1988 harvesting year, concerns regarding the accidental harvesting of females seem to have led to the removal of the female quota and an increase in the male quota to 300 animals sometime in 1988. At this time, it was clearly indicated in the regulations that; “hunting zone J/2 (Southampton Island) was restricted to 300 male caribou.” In 1989 recommendations to increase the TAH to 400 caribou, of which 100 could be female were made. These recommendations were supported by Doug Heard who indicated the proposed increases were based on sound ecological principles (Renewable Resources Official Correspondence 140 007 005 & 150 001 005, October, 1989). Seasons for this new quota were recommended to be from October 1st to October 31st for males and April 1st to May 31st for females. By 1993, and in response to rapid population growth reported by Oullett in 1991, the TAH was removed (Oullett, 1992) (**Table 15**).

From 1993 up until the 2012 harvesting season subsistence harvest was not accurately monitored. In Nunavut monitoring of caribou harvest in the absence of a TAH is not mandatory. Although the 1991 NWMB Harvest Study attempted to assess wildlife harvest through hunter interviews, it is generally agreed that the final estimates are best guesses and may be misleading in some cases. For SHI, however, accurate records of harvest numbers and sex ratios (for most harvests) were kept as part of commercial harvests running consecutively between the harvesting years of 1992 through to 2007 and including 2009.

The first commercial quotas were established in 1992 and were set at 250 animals (gender breakdown unknown) (Junkin, 2003) (**Table 16**) (**Figure 33**). Despite the 1992 commercial allocation, it was not until 1993 that the first five caribou (of unknown gender), harvested for commercial purposes was reported since the herd's reintroduction from Coats to Southampton Island. Commercial quotas continued to rise to 1,000 animals in 1993, 5,000 in 1994 and 6,000 by 1997 (Junkin, 2003). Since 1993 there have been annual commercial harvests up to and including the 2009 harvesting season. Interestingly, a non sex-selective subsistence quota of 1,000 animals was re-instated in 1994 in an effort to offset an increase in the commercial quota from 1,000 to 5,000 over the same period (Junkin, 2003). By 1997, in response to survey results indicating the continued rapid growth of the population to 30,381 animals (Mulders, 1997), concerns about the caribou population having exceeding the Islands hypothesized carrying capacity of 15,000 caribou were being realized (Oullett et al 1994, Oullett et al 1993). In response to these concerns, the wildlife regulations were once again amended to allow an unlimited subsistence harvest and a non sex-selective commercial quota of 6,000 caribou.

Overall the commercial harvest was successful in reducing the population to the estimated carrying capacity of the Island of 15,000 caribou (Oulett et al. 1996). Current concerns however, are that continued high harvest rates, in excess of

6,500 caribou over the 2006 and 2007 harvesting seasons would drive the population too low to sustainably maintain the estimated subsistence harvest rate of 1,500 to 2,000 caribou annually. Additionally, there was the concern of rising Brucella prevalence and its observed impact on the reproductive potential of the SHI herd. The continued decline of SHI caribou following the 2003 survey estimate only heightened these concerns, and by 2007, when the population had dropped further to an estimated 14,389 adult and yearling caribou, discussions on ending the commercial harvest had begun. However, the harvest employed many local people and the political will to continue the harvest was high. Despite these pressures the harvest was cancelled by the Coral Harbour HTO in 2008 and only a small harvest of 843 was undertaken in March 2009. Between 1978 and 2009 an estimated total of 27,400 caribou had been harvested for subsistence purposes and 42,000 for commercial purposes yielding a total harvest of 69,400 caribou, of which 61% were taken for commercial purposes (**Table 16**). Since 2009 there has been no commercial caribou harvest. Results from the 2009 aerial abundance estimate showed no significant change between survey periods suggesting that the cessation of the harvest was having the net effect of slowing or stabilizing the population decline. But, over the same period, annual condition and disease monitoring tracked a steady increase in Brucellosis prevalence and a corresponding reduction in reproductive productivity (**Figure 32**).

Unfortunately the stabilizing effect lasted only a short period and by June 2011 estimates of population abundance dropped further to 8,442 adults and yearlings. With the commercial harvest having been stopped and the subsistence harvest remaining relatively constant at an estimated 1,500 to 2000 caribou annually, the reasons for this rapid decline appeared to now be related to the reported high prevalence of the reproductive disease Brucellosis. By March 2011, Brucellosis disease prevalence had reached a troubling 58.8% and spring pregnancy rates had plummeted to 37% (**Figure 32**). In addition to high rates of disease, around this time and despite the cessation of the commercial harvest, a new method of

selling country foods was gaining popularity, and increasing harvests of SHI caribou. This new harvest pressure was developing from the growing demand for the sale of caribou meat on social media. A ripe market had opened up on Baffin Island where Baffin communities were struggling with declining caribou populations as well. When sales of caribou from SHI on social media began, 24,764 kilograms of caribou meat was sold and shipped from SHI in the first 8 months of sales, representing an estimated 710 caribou (**Figure 34**). Unfortunately the data provided by the airline was cutoff in January 2012 thus removing our ability to assess the internet sales and harvest totals, through export traffic, for the months of heaviest harvest (March, April, and May).

4.51 Harvest Management and Planning 2011 to present:

Meetings in the summer and fall of 2011 between the GN Department of Environment and the Coral Harbour HTO, and additional meetings with all stakeholders in the winter of 2012, led to a formal request by the Coral Harbour HTO to the GN and the NWMB to apply a TAH of 4 caribou per household (1,000 caribou) in an attempt to stabilize the decline through harvest management. Additionally the annual condition harvest of 100 animals, used to asses Brucellosis prevalence and pregnancy rates amongst other health and condition indicators, was discontinued in order to move all harvesting opportunities to local Inuit.

Another product of these meetings was the development of the *Southampton Island Barren-ground Caribou Population Management Plan* (2012),which was submitted to the NWMB for decision in March 2012. The plan outlined an agreement to establish a TAH of 1,000 caribou and a Non-Quota Limitation (NQL) protecting cow/calf pairs. Also in the plan was the specification of continued harvester-supported monitoring, and the continued assessment of SHI caribou population abundance every 2 years. The urgency of the situation lead to the NWMB supported and community requested establishment of a Ministerial

Management Initiative (through the Nunavut *Wildlife Act*) to immediately assign a temporary TAH.

By May 2013, the herd had further declined to an estimated 7,287 adult and yearling caribou, prompting the GN to recommend a further reduction to 2 caribou per household (500 caribou) with 100 caribou held back for the HTO to use as deemed appropriate, for a total of 600 caribou. The community rejected this recommendation, preferring to wait until the May 2015 abundance estimate had been made, to make a final decision. The community based its decision on hunter observations of reduced signs of Brucellosis within their catch and a general thought that herd health and pregnancy rates were improving. Continued reports of healthy caribou, fewer signs of disease, several reports of a possible movement of caribou onto the Island over the winters of 2014 and 2015, and a noticeable increase in calves in June 2014, preceded the May 2015 abundance survey. Consistent with community reports, the 2015 survey estimated a significant increase in adult and yearling caribou. In two years, the population had increased by 5,081 animals to 12,368 caribou, an estimate far higher than could be accounted for by reproduction alone. The community of Coral Harbour was not surprised with the result, attributing the increase to what they believe was the movement of a large group of caribou from the mainland onto the north end of the island. In an attempt to verify these accounts, the GN conducted a genetic analysis using SHI hunter provided tissue samples from 2014 and then comparing them to SHI samples from 2004 and samples collected on the mainland in the vicinity of Naujaat.

Table 14. History of the Southampton Island assigned subsistence harvest quotas (TAH) from 1978 to 1991. Harvest management prior to the first commercial allocation in 1992 (subsistence harvest estimated using government reports, HTO correspondence and personal communications with wildlife staff).

YEAR	Regulated Quotas (TAH)						Total Allowable Harvest (TAH)
	Subsistence				Commercial		
	Female (#)	Male (#)	No Sex Selection (#)	Total (#)	No Sex Selection (#)	Total	
1978	0	25	0	25	0	0	25
1979	0	50	0	50	0	0	50
1980	0	50	0	50	0	0	50
1981	0	50	0	50	0	0	50
1982	0	50	0	50	0	0	50
1983	20	50	0	50	0	0	50
1984	20	50	0	50	0	0	50
1985	20	50	0	50	0	0	50
1986	20	50	0	50	0	0	50
1987	50	250	0	250	0	0	250
1988	0	300	0	300	0	0	300
1989	100	300	0	300	0	0	300
1990	0	400	0	400	0	0	400
1991	0	400	0	400	0	0	400

Table 15. History of the Southampton Island harvest assigned commercial and subsistence Quotas (TAH) from 1992 to present (subsistence harvest estimated using government reports, HTO correspondence and personal communications with wildlife staff).

YEAR	Regulated Quotas (TAH)						Total Allowable Harvest (TAH)
	Subsistence				Commercial		
	Female (#)	Male (#)	No Sex Selection (#)	Total (#)	No Sex Selection (#)	Total	
1992	0	400	0	400	250	250	650
1993	no limit	no limit	no limit	no limit	1,000	1000	no limit
1994	NA	NA	1,000	1,000	5,000	5,000	6,000
1995	NA	NA	1,000	1,000	5,000	5,000	6,000
1996	NA	NA	1,000	1,000	5,000	5,000	6,000
1997	no limit	no limit	no limit	no limit	6,000	6000	no limit
1998	no limit	no limit	no limit	no limit	6,000	6000	no limit
1999	no limit	no limit	no limit	no limit	6,000	6000	no limit
2000	no limit	no limit	no limit	no limit	6,000	6000	no limit
2001	no limit	no limit	no limit	no limit	6,000	6000	no limit
2002	no limit	no limit	no limit	no limit	6,000	6000	no limit
2003	no limit	no limit	no limit	no limit	6,000	6000	no limit
2004	no limit	no limit	no limit	no limit	6,000	6000	no limit
2005	no limit	no limit	no limit	no limit	6,000	6000	no limit
2006	no limit	no limit	no limit	no limit	6,000	6000	no limit
2007	no limit	no limit	no limit	no limit	6,000	6000	no limit
2008	no limit	no limit	no limit	no limit	6,000	6000	no limit
2009	no limit	no limit	no limit	no limit	6,000	6000	no limit
2010	no limit	no limit	no limit	no limit	6,000	6000	no limit
2011	no limit	no limit	no limit	no limit	6,000	6000	no limit
2012	NA	NA	1,000	1,000	0	0	1,000
2013	NA	NA	1,000	1,000	0	0	1,000
2014	NA	NA	1,000	1,000	0	0	1,000
2015	NA	NA	1,000	1,000	0	0	1,000
2016	NA	NA	1,600	1,600	0	0	1,600
2017	NA	NA	1,600	1,600	0	0	1,600
2018	NA	NA	1,000	1,000	0	0	1,000

Table 16. A history of the Southampton Island actual harvest from 1992 to present. Harvest estimates include actual commercial harvest and estimated subsistence harvest (subsistence harvest estimated using government reports, HTO correspondence and personal communications with wildlife staff).

YEAR	Actual Harvest								Total Harvest (Estimated)
	Subsistence (Values Estimated)				Commercial				
	Female (#)	Male (#)	Unknown (estimated)	Total (estimated)	Female (#)	Male (#)	Unknown	Total (#)	
1992	0	400	0	400	0	0	0	0	400
1993	?	?	500	500	?	?	5	5	505
1994	?	?	500	500	500	500	1,000	1,000	1,500
1995	?	?	1,000	1,000	?	?	2,356	2,356	3,356
1996	?	?	1,000	1,000	?	?	1,839	1,839	2,839
1997	?	?	1,500	1,500	2,356	1,009	0	3,365	4,865
1998	?	?	1,500	1,500	2,069	887	0	2,956	4,456
1999	?	?	1,500	1,500	514	580	0	1,094	2,594
2000	?	?	1,500	1,500	1,170	996	0	2,166	3,666
2001	?	?	2,000	2,000	2,070	1,626	0	3,696	5,696
2002	?	?	2,000	2,000	959	2,875	0	3,834	5,834
2003	?	?	2,000	2,000	3,403	1,602	0	5,005	7,005
2004	?	?	2,000	2,000	?	?	3,200	3,200	5,200
2005	?	?	2,000	2,000	2,766	1,272	0	4,038	6,038
2006	?	?	2,000	2,000	2,892	1,136	0	4,028	6,028
2007	?	?	2,000	2,000	1,446	1,129	0	2,575	4,575
2008	?	?	2,000	2,000	0	0	0	0	2,000
2009	?	?	2,000	2,000	322	521	0	843	2,843
2010	?	?	2,000	2,000	0	0	0	0	2,000
2011	?	?	2,000	2,000	0	0	0	0	2,000
2012	?	?	1,000	1,000	0	0	0	0	1,000
2013	?	?	1,000	1,000	0	0	0	0	1,000
2014	?	?	1,000	1,000	0	0	0	0	1,000
2015	?	?	1,000	1,000	0	0	0	0	1,000
2016	?	?	1,600	1,600	0	0	0	0	1,600
2017	?	?	1,600	1,600	0	0	0	0	1,600
2018	?	?	1,000	1,000	0	0	0	0	1,000
Grand Totals				**39,600**				**42,000**	**83,675**

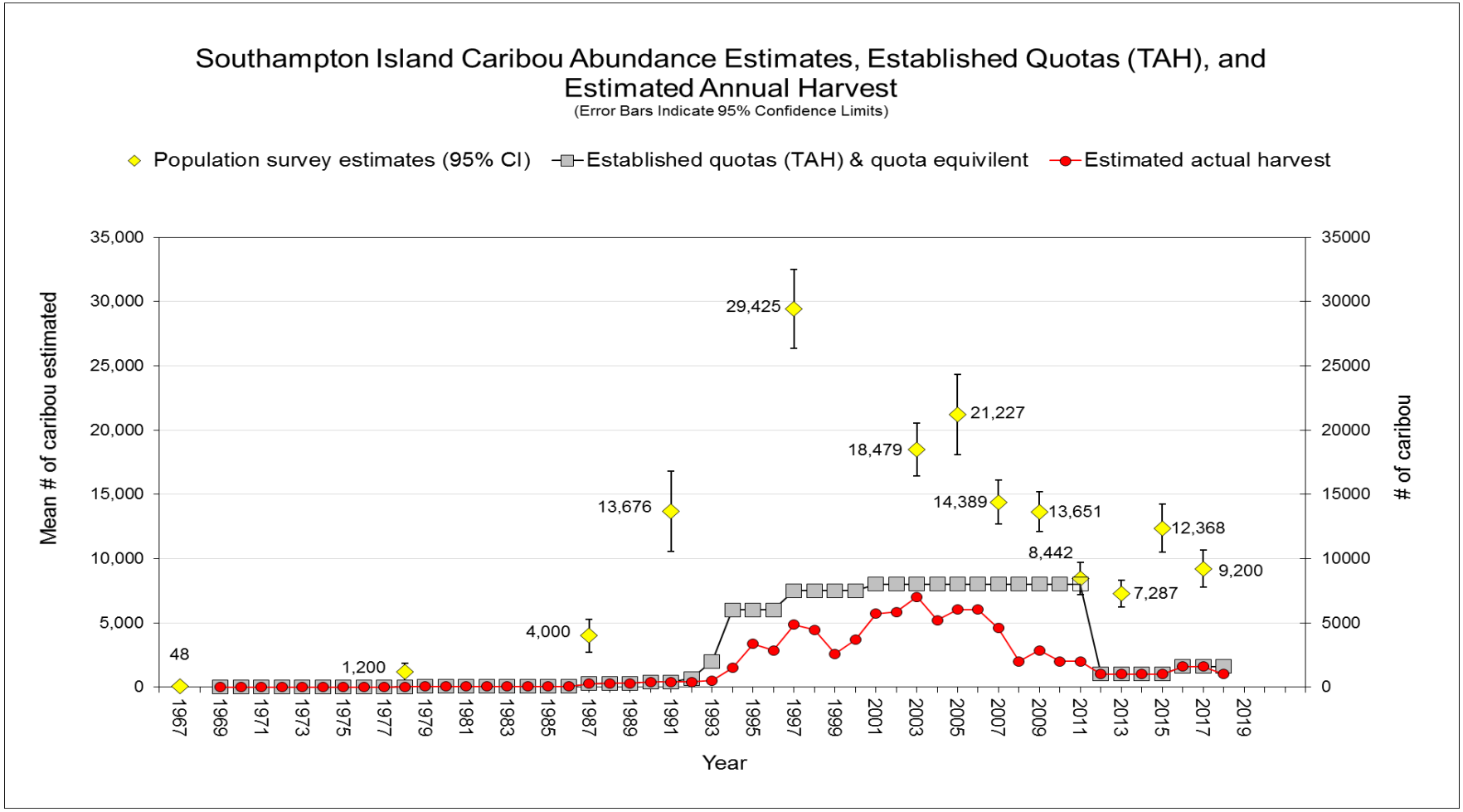

Figure 33. An examination of quota adjustment and actual harvest based on population estimates (Quota equivalents = estimated maximum subsistence harvest substituted for “no-limit” quota allowance values, Tables 1 and 2).

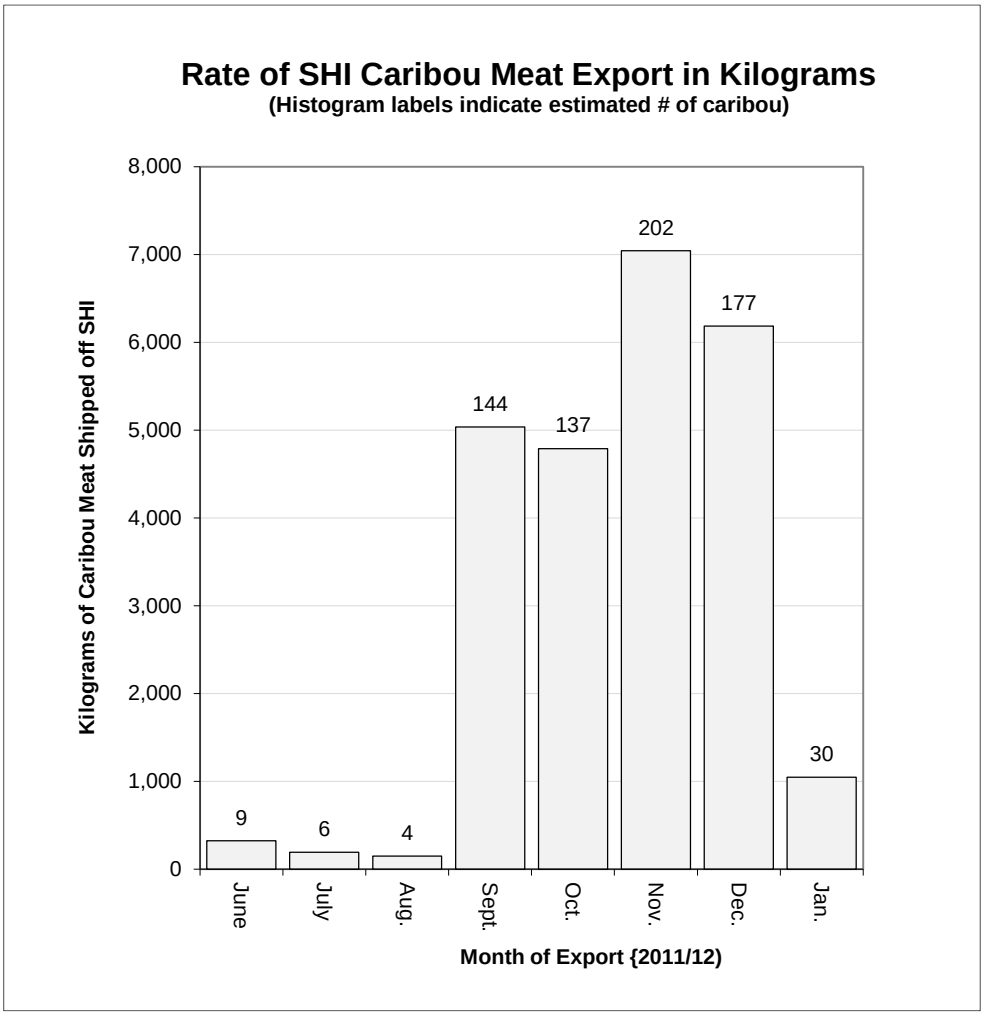

Figure 34. Caribou exports off Southampton Island primarily to Baffin Island communities. Data collected over an 8 month period in 2011/12.

4.6 Population Genetics:

The 2015 abundance survey results showed a statistically significant mean increase of 5,081 caribou from the previous survey in 2013, an increase that cannot be entirely accounted for by reproductive rates alone. The GN, in partnership with the Coral Harbour HTO, set out to try and confirm the possible mechanism of this increase. Based on information collected over two meetings with the Coral Harbour HTO, the primary mechanism forwarded by the HTO was the movement of caribou onto SHI. Hunter reports of many tracks coming onto the Northwest end of the island from across the sea ice suggested immigration was likely a behind the increase in caribou abundance. We sought to verify these observations through a genetic analysis of SHI tissue samples from 2014 (collected just following the reported movement) and 2004 (collected a decade prior to the suspected movement). Both these samples would then be compared with archived Qamanirjuaq caribou samples collected in 2012, and 2015 caribou samples collected in the vicinity of Naujaat, on the Kivalliq mainland. We employed Wildlife Genetics International (WGI) to analyze the results and test the validity of such a movement of caribou onto the Island.

Using Qamanirjuaq and Naujaat (Repulse Bay) samples to represent the mainland population, and starting out by using only the Southampton data from 2004 to start, WGI noted that the dramatic separation of mainland and island populations was not perfectly reflected across all individuals, even in 2004 (Paetkau, 2015) (**Figure 35**). Specifically, Qamanirjuaq individual C45 (partially red bar in group 2) and SHI individual 155 (partially green bar in group 4) were estimated to have ~ 35% ancestry in the 'wrong' population. These unusual individuals were previously dismissed as outliers, but that may have been premature: the stark differences in allele frequencies should have allowed accurate assessments of ancestry using 18 markers (Paetkau, 2015). Upon examining the 2014 samples, WGI found a marked shift between the 2004 and

2014 SHI genotypes, with 3% of the 2004 caribou being estimated to have < 90% SHI ancestry, versus 35% of individuals collected between 2013 and 2015 having < 90% SHI ancestry. Assuming that this shift is not the result of a change in sampling location — the NW region of SHI might show more mainland influence than the south — this change in the genetic composition of the population over the course of a decade is dramatic (Paetkau, 2015).

According to Paetkau (2015), the temporal shift was strong enough to leave little doubt that geneflow had occurred from the mainland to the island. To address the question of ancestry, Paetkau (2015) calculated the likelihood (Paetkau *et al.* 1995 *Mol. Ecol.*) that each genotype in the dataset would have been drawn from either the mainland (using Qamanirjuaq and Naujaat caribou herd DNA samples for allele frequencies) or the Southampton Island group (using 2004 data for SHI) (**Figure 36**). Paetkau concluded that with $P < 0.01$ that any genotype with a lower ratio did not have pure island ancestry, while ratios in excess of -7.8 ($P < 0.01$) had ancestry other than pure mainland.

With consideration to the number of tests conducted and associated hypothesis testing framework, WGI assessed the risk that the outliers are simply Type I errors. Having tested 86 individuals from the mainland, and 58 SHI individuals from 2004, a correction for multiple tests indicated critical values of 0.0006 and 0.0009, respectively, in order to achieve an 'experimentwise' $P = 0.05$, suggesting a genotype with a more extreme *P* than those that would be expected to occur through Type I error in 5% of similar datasets (Paetkau, 2015). The *P*-values estimated by GeneClass2 for C45 and 158 were 0.0003 and 0.0000, respectively, so these 2004 outliers cannot be explained by chance, even after correcting for the number of individuals tested (Paetkau, 2015). Paetkau therefore concluded that the evidence of movement in both directions (onto and off of the mainland) by 2004, was statistically meaningful. Indeed, both SHI individuals are statistically excluded as purebred members of either source

population (mainland or island), indicating that they are members of the F1, or subsequent, hybrid generation (Paetkau, 2015).

Moving forward a decade, Paetkau (2015) found that 19 of the 127 new 2014 SHI caribou had a likelihood of $P < 0.01$ that they were from "pure" SHI caribou as represented by the 2004 samples. According to Paetkau 23 individuals produced a P between 0.05 and 0.01 which individually could be explained as outliers (Type I error). As a group, however, Paetkau believed there were too many outliers to be so easily dismissed, as Type I error for a dataset of 127 pure SHI animals. In total, Paetkau observed 19 individuals beyond the critical ratio for $P = 0.01$, and 42 beyond $P = 0.05$ suggesting a substantial mainland influence present in 2014 but not present in 2004.

Though the results do not support that a pulse of mainland individuals had moved onto Southampton Island recently, they also do not support that genetic isolation of the island herd has been maintained. Paetkau (2015) points out that samples collected on SHI between 2013 and 2015 did not appear to include any F0 (parental generation) immigrants from the mainland. Paetkau concluded that the analysis has documented that a large proportion of 2014 SHI caribou samples (about 1/3 of the current set) are of F1 (offspring generation) or subsequent-generation hybrid ancestry.

One possible explanation of the absence of apparent F0 immigrants from the mainland could be that such individuals arrived at the northwest corner of the island and took a generation or more to reach as far south as the region where the hunter samples were collected, which is more towards the southcentral extents of SHI. This however, cannot explain the statistically significant increase in caribou abundance along with the local reports of mainland caribou migrating onto SHI between the May 2013 and 2015 surveys. Possible reasons for this finding could be related to a sampling bias whereby hunter samples collected

from early 2014 could have missed an immigration event occurring later in the winter. Though unlikely, consideration must also be given to the mainland comparative samples. Most of the samples were collected from areas close to Naujaat creating a second possible sampling bias that could have excluded more northern groups of caribou as potential source populations, such as caribou in the vicinity of Lyon Inlet. Clearly, additional genetic analysis needs to be undertaken to more accurately determine the cause of the hybridization event clearly documented sometime between 2004 and 2015. Overall we suggest that local hunter knowledge, and scientific evidence to date, all point to the arrival of a large contingent of caribou onto SHI from an area or areas not covered by SHI aerial survey extents.

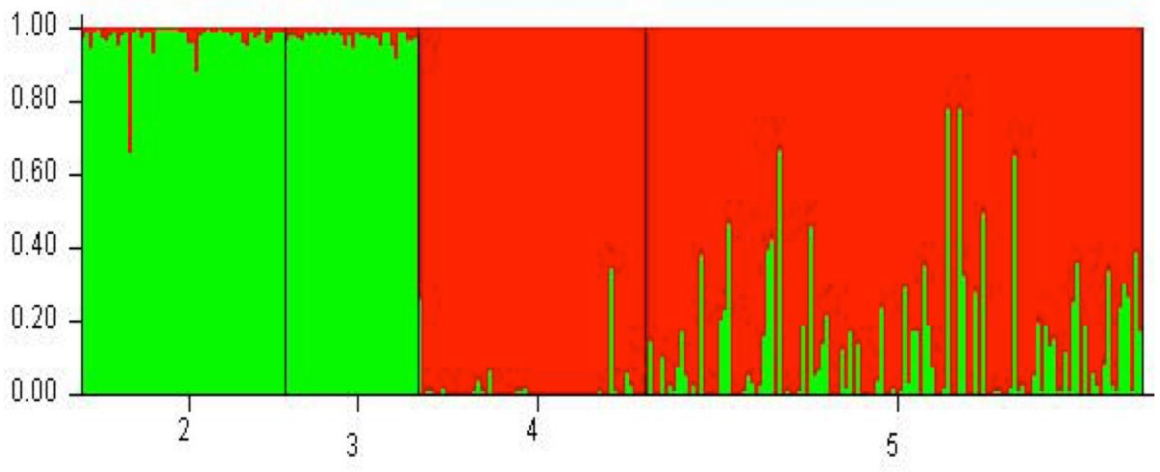

Figure 35. Structure results. Each column represents an individual, with its estimated proportion of mainland ancestry coded green, and SHI ancestry red. The 'populations' are Qamanirjuaq (2; w9741), Repulse Bay (3; g1616), SHI 2004 (4; w9741) and current SHI 2014 (5; g1616) (Paetkau, 2015).

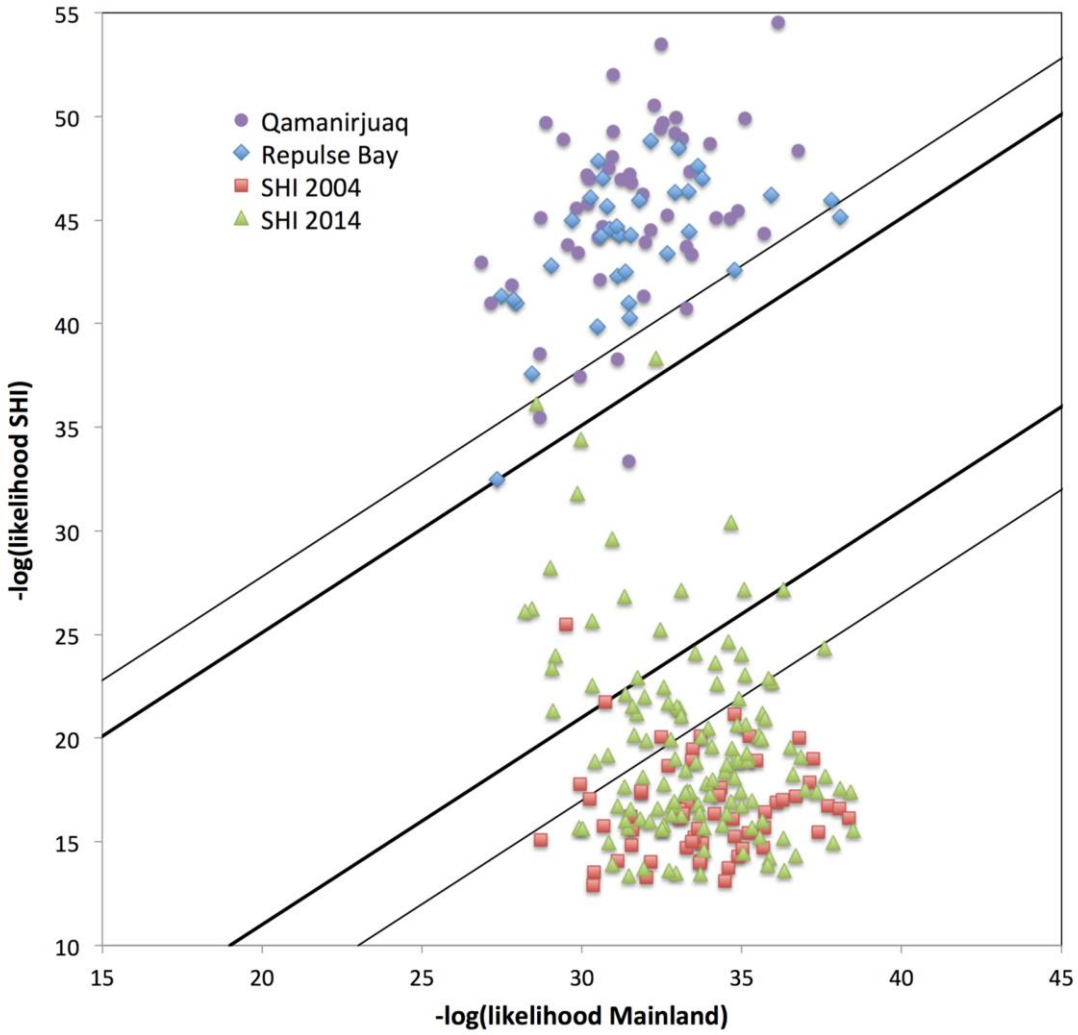

Figure 36. Likelihoods of occurrence based on mainland and (2004) island allele frequencies of caribou according to genetic analysis from different populations and years. Resampling in GeneClass2 indicated that 95% of purebred individuals are expected to have likelihood ratios outside the light lines, while 99% should sit beyond the heavy lines. Individuals between the heavy lines, including C45 (purple circle) and 155 (orange square) have genotypes that are rarer than 99% of individuals of either pure mainland or pure island ancestry. These include seventeen 2013–2015 SHI caribou (Paetkau, 2015).

5.0 Conclusions & Recommendations:

5.1 Aerial Survey Methods:

Overall, survey efforts from 1997 to 2017 were relatively precise (CV = 0.055 to 0.087) and were able to track two decades of decline within the Southampton Island caribou population. Methods changed over the period, namely from single observer pair configurations from 1997 through 2007, to dependant double observer pair configurations in 2009 to 2015, and finally to a composite of dependant double observer pair and distance sampling configurations, in 2017.

The dependant double observer pair configuration proved to be the most advantageous methodology, given that front and rear observers switch positions half way through each survey day, and that both front and rear observers are given the prescribed opportunities (see methods) to see the groups while flying along transects. The method reduced sightability errors common to the single observer pair method, and provides more precise estimates of wildlife populations. This method was the most effective at correcting estimates when the assumption of perfect sightability was violated. The dependant double observer pair method had other advantages. Incorporating more involvement of community members in research builds local support for the method and survey results, increases training opportunities for observers and improves research capacity in the territory, and incorporates co-management partners in research aspects of wildlife management.

Although the addition of distance sampling methods can further improve survey precision, the task of the observers becomes more challenging, and problems can arise when using observers with limited experience. In 2017, distance sampling estimates were higher than dependent double observer pair and strip transect estimates. This may have been due to one of the observer pairs not

putting enough survey effort to the bins near the aircraft (**Figure 25**) as indicated by different shapes of the detection histograms for the 2 observer pairs. This would cause a negative bias in both strip transect and dependent double observer pair estimates. This illustrates a potential issue with distance sampling, observers spending too much time looking out at further bins which are often easier to view than the closer bins rather than surveying one strip more thoroughly. The dependant double observer pair method partially accounted for this by also estimating the sighting probabilities of observers near the survey line. In the 2017 case, the observer was identified using dependant double observer records and the error addressed.

Based on our analyses and experience, we suggest that the dependant double observer pair method is the most appropriate method to meet the rigours of quantitative assessment while promoting collaboration with co-management partners. Distance sampling methods, though exceptional in many respects, should only be deployed when experienced observers occupy all observer positions, and, in combination with the dependant or independent double observer pair configuration. As abundance dwindles on Southampton Island, greater consideration should be given to incorporating distance sampling into survey methods. This may mean working closer with community HTOs to ensure only experienced observers are chosen, to reduce errors which contradict the assumptions of statistical models used in population estimates.

5.2 Herd Trend:

The SHI caribou population peaked sometime between 1995 and 2000 and has, since then, declined by an estimated 9% annually, up until the 2017 survey estimate. A probable immigration event sometime between May 2013 and May 2015 significantly increased abundance by an estimated 5,082 caribou, however, by May 2017 the population trajectory seems to have fallen back into the 9%

annual rate of decline trend that was documented up until 2013. Reasons for the decline are likely related primarily to three separate mechanisms including harvest, Brucellosis prevalence, and icing and its effects on forage availability during some winters. Brucellosis likely had little influence on abundance trend until 2004 when disease prevalence reached an estimated 40%. As a result, we believe harvest was the main mechanism of decline between 1997 and 2004. One must keep in mind, however, that the reduction in abundance was the goal during this period, as the population was believed to be well beyond the island's carrying capacity of 15,000 caribou (Oullet, 1993). Since 2004, both the reproductive disease Brucellosis and harvest were likely the main mechanisms of decline. Unfortunately, at this point we are unable to ascribe which may have had the greater effect on the abundance of SHI caribou. This being said, by 2005, abundance was still above the hypothesized carrying capacity of SHI (Oulett, 1993), so the management goal of reducing abundance remained unchanged. By 2007, herd estimates were below the estimated carrying capacity of 15,000 caribou, however, declines in abundance seemed to slow between 2007 and 2009, based on our surveys. Additionally, Brucellosis prevalence was declining by 2009 and, based on hunter reports, general condition was increasing. As Brucellosis prevalence had been steadily decreasing from 2006 through 2009, and the declines over the same period were slowing, the management goals were amended by the Coral Harbour HTO to reduce the Islands commercial harvesting. Agreement was reached amongst all co-management partners to suspend the commercial harvest after 2009, in an attempt to further stabilize the decline and maintain an abundance that could support the subsistence harvest. Between 2009 and 2011, however, the caribou population significantly dropped by 5,209 animals, the greatest observed decline over any 2-year period. During this period trends in Brucellosis prevalence reversed and climbed to the highest recorded, and pregnancy rates dropped to below 40%, the second lowest recorded since 2000. Additionally, the unanticipated sale of caribou meat through social media, a new form of commercial harvesting protected as a right under the Nunavut Agreement,

beginning in 2010, reached levels estimated to have exceeded the subsistence harvest over the 2011/2012 harvesting season. It appears that during this period, disease and harvest together were driving the population down. With the formal commercial harvest already stopped in 2009, the Coral Harbour HTO and GN had little option but to apply a TAH to reduce the subsistence harvest as an attempt to control the sale of caribou meat, primarily to Baffin communities, through social media.

The statistically significant increase in the SHI caribou population between May 2015 and May 2017, and subsequent decline of an estimated 9% between 2015 and 2017, has been difficult to explain. Genetic studies conducted as a follow-up to hunter observations suggesting a large group of mainland caribou had come onto the island from, the mainland sometime between 2013 and 2015, have yet to provide a conclusive answer regarding whether a migration event was the key mechanism of the increase. However, the genetic work did indicate that sometime between 2004 and 2015, a significant mixing of mainland and SHI caribou occurred. More analysis comparing consecutive years of SHI genotypes, with a more geographically broad collection of caribou genetic samples from coastal areas bordering SHI, will be necessary in order to more effectively explore possible mainland connections and reduce potential sampling bias that may be masking actual events. Although it is only a remote possibility, we believe that SHI caribou reproductive potential alone is unlikely to have accounted for the 41% increase estimated between 2013 and 2015.

5.3 Future Management:

Another survey planned for May 2019 will further assess the most recently observed decline in SHI caribou, however, based on our trend analysis we expect to observe further declines. Should a continued decline be confirmed, discussions with the Coral Harbour HTO and other stakeholders regarding the

consideration of a substantial reduction in TAH will have to be arranged shortly following the surveys completion, in an attempt to try and safeguard against further decline and associated hardship to the residents of Coral Harbour.

The mechanisms driving the decline are multiple and difficult to isolate, suggesting that further research is required. It appears that the main drivers have been the disease *Brucella suis* Type IV, harvest (with emphasis on the sale of caribou meat through social media), and potentially poor winter weather in some years. Clearly the need to continue monitoring disease prevalence in SHI caribou is required if we are to understand present day infection rates and associated productivity for the herd. Recently, hunters have reported fewer caribou with signs of disease, and a noticeable increase in the number of calves observed in 2015 and 2016 suggest that the disease prevalence may be decreasing. If this is the case, and Brucellosis no longer represents a primary mechanism of decline, then harvest, along with weather and condition monitoring should become the focus of future monitoring for the SHI herd. Additionally, more effective means of monitoring the harvest, and any exports of caribou meat off the island, will be critical in understanding the true extent of the harvest. At present these tools are not available to enforcement officers within Nunavut, suggesting that further thought and required amendments to current harvesting regulations should be seriously considered by wildlife management organizations. Attempts to control the sale of caribou meat through social media have failed and consideration should be given to addressing this issue through amendments to legislation. In recent consultations with Kivalliq community HTOs, all communities expressed a willingness to address the problem in this way, suggesting that some mutual agreement could be reached to more permanently address this issue. If nothing is done to monitor this novel and growing mechanism of caribou meat sales, we fear the problem will grow more serious as more and more caribou populations within Nunavut are managed through the establishment of a TAH.

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ᐱᓕᕆᓚᐅᓂ ᐊᓯᐊᓕᓕᓴᐅᔪᑦ, ᓄᓇᕗᓂ ᒪᓯᒪᐅᔪᕐᓂ ᐊᖏᓚᐊᓂ, ᐊᖏᓚᐊᓂ ᓄᓇᕗᓂ

ᐊᐅᓪᓚᕆᔨᓄᑦ ᓇᐃᓐᓇᕐᓂᒪᔪᖅ

ᓄᓇᕗᑦ ᑲᑭᒪᑯᓘᑎᓪᓗ, ᐱᓕᕆᓚᐅᖅᑕᒥ ᐊᐅᑎᓕᕆᔨᑯᓘᑦ ᑭᐅᖅᓕᒥ ᖃᐅᔨᓴᐃᓚᐅᖅ ᐱᓕᕆᔨᐅᑦᑐᑦ ᐊᐅᖅᓯᓚᐅᕐᓇᒪᔪᑦ ᓴᖅᓯᓄᑦ ᑲᑎᒪᔨᖃᑎᖃᕐᓂᖅᓯᒪᓕᕐᓂᖅᓴᖅᑐᑦ ᑕᑯᓐᓇᖅᑕᐅᑎᓗᑎᑐᖅ ᑲᑎᒪᔨᖃᑎᖃᖅᑐᓂᖅᓴᖅᑐᓂᒃᓴᓕᒫᓂ ᓄᓇᕗᑦ ᐊᖏᓇᓱᑐᓕᕆᔨᑯᓘᑦ ᑎᒥᖅᑯᑎᐅᓐᓂᒃ (HTOs). ᑕᐃᔨᓱᒪᓂ ᑲᑎᒪᓂᖅᖃᖅᑎᓯᓚᑦ ᑐᓴᕆᑎᕐᑕᐅᖅᑐᑦ ᓴᓯᓂᖅ ᖁᑭᖅᑕᖏᓐᓄᓂ ᑐᓴᑦ ᓇᓕᓐᓂᖅᑐᓂᒃᓴᓕᒫᓂ 2017-ᖏᑎᓯᓚᔪ, ᐊᒻᒪᓗ ᑭᐅᔨᓚᑦ ᐊᓱᖏᓐᓂᒃ ᐱᔨᑦᑕᐅᔨᓂᒃ ᐃᓱᒪᓚᑐᐅᔨᓂᒃ ᑐᓴᑦ ᐱᓕᕆᔨᑯᓘᑦ ᐃᓱᒪᓚᑐᐅᔨᓂᒃ ᐊᐱᖅᑯᑎᐅᔨᓂᒃᓴᖅᑐᑦ ᐅᓄᖅᑐᓂ, ᐊᒻᒪᓗ ᑭᐅᔨᓚᑦ ᐱᓕᕆᔨᑯᓘᑦ ᐊᐱᖅᑯᑎᐅᔨᓂᒃᓴᖅᑐᑦ ᓄᓇᕗᑦ ᐊᖏᓇᓱᑐᓕᕆᔨᑯᓘᑦ ᑎᒥᖅᑯᓂᓂᒃ ᐃᓕᓴᓂᒃ ᐃᓱᒪᓚᑐᐅᔨᓂᒃᓴᖅᑐᓂᒃ ᓄᓇᕗᑦ ᐊᖏᓇᓱᑐᓕᕆᔨᑯᓘᑦ ᐊᐱᖅᑯᑎᐅᔨᓂᒃ ᐃᓕᓴᓂᒃ. ᐅᓂᒃᑲᓕᕆᔨᐅᑎᓯᓚᐅᖅᑐᑦ ᐃᓕᓴᓂᒃ ᐊᐱᖅᑯᑎᐅᔨᓂᒃ ᐊᐱᖅᑯᑎᐅᔨᓂᒃᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᐅᔨᓂᒃ. ᐊᖏᓪᓗᒍᑎᓯᒪᔪᑦ ᐃᓕᓴᓂᒃ ᐊᐱᖅᑯᑎᐅᔨᓂᒃᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᐅᔨᓂᒃ ᐊᐱᖅᑯᑎᐅᔨᓂᒃᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᐅᔨᓂᒃ ᐊᐱᖅᑯᑎᐅᔨᓂᒃᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᐅᔨᓂᒃ. ᐊᖏᓪᓗᒍᑎᓯᒪᔪᑦ ᓴᐃᔪᑎᑦᑎᓯᒪᓕᕆᔨᑯᓘᑦ ᑕᐃᒪᓕᑦ ᓱ ᐃᓕᓴᓂᒃ ᓴᓇᔨᓴᖅᑐᑦ ᐃᓕᓴᓂᒃ ᓴᓇᔨᓴᖅᑐᑦ

Δ⊃⊂ᖏ⊂ ⊆⫟⊂ᑎᓯᖏ⊂

ᐊᐅ⊂⊂ᓯᔨ⊃ᓂ ᐅ⫟ᖁ⊃⊂ᐊᖍ ..... Error! Bookmark not defined.
ᐱᑎᐊᖍ⫟ᖏᖅ ..... Error! Bookmark not defined.
Δ⊃⊂ᖏ⊂ ⊆⫟⊂ᑎᓯᖏ⊂ ii
1.0 ᐅ⫟ᖁ⊃⊂ᐊᐟ ᐱᐟᔪᑎᖅ ᐊᒻᒪ⊃ ᐋᑲᕈᑫᓯᓂᖅᓯᒪ⫟ᖅ 3
2.0 ᐱᐟᔪᑎᖅ >⊂ᕀᵒ⫟ᐅᐟ 3
2.1 ᐋᑲᕈᑫᓯᒪ⫟ᖏ⊂ ᖃᑎᒪ⫟ᐅᔪ⊂ 3
3.0 ᕿᑦ⊂ᕀ⫟ ᖃᑎᒪᐟᔪ⊂ᐅ⊂ᐅᖍᖅ⊃⊂ ᓇᐃᓰᖍᓂᓯᔪ⊂ 4
4.0 ᕿᑦ⊂ᕀ⫟ ᐊᖏᔪᓂᓴ⊃⊂ᓭᓴᖅᑯᓂ ᑎᒥᖏᓪᓕᕐᓂ⊂ ᖃᑎᒪᐟᔪᑎᐃᓇ⫟ᖏ⊂ 12

1.0 ᐅᓂᖅᑲᓕᕆᑦ ᐱᓱᐊᑎᖓ ᐊᓚᒪ ᐊᑯᖏᓀᓕᓂᖓ

ᐅᑯᐊ ᐅᓂᖅᑲᓕᕆᓂ ᐊᑯᖏᑎᓀᕐᓂᐅᓕᒫᓂ ᐊᓚᒪᓗ ᓇᐃᓵᖅᓂᕐᓂᐅᓕᒫᓪᓗᑎᒃ ᐅᖅᑲᓂᐅᓕᒫᓂ, ᐊᐱᖅᑯᑎᐅᓕᒫᓂ, ᐃᓕᓴᐃᓚᐅᓕᒫᓂ ᓄᓇᓗᓂ ᐅᓂᖅᑲᓂᐅᓕᒫᓂ ᐅᓂᖅᑲᖏᓐᓄᑦ ᒥᐅ ᑲᑦᓴᐅ ᑭᖑᓪᓕᖅᑎ ᐅᒪᔨᐅᓂ ᖃᐅᔨᓴᖅᑎᑐ ᓄᓇᕗᑦ ᓚᖓᒪᑦᖁᖓᓂᐅᓂ. ᓴᓪᓕᓂᐅ ᓄᓇᓕᓂᐅᓂ ᐊᐱᖅᑯᑎᐅᓕᒫᓂ ᐅᓂᖅᑲᓂᐅᓕᒫᓂ ᐅᓂᖅᑲᓂᐅᓕᒫᓂ ᐅᓂᖅᑲᓂ

3.0 ᓴᓪᓂᓂᓄ ᑲᑎᒪᓄᐅᓪᓚᐅᖅᑐᒍᑦ ᓇᐃᓐᓇᖅᓯᒪᔪᓂ ᑲᑎᒪᔨᔪᑎᑕᐃᓄᑦ

ᐅᓪᓗᐊᑦ: ᓱᵒᓄᐊᕆᓚ 27th, 2019

ᑭᖑᓕᖅᑐᐃᐅᐅᔪᓂ:

- ᓄᓇᕗᑦ ᑕᖏᒪᵇᐞᖏᓂ - ᐱᓕᕆᔨᖑᓕᑦ ᐊᖏᑎᓕᕆᔨᵇᐞᑦ, ᒥᐣ ᑲᒪᐳᓪ, ᑭᖏᓪᓕᖅ ᐅᒪᔪᓕᕆᔨᓂ ᐅᒪᔪᓄᵇ ᖃᐅᐅᔨᕐᖑᓕᑦ
- ᓄᓇᕗᑦ ᑕᖏᒪᵇᐞᖏᓂ - ᐱᓕᕆᔨᖑᓕᑦ ᐊᖏᑎᓕᕆᔨᵇᐞᑦ, ᑭᓇᵒ ᓕᵒᑎᐅᓪ, ᑭᖏᓪᓕᖅ ᐅᒪᔪᓕᕆᔨᓂ ᖃᐅᐅᔨᕐᐃᐅᓂᖅᓗᑦ ᐱᓕᕆᔨᓂ
- ᓄᓇᕗᑦ ᑐᵒᖑᒪᓚᵇ ᑎᒥᖑᓂ, ᐅᒪᔪᓄᵇ ᖃᐅᐅᔨᕐᖑᓕᖅᑎ: ᑕᐃᓚᑦ ᓪ
- ᓇᐅᑦᓯ ᐊᖓᓇᓕᕆᔨᖑᓕᕆᔨᑭᓄᵒᓄᖑᓐᓂᖅᑐᖅ ᑲᑎᒪᔨᐅᔪᑦ ᐅᑯᖑᓐ
  - ᐃᵇᓯᖖᐅᑕᖅ, ᒧᓯᓯ ᓇᑯᓪᖅ
  - ᐃᵇᓯᖖᐅᑕᐅᑦ ᑐᖏᓕᖑᓐ, ᑕᓯ ᑲᑯᓪᖅ
  - ᐊᖑᓐᓚᔪᵇᑲᖅ, ᓇᑕᓴ ᐅᑕᑐᕈ
  - ᑲᑎᒪᔨᐅ, ᑕᓄ ᐋ
  - ᑲᑎᒪᔨᐅ, ᐊᕐᓇᐅᓴᖅ ᐃᒥᖅᑕᐅᑦ
  - ᑲᑎᒪᔨᐅ, ᓗᑲᓯ ᓇᑯᓪᖅ
  - ᑲᑎᒪᔨᐅ, ᓴ ᓴᖃᐃᕐᔪᵇ

ᐊᑯᓄᐅᓄᖑᓐᓄ:
2.10 ᐃᑲᕐᑦ

<u>ᖃᓇᐃᓕᖑᔪᕐᒪᖑᓐᓇᑦ ᐅᔪᑲᑕᐅᔨᓂᑦ ᐊᒪᒪᓗ ᐊᐱᖅᑯᑎᑦ:</u>

ᐃᓱᒪᓗᐊᑕᐅᓵᓗᐊᓕᐅᖅᑐᖅ ᑐᑲᑦ ᓇᓴᐊᔪᓗᔪᑲᑦᑕᔪᓄᖑᓕᑦ. ᓄᓇᖅᑲᖅᑐᑦ ᖃᐅᐅᔨᔭᒪᓕᐅᖅᑐᑦ ᖃᓇᐃᑎᑦ ᐋᵒᓂᐊᔪᕐᒪᖑᓐᓇᑦ, ᖃᓇᔪᓗ ᑐᑲᑦ ᐋᵒᓂᐊᔪᖅᑕᔪᖅᑐᐃᐃᓂᐅᖑᕐᒪᖑᓐᓇᑦ, ᖃᓇᔪᓗ ᖃᐅᐅᔨᖅᑎᐅᵒᓇᖏᕐᒪᖑᓐᓇᑦ ᑐᑲᑦ ᓇᓴᐊᔪᓗᑦᑐᕐᑕᑦ ᐊᖏᒪᓚ ᖃᓇᐃᓕᖑᔪᕐᒪᖑᓐᓇᑦ ᓂᓪᓕᖅᑐᓂᖅᑐᓂᑦ ᓇᓴᐊᔪᓗᑦᑐᕐᑕᑦ ᐊᖏᒪᓚ 2018-ᖓᑎᓪᓗᒍ. ᑲᑎᒪᔨᐅᔪᖅᑕᐅᓕᕆᔨᑦ ᐅᔪᖅᑕᐅᔪᖅᑐᑦ ᐊᖓᓇᕐᒪᖑᓐᓇᑦ ᓂᓪᓕ ᑐᑲᑦᖃᑦᑕᔪᓄᖑᓐᓂᵒᓄᵇ ᓇᓴᐊᔪᓗᑦᑐᕐᑕᑦᑎᓄᵒᓄᵇ ᑭᓯᐅᓂᑦ ᐃᓕᓴᕐᓂᖅᑕᐅᓕᕆᔨᖑᓕᕆᔨᑭᓄᵒᓄᖑᓐᓇᑦ. ᓂᓪᓕ ᑭᓱᐊᓂ, ᐊᖓᓇᕐᒪᖑᓐᓇᑦ ᑐᑲᑦᑎᖃᑦᑕᔪᓄᖑᓐᓂᖅᑐᑦ ᑐᓴᓂᖃᓂᔪᓄᖑᓐᓇᑦ ᐊᖏᒪᓚ ᐃᓱᒪᓗᐊᑕᐅᓵᓗᐊᓕᐅᖅᑐᖅ ᑲᑎᒪᔨᐅᔪᖅᑕᐅᓵᓗᐊᓕᐅᖅᑐᖅ ᐊᖏᒪᓚ ᐃᓱᒪᓗᐊ

4.0 ᓴᓪᓕᓂᓂᓄ ᓄᓇᓕᖅᓂᒃ ᐊᖏᓚᓯᖏᑐᓕᕐᓗᒃᑯᑦ ᑎᒥᐅᑎᖅᑕᒥ ᑲᑎᒪᔨᐅᑎᒪᓂᖅᓯᒪᓂᖅ:

ᓄᓇᖅᑐᑦ ᑲᖏᒪᒃᑯᖅᓯᒪᓂ

ᐱᓕᕆᓂᖅᓴᓂᑦ ᐊᖏᑎᓕᕐᓗᒃᑯᑦ

ᐅᒪᔨᓕᕆᓂᕐᒧᑦ ᐱᓕᕆᓂᖅ
ᑎᑎᖅᑲᓂᓕᖅᐸᒃ 120
ᐊᖅᖃᐊᑦ, ᓄᓇᖅᑐᑦ X0C 0E0
ᐅᖅᑲᓗᑦ (867) 857-2828
ᓯᑲᔨᒃᑯᑦ (867) 857-2986

ᑲᑎᒪᔨᐅᑎᖃᓕᖅᑐᑦ

ᐱᓕᕆᔨᑦ: ᐳᖏᑦᓕᕐᕈᑦ ᖃᐅᔨᓴᐃᓂᕐᒧᑦ ᑐᓴᕋᓂᕐᒪᑦ: ᓴᓪᓕᓂᓂᓄ

ᐅᖅᑳᑕᐅᔨᖅ: ᓴᓪᓕᓂᖅ ᖃᐳᖅᑕᖏᓚᓂᓂ ᑐᓂᑐᓂᒃ ᖃᐅᔨᓴᐃᓂᖅ ᐊᓱᒪᓗ ᖃᐅᔨᓴᐃᓂᕐᕈᑦ ᖃᐅᔨᓴᐅᐅᔨᓂᒃ ᑐᓴᖅᑎᑦᓯᓂᖅ ᐊᓱᒪᓗ ᐊᓱᒍᖅᑐᖅᓯᒪᓂᕐᕈᑦ ᖃᐅᔨᓴᐃᓂᕐᒧᑦ ᑐᓴᖅᑎᑦᓯᔨᓪᓗᒋᑦ.

ᓇᓂᓐᓂᖏᓚ: ᓴᓪᓕᓂᖅ

ᐅᓪᓗᐊ: ᑲᓐᓂᐊᖅ 27th, 2019 **ᐃᒃᓕᓐᓂᑦ:** 6:50 – 9:00

ᐊᒍᓂᐅᓂᖏᓚ: 2.10 ᐃᒃᓕᓐᓂᑦ

ᑲᑎᒪᓴᖅᑐᖅᑕᐅᔨᒪᔨᓪᓗ

ᐊᑎᖏᓚ	ᐃᓴᖅᓇᐃᓴᓴᕐᓚᓂᖏᓚ
ᒥᐣ ᑲᓪᓚᐅᓪ	ᓄᓇᖅᑐᑦ ᓴᓪᓕᓂᖅᓯᒪᓂᓐᓂ ᐅᒪᔨᓂᒃ ᖃᐅᔨᓴᖅᑎᖅᑎᑐ
ᑭᓇᓐ ᓪᓐᑎᐅᑦ	ᓄᓇᖅᑐᑦ ᓴᓪᓕᓂᖅᓯᒪᓂᓐᓂ ᐅᒪᔨᓂᒃ ᖃᐅᔨᓴᐃᓂᕐᒧᑦ ᐱᓕᕆᔨ
ᒡᓗᓗ ᓇᑯᓪᖅ	ᓴᓪᓕᓂᖅ ᓄᓇᖅᑐᑦ ᐊᖏᑎᓕᕐᓗᒃᑯᑦ ᑎᒥᐅᑦ ᐃᓕᓴᕐᓂᖅᐹᒥᒃ
ᒋᓗ ᓇᑯᓪᖅ	ᓴᓪᓕᓂᖅ ᓄᓇᖅᑐᑦ ᐊᖏᑎᓕᕐᓗᒃᑯᑦ ᑎᒥᐅᑦ ᑐᖓᓕᖓᓄ ᐃᓕᓴᕐᓂᖅᐹᒥᑦ

ᓇᒋᓴ ᐅᑦᑐᖅᑭ	ᓴᖅᑭᓂᖅ ᓄᓇᕗᒻᒥ ᐊᖏᓇᓱᐊᕐᑐᓂᓕᕆᔨᒃᑯᑦ ᑎᒥᖓᑦ ᐊᖏᖓᔪᖅᑲᖅ ᐊᑦᓚᐃᖏᒥ
ᐊᖅᓇᐅᔭᖅ ᐃᒥᖅᑕᐅᑦ	ᓴᖅᑭᓂᖅ ᓄᓇᕗᒻᒥ ᐊᖏᓇᓱᐊᕐᑐᓂᓕᕆᔨᒃᑯᑦ ᑎᒥᖓᑦ ᑲᑎᓚᔨ
ᑕᒥ ᐋ	ᓴᖅᑭᓂᖅ ᓄᓇᕗᒻᒥ ᐊᖏᓇᓱᐊᕐᑐᓂᓕᕆᔨᒃᑯᑦ ᑎᒥᖓᑦ ᑲᑎᓚᔨ
ᔪᑲᓱ ᓇᑯᒋᖅ	ᓴᖅᑭᓂᖅ ᓄᓇᕗᒻᒥ ᐊᖏᓇᓱᐊᕐᑐᓂᓕᕆᔨᒃᑯᑦ ᑎᒥᖓᑦ ᑲᑎᓚᔨ
ᓱᐧ ᓴᖃᐊᖅᓯᒃ	ᓴᖅᑭᓂᖅ ᓄᓇᕗᒻᒥ ᐊᖏᓇᓱᐊᕐᑐᓂᓕᕆᔨᒃᑯᑦ ᑎᒥᖓᑦ ᑲᑎᓚᔨ

ᑎᑎᕆᖅᑎ: **ᑭᓇᐤ ᓪᓗᑎᐅᓪ**

***** ᑭᓇᐤ ᓪᓗᑎᐅᓪ ᓂᐱᓪᐅᑦᓚᐅᔭᖅᑐᖅ ᑲᑎᒪᓂᐅᔨᓂᑲ ᑲᑎᒪᓴᐊᑎᐱᒪᕆᔨᐊᖅᑯᓪᓗᑎᓪ, ᓂᐱᓪᐅᖅᑕᐅᔨᓱᒪᔨᓪ ᓄᓚᕆᖅᑕᐅᔭᖅᑐᖅ ᑭᓚᕆᑐᖅ ᓄᓇᕗᒻᒥ ᐊᖏᓇᓱᐊᕐᑐᓂᓕᕆᔨᒃᑯᑦ ᑎᒥᖓᑦᓯᑎ ᐊᓚᒪᓗ ᐊᓱᓗᑐᓄᓪᓕᓪ ᐊᑐᖅᑕᐅᑲᖏᖓᑦᓯᑎᓂᓪᓗ ᓂᐱᓪᐅᖅᑕᐅᔨᓱᒪᔨᓪ. *****

ᑲᑎᒪᓴᐊᑎᐱᒪᓕᒃᓂᓪ

ᑲᑎᒪᓂᖅ ᒪᑐᐃᖅᑕᐅᔭᐅᔭᖅᑐᖅ ᑐᓴᓇᕆᓂᖅ 6:50

ᑭᓇᓱᑯᓂᖅᓯᓂᒃ ᖃᐅᔨᒃᑲᓂᖅ 6:52

ᒥᐣ ᑲᓪᓚᐅ ᐅᓂᓱᖅᑎᓪᓗᑐᓗ ᑭᓴᐅᓪᓕᖅᑎᓪᓗ ᖃᐅᔨᓴᐃᐃᓂᔪᓪ ᑐᓴᕆᓇᐊᖅᑐᖅᓂᐅᓂᓂᐅᓂᓂᐅᓂᐅᓂ ᒥᓱᓴᓪᓗ 7:00 – 8:45

ᖃᑲᒪᓂᕆᐊᓂ ᑐᓴᐅᓯᓂᓪᓗ ᑲᑎᑲᓂᖅ

ᑎᓴᓕᐅᓂ ᑐᓴᐅᓯᓂᓪᓗ ᑐᓴᑲᓂᖅ

ᖃᓂᖅᑕᐅᓂᓪ ᑐᑭᖅᑕᖅᓱᒪᓂ ᑐᓴᐅᓂᓪᓗ 2017-ᖓᔪᓕᓪᓗᐊ ᓴᓴᐃᓂᐊᓂᔪᓪᓗ ᑐᓴᑲᓂᖅᑎ

- ᑕᑯᓴᐃᐅᑎᓯᓪᓗᑐᒃ ᓄᓇᕐᖓᐊᑎᑐᓪᓗ ᓴᓂᒪᒪᓱᓕᒃ ᑐᓴᑐᓂ ᖃᖅᖏᑕᓯᐅᒃᑯᑦ ᓴᓴᖅᑕᐅᐃᑎᓪᓗᑕᑐᓂ 2005-ᒥ ᐊᒪᒪᓗ 2017-ᒥ ᐊᔪᐃᑦᖏᓕᕐᓂᖓᒃ.
- 2017-ᖓᔪᓕᓪᓗᐊ ᑐᓴᑐᓂ ᑕᕐᓂᓂᕆᐊᕐᓴᖅᑐᓂ ᐃᓄᐃᐊᓕᕐᓂᖅᑎᓯᐅᒪᔨᖅᑐᓂ.
- ᓴᓪᓗᓴᐃᖅᓯᓪᓗᑐᒃ ᐃᑲᔪᖅᓂᖏᓐᓂᖏᓂᒃ ᒪᕐᒋᖅᓂᒃ ᐃᓗᓪᓚᑐᖅᑐᓂᒃᑭᒃ ᓴᓴᐃᐃᔨᖅᑎᓪᓗᑎᒃ ᐊᑲᐅᓂᖅᑎᓯᐅᒪᔨᖅᑐᓂᖏᓂᒃ.
- ᑎᕆᒪᓂᒃ ᓴᓴᐃᐃᔨᖅᑕᐅᔭᖅᑐᓂ ᐱᑕᓱᔪᒪᓂᖅᑎᓯᓯᐅᒪᔨᖅᑐᓂᓂᒃ, ᒪᕐᒋᖅᓂᓗ ᑎᑎᕆᖅᑭᑦᑕᔪᐊᔨᖅᑎᓯᓪᓗᑎᒃ.
  - ᓴᓪᓕᖅᑯᓂᑎᔪᒪᓂᖅᑎᐅᑦᓚᑎᒃ ᒥᖅᑎᐅᖅᑎᓯᐅᓪᓕᖅᑕᓪᓕᖅᑐᓂ.
  - ᐅᑭᐅᓂᒃ ᑭᑭᑕᑎᖅᑐᓂᖅᑎᐅᑦᓚᑎᒃ ᐃᑲᔪᖅᑎᑎᖅᑐᓂᒃ ᐊᐅᑕᓂᖅᑎᓯᓪᓗᑎᓪᓗᑕᑐᓂ.
  - ᐋᖅᑭᒃᑯᔪᓂᓴᓐᓂᖅᑐᓂ ᖃᓄᐃᓂᒃ ᐅᑭᐅᓂᒃ ᑭᑭᑕᑎᓂᖅᓯᒪᔨᖅᑕᒥ ᑕᒪᑲᒃᑯᒃ ᐊᓚᐅᔭᐅᔪᓂᒃ ᑎᑎᕆᐅᑎᔪᒪᓂᖅᑎᓯᓂᓴᓐᓂᖅᑐᓂᒃ ᑐᑭᖅᑕᐅᔭᐅᔪᓂᒃ ᑎᑎᕆᐅᑎᔪᒪᓂᖅᑎᓯᓂᓴᓐᓂᖅᑐ

- ᑐᒃᑐᑦ ᓇᓗᐊᖏᓂᖅᓯᒪᔪᖅᑐᑦ ᐊᒻᒪ ᓯᖏᓪᓗᓴᐅᑦ ᐅᓄᖏᓕᕐᓂ ᐊᑯᓂᓕᕐᓂᓕ ᐊᖏᕐᕋᔭᐅᑦ 1966 – 2014 ᓇᓂᓇᐃᖅᓯᒪᔪᑦ.
- ᖃᐅᔨᓴᐃᓪᓚᐅᓕᑦ ᖃᐅᔨᓴᐅᔨᐅᓚᑦ ᐊᖏᕐᕋᔪᑦ ᐊᑯᓂᓕᕐᓂᓕ 1966-ᒥᑦ ᑎᑭᑦᓂᔪ 1999 ᓇᓗᐊᖏᓂᖅᓯᒪᔪᖅᑐᒥᐅᔪᖅᓯᑐᑦ ᑐᒃᑐᑦ. ᓇᓗᐊᖏᓂᒃᑐᑐᓄᑦ ᑕᖏᑉᖃᖅᑕᒥᓂᐅᔪᖅᑐᑦ ᓯᔾᔪᓕᖅᑭᖅᑎᑦ ᐊᖏᕐᕋᔪᓕᓄ 2000-ᒥ ᐊᒻᒪ ᑭᖑᕐᖏᓗᔪᑦ ᐅᓄᖏᓕᕐᓂᓕᒃᑐᓂᓐᓂᑎᑦ ᐃᓐᓄᐊᖏᖅᑐᑦ ᖃᑎᖅᑕᒥᑦ.
- ᓇᓗᐊᖏᓂᖅᓯᒪᓕᖅᑎᑦ ᐃᓐᓄᐊᖏᖅᑐᑦ ᐃᓄᐃᓗᓕᕐᓕᐊᖅᑐᑦ ᒪᑕᓗᓂᑦ ᐊᖑᓇᓯᑎᑐᓂ ᑕᓯᔾᔪᑎᖅᑐᓂᑦ.
- ᖃᐅᔨᒪᔪᑎᑦ ᑕᒪᑐᐊᖓ ᖃᓄᐃᓕᐅᑎᑦ ᐊᖑᓇᓯᑎᑐᓂ ᑕᓯᔾᔨᕐᓂᐊᕐᓂᒪᖓᓗᒃ ᑕᑐᐃᓐᓇᕐᓂᓕᒃᑐᓂᓐᓂᑦ ᐊᑐᐃᓐᓇᑦᑎᐊᓕᒃᑐᑦ.

MC – ᐊᑐᐅᓂᐊᕋᖏᖅᑐᑦ ᓄᓇᕗᒻᒥ ᐊᖑᓇᓯᑎᑦᓯᓚᐅᖅᑎᓯᑦ ᐃᓱᒪᒋᔭᖏᓐᓂᒃ ᑐᓴᕈᑎᒃ ᐅᑦᓗᒥᐅᔨᖅ ᖃᓄᐃᓕᕐᓂᒪᖓᓗᒃᑐᑦ ᑐᒃᑐᑦ ᐃᓐᓄᐊᖏᖅᑐᑦ ᐊᖑᓇᓯᑎᑐᓂ ᐱᑦᓗᒥᐅᔨᖅ ᖃᓄᐃᓕᕐᓂᒪᖓᓗᒃᑐᑦ ᐊᒻᒪᓗ ᐅᓄᖏᓕᕐᓂᓕ ᖃᐅᔨᔭᓕᕐᓂᒪᖓᓗᒃᑐᓂ ᓴᓪᓕᓂᑦ ᖃᑭᖅᑕᒥᑦ.

MN – ᑐᓴᓗᐊᓕᐅᖅᓯᒪᓐᖏᑕᕋ ᓇᓗᐊᖏᓗᓐᓇᖅ ᐊᑯᓂᒃᑎᓕᒃ. ᖃᐅᔨᓴᒪᓐᖏᑐᑐᓴᓕ ᐊᓯᖅᑲᓂᒃ. ᐃᓚᒪᖏᒃ ᐅᓪᓗᒥᐅᔨᖅ ᖃᓄᐃᓕᐅᓐᓂᖅᓯᒪᔪᖅᑐᑦ. ᐊᖏᕐᕋᔪᑦᑐᓂᓯᓂᒃᑐᑦ ᐃᓕᓐᓂᒪᒃᑯᑦ ᐃᓐᓄᐊᖏᖅᑐᑦ ᐅᓄᖏᓕᕐᓂᒪᖓᓗᒃᑐᑦ ᐊᒻᒪ ᑐᓴᓗᒃᑲᓂᓗᒃᑐᑦ ᐃᓐᓄᐊᖏᖅᑐᑦ. ᐃᓕᓐᓂᒪᒃᑯᑦ ᐃᓐᓄᐊᖏᖅᑐᑦ ᐃᓕᓐᓂᒪᒃᑯᑦ ᐊᑐᐃᓐᓇᕐᓂᓕᒃᑐᑦ ᐊᑐᐃᓐᓇᑦᑎᐊᓕᒃᑐᑦ ᐊᒻᒪᓗ ᐃᓕᓐᓂᒪᒃᑯᑦ ᐊᑐᐃᓐᓇᕐᓂ

- 30,000 ᑐᒃᑐᑦ ᖃᐅᖅᑕᒧᑦ ᐅᓇᓗᐊᓕᐅᖅᓯᒪᑕ, ᐊᖏᔪᓱᒃᑕᐅᕐᓂᒃᑯᑎᒃ ᓇᖏᔨᖅᖁᖅᑐᙵᐊᓂᐅᔪᑦ ᐊᒃᓱᒃᖳᓂᐅᔪᖅᖁᖅ.
- ᐊᒃᐸᓱᖏᓂᖅᐸᖏᓪᓕᐅᕈᖅᓯᒪᔪᑦ ᖃᒃᓂᕐᔪᓐᓇᓂ 2013 ᐅᓇᔨᓂᖅᑎᓕᐅᕈᖅᓯᒪᔪᑦ ᑎᑭᓪᓗᐊᑐ 7,200-ᓇᑦ ᑭᓱᐊᓂᓕ ᐅᑎᕐᕋᒥᓕᐅᖅᑐᑦ. ᓇᓇᓕᖏᓂ ᐊᖏᔪᓱᒃᑐᑦᓕᓚᐅᓴᒃᑯᑦ ᑎᒥᕐᑯᑎᖓᑦ ᐅᕐᖃᓕᐅᖅᑐᑦ ᑐᒃᑐᑦ ᖃᐅᖅᑕᒧᐊᕐᖃᑦᑕᕐᓂᖏᓪᓂᑦ ᖃᐅᖅᑕᐅᐅᕐᓂᑐᒥᑦ ᓇᓇᒥᖏᓐᓄᑦᑐᑦ ᓇᓇᕐᐳᖅᑎᓂᑦ ᐊᒻᒪ ᑕᑯᖃᑦᑕᕐᓂᕐᖁᓕᕐᓂᖏᓪᓂᒃ ᓇᕐᖃᕐᖁᖅᑐᓂᒃ. ᖃᐅᐅᔨᓴᐃᓕᐅᕐᓂᑦ ᐊᐅᖏᑎᒍᑦ ᑐᒃᑐᓂᒃ ᐊᒻᒪ ᖃᐅᐅᔨᐅᕐᓂᖅᑐᑦ ᓴᓕᑦᑎᓐᓂᑦ ᑐᒃᑐᑦ ᐊᒃᓱᒃᓴᓕᒃ ᐊᒃᑐᖅᑕᐅᓱᒪᔪᑦ ᑐᒃᑐᓂᒃ ᖃᐅᖅᑕᐅᐅᕐᖁᓕᕐᓂᖏᓪᓂᑦᓯᐊᓗᑎᖅᑐᓂᒃ.
- ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᓯᒪᔪᑦ ᐊᖅᖁᒍᐊᓂ 2017 ᐅᓇᔨᓂᖅᑎᓕᐅᕈᖅᓯᒪᔪᑦ ᑎᑭᓪᓗᐊᑐ 8,300-ᓇᑦ ᑭᓱᐊᓂᓕ ᐅᐱᐳᖅᑐᒍᑦ ᐊᖏᕈᓐᓇᖅᑐᓴᓂᓂᐅᔨᓂᒃ ᐅᓇᔨᓂᖅᑎᖅᐸᑦ 15,000-ᓂᒃ.
- ᑕᐃᒪᓕ ᑐᒃᑐᑦ ᐅᓇᔨᓂᖅᑎᖅᐸᑦ ᒥᖏᐅᖏᖏᑕᐅᓯᒪᔪᓇᑦ 15,000-ᓂᒃ, ᐅᖃᖅᖃᑎᒋᐊᖅᐸᓪᓕᐊᓕᕐᓂᖓᓕᖅᑐᑦ ᑲᑎᓪᓗᓴᓂᒃ ᑐᒃᑐᑕᐅᔪᓂᓕᖅᑐᑦ ᑭᒃᓕᕐᖏᓪᓂᒃ ᐊᒻᒪᓕ ᐊᖏᔪᓱᒃᓴᓕᒍᑦ ᒪᓕᕐᓂᑐᒃᓴᓕᒃ.
- ᓯᘁᒧᐊᓯᔪᐊᓂᖅᑎᖅᐸᑦ ᐊᑦᖢᑭᐊᖅᑕᐅᓂᕆᔨᖓ ᑲᑎᓪᓗᓴᓂᒃ ᐅᒪᔪᓴᓕᐅᕈᓐᓇᖅᑐᑦ ᑭᒃᓕᕐᖏᓪᓂᒃ ᐅᖃᓕᕐᓂᐅᔪᖅᑐᑦ ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᓯᒪᔪᓇᓂᒃ ᐅᑎᕐᕋᐅᐅᔪᖅ ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᓯᒪᔪᓇᑦ ᑐᒃᑐᓂ ᓂᕐᔪᓐᓇᖅᑐᑦ ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᓯᒪᔪᓇᑦ ᑐᒃᑐᑦ ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᓯᒪᔪᓇᑦ ᑐᒃᑐᑦ ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᓯᒪᔪᓇᑦ ᐊᒻᒪ ᐅᑎᖅᑎᒋᓗᑎᒃ.

LN – ᑐᒃᓱᔾᔨᓂᐊᕐᓯᒪᔪᖅᑐᒃ. ᓂᖅᑲᒃᖏᓐᓂᖅᓗᖅᑕᑐᖅᖁᓂᑦ ᑐᖅᑯᖃᓕᐅᕈᖅᑐᑦ. ᓇᓇ ᓱᑯᐊᖅᑐᐊᑦᑕᓕᐅᕈᓂᖅᖁᓇᓂ ᐱᔪᓀᖅᑲᓕᐅᕈᖅᑎᒋᔪᑦ. ᒪᖅᑯᓕᐅᕈᖅᓯᒪᑦ ᖃᑯᐊᖅᐸᖅᓯᓕᐊᓕᓐᓂᒃ ᐅᓇᖅᑐᐊᓕᐅᑎᓂᒃ ᐃᓇᓇᓴᓕᑎᐊᖅᑎᓕᐅᕈᖅᑐᑦ ᑐᒃᑐᑦ ᑐᖅᑯᖃᓕᐅᕈᖅᑐᑦ.

JS – ᑐᒃᑐᑐᖅᑲᓕᐅᕈᖅᑐᖅᖁᓂᒃ ᓇᓗᔪᐊᖅᓇᒃᑐᒥᒃ ᐱᓇᓱᐊᕈᓱᓚᓕᐅᕈᖅᑲᑕᑎᓐᓂᒃᓂᓕ.

MC – ᐅᑭ, ᓱᓕ ᑕᑯᖃᓕᑲᖅᐱᓱᐅᒃ ᓇᓗᔪᐊᖅᓇᒃᑐᒥ ᐊᐅᑦᓕᖅᓱᒪᑎᑦᓇᓱ?

JS – ᐃ ᐃᓕᓐᓂᒃᑯᑦ, ᑭᓱᐊᓂ ᐅᓇᓐᓂᓂᖅᓴᐅᓕᖅᑐᑦ ᑕᐃᔨᓱᒪᓂᓂᓕ.

MC- ᑐᒃᑐᑦ ᓇᓇᖅᐱᓕᐅᓴᒃᑐᑦ ᓇᓗᔪᐊᖅᓇᒥᖓᒪᑕᑦᑕᐅᖅ. ᐊᓐᓇ ᐊᓐᓂᐊᓚᖅᐅᙱᖅᑐᖅ ᑐᒃᑐᕆᐃᓇᓂᑦ. ᐃᓱᒪᓴᔪᑦᓕ ᑲᑎᓪᓗᐊᑐ ᐊᒃᐸᑐᔨᓴᐃᓐᓇᖅᑐᖅᑐᑦ ᐊᓐᓂᐊᖅᑲᓂᕐᑐᑦ ᖃᐅᔨᓴᐃᓐᓇᖅᑐᖅᑐᑦ. ᖃᐅᖅᑕᐅᓚᓇᓂᑦ ᓴᖅᑭᑎᓚᐅᖅᑐᑦ. ᖃᐅᖅᑕᐅᓚᓇᓂᑦ ᐊᓐᓂᐊᖅᑲᓂᕐᑐᑦ ᐊᓐᓂᐊᖅᑲᓂᕐᑐᑦᓇᓱᒃᑐᒥᒃ ᐅᖅᖁᓗ ᓇᓇᖅᐱᓕᐅᓴᒃ

MC – Δ Þ^{ᓕᒃ}ᑭᑎᐊᖅᑐᓗᓂ Þᓴᓕᐅᑎᐊᖅᓯᒪᔭᑦ, ᑐᖅᑯᓯᓴᓂᓇᖅᑐᖅ ᐊᓂᐅᐊᓇᖅᑎᒃ ᐊᒪᒪ ᓂᓚᔨᖏᓕᐅᓪᓗᓂ. ᑕᐃᒪᓕ ᑐᑲᓂᓂᑭᐃᓕᓂ ᓇᓱᐊᕐᓄᑦ ᐅᕐᖅᑕᓂᓂ ᐃᖏᓪᓂᓂ ᐃᓱᓴᖅᑯᑐᖅᑕᓯᓯᒪᓴᑦ, Þᖅᐅᑎᑕᔨᖅᑲᖅᑕᖅ ᐃᐅᓂᓇᖅᑕᐅᖅᑯᓪᓗᐃ. ᐃᓯᒪᖅᓯᖅᑭᒪᓕ ᓂᖅᑭᖓ ᐱᒪᔨᓂᕐᑎᑕᖅ. ᑐᑭᓯᑎᓯᑎᐊᑭᑎᓂᓂᒃ Þᖅᑲᓕᒪᔪᖅᑕᖅᑕᖅᑐᖅ ᓇᓱᐊᖅᑐᓂᓇᐅᔨᑦ ᒥᒃᓴᓂᓕ ᐅᒪᔪᓕᑦᓴᓂᑯᓂ ᑎᑎᕆᖅᓇᖓᓂᓂᒃ ᐱᔨᐅᔪᓇᖅᑐᓂ.

AE – ᖃᓄᖅᓯᓂᑯᐊ ᐋᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕᑦ?

MC- ᓯᖏᒪᐃᔨᐅᔪᑦ ᐃᒪᖅᓯᓚᖓᑎᓂ ᓂᓚᔨᐅᑎᑦᓗᑎᓂ ᐱᔨᐅᖕᐅᖅᑐᓂ ᐅᖕᖕᓂᓯᓂᓂ ᒪᑐᐃᖏᓴᓂᓂ. ᑭᓪᖅᓯᒪᖏᑦᑕᑦ ᓂᖅᔪᑎᑎ ᐋᓂᓂᐊᖅᑕᑭᕆᔨᖏᑎᑐᑐᖅ, ᑕᐃᒪᓕᕆᐃᓂᓇᖅᑕᐅᖅᑐᖅ ᐃᓄᐅᓂᒃ ᐃᓂᓯᒪᓂᑦ. ᐋᖏᓴᓇᓗᑎ ᐋᓂᑐᐃᑎᑦᓗᑎᓂ ᑐᓗᑦ, ᑕᐃᒪᓕᕆᐃᓂᓇᖅᑕᖅᑕᓂ ᐅᐃᓂᖏᓂ ᐊᖅᖏᑦᖅᓯᒪᔭᑦᑕ, ᖃᓄᖅᓯᒪᔭᑦᑕ, ᑭᓪᖅᓯᒪᔭᑦᑕᑦᓯᓂᓂᒃ ᐋᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕᑦ ᐋᐅᖏᓪᓂᒃ ᐋᓂᓂᐊᖅᑕᑭᕆᔨᖏᑎᑐᑐᖅ ᐃᓄᐅᓂᒃ ᐋᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕᑦ.

DL – ᓴᖅᑭᑎᐊᖅ ᐱᐅᔨᒥᒃ ᐱᓴᔨᑎᑦᓗᑎ ᐅᓄᖅᑐᑦ ᐊᖏᓪᓗᓯᓚᐅᔨᑦ ᑐᖅᑐᓂ ᐃᒃᓯᒪᔨᖅᑎᑦᑎᓪᓗᑎᑦ, ᐅᓄᖅᑐᓄᑦ ᓂᓚᔨᐅᑎᓂᓂᑦᓗᑎᒃ ᐊᔨᖅᑭᒃᓄᑦ ᓂᓚᔨᐅᑎᓄᑦ ᔨᐊᓕᒪᖅᑎᑭᒃᓕᑦᓕᑦ. ᑕᐃᒪᓕᓕᑦ ᐊᐱᓚᔨᖅᑕᐅᓂᖅᑐᖅ ᐊᖅᑭᑦᑕᐅᖅᑐᖅᑕᖅᑕᖅᑕᖅᑕᑦ ᐅᓂᒃᑯᓂᓂ ᐃᑯᐊᓕᑦ?

MC – Þᖅᑲᕐᓴᖅᓴᖅᑐᓪᓚᓕᓗ ᑐᖅᑐᑕᐅᓯᑦ ᓇᓱᐊᖅᑐᓗᑐᕐᓴᐅᖅᑐᓯᑕ ᓄᓇᖅᑲᖅᑐᓂ ᖃᖅᓂᑭᓴᓪᓚᓄᓇᖅᑕᐅᖅᑯᓇᓕᑎᓂ.

AE – ᐃᓄᐃᑦ ᑐᓴᓪᓕᐅᖅᑲᓯᒪᓕᒪᓕᑕ ᑕᐃᒪᐃᑦᑐᓂᒃ, ᑐᖅᑐᑐᐃᓇᐅᕐᓴᕐᑕ ᖃᐅᔨᖅᓂᒃ ᐊᒪᒪ ᒥᑭᑌᓄᑦ ᐊᖅᑭᑦᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕᑦ ᐃᓄᐃᑦ ᐊᖅᑭᑦᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕᑦ ᓯᓚᑦᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕᑦ. ᐃᓕᓴᐅᓕᖅᑎᑦᑎᓪᓗᑎᑦ ᐃᓄᐃᑦ ᐊᖅᑭᑦᓂᓂᐊᖅᑕᖅᑕᖅᑕᖅᑕ

MC –ጋϟϧΓᓇᖅᑐᖅ ᐃᓄᐃᑦ ᖃᐅᔨᓴᖅᑕᑦ ᑕᒪᑐᒍᖏᑦ ᐊᒻᒪ ᓄᖓᑕᐅᔪᓂᖅᓯᓂᖅ ᑎᑎᒥ ᐱᔪᓐᓇᕆᓇᕐᓂᖅᓯᓂᖅ ᐋᓐᓂᐊᖅᑕᐃᓕᐅᕆᓚᐅᖅᑐᓚᐅᖅᑎᑎᖅ ᓇᓚᖅᑐᖅᑕᐃᒪᐃᓐᓇᖅᑐᖅ. ᐊᐅᐅᓇᖅᑕᑦᐅᕐᑐᖅᓯᓕᒥ ᓇᖏᖅᓯᐊᕆᓚᓇᖅ ᐊᑦᓂ ᖃᐅᔨᓴᐅᓇᖅᑕᖅᑐᓚᐅᖅᑎᑎᒥᓚᒍᒍ ᑎᑎᖅ ᓯᓚᖅᑎᔪᓇᖏᑦ ᓯᕗᑦᔨᒧᑦᓗ ᐊᐅᖅᐃᓂᖅᑐᓇᖅᑕᖅᑐᖅ ᓯᓚᖅᑎᓚᓂ ᐅᓯᐊᓂᓕ ᐊᖏᔪᓕᐅᖅᑎᑎᖅᑐᑦᓯᒪᓕᖅᑐᖅ ᐱᔪᓐᓇᕆᓇᕐᓂᖅᓯᓂᖅ ᐊᒻᒪ ᐊᓯᐅᓚᓂ. ᑕᐃᒪᓕ ᐋᓐᓂᐊᖅᑐᖅᑕᓚᔨᐅᑦ ᖃᐅᔨᓴᓕᕐᒪᑕ ᑕᒪᑐᒍᖏᑦ ᐃᓄᐃᑦ ᐅᑎᖅᑕᑦ, ᖃᐅᔨᓴᖅᑕᕐᑯᒪᔭᖅᑕᑦ ᐊᐅᖅᓯᖅᑕᐅᓚᓂᖅᑕᑦ. ᐅᖃᐅᔨᔨᓇᐅᔨᔭᖅᑕᖅᓯᒪᓂ ᐃᓄᐃᑦ, ᑐᐅᑎᑦ ᐋᓐᓂᐊᖅᑐᖅᑎᑦ ᓇᖏᖅᓯᐊᕆᓚᓇᖅᑎᑦ ᐊᐅᖅᑕᐃᓐᓂᕐᒪᑦ, ᐅᐙᓴᓐᓂᑦ ᖃᐅᔨᓴᖅᑕᕐᒪᑦᓯᒪᓕᖅᑐᒪᑦ, ᖃᐅᔨᓴᖅᑕᕐᑯᐊᓘᖅᑐᓕᕐᓂᖅᑕᑦ. ᑕᐃᒪᓕ ᐱᑕᖅᖢᑦᑦ, ᑕᖅᑭᒃᑲ ᒪᕐᒉᒃ ᐊᓂᔪᓕᐅᖅᑎᑦᓚᓂᖅᑕᑦ ᐅᑎᕐᓚᓂᑦ ᐊᕐᑯᔨᓴᖅᑕᐅᒃᐅᓂᕐᓗᓂᑦ ᐱᔭᕇᑎᑦᓚᒍ ᓇᓚᖅᑎᔪᓇᐅᔨᑦ ᐋᓂᔪᓂᖅ ᐱᖓᓱᖏᑦᑐᖅ ᐱᔪᓐᓇᕆᓇᕐᓂᖅᓯᓂᖅ ᐊᐅᖅᓯᖅᑕᐅᓚᓂᖅᑕᑦ.

AE – ᕿᑯᓂᓇᒃᑲᒃ.

ᑐᒃᓯᕐᕈᑎᑦ ᐱᓕᓚᐊᒃᓴᓇᕈᒪᔨᓪᓗᑎᒃ

ᐃᓇᓇᐃᑦ ᖃᐅᔨᓴᓕᑐᖅᖁᓂᒃᑯᓂᓂᒃ ᖃᐅᔨᔨᓴᐃᓂᖅᖁᓂᒃ ᑐᒃᑐᓂᒃ

· ᓇᓇᕐᑦ ᑲᕈᒪᒃᑯᖓᓂ ᕿᓇᐅᔨᕐᕋᑎᖅᖁᖅᑐᑦ ᕿᓇᐅᔨᕐᖁᖅᑎᓂᓯᓯᕐᔪᐊᑎᓂᒃ ᐃᓇᓇᐃᑦ ᖃᐅᔨᓴᓕᕐᑐᖅᖁᓂᒃᑯᓂᓂᒃ ᑐᒃᑐᓂ ᐱᓪᓗᓂ.
· ᐋᓪᓕᓇᓕᓚᓱᐊᑎᖅᓂᖅᓂ ᑲᒪᓂᕐᐅᓂᑦ ᕿᓯᐊᓂᓚ ᐊᐅᓪᓚᕆᐅᖅᓂᖅᑐᑦᓂᖅ ᓇᓇᕐᑦ ᑲᕈᒪᒃᑯᖓᓂᓇᓂ, ᐊᐅᓪᓚᕆᐅᔨᓇᓇᖅᑐᖅᑎᑦ ᐃᓕᓴᖅᓯᕕᐅᓇᓂᒃ ᓇᓇᓪᓂᒃ ᐋᖅᓯᒪᔨᖅᑐᓚᐅᖅᑎᑦ ᑎᑎᕐᓂᖅᓂᓇᓂᑦ ᐊᒻᒪ ᕿᓇᒧᑦ ᑐᓄᕐᔨᐅᔨᒪᓂᖅᓂᖅ ᐃᓕᓴᖅᓯᕕᐅᓇᓂᒃ.
· ᑲᖅᓂᓀᓂᕆᐅᑦ ᐱᑎᕐᓯᓚᓂᖅᑐᑦ ᐱᓕᕐᓯᖅᑐᓂᒃ ᐃᓂᕐᓯᑎᖅᑐᓚᐅᓂᕆᖅᑐᑦ, ᐃᓪᓚᖅᑦ ᐃᓇᓇᐃᑦ ᖃᐅᔨᓴᓕᕐᑐᖅᖁᓂᒃᑯᓂᓂᒃ ᐊᑐᖅᑕᐅᓂᖅᑐᑦᓂᒃ ᕿᑎᓕᖅᑐᓂᑦ ᐊᐅᓪᓚᕆᐅᔨᓇᓇᖅᑐᖅᑎᑦ ᓇᓇᕐᑦ ᑐᓄᕐᔨᐅᔨᒪᓂᖅᓂᖅ ᓇᖏᖅᓯᐊᕆᓚᓇᖅᑎᑦ ᓇᓇᕐᑦ ᑲᕈᒪᒃᑯᖓᓂᓇᓂᒃ.

LN – ᐅᖃᐅᔨᓴᖅᓯᐅᖅᓯᒪᓂ ᑕᒪᓐᓇ ᐃᓕᓴᕐᕕᐅᓕᕐᓂᖅᑐᐃᓚᐅᑎᓂᐅᑎᓚᓂᒃ.

MC –ᐊᖅᐅᓪᓚᐊᕐᓴᖅᑐᖅ ᐃᓇᓇᐃᑦ ᖃᐅᔨᓴᓕᕐᖁᓂᒃᑯᓂᓂᒃ ᓇᐊᕐᓯᖅᑕᑦ ᐊᐧᑎᑎᖅᑳᖅᓴᖅᑐᖅ, ᐱᓕᕐᓂᖅᑕᖅᑐᓚᐅᖅᑎᑦ ᐅᖃᐅᔨᓴᓕᕐᑐᖅᖁᓂᒃᑯᓂᓂᒃ ᐃᓇᓇᐃᑦ ᖃᐅᔨᓴᓕᕐᑐᖅᖁᓂᒃᑯᓂᓂᒃ ᐃᓕᓴᖅᓯᕕᐅᓇᓂᒃ ᑐᓄᕐᔨᐅᔨᒪᓂᖅᓂᖅ ᓇᓇᕐᑦ ᑲᕈᒪᒃᑯᖓᓂᓇᓂᒃ ᐃᓕᓴᖅᓯᕕᐅᓇᓂᒃ ᐃᓇᓇᐃᑦ ᖃᐅᔨᓴᓕᕐᑐᖅᖁᓂᒃᑯᓂ

ᐊᐅᓪᓚᖅᑎᒍᓐᓇᖅᑕᓯ ᐱᓕᕆᔨᕐᑎᓐᓄᑦ ᐊᓗᒪᓗ ᓄᓇᕗᒥ ᐅᒪᓚᔪᓐᓇᕐᓂᐊᕐᓂ ᑲᑎᒪᔨᖏᑐᓄᑦ ᐊᓯᔪᖅᑕᐅᔪᓂᑕᒃ ᑲᑎᒪᔨᖏᑐᓄᑦ ᐊᓯᔪᖅᑕᐅᔪᓂᑕᒃ ᐅᑎᓪᓗᒍ 1000-ᓄᑦ.

- ᐃᑲᔪᐳᓇᖅᑐᒍᑦ ᑎᑎᕋᔪᓇᓂᕐᔪᓐ ᑕᐃᒪᓐᓇ.
- ᖃᓄᐃᓇᓕᔪᐊᖅᑐᖅ ᓇᖏᓂᖅ ᐱᓕᕆᓂᐅᕋᔪᓐ ᐳᔪᐊᓂᓕ ᐊᔪᕐᓇᓐᓄᑦ ᐊᐸᔪᖅᑐᐅᕋᔪᓂᕐᓂᔪᓐ ᐱᔪᕋᓐᓂᕐᕋᕆᔪᓐ.
- ᑕᒪᓐᓇ ᑐᓂᔪᐅᔪᒪᔨᖅ ᐃᓕᐃᔨᒪᓕᒫᑎᓂᑐᖅ ᑐᔪᕐᕿᔭᕐᑎᑎᐅᔨᖅ ᐊᖏᕐᖅᑕᐅᓕᐅᖏᑦᑐᓂᒃ ᐃᑲᔪᖅᑐᖅᑕᐅᖏᓐᓂᕐᒥᓕᐅᔪᓂᒃᑯᑦ ᐊᔪᕐᕈᑐᑦ ᐅᒪᔨᓄᑦ ᐃᓱᒪᓕᐊᑕᐅᔨᓄᑦ ᐱᔪᐊᓂᑕᐅᖅᑎᑦᑎᓂᒃᑯᑦ ᐊᔪᕐᕈᑐᑦ ᐅᒪᔨᓄᑦ ᐃᓱᒪᓕᐊᑕᐅᔨᓄᑦ ᐱᔪᐊᓂᑕᐅᖅᑎᑦᑎᓂᒃᑯᑦ ᐊᔪᕐᕈᑐ

DL – ᑕᒪᐢᓇ ᓴᐢᖅᑭᑦᓯᓂ ᐱᔨᕈᐅᔨᐢᓇᖏᑐᖅ ᓄᓇᕗᖅᓂ ᐊᐢᖃᓂᕆᔨᐅᑉᐹᖅᑯᑦ ᑎᒥᖅᓯᐢᓄᑦ ᐅᖅᑲᐅᖅᑎᑦᓯᓂᕐᒥᑦ ᓄᓇᕘᑦ ᓱᖏᒪᖅᑯᖅᑎᑦ ᒪᕗᔪᐢᐅᖅᓯᖅᑎᖅᓂᕐᒥᓂᖅ ᒥᐢᖃᓂᖅᑭᔨᕐᓴᖅᑕᑦ. ᑕᖅᑯᐊᑐᐊᑦ ᑐᖅᑐᑦ ᐊᑐᓕᐊᖅᖄᑎᕈᐅᖅᓯᒪᓕᕐᓂᖅ ᖃᐅᔨᖅᑎᒃᓴᖅᑐᑦ ᐅᒪᓕᕐᕈᐅᔨᓇᖅᑐᑦ ᑭᓱᓕᕐᕈᐅᖅᑎᕈᐅᔨᓂᒪᓕᐅᖅ ᐊᒻᒪᓗ ᓄᓇᕘᖅ ᐱᓕᕐᓇᖅᖄᑦ ᒥᐢᖃᓂᖅᑭᔨᕐᓴᖅᑕᑦ ᐅᒪᓕᕐᕈᐅᔨᓇᖅᑐᖅ ᐊᒻᒪᓗ ᒪᓕᕐᕈᐅᖅᑎᒃᓴᖅᑐᑦ ᐅᒪᓕᕐᓂᕐᒥᑦ ᑭᔨᔨᕐᓂᖅ ᓄᓇᕗᑦ ᓱᖏᒪᖅᑯᖅᑎᑦ ᒪᓕᕐᕈᒪᓕᖅᑎᑦ ᐱᓕᕐᓇᖅᖄᑦ ᐊᒻᒪᓗ ᑐᓂᔨᕐᓂᖅᑐᖅ ᐊᒻᒪᓗ ᑐᖅᑐᓂᔨᕐᓂᖅᑐᖅ ᐅᖃᓕᕐᓂᖅᑐᖅ ᐅᖃᐅᖅᑯᓯᒪᔨᖅᑕᐅᓂᖅᑐᖅ ᑐᖅᑐᓗᓂᑦ.

MC – ᐊᔪᐅᖅᑐᐃᓐᓇᐅᖅᓯᐊᖅᓴᔪᑦ ᑭᓇᐅᔪᓂᖅ ᐱᓚᑦᓴᑐᑦ ᑭᑦᓯᐊᓂᑦ ᐃᑲᔪᖅᑕᐅᔪᓚᐊᖅᑲᓴᖅᑐᖅ. ᐊᑐᐃᓐᓇᐅᖅᑎᑦᓯᐊᖅᑲᓴᖅᑐᑦ ᐃᓚᖅᑎᖅᑐᑦ ᐱᓚᑦᓴᑐᑦ ᐃᓕᓴᕐᓂᐊᕐᓇᖅᑐᑦ ᐊᒻᒪᓗ ᑐᖅᑐᓂᔨᕐᓂᖅᑐᖅ ᐊᒻᒪᓗ ᑐᖅᑐᓂᔨᕐᓂᖅᑐᖅ ᐃᓕᓴᕐᓂᐊᕐᓇᖅᑐᑦ ᑭᓱᓕᕐᓂᖅ ᑭᑦᓯᐊᓂᑦ ᑭᓱᓕ

AE – ᐃᓚᒪᖏᑲ ᓯᓚᐅᐸ ᐊᓯᔨᖅᓯᖅᓂᓕᐅᓂᖅᓴᓗᓂ?

MC –ᐋᖅᔪᑐᐃᓐᓇᕆᐊᕐᖏᖅᑐᖅ ᐃᓕᖅᓴᔪᓂ, ᑭᓯᐊᓂᓕ ᐊᔨᐱᓂᓴᓂᓕ ᖃᑭᓂᐅᑎᐅᔪᓂ ᑕᐃᒪᐃᓴᔪᖅᓴᒪᑕ ᐃᒪᖏᑲᓴᐊᒍᓂ ᐱᓇᐃᓴᕐᓂᓕᖅᑐᑦ ᐃᓇᐃᓴᕐᓂᓕᖅᑐᑦ ᑕᐃᒪᐃᓕᓂᐅᕐᖏᖅᓯᖅᑐᑦ.

DP – ᓇᓴᐃᓕᐅᖅᐱᓕ ᖃᑭᖅᑕᓗᖅᒥ ᐊᓪᒪ ᖃᑲᓯᐅᓕᐅᖅᓯᓂ?

MC –ᓇᓴᐃᓕᐅᖅᓯᒪᔪᔪᑦ ᐊᖏᖅᔫᐊᓂ 2014-ᒥ ᐊᓪᒪ ᐅᓇᖏᓂᖃᓕᐅᖅᑐᑦ 5,500-ᓂᒃ ᖃᑭᖅᑕᓕᓕᒫᖏᑦ ᐃᓪᒪᖏᑲᒃ 1,700-ᖅᔪᓕᐅᖅᑐᑦ ᖃᑭᖅᑕᖅᓴᓂ Prince Charles Island. ᑕᐃᒪᓐᖅᓂᑦ ᐅᑎᖏᐅᓕᐅᖅᓯᓯᖅᑕᒪᓂᑦ ᓇᓴᐃᖅᑭᖅᑐᖏᓗᑎᒃ. ᑐᓴᓂᒃᓴᑦ ᐱᓯᒪᔭᕐᑦ ᒪᓕᓪᓗᑎᑦ ᐅᖏᖅᑐᑦ ᐅᓐᖓᑎᑐᓐᓇᓂᑦ ᐅᓇᓴᓂᕆᔨᖅᑎᑦ ᓯᓂ ᕿᓗᐊᓐᖅᑎᑦᑐᑦ ᓂᑭᐊᓂ ᖃᑭᖅᑕᓗᒍᓵᓐ ᑐᑐᑦᑐᓂ ᐅᓇᖏᓂᖃᓕᐅᖅᑐᑦ ᑭᓯᐊᓂᓕ ᒧᓐᓇᐃᓐᓂᓕᖅᑐᑦ ᑕᐃᒪᐃᓕᓂᐅᕐᖏᖅᓯᖅᑐᑦ ᑭᓯᐊᓂᓕ ᐃᓕᓴᕐᓂᓕᖅᑐᑦ ᐱᓕᕆᔨᒃᑯᑦ ᐃᓚᖅᑎᑦᑎᓂᒃ ᐊᓯᔪᖏᑦᑐᖅ ᐃᓇᐃᓴᕐᓂᓕᖅᑐᑦ ᐅᓪᓗᒥ ᐃᓚᖅᑎᑦᑎᓂᒃ ᐊᓯᔪᖏᑦᑐᖅ ᐃᓇᐃᓴᕐᓂᓕᖅᑐᑦ ᐱᓕᕆᔨᒃᑯᑦ ᑭᓯᐊᓂᓕ ᐃᓚᖅᑎᑦᑎᓂᒃ ᐊᓯᔪᖏᑦᑐᖅ ᐃᓇᐃᓴᕐᓂᓕᖅᑐ

Yes sorry, no harvesting females with calves.

Natasha Hattie Ottokie
Aiviit Hunters' and Trappers' Organization
Manager
Coral Harbour, NU
X0C 0C0
(T) 867-925-8622

From: Denis Ndeloh [mailto:DNdeloh@nwmb.com]
Sent: Thursday, May 23, 2019 4:44 PM
To: Aiviq HTA <aiviit@kivalliqhto.ca>
Subject: RE: Support

Hi Natasha,

Did you mean NOT harvesting females with calves? That is what the GN recommendation is.

From: Aiviq HTA <aiviit@kivalliqhto.ca>
Sent: May 23, 2019 5:31 PM
To: Denis Ndeloh <DNdeloh@nwmb.com>
Subject: Support

Hi Denis,

The Aiviit HTO supports the TAH recommendation to decrease to a 1000 for Coral Harbour for caribou harvesting and harvesting of females with calf.

Natasha Hattie Ottokie
Aiviit Hunters' and Trappers' Organization
Manager
Coral Harbour, NU
X0C 0C0
(T) 867-925-8622

ⵋⵔⴷⵛⴷⵗⵖ

ⵔⵇⵚⵍ ⵔⵖⵀⵏⵔ ⵃⴷⵛⵛⵏⴻⵒⵔ

ⴷⵀⵔⵙⵠ

ⵋⵚⵍⵛⵃⵏ: ⵄⵔⵍⵛⵖ: X

ⵑⵑⵀⵛⴷⵗⵖ: 2017-ⵛ ⴷⵛⵠⵍⵄⵛ ⵇⵊⵠⵖⴷⵔⵔⵛ ⵃⵍⵊⵔ ⵑⵛⵖⴱⴷⵖⵔⵔⵛ, ⴷⵛⵠⵍⵠⵔ ⵃⴷⵛⵛⵏⵔⵖ MX-08, Boothia ⵔⵚⵃⵔ

ⵑⵑⵀⵛⴷⵗⵖ

- Boothia ⵔⵚⵃ ⴱⵠⵋⵛⴷⵗⵖ ⵄⵔⴷⵑⵔⵔ ⴷⵛⵠⵍⵄⵛ ⴱⵏⵑⵑⵛⵔⵃⵙⵖⴱⵠⵔⵖⵛⵛⵖⵛⵖ ⵄⵔⵔⵋⵖⵇⵛⵔⴷⵖⵔⵔⵖ. 1985-ⵛ, Boothia ⵔⵚⵃ ⵖⴱⴷⵑⵍⵖⴷⵛⴷⵖⵔⵔⵋⵖ ⴷⵛⵠⵍⴱⵛⵖⴱⵛⴷⵠⵔⵠⵔⵛⵍⵔ.

- 1995-ⵛ, ⴷⵛⵠⵍⵠⵔⴱ ⴰⵑⵍⵄⵛⴷⵖⵔⵔⵛ ⵇⵛⴷⵛⵒⵔⵖⴱⵛⵖⵔⵏⴱ 554 ± 205 (S.E)-ⵔⴱ. ⵛⵄⵍⵠⵍⵔⵔ, ⵛⵔⵔⵑⵑⵖⵛⴷⵛⵄⵛ ⴷⵖⴱⵖⵒⵍⵑⵛ ⵑⵛⵛⵖⵔⵑⵔⵠⵔⵔⵠⵔⵛ ⴷⵛⵠⵍⵄⵛ.

- ⵛⵔⵔⵑⵑⵖⵛⵛ ⵍⵖⴱⵄⵏⵛ ⵄⵒⵍⵑⵛ ⴷⵛⵠⵍⵄⵛ ⵑⴷⵠⵒⵛⵔⴱⵒⵛ ⵑⴱⵔⵒⵔⵖⵔⵠⵔⵔⵠⵔⵛ ⵔⴱⵔⵄ ⵃⵖⵇⴷⵖⵔⵠⵔⵔⵠⵔⵛ Boothia ⵔⵚⵃⵠⵛ, ⴷⵛⵠⵍⵄⵛ ⵄⵇⵠⵏⵔⵑⵖⵍⵛ ⵔⴱⵔⵄ ⵄⵔⵔⵖⵛⵖⵛⵖⵔⵠⵔⵔⵠⵔⵛ.

- ⵊⵠⵇⴷⵑⵗⵖ ⴱⵏⵔⴱⵏⵛⵖⵔⴷⵖⵔⵔⵛ ⵍⵖⴱⵄⵛⴷⵑⵗⵇⵖⵔⵔⵛ (TAH) 66-ⵔⴱ MX-08-ⵔⵛ ⵊⵖⴱⵖⵛⴷⵑⵔⴷⵖⵔⵔⵖ ⵑⵛⵄⵔ 1, 2014-ⵛ ⵍⵛⵔⴱⵔⵔ ⴷⵛⵠⵍⴱⵛⴷⵑⵗⵇⵖⵔⵔⵛ ⴱⵏⵖⵔⵔⵛ ⵑⵖⵛⵗⵔⵖ, ⵊⵍⵍⵔ ⵊⵖⵏⵔⵊⵛⵊⵍⵖ ⵚⵓⵛⴷⵑⵗⵖ ⵛⵄⵍⵠⵇⵄⵑⵠⵇⵛⵖⵔⵔⵠⵔⵛ.

- ⵊⵋⵏⵛⵛⵑⵔⵠⵔ ⵑⵛⵛⵑⵖⵏⵖⵛⵑⵔⴷⵖⵔⵔⵛ ⵛⵔⵔⵑⵑⵖⵛⴷⵛⵔⴷⵛⴱ ⵊⵍⵍⵔ ⵊⴷⵛⵛⵏⵔⵑⵔⵠⵖⵛⵖⵛⵛⵛ

ᐃᵒᓇᐃᶜ ᑎᑎᕃᐅᓪᓕᐅᕆᓂᖅᑐᑦ ᓇᑎᐅᑦᑕᓚᐸᓗᑎᒃ 3,649 ± 316 (S.E.)-ᖅᓱᐸᑦ ᓂᕐᔪᐊᑦ MX-08-ᒥᑦ. ᖃᐸᓯᐅᓂᖏᑦ MX-08-ᒥᑦ ᐊᒥᓯᖅᓚᐊᖅᓯᒪᔪᑦ ᑕᐃᒪᖅᓄᓂᑦ ᑭᖅᓱᓪᓕᒥᑦ ᐅᓴᐃᖅᒪᑕ.

- ᐅᑭᐅᖅᓱᐸᖅ ᑕᓗᕐᔪᐊᖅᒥᐅᑦ ᐊᒧᐸᒃᓴᐃᕈᑎᓕᐅᕆᓂᖅᑐᑦ ᐃᓯᓕᖅᑲᖅᑎᓵᓇᒧᑦ ᒪᖃᐃᑦᑐᖅᓇᖅᑎᓂᕐᔪᐊᖅᐳᖅ. ᐃᓵᓇᐅᔪᖅ ᐅᒥᖅᒪᑦᑕᐅᔪᖅᓇᖅᑐᑦ, ᐱᖁᐅᓯᒪᓕᖅᑐᑦ, ᐊᓚᓗ ᑐᓱᓯᖃᑲᓵᓂᖅᑐᑦ ᐊᒧᐸᒃᓴᖅᑕᒃᑲᓵᓂᕈᓯᒪᔨᓗᑎᒃ.

ᐅᖃᖅᑎᖅᑎᓂᖅ:

- ᐱᓕᑦᓂᐊᓂᖏᓐᓄ ᖃᑎᐅᔨᓴᐃᓂᐅᔨᑦ ᐊᓚᓗᓗ ᖃᑎᐅᔨᐱᖃᑎᑦᑕᖅᓂᐅᔨᑦ ᑐᓴᖅᑎᑕᐅᖅᑎᑦ ᖃᑭᑎᑕᐅᓂᖅᓴᑦ ᐊᐅᓚᑦᑎᓂᖅᓴᑦ ᐸᖅᓴᕐᓗᒍᑎᓪᓗᑎᑦᓗᑕᑦ ᑲᑎᖅᑎᑦᑕᐅᖅᑎᑦ ᓂᕐᔪᐊᑦ ᐃᒪᓕᒃᑯᓂᖅᑐᑦ ᖃᑎᒌᒍᒪᔪᖅᑐᑦ (KRWB ᐊᖅᑭᒍᒪᔭᖅ ᖃᑎᒪᑎᑦᑎᓂᖅᖢᖅ) ᓯᑎᐱᓇ 2017-ᒥᑦ ᐊᓚᓗᓗ KRWG ᐊᖅᑭᒍᒪᔭᖅ ᖃᑎᒪᑎᑦᑎᓂᖓᓐᓂᑦ ᓯᑎᐱᓇ 2018-ᒥᑦ.

- ᐃᐳᓇ 23, 2019-ᒥᑦ, ᐅᖅᑐᑎ ᐅᓂᐸᑲᖅ ᐊᓚᓗᓗ ᐊᐅᓪᓚᑎᓂᖅᓴᑦ ᐃᒪᓐᓇᖅᑐᑐᐅᔨᑦ ᐅᖃᖅᑐᓗᓴᕆᐅᑕᐅᕆᓂᖅᑐᑦ ᑕᓗᕐᔪᐊᖅᒥᑦ ᐅᒪᔨᓪᓚᐊᓯᒃᑯᓐᓄᑦ ᐊᓚᓗᓗ ᓄᓇᓕᖕᒥᐅᓄᑦ. ᐅᒪᔨᓪᓚᐊᓯᒃᑯᖅᑎᑦ ᐊᒥᓯᖅᓚᐊᖅᑯᓕᐅᕆᓂᖅᑐᑦ ᓯᓐᓇᐅᔨᖅ TAH-ᖅᓱᐸᖅ ᐅᖃᖅᑎᖅᓴᑎᓗ ᒪᖃᖅᐃᑕᐅᖅᑎᑦᑕᖅᓂᖅ 6%-ᒥᒃ (218 ᓂᕐᔪᐊᑦ ᒪᓪᓴᑎᑦ ᓄᑦᑖ ᖃᐸᓯᐅᓂᖏᓐᓄᑦ ᓇᓪᓕᐅᑦᑖᖅᑕᐅᔨᓪᓕ) ᐃᓗᒪᖏᓐᓄᓂᕆᓄᑦ ᐊᒥᓱᓂᖏᓐᓄᑦ ᐅᒥᖏᓪᐃᑦ.

ᐃᓂᓪᐅᓚᓂᖅ:

- ᑕᐃᒪᵒᓇᒧᑦ ᐅᖃᖅᑎᑎᕃᓪᓕᐅᕆᓂᖅᑐᑦ, ᐊᓯᐊᖏᓱᖅᑎᓚᐸᓗᑎᒃ ᐅᓂᐸᑲᒧᑦ ᐃᓪᓕᐅᑎᵒᓂᐊᖅᓗᑎᓗᑎᑦ ᓄᓇᓕᖕᒥᑦ ᐊᓚᓗᓗ ᐱᖅᑯᓯᓯᑯᓂ ᖃᐅᔨᓴᖅᑎᑦᑕᐅᖅᑎᓯᒪᔨᓗᑎᒃᓴᑦ, ᐃᓪᓕᐅᔨᓴᐃᓂᑦ:

    - ᑐᑭᓯᐅᒪᔭᕐᓂᖅᑎᓂᒃᓴᑦ ᐅᒥᖏᓪᓕᐃᑦ ᓅᑦᑕᖅᓂᖏᓐᓄᑦᓄᑦ - ᓂᕐᔪᐊᑦ ᓅᑐᐃᓇᖅᑎᑦᑕᖏᑐᑐᐃᓇᓚᐊᖅᑎᖅᑐᑦ Prince of Wales-ᒥᑦ ᐊᓚᓗᓗ Somerset ᖃᑭᖅᑕᒥᑦ, ᑕᒪᐅᖏᓪᓕᖅᑐᑦ ᐊᓯᖏᓐᓄᖓᓐᓄᑦ.

    - ᐅᖃᖅᐅᑎᓴᐅᓪᓕᐅᕆᓂᖅᑐᑦ ᖃᓄᖅ ᐅᒥᖏᓪᓕᐃᑦ ᑐᓴᐃᓪᓗ ᖃᐸᓯᐅᓂᖏᓐᓄᑕ ᖃᓄᓇᐃᓕᐅᑦᓂᖏᓐᓄᑦ ᐊᓴᓵᖏᓐᓄᑦᑐᓴᒃᑯᓂᖅᑎᑦ ᐱᓕᓓᖏᓐᓄᓂᕆᓄᑦ, ᐊᓚᓗᓗ ᐅᖃᖅᐅᑎᓴᐅᓪᓚᐸᓗᑎᒃ ᑐᓴᐃᓪᓗᑎᑦ ᖃᐸᓯᐅᓂᖏᓐᓄᑦ ᖃᓄᓇᐃᖏᑎᓪᓚᐸᓂᖏᓐᓄᑦ ᐊᓚᓗᓗ ᓄᓇᓕᖕᒥᐅᓄᑦ ᖃᓄᐃᓪᓚᐸᓂᖏᓐᓄᑦ.

- ᐊᖏᑎᓕᓓᖅᑯᓂᑦ, ᑕᓗᕐᔪᐊᑦ ᐃᒪᓐᓇᖅᑐᑐᐃᓇᖅᑯᓂᑦ, ᐊᓚᓗᓗ ᓄᓇᓕᖕᒥᐅᓂᑦ ᐊᖏᓕᖅᑎᑎᑲᓂᖅᓴᐃᓇᖅᑎᓂᖏᓐᓄᑦ ᐊᒥᓱᓂᖏᓐᓄᑦ ᐃᓵᓇᐅᔨᖅ ᐅᒥᖏᓪᓕᐅᕆᓂᖅᑐᑦ, ᐃᒪᓐᓇᖅᑐᑐᐃᓇᖅᑯᓂᑦ ᐃᑲᔪᖅᑐᐃᓐᓇᖏᓐᓂᑦ ᒪᖃᐃᑦᑕᐅᖅᑎᑦᑕᖅᓂᖏᓪᓕᖅᑐᑦ 6%-ᒥᒃ (TAH 218-ᓂᒃ) ᐃᓗᒪᖏᓐᓂᑐᓚᐅᖅᓪᓕᑦ. TAH ᐱᓯᒪᔪᓚᖅᑐᑦ, ᑭᓯᐊᓂ ᐊᒥᓱᓂᖏᓐᓄᑦ ᒪᖃᐃᑦᑕᐅᔨᖅᓴᖅᑐᑦ 7-8%-ᒧᑦ, ᑕᐃᒪᵒᓇ TAH 255-288-ᖅᓱᑦᓗᓂ.

ᐃᒪᓐᓇᖅᑐᑦ

- *ᐊᐅᑎᓪᓗᓈᑦ ᓄᓇᕗᒻᒥ ᓂᕆᔨᑎᓂᒃ ᐊᐅᓪ ᕿᑎᓈᑦ ᐊᒥᓲᖑᓗᐊᕆᐊᖃᖅᑎᑦ TAH 66-ᓂᑦ 275-ᓄᑦ ᐅᒥᖓᒪᐃᑦ ᐅᒥᖓᒪᖓᓄᑦ ᐊᐅᓪ ᕿᑎᓂᒃ MX-08-ᒧᑦ.*

ᐅᒥᖏᒪᐃᑦ ᓇᒧᖏᑐᐅᑲᑕᕐᓂᖏᑦᑐᑦ ᐊᓚᒪᓗ ᐱᑕᕐᑲᐅᕐᓂᖏᑦᑐᑦ, ᐅᒥᖏᒪᖏᓂᒃ
ᐊᐅᓪᓗᓂᒃ MX-08, Boothia ᓄᑭᐊ, ᐊᒻᒪᓗ 2017.

ᐅᓂᒃᑲᖅ ᐅᒥᖕᒪᐃᑦ ᓇᒧᖕᓱᐅᕆᑦᑐᖕᖏᓐᓂᑦ ᐊᒻᒪᓗ ᐱᑕᖅᐅᕆᖕᖏᓐᓂᑦ, ᐅᒥᖕᒪᖕᓂᒃ ᐊᐅᓚᕋᑎᓂᖅ MX-08, ᐋᒋᓗ 2017

ᓇᐃᓴᖅᓯᒪᔪᖅ

ᒥᒃᓇ ᓇᐃᑦᑐᖅ ᑎᑎᖅᑲᖅ ᑐᑭᓯᐅᒪᔭᕐᑎᑦ ᓇᐃᓴᖅᓯᒪᔪᑦ ᐅᓂᒃᑲᒋᑦᑐᑦ ᐊᑎᖕᒪᑦ: “ᐅᓂᒃᑲᖅ ᐅᒥᖕᒪᐃᑦ ᓇᒧᖕᓱᐅᕆᑦᑐᖕᖏᓐᓂᑦ ᐊᒻᒪᓗ ᐱᑕᖅᐅᕆᖕᖏᓐᓂᑦ, ᐅᒥᖕᒪᖕᓂᒃ ᐊᐅᓚᕋᑎᓂᖅ MX-08, ᐋᒋᓗ 2017.”

ᒥᒃᓇ ᐅᓂᒃᑲᖅ ᑎᑎᖅᑲᐅᔪᖅ ᖃᐅᔨᓴᐃᓂᕐᒥᒃ ᐋᖅᑭᒃᓯᔪᖅ, ᐃᓕᑕᖅᓯᔪᓄᓂ ᐊᑐᖅᑕᐅᔪᓂᒃ ᖃᐅᔨᓴᐃᓂᕐᒪᑦ, ᐅᖃᐅᓯᖅᑲᖅᓵᓂ ᖃᓄᐃᖏᓂᐅᔪᓂᒃ, ᑐᓂᓯᔪᓄᓐᓄ ᓯᕗᓂᖏᓐᓄᑦᑐᖅᑐᓂᒃ ᖃᐅᔨᓴᐅᕈᖅᑕᕐᓂᕐᒪᑦ ᐊᐅᓚᕋᑎᓂᖅᑐᓂᒃ ᐃᓚᖓᔾᔪᑎᒃᑲᓂᕐᒪᑦ ᐊᐅᓚᕋᑎᓂᕐᒪᑕ ᐃᓚᖓᔾᔪᑎᒃᑲᓐᓂᒃ.

ᓄᓇᕗᒻᒥ ᒪᓕᒐᓱᑦ ᐃᓂᖅᑲᖅᑐᑦ ᐊᐅᓚᕋᑎᓂᕐᒪᑦ ᐅᒥᖕᒪᖕᓂᒃ ᓄᓇᕗᒻᒥᑦ ᐊᒻᒪᓗ ᐅᖅᑲᖅᑎᖅᑲᖅᓂᓱᑦ ᐊᐅᓚᕋᑎᖅᑎᑎᕐᓯᒪᓂᒃ, ᐱᖓᓱᐅᑎᖅᑲᖅᑐᓂᒃ ᖃᐅᔨᓴᐃᓂᕐᒪᑦ ᐊᒻᒪᓗ ᖃᐅᔨᓴᐅᕈᖅᑕᕐᓂᕐᒪᑦ (ᖃᐅᔨᓴᐅᓂᖕᓂᓂᒃ ᐅᓱᐃᐅᓂ), ᐊᑐᖅᓂᐊᖅᑕᐅᑦ ᐊᐅᓚᕋᑎᐅᓪᓗᑎᓂᕐᒪᑦ. ᐅᓂᒃᑲᖅ ᑐᓂᓯᔪᖅ ᓯᓇᑐᐃᓐᓇᓚᓂᕐᒪᑦ ᑐᑭᓯᐅᒪᔭᕐᑎᑐᓂᒃ ᐊᒻᒪᓗ ᐃᓕᖅᑲᖅᑐᑦᑕᐅᔪᓂᒃ ᐃᓐᓇᕐᓂᐊᕐᓂᒃ ᐊᐅᓚᕋᑎᓂᐊᖅᑕᐅᑎᒃ ᐅᒥᖕᒪᖕᓂᒃ ᑕᓪᓕᐅᕐᑕᒥᑦ ᐊᐅᓚᕋᑎᓂᐅᔪᓂᒥᑦ.

ᓇᐃᓴᖅᓯᒪᔪᖅ ᐊᑐᖅᑐᖅ ᑐᑭᓯᐅᒪᔭᕐᑎᑐᓂᒃ ᖃᓄᐃᖏᓐᓄᑦᑐᖅᑐᓂᒃ ᐅᓂᒃᑲᖅ ᐅᒥᖕᒪᐃᑦ Boothia ᓄᓕᐊᓂᕐ ᐱᓕᕆᐊᖑᓚᐅᖅᑐᑦ ᐋᒋᓗ 2017-ᒥᑦ. ᓯᓚᑦᓚᐅᓚᐅᖅᑐᑦ ᑎᑎᕆᖅᑕᐅᓯᒪᔪᑦ ᖃᓄᐃᖏᓐᓄᑦᑐᖅᑐᑦ ᐊᔾᔨᖓ ᐅᓂᒃᑲᒥ ᐃᓚᐅᔪᖅ ᑕᑯᑎᕆᔨᐅᖅᑕᐅᕆᕐᓂᐊᓂ.

ᑐᑭᓯᐅᒪᔭᕐᓂ

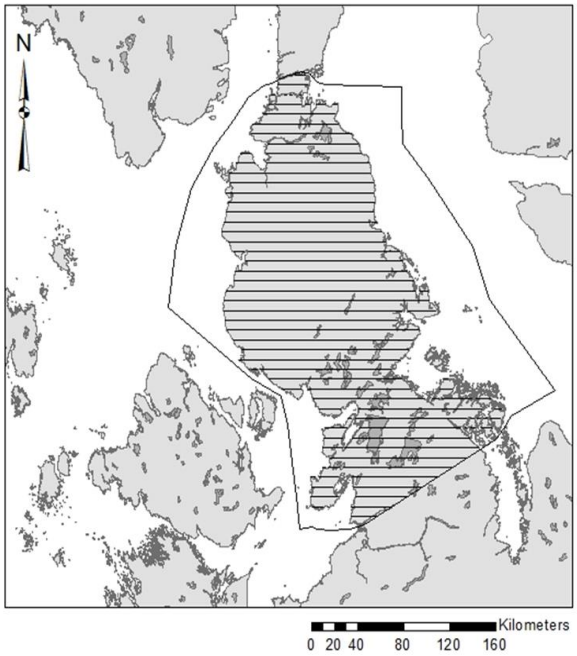
Boothia ᓄᖅᐊ ᐅᐸᑐᑕᐅᔪᖅ ᐅᒥᖓᒪᐃᑦ ᑲᑎᖅᓂᕆᔨᖅᑕᕐᓂᖅᓴᓕᓂᕐᓂᖓᓂᓄᑦ ᐃᓕᑐᖅᓴᖓᓕᒫᓂᕐᓂᖓᓂᓄᑦ 1985-ᒥᑦ, Boothia ᓄᖅᐊ ᖃᑲᐅᓚᕿᐅᓕᕆᐅᕙᖅᑐᖅ ᐅᒥᖓᒪᑕᖃᖅᓴᓕᐅᕙᓪᓕᐅᖅᑐᖅᓯᒪᓄᑦ. ᐅᑭᐅᑦ ᖃᐅᓪᓕᑦ ᑭᓯᐊᓂᖅ, 61 ᐅᒥᖓᒪᖕᓂᒃ ᑕᐅᖏᐃᐅᓕᕆᐅᕙᖅᑐᖅ ᑲᑎᕐᓂᖅᒥᑦ ᐋᓚᐃᓂᖃᖅᑎᓪᓗᑎᑦ ᓇᓕᐅᑦᑕᖃᖅᑕᐅᔪᑦ 544-ᓂᒃ ᓂᕿᐱᑕᖃᖅᓱᓂᒃ, ᐊᖏᓪᓗ ᐊᒥᓲᖕᖅᓴᓪᓕᐊᑭᐅᓇᖅᑐᓂᒃᓴᖅᓱᑎᒃ. ᓯᓪᓕᐅᕐᓂ ᖃᓄᐃᓇᓂᖓᓐᓄ Boothia ᓄᖅᐊᓂᒃ ᐋᓪᓚᑲᓇᖅᑐᖅ ᐅᒥᖓᒪᓂᒃ ᐱᓇᓱᖅᓴᓪᓕᐊᑎᑦᑎᓪᓗᒍ.

ᑕᓗᕕᐊᖅ ᓄᓇᓕᑐᐊᖓᔪᖅ ᒪᖃᐃᓴᖃᑦᑕᖅᑐᑦ ᑕᒡᔭᐅᓂᑐᖅ. ᑕᓗᕕᐊᖏᑦ ᒪᖃᐃᑎᑦ ᐅᖃᖅᖢᓪᓚᕆᔨᑦ ᐅᒥᖓᒪᓂᒃ ᐊᒥᓱᓂᖅᓴᐅᓕᖅᑐᖅ ᑕᑯᖃᑦᑕᖅᓱᓂᒃ. ᐃᓱᒪᓕᐅᑎᖃᖅᑐᑦ ᐅᒥᖓᒪᐃᑦ ᐱᐅᔭᕐᑕᑐᑯᑦ ᐊᐅᑐᖅᓴᖃᖅᓯᐊᓂᔭᕐᓂᖓᓂᓄᑦ ᑐᑦᓄᖅᑕᐅᔪᑦ Boothia ᓄᖅᐊᓂᑦ. ᐃᓄᐃᑦ ᐱᖅᑯᓱᖅᑯᑦ ᖃᐅᔨᓴᖅᓱᑦ ᐃᓂᐅᔪᒪᑦ ᐅᒥᖓᒪᐃ ᐃᓇᖅᑎᓱᒪᑦ ᑐᑦᓄᖅᑐᖅ ᐃᓂᐅᑦᑕᕐᓂᖓᓂᓄᑦ.

ᐊᑐᖅᓯᒪᓂᒃ ᓄᓇᓕᒥᐊᖅᑐᖅ ᐃᓂᖅᑐᖃᑦᑕᖅᑐᑦ ᑕᑯᓀᓂᑐᓄᑦ. ᐊᑐᖅᓯᒪᓂᖅᑎᑦ ᐃᓂᖅᑐᖃᑦᑕᖅᑐᑦ ᐊᑐᖅᓯᒪᓂᖅᑎᑦ ᐃᓂᖅᑐᖃᑦᑕᖅᑐᑦ ᑕᑯᖃᑦᑕᖅᓱᓂᒃ. ᐃᓱᒪᓕᐅᑎᖃᖅᑐᑦ ᐅᒥᖓᒪᐃᑦ ᐱᐅᓂᕐᑐᐃᓐᓇᖅᑐᓂᒃ ᐊᑐᖅᑐᖃᑦᑕᖅᑐᑦ ᑐᑦᓄᖅᑕᐅᔪᑦ Boothia ᓄᖅᐊᓂᑦ. ᐃᓄᐃᑦ ᐱᖅᑯᓱᖅᑯᑦ ᖃᐅᔨᓴᖅᓱᑦ ᐃᓂᐅᑦᑕᕐᓂᖓᓂᓄᑦ ᐅᒥᖓᒪᐃ ᐃᓇᖅᑎᓱᒪᑦ ᑐᑦᓄᖅᑐᖅ ᐃᓂᐅᑦᑕᕐᓂᖓᓂᓄᑦ ᑕᑯᓀᓂᑐᓄᑦ.

ᐊᑐᖅᓱᓂᒃ ᓄᓇᓕᒥᒃᑐᑦ, ᖃᑭᒥᖅᑭᓂᖅᑕᕐᑐᖃᖅᑐᖅᓯᒪᓂᖅᑐᖅᓴᖅᓱᑎᒃ ᑕᐃᒪᐅᔪᓂᖅ ᐅᒥᖓᒪᓕᖅᓴᐅᓕᖅᑐᖅ (TAH) 66-ᖅᓯᔪᓂᒃ, ᐊᓐᓇᐅᔪᖅ ᐊᐅᑕᓂᑎᓪᓗᑎᑦ ᑐᖏᖅᓯᔪᓂᑦ. ᑕᓗᕕᐊᖏᑦ ᒪᖃᐃᑎᑦ ᐊᒥᓲᖕᖅᓴᓪᓚᖏᑐᔪᑦ ᐅᒥᖓᒪᑕᖃᖅᑕᕐᑐᖅᓴᖅᓱᑎᒃ ᐊᒥᓲᖕᖅᓴᑐᐊᓕᖅᑕᐅᓚᑦ ᐊᖏᓪᓗ ᑐᑦᓄᖅᑕᐅᔪᑦ ᐃᓂᖅᑐᖃᑦᑕᖅᑐᑦ ᐊᓯᐅᓕᑕᒪᓕᓚᑦ. ᖃᐅᔨᓴᖅᓱᐃᓐᓇᖅᑐᓂᒃ ᐊᓯᐅᓚᑦ ᐅᒥᖓᒪᐃᑦ ᐱᐅᓂᕐᑐ

ᐊᖑᑎᒋᔭᖅᑎᓂᐅᕐᓗᒍ ᖃᐅᔨᓴᐃᖏᕐᕕᖏᓐᓂᖅ, ᓂᓚᐅᑎᕐᐅᕐᓴᑦ ᐅᒥᖓᒪᐃᑦ ᓇᒧᖓᐅᖃᑎᒌᓂᖓᑎᑎᑦ ᐱᕐᐅᓪᓚᐅᖅᑐᑦ ᖃᐅᔨᓴᐃᖏᕐᕕᐊᓂᐅᑐᓂᑦ, ᒪᖃᐅᑦᑏᑦ ᖃᐅᔨᓯᒪᔪᖑᑎᓂᑦ, ᐊᓚᒪᓗ ᐃᓄᐃᑦ ᐱᖃᑯᓯᑯᓂᑦ. ᐃᓄᐃᑦ ᖃᐅᔨᒪᓂᖖᑎᒍᑦ, ᐅᒥᖖᒪᐃᑦ ᐊᒥᓯᖓᖃᓯᒪᔪᑦ ᐊᓚᒪᓗ ᓇᓂᑐᐃᓪᓕᓚᖅᑐᑦ Boothia ᓄᙳᐊᓂᑦ. ᒪᑉᓴᒍ ᐊᓯᐊᖏᓴᓂᐅᔪᖅᑐᖅ, ᑕᒪᐃᓐᓂᑦ ᐊᐅᑕᕐᑎᓂᖅ ᐃᓕᓴᕐᐅᑕᖅᑐᖅ 20%-ᖓᔪᒥᑦ, 8 ᑭᓕᒥᑕᓂᒃ ᐃᓂᕐᑕᖅᓴᑎᒃ ᐊᑯᐃᓂᖖᑎᓂᑦ ᑲᑎᒪᔨᑦ ᐊᒥᑕᖅᑐᖅᓴᑎᒃ, ᐊᓯᐊᓂᒃ ᐱᑕᖅᖏᑦᓴᒍ ᐊᔪᐅᖖᑕᖓᓂᕐᒪᑦ (ᐊᔪᓄ 1).



ᐊᔪᓄ 1: ᑲᑎᒪᔨᓂᑦ ᑎᖏᓕᓄᑦ ᐊᑎᓯ 2017-ᒥᑦ, ᑕᑯᐱᐊᔨᑦ 20% ᖃᐅᔨᒪᖅᑕᐅᔨᑦ, ᐅᒥᖓᓐᓂᒃ ᐊᐅᑕᕐᑎᓂᖅ MX-08-ᒥᑦ.

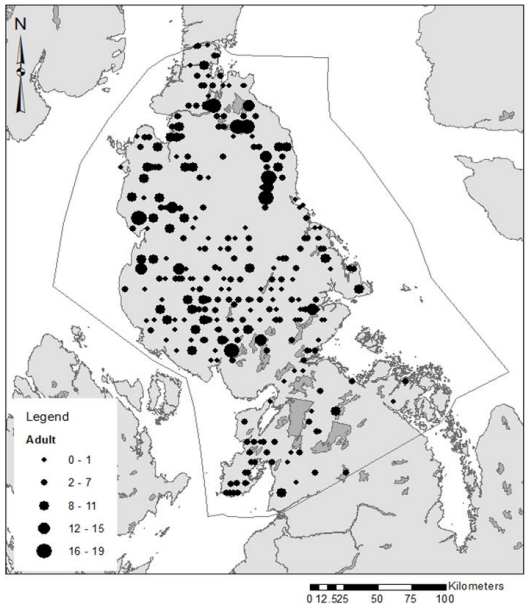
ᑎᖏᓕᓄᖅ ᐊᑐᖅᑕᐅᔪᖅ

ᐊᑦᖁᐅᒪᔪᑦ ᑲᑎᒪᔨᓂᕐᓂᑦ ᖃᐅᔨᒪᓕᐅᓂᑦ ᑎᖏᓕᓄᑦ ᐃᓄᐃᑦ ᐊᑎᓯᓄᓕᖅᑐᑦ ᐃᓚᖓᑕᖓᓐᓄᑦᑐᑦ; ᐃᓚᖓᑎᒪᔨᓕᑎᒋᐊᖏᖅ. ᑲᑎᒪᔨᓂᖓ ᖃᐅᔨᒪᖅᑕᐅᓕᐅᖅᑐᑦ 160 ᑭᓕᒥᑕ/ᐃᑲᖅᓕᒥᑦ ᖃᔪᑎᖓᓂᖅᖃᖅᓴᓂ 121 ᒋᑦᓯᓗᐊᑦ ᓄᓇᐅᑦ ᖃᔪᓕᓂᑦ. ᑲᑎᒪᔨᓂᖓᓂᑦ ᐃᓕᐅᓕᐅᖅᑐᑦ 800 ᒋᑕᐃᑦ ᐊᑐᓂ ᓴᓂᖓᑐᓂᑦ ᑎᖖᒥᓵᑦ. ᑕᐅᑐᒃᑐᑦ ᑕᒪᐃᖏᓂᑦ ᓴᓂᖓᓂᑦ ᑎᖖᒥᓵᑦ ᖃᑭᓂᕐᒪᖅᓯᒪᔪᖅᓄᑦᓯᐊᕐᒪᐃᖅᑐᖅᑐᑦ, ᓇᓂᓵᓗᑎᒃ, ᐊᓚᒪᓗ ᐃᓴᓐᓇᐅᑦ ᐅᒥᓴᑦ, ᓇᓂᐃᑦ, ᐊᓚᒪᓗ ᐊᒪᑭᐃᑦ ᑎᑎᕋᐅᓵᑐᖅᑐᖅᑐᑦ.

ᖃᓄᐃᖖᑎᓂᑦ

ᓇᓗᖖᑦᑐᓵᖖᑎᓂᑦ

ᐃᓂᓴᓂᖃᑎᒌᓂᑦ ᑕᑐᕐᔪᐊᖏᒥᑦ ᐋᒋᓵ 7-ᒥᑦ 12-ᓗᑦ, 2017-ᒥᑦ. ᐃᓂᓴᑎᑦᑐᒋᑦ, 170 ᑲᑎᒪᓕᐃᑦ ᐅᒥᖖᒪᐃᑦ ᑕᑯᔪᐅᓕᐅᖅᑐᑦ, ᑲᑎᒪᔨᓂᖏᒥᐅᖖᑦᑐᓂ. ᐊᒥᓵᓂᖃᖅᓴᐃᑦ ᐅᒥᖖᒪᐃᑦ, 16-19-ᖓᔪᓕᐅᖅᑐᖅᑐᑦ ᐃᓱᓇᐃᑦ, Cape Farrand-ᒥ ᑐᑐᐊᖓᔪᓕᐅᖅᑐᖅᓴᖅ Abernethy Bay-ᓗᑦ 40 ᑭᓕᒥᑕᓂᒃ ᐅᖖᓚᔪᑎᑎᕐᑐᑎᒃ ᓚᐱᒃᖖᓚᑦ. ᓚᐅᖅᓯᖃᖅᐸᒥᑦ ᐅᒥᖖᒪᐃᑦ ᑎᑎᕋᐅᓚᔪᓕᐅᖅᑐᑦ ᓂᒋᐊᓂᑦ Cape Cambridge-ᒥᑦ, ᖃᓄᑦᑐᓴᖅ Acland Point-ᓗᑦ ᓂᒋᖅᐸᓚᓕᖓᔪᒥᑦ ᑲᑎᒪᓴᑦᑐᒍᑎᒃ (ᐊᔪᓄ 2).



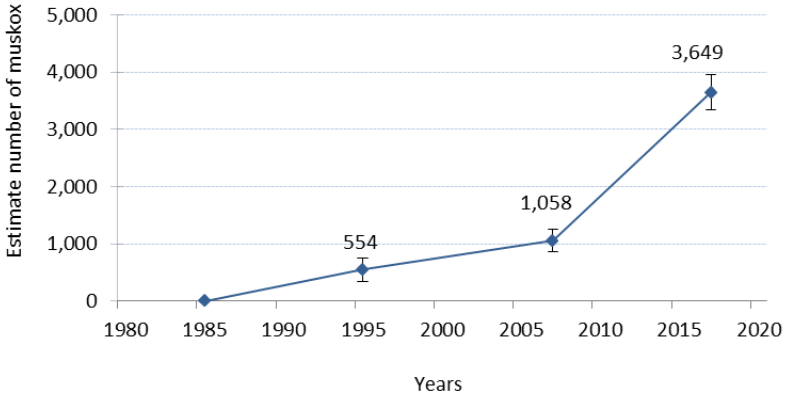
ᐊᒃᓴ 2: ᐅᒥᖏᒪᐃᑦ ᓇᒪᖏᑐᒪᓂᖅᑎᑦ, ᑲᑎᕐᓯᑎᓂᒥᑦ ᐊᓚᒪᓗ ᑲᑎᕐᓯᑎᓂᒥᐅᖏᖅᑎᑦᑐᖅ, ᐊᐅᓪᓚᑎᓂᖅ MX-08-ᒥᑦ ᐅᓴᐃᓂᒥᑦ, ᑲᑎᖏᒪᕐᒪ ᓂᕐᔪᑦᑎᑦ ᐃᓂᓪᑕᐅᖅᑕᐅᓪᓚᐅᖅᑐᑦ ᑲᑎᖏᒪᓪᓗᑎᑲ 0-1, 2-7, 8-11, 12-15-ᓄᑦ ᐊᓚᒪᓗ 16-19-ᓄᑦ ᓂᕐᔪᑦᑎᑦ.

ᑲᑎᒪᔨᑦ ᖃᓄᐃᖏᓂᖅᑎᑦ ᐊᓚᒪᓗ ᓇᓪᓕᐅᑦᑕᐅᓪᓚᐅᖅᑐᑦ

ᐊᒥᓱᓂᖅᓴᐃᑦ ᑲᑎᒪᔨᑦ ᑕᑯᐅᑦᑕᐅᓪᓚᐅᖅᑐᑦ ᐊᒥᓱᓕᕆᑎᑦᑕᐅᓪᓚᐅᖅᑐᑦ 2-ᓂᒃ 7-ᓄᑦ ᐃᓐᓇᐃᑦ. ᐊᒥᓱᓂᑦ ᐊᑕᐅᓱᓕᖅᓯᒪᔪ ᐃᓴᓴᐅᑎ ᖃᓄᐊᓂᓂᖅᑎᑦ ᐃᓐᓇᐃᑦ (+1 ᐅᑭᐅᓪᓕᑦ ᐅᓕᒐᓄᑦᓗ) ᑲᑎᒪᔨᓂᑦ 5 ± 4.45 (S.D.)-ᓕᒫᓪᓚᕐᓂᖅᑐᑦ. ᐃᓕᓴᐃᑦ ᐃᓴᓴᓂᖅᑲᖅᑎᑦᓗᑎᑦ 702 ᐃᓐᓇᐃᑦ ᐅᒥᓕᐃᑦ ᐃᓴᓴᖅᑕᐅᓪᓚᐅᖅᑐᑦ ᑲᑎᕐᓯᑎᓂᒥᑦ. ᑕᐃᒪᓐᓄᑦ, ᐅᒥᓕᐃᑦ ᐊᒥᓱᑎᑎᓂᓕᑦ ᖃᑲᐅᔪᓕᐃᖅᑲᐅᔪᒥᑦ ᐊᑦᑎᓐᓂᓂᖅᑕᐅᓪᓚᐅᖅᑐᑦ 0.084 ᐅᒥᓕᐃᑦ/ᑭᒃᒥᑕᐃᓕᒥᑦ. ᐃᓕᐅᑦᑕᖅᑕᐅᔪᑦ ᖃᑦᔪᐅᓂᓐᓄᓄᑦ ᐅᒥᓕᓐᓂᑲᖅ ᐊᐅᓪᓚᑎᓂᖅᑲᖅ MX-08-ᒥᑦ 3,649 ± 316 (S.E.)-ᓕᒫᔪᑦ (ᐊᒃᓴ 3). ᑕᑯᑲᓴᐅᑎᑦᑎᔪᖅ ᑭᓇᓪᑕᖅᑕᒥᑦ ᐃᓕᓴᐅᑦᑕᖅᑕᐅᓕᒡᒪ ᐊᒥᓱᓇᓗᐊᕐᔪᐊᖅᓯᒪᔪᑦ. ᖃᖅᑎᐅᓴᐅᔪᑦ ᒪᖃᖅᐃᑎᓄᑦ ᐃᑲᔪᖅᑕᐅᔪᑦ ᖃᖅᑎᐅᓴᐅᔪᓄᑦ ᐃᓴᓴᐅᑎᑦᑎᓂᖅᑐᑦ.

ᐃᒪᓐᓇᖅᑐᑎᑦ

ᐊᑐᖅᓯᒪᓂᑦ ᖃᖅᓄᐃᓐᓇᓂᐅᔪᑦ ᐊᓚᒪᓗ ᐅᖃᖅᑎᑲᖅᑎᖅᓴᓂᑦ ᑕᓗᖕᔪᐊᕐᒥᐅᓂᑦ ᐊᓚᒪᓗ ᑕᓗᖕᔪᐊᕐᒥᑦ ᐅᓪᓚᔪᓕᐅᑦᑕᐅᓂᓕᒫᖅ ᐃᒪᔪᓯᓚᐊᓇᖅᑐᑦ, ᐊᑐᖅᓯᒪᓂᓚᐅᖅᑐᑦ ᐃᒪᓐᓇᖅᑐᑎᑦ: ᓄᓇᕗᒻᒥ ᓂᕐᔪᐊᓂᓂ ᐊᐅᓚᑦᑎᐅᖅᑐᑦ ᐊᒥᓱᓇᓗᐊᖅᓯᒪᓕᖅᑐᖅ TAH-ᒥᑦ 66-ᓂᑦ 275-ᓄᑦ ᐅᒥᓕᑲᑕᐅᔪᓐᓇᖅᑐᓄᑦ ᐅᒥᓕᓐᓂᓂᖅ ᐊᐅᓚᑦᑎᓂᖅᑲᖅ MX-08-ᒧᑦ.



ᐊᒃᓴ 3: ᐅᒥᓕᐃᑦ ᖃᑦᔪᐅᓂᓂᖅᑎᑦ ᐃᓚᐅᑦᑕᖅᑕᐅᔪᑦ MX-08-ᒥᑦ ᐅᔨᓗᐃᑦ ᐊᓂᒍᖅᑦᓚᑕᐊᔪᓂᑦ, ᐃᓚᐅᑦᑕᖅᑕᐅᔪᑦ ᑎᓕᒥᓱᓄᑦ ᐃᓕᓴᐃᓂᓂᑦ 1985-ᒥᑦ 2017-ᒧᑦ

MUSKOX (*Ovibos moschatus*) DISTRIBUTION AND ABUNDANCE, MUSKOX MANAGEMENT UNITS MX-08, BOOTHIA PENINSULA, AUGUST 2017.

Lisa-Marie Leclerc[1]

Version: February 2019

[1]Wildlife Biologist Kitikmeot Region, Department of Environment

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NUNAVUT DEPARTMENT OF ENVIRONMENT

WILDLIFE RESEARCH SECTION

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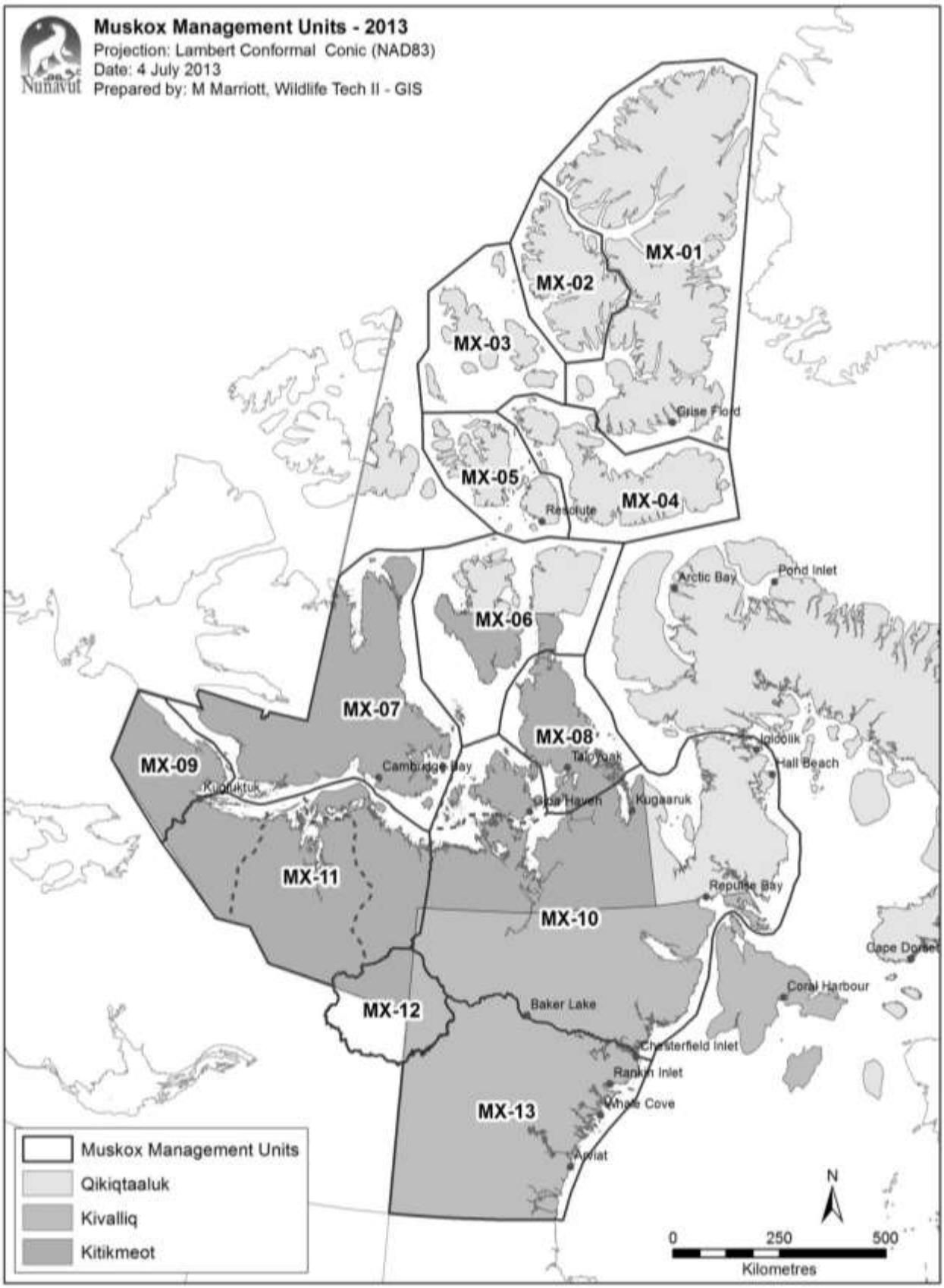
Muskox Management Units - 2013
Projection: Lambert Conformal Conic (NAD83)
Date: 4 July 2013
Prepared by: M Marriott, Wildlife Tech II - GIS
Nunavut
MX-01
MX-02
MX-03
MX-04
MX-05
MX-06
MX-07
MX-08
MX-09
MX-10
MX-11
MX-12
MX-13
Grise Fiord
Resolute
Arctic Bay
Pond Inlet
Cambridge Bay
Kugaaruk
Repulse Bay
Cape Dorset
Coral Harbour
Baker Lake
Chesterfield Inlet
Rankin Inlet
Whale Cove
Arviat
Hall Beach
Muskox Management Units
Qikiqtaaluk
Kivalliq
Kitikmeot
N
0
250
500
Kilometres

Executive Summary

A systematic strip transect survey of the Boothia Peninsula was undertaken in summer 2017 to determine the abundance and distribution of muskox. The survey took place from August 07 to August 12. A total of 8,317.71 km^2 were flown, representing 20% coverage of the total study area (43,238 km^2). During the survey, 702 adult muskoxen were recorded on transect resulting in a population estimate of 3,649 ± 316 (S.E.). Calves represented 14% of the adult muskox seen and the average number of adults per group was small, 5 ± 4.45 (S.D.) The muskox density was of 0.084 muskox / km^2 in the management unit. This is an increase of muskoxen in MX-08 from what have been estimated previously, and it is consistent with the reported local knowledge. Thus, an increase in the current harvest rate could be supported by this current population estimate, as well as continuing the monitoring, harvest reports, and health monitoring program. A survey cycle of 5 years is advisable for this Muskox Management Unit or sooner if traditional knowledge indicates a significant change in the population trend, so harvest rate could be review.

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Introduction

The mechanism driving muskox population dynamics is not well understood. In the early part of the 1900s, hunting pressure drove muskox numbers close to extinction and reduced the distribution to a few limited pockets in the Canadian Arctic (Spencer 1976; Gunn 1984). To stimulate recovery, the existing muskox populations in Nunavut were managed to foster a continued colonization of the historic range (Gunn, 1983). While muskox are currently re-colonizing historical habitat in the central and eastern Canadian Arctic, most muskox populations have been increasing in the last few decades, inhibiting the monitoring of long-term population variation. Therefore, there is limited information available to determine how muskox populations naturally cycle.

The Boothia Peninsula is an example of a location where muskox is re-colonizing their historical range. Previous surveys on the Boothia Peninsula were conducted in 1985, 1995, and 2006. In 1985, the Boothia Peninsula was known to be devoid of muskox (Spencer Bay HTO pers. Comm; Gunn and Ashevak 1990). A decade later, 61 muskoxen were seen on transect providing an estimate of 554 ± 205 (SE) animals (Gunn and Dragon, 1998). According to hunter observations, muskox numbers around Taloyoak have been increasing since 1995. During the latest population survey, in 2006, muskox abundance for Boothia Peninsula was estimated at 1,100 ± 253 animals from the 562 adult muskoxen seen on transect (Dumond, 2007). Based on the location of sightings between the 1995 and 2006 surveys and local knowledge, muskox appear to occupy the Boothia Peninsula from Somerset Island, north of Amittaryouak Lake, moving southward reaching a southern limit at Cape Cambridge (Gunn and Dragon, 1998; Dumond, 2007). In 2006, areas of higher muskox density were found in the vicinity of Murchison Promontory and Pasley Bay. Thus, the environmental conditions on the Boothia Peninsula seem to be optimal to promote muskox population growth.

Taloyoak Hunters have commented on the higher numbers of muskoxen sightings. They fear that muskox will start impacting negatively the caribou calving grounds on the Boothia Peninsula. It is part of Inuit traditional knowledge that muskoxen displace caribou from their habitat. Muskox feeding pits or the destruction they cause to ground cover might prevent caribou from feeding in the area or the strong muskox musk might deter caribou. However, muskox and caribou have been co-habiting in the Arctic for thousands of years, where their ranges overlap temporally and spatially. Even today, there is no clear scientific evidence determining an negative impacts related to the muskox-caribou relationship, as this inter-species relationship is difficult to isolate from confounded variables in the wild. Thus,

traditional knowledge might be a more powerful tool to understand in inter-species relationship.

In 1995, there was no muskox quota in the area that was previously part of MX-09 due to the very low number of muskoxen on the Boothia Peninsula. However, a quota of 20 tags for Prince of Wales Island and 12 tags for Somerset Island were assigned to these two harvest zones (Gunn and Dragon 1998). After the 2006 population survey was completed, a Total Allowable Harvest (TAH) of 20 was set. Assuming that this quota was filled on a yearly basis, a harvesting rate of 6% would have led to a slow decline, as the harvesting rate will be higher that the population yield. (Tener, 1965). Despite this risky management approach taken with limited knowledge on the population demographics, recent local knowledge has indicated that the muskox numbers have still continued to increase.

Based on local knowledge, there is a need to re-evaluate the existing TAH relative to the management goal. Taloyoak hunters are requesting an increase in harvesting opportunities to keep the muskox population relatively low so they can preserve the caribou calving grounds. A reassessment of the muskox population in MX-08 is necessary to revisit the TAH. It is also important to make sure that there is enough incentive in place to reach the harvesting rate in order to achieve the goal of keeping the muskox population relatively low. Thus, this project aims to first provide an update of the current muskox population in the muskox management unit MX08. Consistent with other muskox surveys, the Nunavut wide monitoring approach will be used. This scientific information will be provided and paired with traditional knowledge to review existing management strategies and promote a sustainable harvest of muskox for future generations of Inuit allowing for the co-habitation of caribou and muskox on the Boothia Peninsula.

Objectives

This project aims to address the concerns and requests of Inuit hunters, as well as to provide up to date scientific information for management purposes. Therefore, the main objectives of this study are:

1. Determine the estimated number of muskox;
2. Determine muskox distribution and density;
3. Determine calf crop and group size.

By doing so, it will be possible to have better information on current muskox abundance and distribution in the muskox management unit MX-08. Information on group structure, calf

production, group size, and density is essential to gain insight on the relation between these variables and population dynamics.

Materials and Methods

Study Area

The study area is the muskox management unit MX-08, which includes the Boothia Peninsula and a portion of the mainland. The area lies between M'Clintock Channel to the west and the Gulf of Boothia to the east, and is separated from MX-06 to the north by Bellot Straight and the southern boundary is shared with muskox management unit MX-11. The Boothia Peninsula is the northernmost extension of the Canadian mainland and the North American Continent. The area is rich in topography with plains, lowlands, plateaus, and rolling bedrock hills (Dyke, 1984).

This management unit is part of the Northern Arctic Ecozone, which has two Northern Arctic Ecoregions. The southeast and the north part are characterized by Boothia Peninsula Plateau and a small portion of the southwest by the Victoria Island lowlands (Environment Canada, 1995). Due to the spatial heterogeneity of the area, the Arctic tundra vegetation cover is influenced by the soil moisture, nutrient availability, snow cover, wind exposure, and microclimate differences defining dwarf-shrub health or moist to wet sedge meadows (Laidler *et al.,* 2008).

Vegetation covers in the Victoria Island Lowlands are dominated by *Saxifraga oppositifolia, Dryas integrifolia*, and *Sali*x spp., and the wet areas are characterized by sedges, cottongrass, saxifrage, and moss (Walker, 2000; Environment, 1995). Remaining upland areas are part of the Boothia Peninsula plateau, which have a mid-arctic eco-climate. In the upland the vegetation is discontinuous, and dominated with tundra species (Environment, 1995). Vascular plants are found in bedrock cracks and depressions where it is well irrigated by runoff and protected from winds (Walker, 2000).

Survey Area

Prior to survey, no reconnaissance survey was undertaken to maximize the coverage area investigated. Instead, anticipated muskox distribution patterns were obtained from past ground surveys, hunter observations, and Inuit Traditional Knowledge/*Inuit Qaujimajatuqangit* (IQ). According to *IQ,* muskox has increased in numbers and they are now uniformly distributed over the entire Boothia Peninsula with no specific aggregation. Based on this change in distribution,

the whole management unit MX-08 was surveyed at 20% coverage with no strata of different effort allocation (Figure 1).

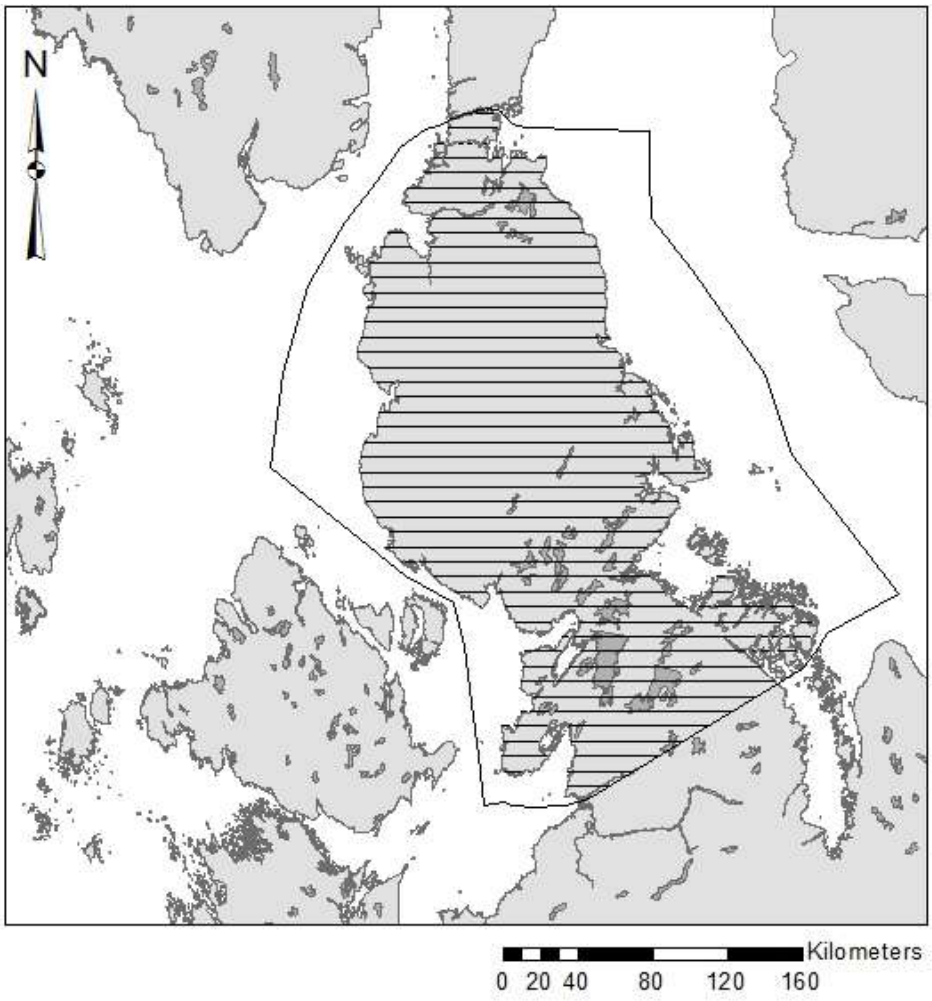

Figure 1: Transect lines representing 20% coverage of the muskox management unit MX-08.

To increase the precision of the survey areas, ESRI'S ArcGIS software with an adapted survey design tool was used to randomly plot the transect lines until the desired percentage of coverage was achieved. The tool allows the user to determine the precise number of transects and the distance between each transect line required in function of the transect strip width and the total area of the management unit. Orientation of the transect lines within the stratum was determined in function to have the most homogeneous and shorter transect line length under the assumption that muskox are randomly and uniformly distributed on the landscape (Figure 1).

Table 1, below, summarizes the total area, the percentage of cover, the total number of kms of transects of different length, the number of lines, the resulting distance between each transect line and the orientation of the transect lines. In sum, the management unit, MX-08, of 43,238 km^2 was surveyed with a total of 5,198 km of transect lines, which represented 46 transect lines of different length at a spacing of 8 km (Table 1).

Table 1 Characteristic of the study area and the transect lines per stratum in the Management Unit MX-08.

Stratum	Total area (km^2)	Percentages (%)	Total transect lines (km)	Number of lines	Distance between transect line (km)	Orientation
1	43,238	20	5,198	46	8	East-West
MX-08	43,238	----	5,198	46	----	----

Aircraft configuration

A systematic transects line survey was flown with a fixed-wing single engine turbine aircraft, a grand caravan. The transect lines were surveyed at a speed of 160 km/hr and the survey altitude of about 152 meters, which was mostly maintained following the relief of the study area using a radar altimeter. The pilot responsibilities were to monitor this air speed and altitude while following the pre-programmed transect on a Geographic positioning system (GPS). The strip transect was 800 meters on each side of the aircraft, for a total transect width of 1.6 kilometers. The pre-determined transect width of 400 meters was set on each wing based on calculation using the formula of Norton-Griffiths (1978) and others (Gunn and Patterson 2000; Howard 2011).

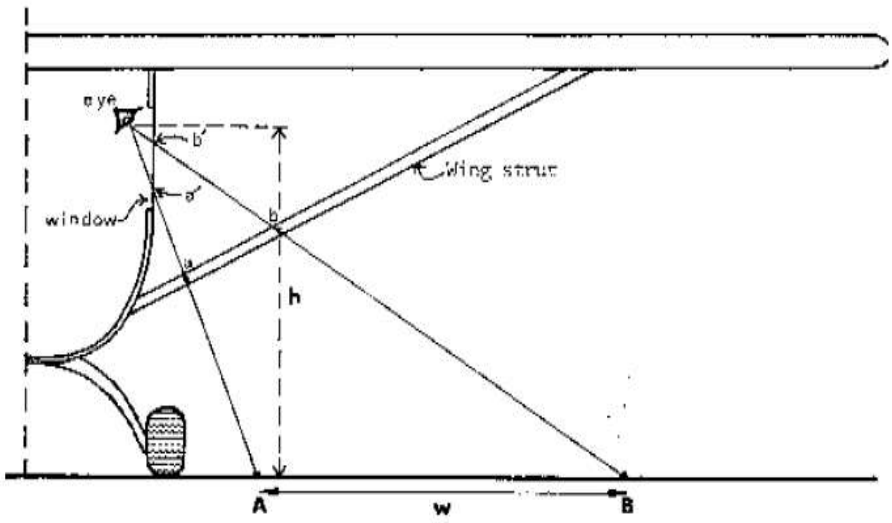

Figure 2: Schematic diagram of aircraft configuration for strip width sampling North-Griffiths (19878). W is marked out on the tarmac, and the two lines of sight a'-a-A and b'-b-B establish, whereas a'- and b' are the window marks.

$$w = W*h/H$$

Where, W= the required strip width; h= the height of the observer's eye from the tarmac; and H= the required flying height.

The entire survey was set up with an observer/recorder crew: two recorders, one left side observer and one right side observer. Each left and right observer and a recorder were divided into a team. Observers were responsible to continuously searched for and counted muskox; the number of calves (5-6 months old) were counted when they were conspicuous while on transect. No sex and age classification count were systematically attempted. The data recorded included the number of muskox and GPS locations. Only counts of adults were used in the final population estimate. Even if this survey focused on muskox, additional sightings of other species were also recoded, such as caribou, polar bear, and wolf.

Analyses

As this survey focused mainly on obtaining an estimated number, only unambiguous classification criteria were used to determine the number of calves and adults. The group was then broken down into adults (female/male) and calves (Howard 2011). The flying height and speed did not allow for accurately distinguishing male from female muskox from horn size and shape. Therefore, the proportion of calves per female cow was not determined, and no information on the recruitment or productivity was generated. The group structure was

however described such as calf:adult ratio, mean group size, and the number of single lone bulls encounter.

To determine the number of muskoxen in the study area, only the adult muskoxen sightings recorded on transect were analyzed using Jolly's Method 2 for unequal sample sizes (Jolly 1969) using a coefficient limit of 95%. Such methodology was previously used for the survey of Boothia Peninsula in 2006. The population estimates for fixed-width strip sampling using Jolly's Method 2 for uneven sample sizes (Jolly 1969; summarized in Caughley 1977) are derived from the following equation:

$$\hat{Y} = RZ = Z\frac{\sum_i y_i}{\sum_i z_i}$$

Where $\hat{Y}$ is the estimated number of animals in the population, R is the observed density of animals (sum of animals seen on all transects $\sum_i y_i$ divided by the total area surveyed$\sum_i z_i$), and Z is the total study area. The variance is given by:

$$Var(\hat{Y}) = \frac{N(N-n)}{n}\left(s_y^2 - 2Rs_{zy} + R^2 s_z^2\right)$$

Where N is the total number of transects required to completely cover study area Z, and n is the number of transects sampled in the survey. s_y^2 is the variance in counts, s_z^2 is the variance in areas surveyed on transects, and s_{zy} is the covariance. The estimate $\hat{Y}$ and variance $Var(\hat{Y})$ are calculated for each stratum and summed. The Coefficient of Variation (CV = σ/$\hat{Y}$) was calculated as a measure of precision.

Density, the number of muskoxen per unit area (muskox/km^2), will be determined using the number of adult muskoxen seen on transect divided by the total area of the study area. Lakes and stream areas will be not subtracted from the total area calculations used in muskox density (Statistical analysis based on Campbell and Setterington (2001)).

The area occupied by muskox and the time of the survey within the study area was determined. Thus, the distribution of muskox was illustrated by plotting each muskox sighting on and off transect, based on their precise geospatial position captured with GPS. In addition, the number of animals composing each group was highlighted using an increasing size of circles to represent groups of 0-1, 2-7, 8-11, 12-15, and 16-19 animals.

Given the importance of predators, Polar bear (*Ursus maritimus*) and Arctic Wolf (*Canis lupus arctos*), we collected standardized information of predator sightings in the management units

using the predator index (Heard, 1992). The predator index reports all predator sightings per species against the total number hours flown, also including the ferry time in this case. It is then possible to have a yearly trend, as the number of predators observed is expressed per 100 hours for this particular time of the year.

Results

The survey was conducted out of the community of Taloyoak from August 07th to August 12th, 2017. The management unit was surveyed in 40 hours, including on transect and ferrying flights from Taloyoak airport to the start of the transect lines. Low ceiling and fog prohibited the ability to survey continuously from the North to the South of the study area. Therefore, some sections were left to be completed at a later time, when the weather was permitting. The sedentary muskox behavior (Adamczewski *et al*., 1997) reduces the probability that an individual will move any great distance within the short survey time frame.

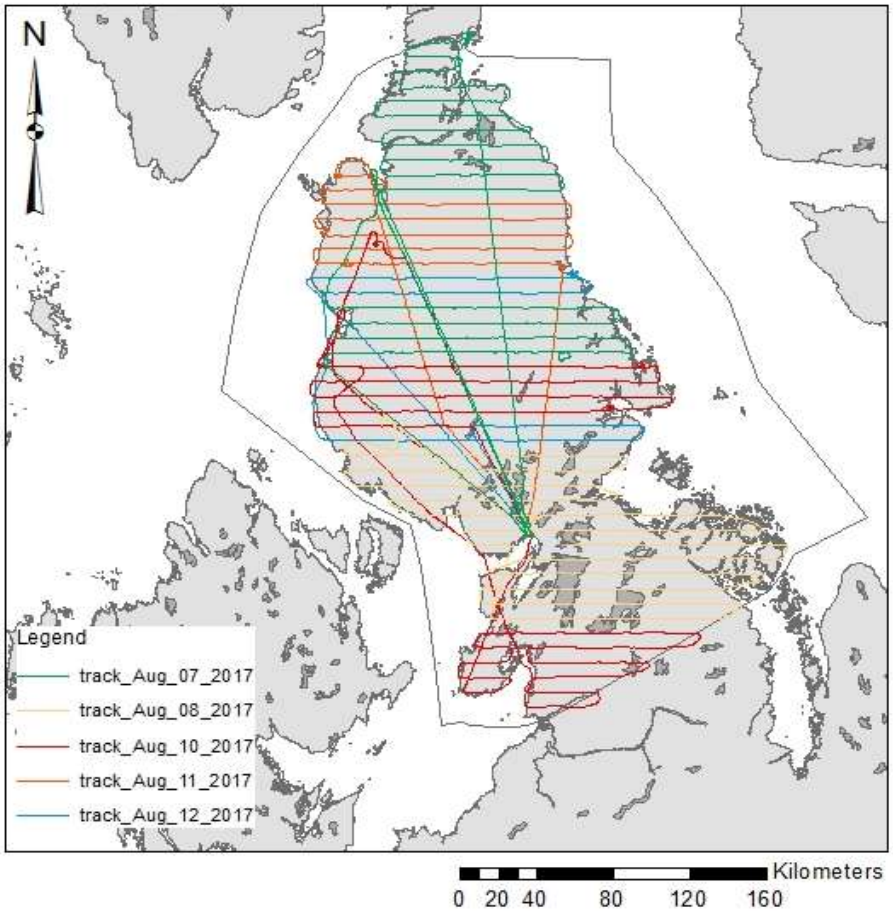

Figure 3: Daily tracks completed to cover 20% of the muskox management units MX-08 from August 7th to 12th.

Distribution

The adult muskox distribution in the management unit is represented in Figure 5 below. During the survey, 170 groups of muskox were seen on and off transect. The large groups of muskox, 16-19 adult animals, were distributed from Cape Farrand to Abernethy Bay within 40 km from coast. Additional muskox aggregations were found around the Wrottlesley River valley in the small portion of the southwest by the Victoria Island lowlands. Very few groups were located at high elevation (594 meters) in the central north part of the Peninsula, as muskoxen appeared to avoid the Boothia Peninsula Plateau. It was the first time that muskox observations was recorded south of Cape Cambridge, now reaching close to Acland Point at their southernmost distribution.

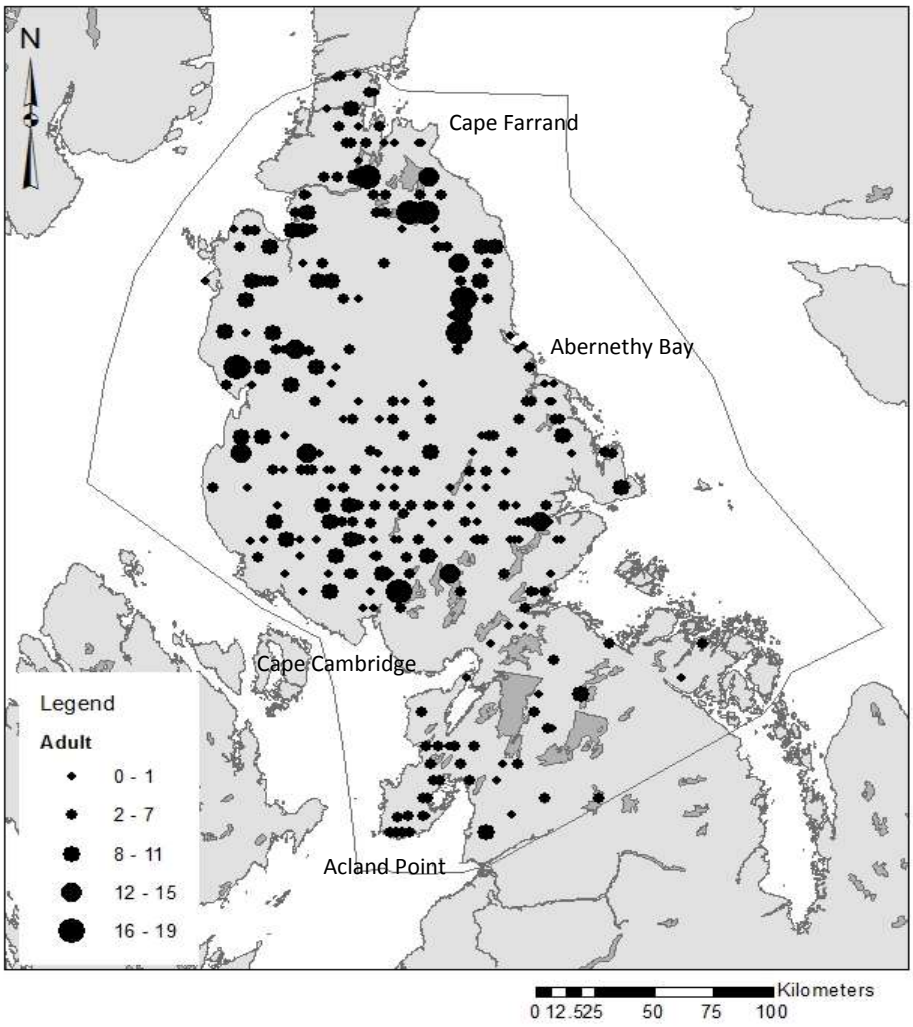

Figure 4: Muskox distribution on and off transect in the management unit MX-08 during the survey where the number of animals per group was grouped as 0-1, 2-7, 8-11, 12-15, and 16-19.

Group Characteristic

During the survey, 139 groups of muskox were recorded on transect, where 26 were single lone bulls. The majority of the groups (41%) were very small groups of 2 to 7 adults (Figure 5). The average number of adults (+1 year and older) per group was 5 ± 4.45 (S.D.) where calves were not included in the group size. The highest number of adults counted in one group was 19.

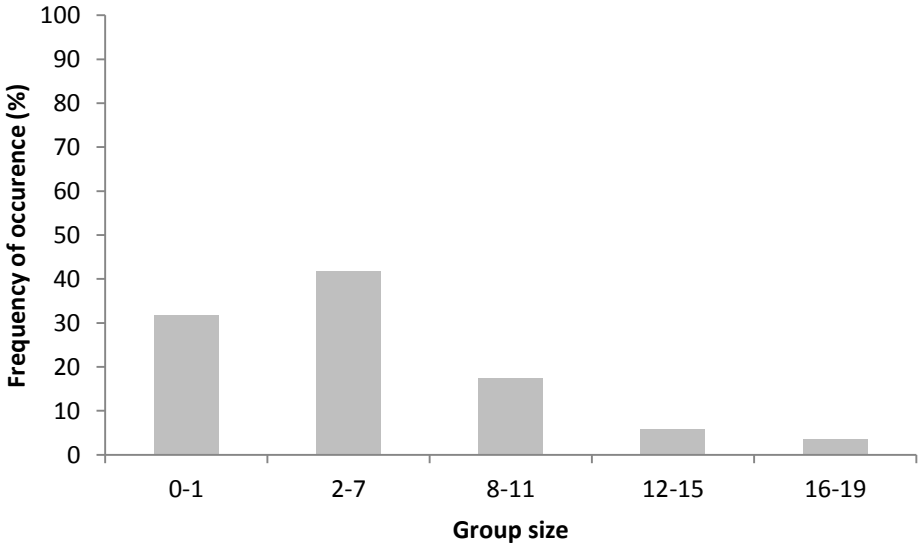

Figure 5: Frequency of occurrence (%) of adult muskox number per group size, grouped as follow 0-1, 2-7, 8-11, 12-15, and 16-19.

Nonetheless, the calf to adult ratio was determined for each group of muskox seen on transect. Since the identification was done from a fixed-wing, it was impossible to distinguish the sex of the adult or yearling based on the horn shape and length. Close to half the group seen (45%) did not have any calves. For the group that had calves, most of them had at least one calf, but some larger groups had up to 4. Most of the groups that had a larger proportion of calves were located north of the Boothia Peninsula, south of Murchison Promontory and west of Cape Farrand, and in the Wrottlesley River valley (Figure 6). The overall proportion of calves to adults was 14%.

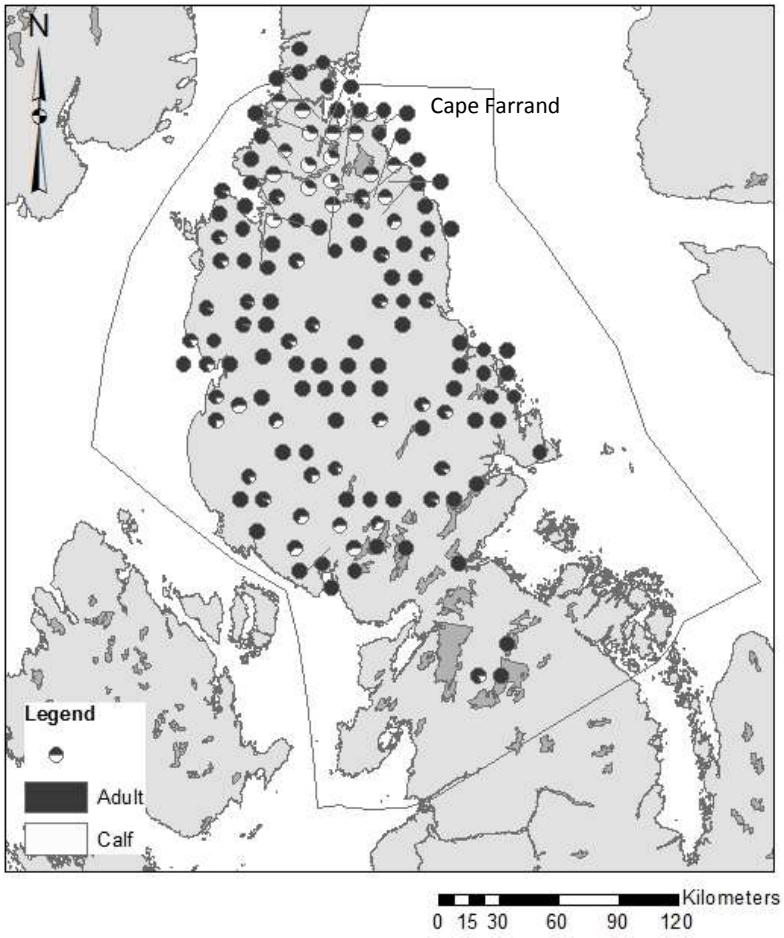

Figure 6: Proportion of calves per adult muskox in each group observed on transect in the Muskox Management Unit MX-08.

Estimate

The percentage of the overall cover of the management unit surveyed with 8,317 km^2 represented 20% of the total study area (43,238km^2). During the survey, 702 adult muskoxen on transect were recorded. The estimated number of muskox in the management unit MX-08, totaled 3,649 ± 316 (S.E.) ($p<0.005$, $t = 1.676$, $N = 184$ and $n = 46$). For this estimate, the total number of transect at 100% coverage was 184 (N) and 46 (n) transect lines were surveyed (Table 2). Overall, the muskox density of the management unit was 0.084 muskox/km^2.

Table 2 Muskox estimate in the Muskox management Unit MX-08

Stratum	Area Survey (km^2)	Total area (km^2)	Muskox on Transect	Estimate	Standard error (S.E.)	95% CL (±)	CV
MX-08	8,317	43,238	702	3,649	316	530	0.09

* $p<0.005$, $t = 1,676$, $N = 184$ and $n = 46$

Predator sighting (wolves, polar bear and grizzly bear)

In 2017, during the 40 hours of flying within the management unit, 2 wolf and 7 polar bear sightings were recorded. The wolves were found on the southern part of the study area; 1 wolf west of Josephine River and south of Netsilik Lake where four wolves constituted the pack. These two locations were overlapping with caribou sightings (Figure7). Being located between two polar bear management units, the M'Clintock Channel and the Gulf of Boothia, it was probable to observe polar bears. Indeed, 1 female and two cubs were observed on the Cape Hobson and a lone adult was seen a few kilometers inland off the shore of M'Clintock Channel. The remainder of the sightings were on the Gulf of Boothia, with seven polar bears and a female with a pair of cubs observed between Abernethy Bay and Mary Jones Bay. No grizzly bears were seen during the survey. Predator sightings, using the predator index, (Heard, 1992) reveled 13 wolves/100 hours and 25 polar bears/100 hours.

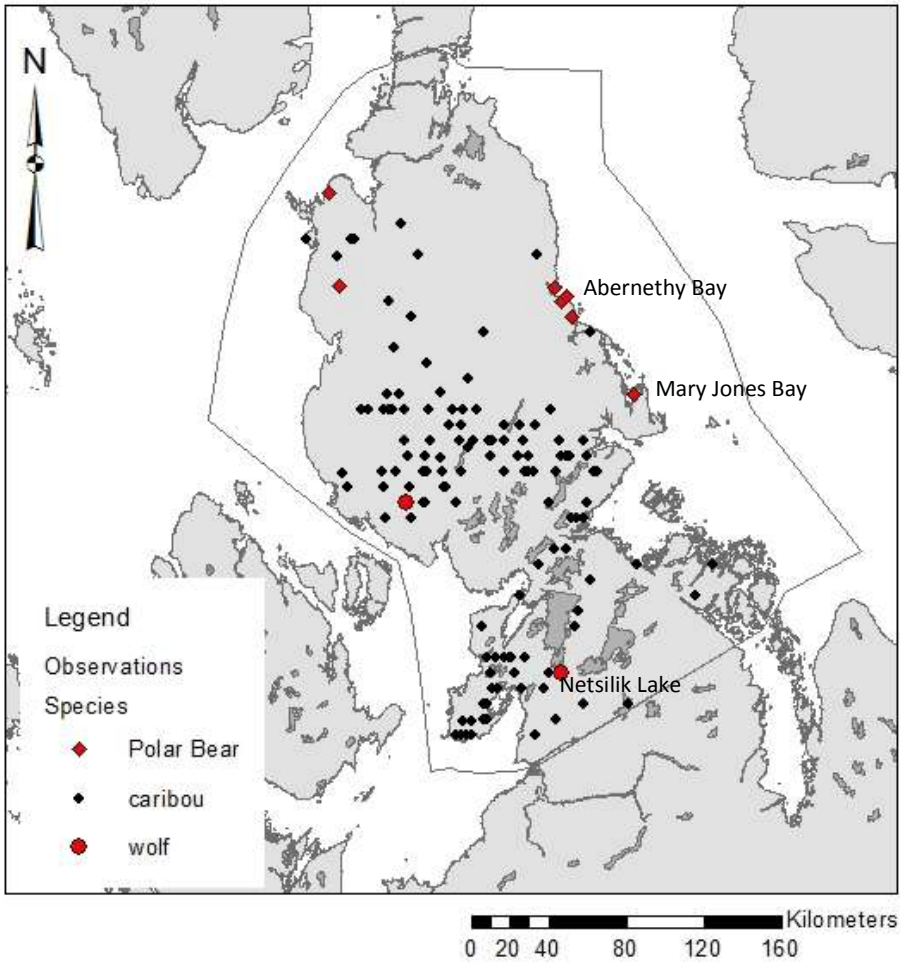

Figure 7: Locations where Polar Bears and Wolves were observed in the Muskox Management Unit MX-08 in relation to caribou distribution.

Discussion

Distribution

Despite the fact that traditional knowledge indicates that muskox inhabited the Boothia Peninsula long ago, it is common knowledge that muskox were not found in this management unit in the 1980s (Gunn & Dragon, 1998). Early in the 1990s, muskox started to move south of Somerset Island colonizing the northern part of the Peninsula, where their distribution remained consistent for a decade (Dumond, 2007). In 2017, although the majority of the muskox were still located north, they re-colonized most of the Peninsula reaching a new southern limit. Favorable environmental conditions and adequate forage, low number of predator (no grizzly bear) would have either contributed to increase the immigration rate or herd productivity. Muskox are now found in close proximity to the community of Taloyoak, as far south of Arcland Point, which can provide new harvesting opportunities. If these muskoxen are from the Arctic Archipelago, a subspecies genetically and morphologically distinct from the mainland muskox, it will be interesting to track their expansion and monitor the exchange between these two subspecies once their ranges overlap.

Group Characteristic

The number calves represented 14% of the total number of adult muskox observed on transect. This ratio is normally associated with a population that would be increasing, since it has been establish that a calf recruitment rate of 10.5% is necessary to keep the muskox population stable (Freeman, 1971). Since the calf ratios have known to vary greatly between years, long term data is needed in order to determine a trend (Reynolds, 1998).

Small groups of muskoxen was characteristic of the Boothia Peninsula, with an average group number of 5 ± 4.45 (S.D.), which is consistent with the mean group size of 6 established in 2007 (Dumond, 2007). Comparatively to King Williams Island where the muskoxen are known to increase in number rapidly, larger groups, 13 ± 8.40 (S.D.), were observed at the same period of the year (Leclerc, 2015).

Abundance Estimate

The extent of the harvest zone MX-09 remained relatively the same after the creation of MX-08 when each muskox management unit in Nunavut was reviewed in 2015. The major change, is that the new unit, now called MX-08, did not include the southern portion of Somerset Island, but starts at the northernmost coastline. This similarity allowed the 2017 study to be compared with previous population estimates.

Surveys of the Boothia Peninsula have occurred sparsely since the 1980s. In 1985, during a survey, no muskoxen were seen. Local knowledge has indicated that the presence of disease or parasites caused an abrupt decline in the muskox number. It is only in 1995 that muskox started to re-colonize the area and resulted in an initial population estimate of 554 ± 205 SE. The community members of Taloyoak mentioned a consistent increase in the muskox number. This observation triggered a second survey, in 2007. This survey confirmed the local knowledge and the muskox numbers were then estimated at 1,058 ± 198 SE. Recently, with this continuous increase, muskox have been recorded to be close to the community of Taloyoak and they would like to re-adjust the Total Allowable Harvest. In 2017, the estimated number of muskox reached 3,649 ± 316 SE, which is the highest number to date. There will likely be a recommendation for the NWMB to consider an increase to the current harvesting rate after consultation and discussion with the Hunter and Trapper Organization in the affected community of Taloyoak.

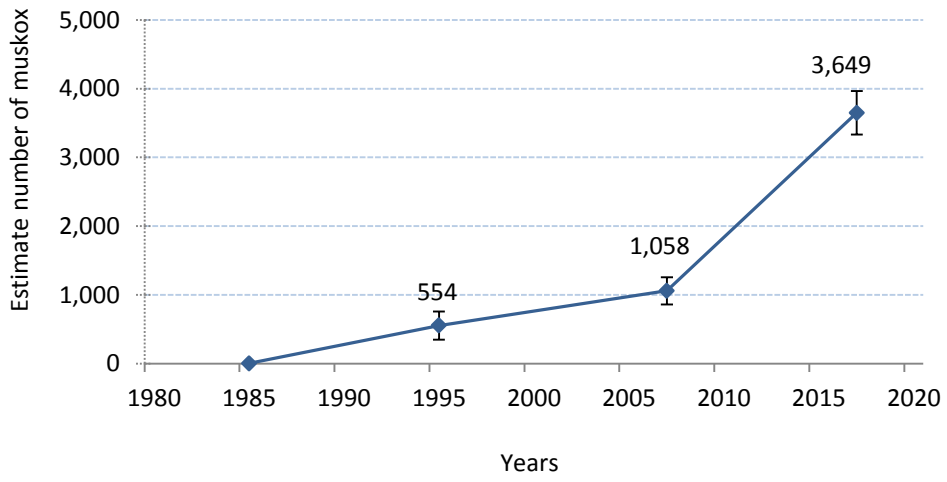

Figure 8: Muskox population estimate for MX-08.

Density

Muskox density was 0.084 muskoxen / km^2 on the Boothia Peninsula. In this management unit, the muskox has a density higher to the overall density on Victoria Island, 0.074 muskox / km^2, but is lower to that of King William Island where the muskox population has also increased (0.1123 muskox/km^2) (Leclerc, 2015). The mechanism driving muskox density is still not fully understood. Heard (1992) noted that group size in not generally related to muskox density. These qualitative comparisons between areas, highlight that density might fluctuate spatially and temporally.

Predator sighting (wolves, polar bear)

The number of known muskox predators was minimal. Only two wolves were observed, and at close proximity of caribou, their main prey. Wolf predation on muskoxen is common, with packs

or single wolves observed following and killing muskox. No Grizzly bears were found on the Boothia Peninsula. The relatively low abundance of predators would benefit calf recruitment and female survival, promoting an increase in the population. 25 polar bears 100 hours were seen, but polar bear are known to feed on a mainly marine based diet and are not known to predate on muskox and should not contribute to regulate muskox population.

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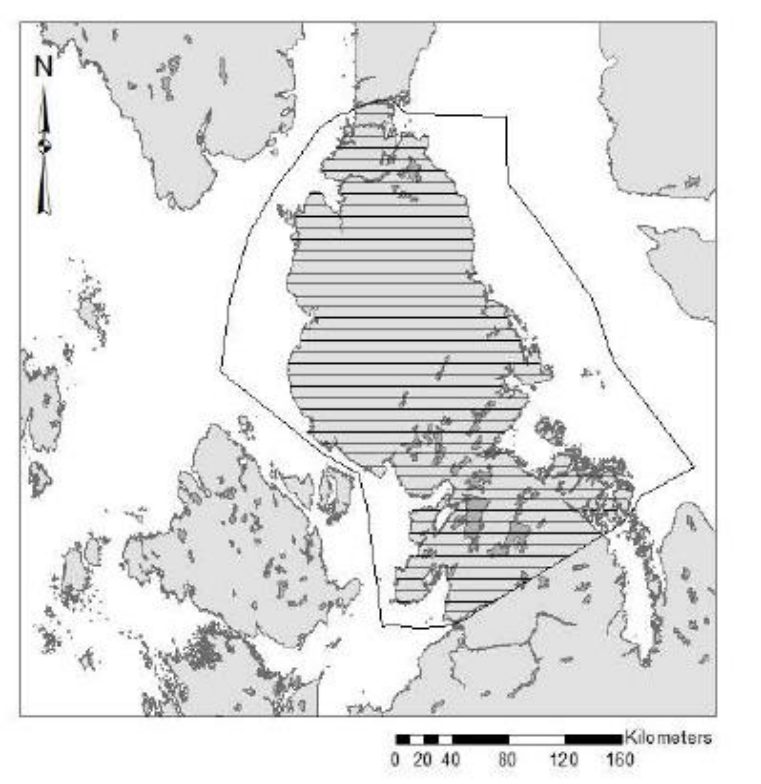

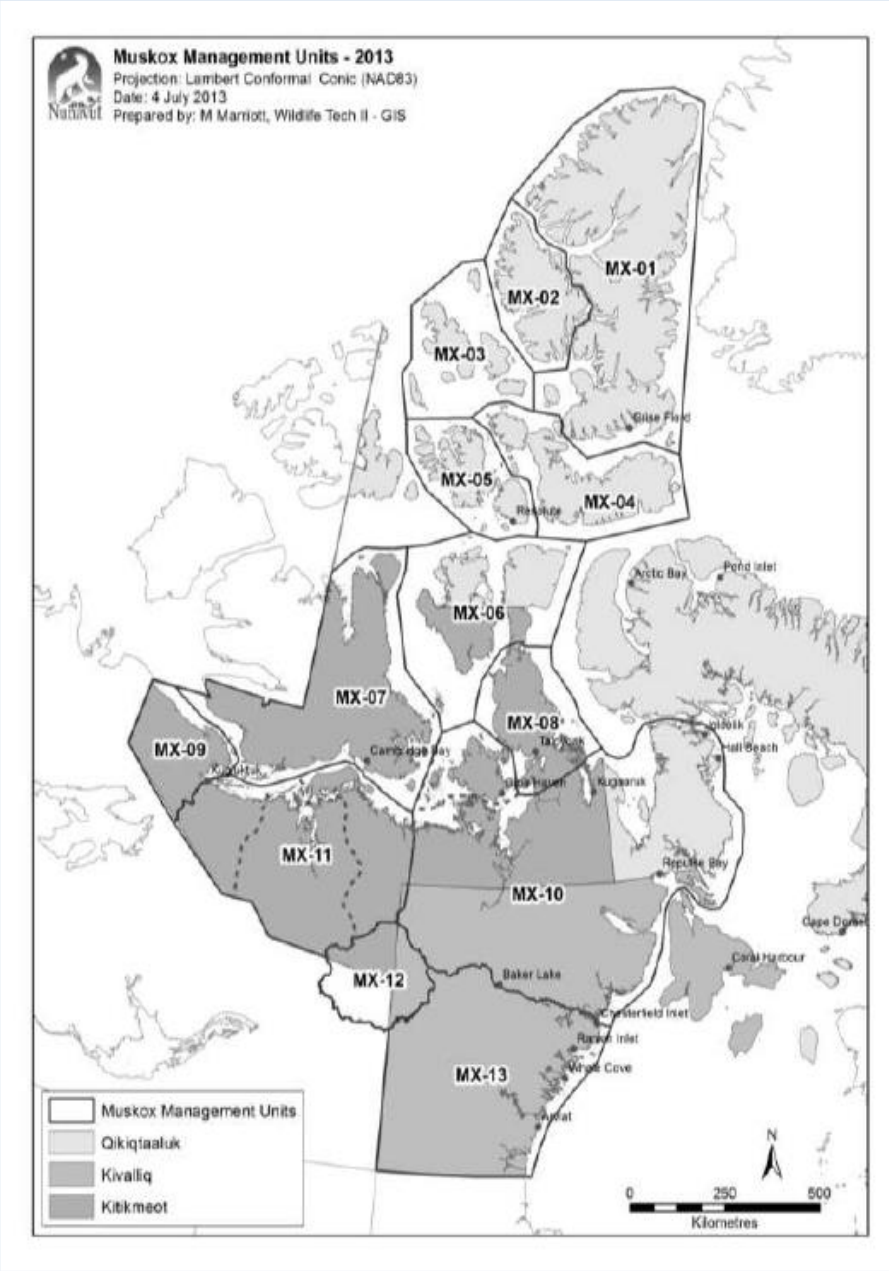

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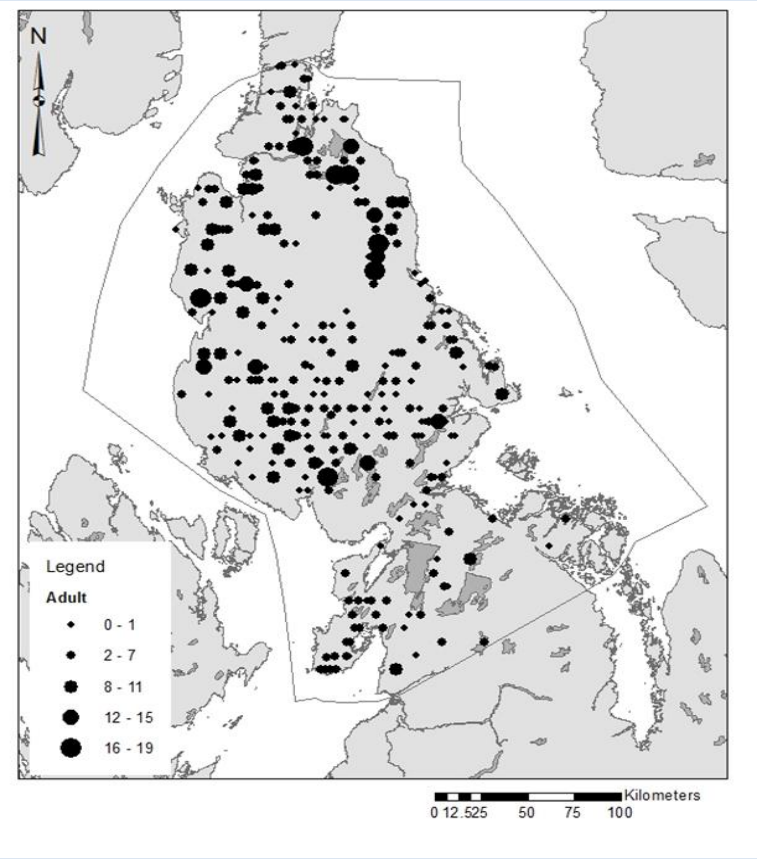

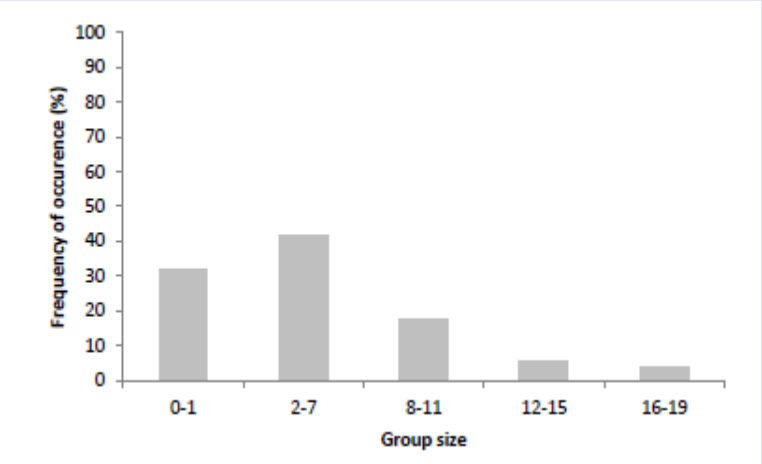

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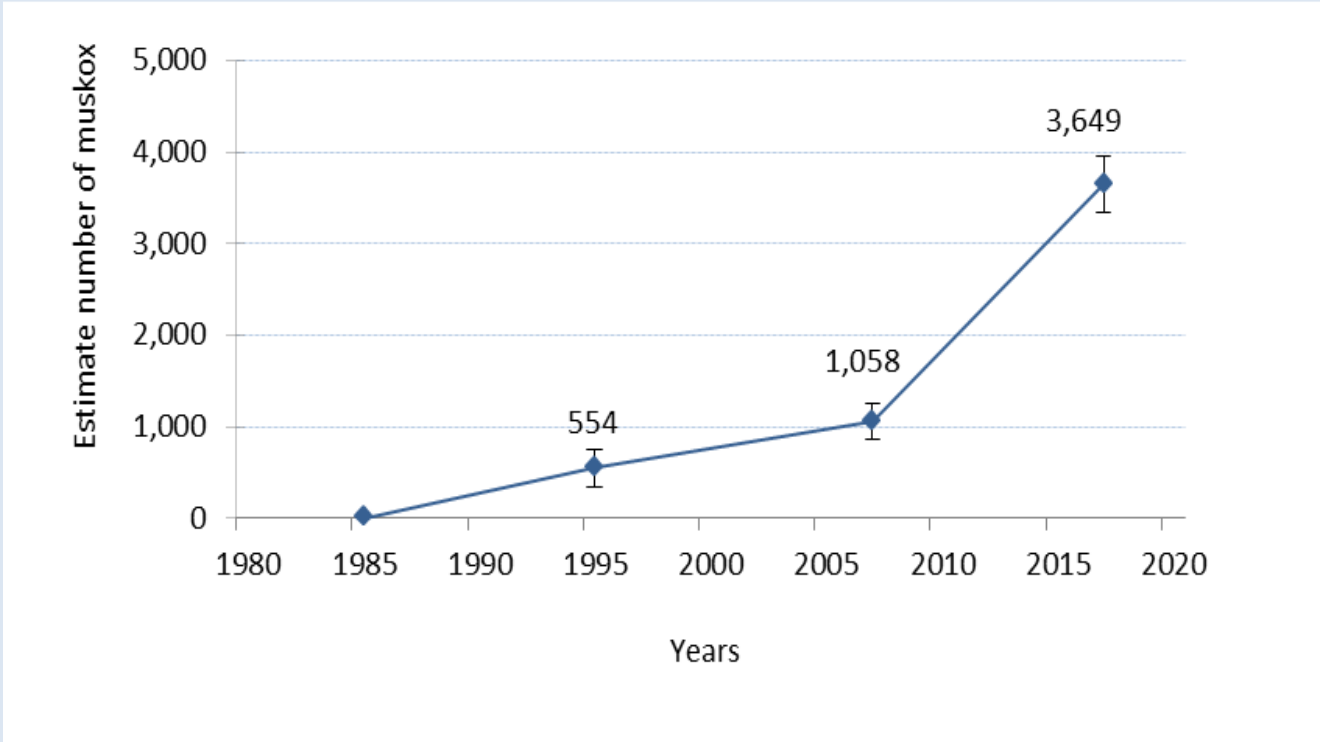

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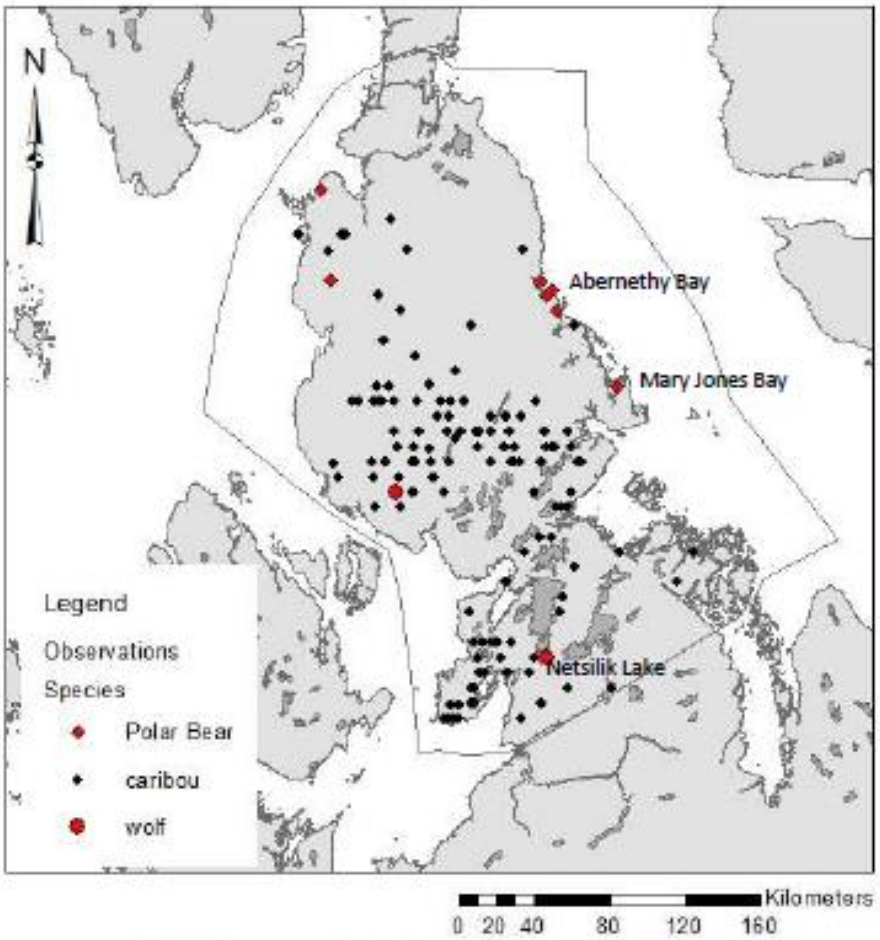


Figure 7: Locations where Polar Bears and Wolves were observed in the Muskox Management Unit MX-08 in relation to caribou distribution.

ᐊᐅᓪᓗᑎᒋᖅ - ᐅᖅᑲᓪᓚᓱᒃᓴᒋᖅ

ᓯᓇᐅᒡᓕ ᐅᒪᔨᓪᓕᓗᔨᓴᒃᑯᓪᓕ ᐊᐅᓪᓗᑎᒋᔭᖏᑦ ᐱᖃᓴᖅᑕᒡ?

ᐊᒧᖃᓴᓄᑦ ᐱᖅᐅᓯᒪᒡᓕ?

ᓯᓇᐅᒡᓕ ᓄᓇᓪᓐᓗᑦ ᐱᖅᐅᔨᒪᔨᕐᖓ?

ᑲᑎᒪᓗᒍ ᓄᓇᕗᑦ ᐱᖃᓐᓇᕐᓂᖅᓯᒪᔪᑦ
Building Nunavut Together
Nunavuliuqatigiingniq
Bâtir le Nunavut ensemble

ᓄᓇᕗᒻᒥ ᑲᑭᒪᑲᓂᑦ ᐃᒪᓐᓇᖅᑐᑎᓐᓄᑦ

ᐊᑐᖅᓴᔪ ᓇᓴᐃᓂᕐᒧᑦ ᖃᑲᓇᐃᓐᓄᑐᐊᕐᑦ ᐊᒻᒪᓗ ᓄᓇᓕᓐᓂᐅᓂᒃ ᐊᒻᒪᓗ ᐅᒪᔨᓕᕆᔨᓴᑲᓂᒃᓴᓂᒃᑯᓂᖅ, ᐅᑯᐊ, ᐃᓐᓇᐅᔨᖅᑲ ᐃᒪᓐᓇᖅᑎᑎᓕᕆᔨᑦ ᐊᐅᓚᑦᑎᓂᕐᒧᑦ ᐅᒥᐊᓐᓂᒃ ᐅᕐᓂ ᐅᒥᐊᓐᓂᒃ ᐊᐅᓚᑦᑎᓂᖅ MX-08-ᒥᑦ:

ᐊᑭᑎᓕᓵᓴᒃᑯᑦ ᓄᓇᕗᒻᒥ ᓂᕆᔨᑎᓂᒃ ᐊᐅᓚᑦᑎᓴᒃᑯᑦ ᐊᖏᕐᓂᕈᐊᖅᓯᕆᑯᓴᑦ ᐃᓐᓇᐅᔨᖅᑲ ᐅᒥᐊᓕᕐᑕᐅᓴᕿᓂᒃ 66 ᐱᓴᐅᔨᖏᓪᓗᓂᐅᕙᖅᑐᓂᒃ 275-ᓄᑦ ᐅᒥᐊᓐᓂᒃ ᐊᐅᓚᑦᑎᓂᖅ MX-08-ᒧᑦ.

ᐅᒪᔨᓕᕆᔨᒃᑐᖅ ᐅᖃᖅᑐᑎᖅᑎᓂᖅ ᐅᒥᖏᒪᐃᓕ ᓇᓂᖏᓂᖅᓯᓕᓂ ᐊᓚᒪᓗ ᐱᑕᖃᐅᖅᓯᕐᓂᒃ ᒥᑦᑐᐅᓂᖅ, ᐅᒥᖏᓐᓂᖅ ᐊᐅᓚᖅᑎᓂᕐ MX-08

ᐃ>ᓗ 2019

ᐊᖏᑎᓕᓗᓴᖅᑯᑦ, ᓄᓇᕗᒻᒥ ᑲᖏᒪᖅᑯᑦ
ᐃᖃᓗᐃᑦ, ᓄᓇᕗᑦ

ᓇᐃᓐᓴᖅᓯᒪᔪᖅ

ᓄᓇᕗᒻᒥ ᑲᕐᒪᑲᑯᓂᓪ, ᐊᕐᑎᑕᕆᔨᑲᑯᓄᓪ ᑭᒍᖅᑐᐃᔨᐅᐅᔪᓪ ᐅᖏᖅᖏᑎᖏᓕᐅᖅᑐᓪ ᑕᓗᖏᔪᐊᖏᑦ ᐅᒪᔪᑦᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᓄᓇᓕᖅᖁᐅᓂᓪᓗ ᐃᐳᕆ 23, 2019-ᒥᑦ. ᐱᔭᔪᑕᐅᓗᖅᑐᖅ ᐅᖏᖅᑎᑎᖏᓂᐅᔪᓄᓪ ᑐᐱᐅᒪᑎᑲᖅᓂᕈᒪᖏᓪᓗᑕ ᐅᒪᔪᑦᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐊᒻᒪᓗ ᓄᓇᓕᖅᖁᐅᓪ ᖏᓄᐃᖅᓯᐅᔪᓂᒃ 2017-ᒥᑦ ᐱᑕᖏᐅᖏᓂᖅᓄᓪᓕᓪ ᓇᓕᐅᑦᑕᖅᑕᐅᓕᐅᖅᑐᑦ MX-08-ᒧᑦ ᐅᖏᖅᑦᓕᖏᑎᑎᕈᑦᓗ ᐊᐅᓕᕆᓂᖏᓐᓇᒥᑦ ᐃᒪᓐᓴᖅᑐᑎᑕᐅᓕᐅᔪᑦ ᐊᑐᓕᖅᑎᑕᐅᖏᖅᓯᖏᓕᖅᑐᑦ. ᐅᖏᖅᑎᑎᖏᖅᑲᓂᖏᓂᖏᓪᓗᑕᐅᓕᐅᖅᑐᑦᑕᐅᖅ ᐱᖅᓂᖏᓂᐊᖏᓪᓗᑐ ᑲᑎᒪᔨᖅᑐᑦᑕᐅᖅᑐᓪ ᓄᓇᓕᖅᑐᑦ ᐊᒻᒪᓗ ᐱᖅᑯᓯᖅᑐᑦ ᖃᐅᔨᓴᖅᑐᑦ.

ᑕᓗᖏᔪᐊᖏᑦ ᐃᒪᔪᑦᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐃᓚᒪᕿᖅᑐᑦ ᖃᐅᔨᕿᐃᓂᖏᓪᓗᑕ ᖃᐅᔨᓴᖅᑐᑦᑕᐅᖅᑐᑦ, ᑕᑯᕿᕆᐅᑎᓂᕐᑎᔪᑦ ᖃᐅᖅᓯᐅᓂᖅᑐᑦ ᐅᒥᖅᓯᒪᑦ ᐊᒥᓱᓕᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐅᒥᖅᓯᒪᔨᐅᓕᖅᑐᑦ 66-ᖅᓯᔪᑦ, ᐅᕐᓀᓗᓐᓂᑦ ᐅᒥᖅᓯᒪᔨᐅᓕᖅᑐᑦᓄᓪ ᐃᓚᐅᑎᑦᑎᓯᖅ 1.8%-ᖅᓯᔪᖅ ᐃᓐᓇᐅᔨᔪᖅ ᖃᐅᔨᓴᖅᑐᖅᓂᒃᓄᓪ ᓇᓕᐅᑦᑕᖅᑕᐅᔨᓂᑦ 3,469 ᐅᒥᖅᓯᒪᓄᑦ. ᓇᓂᓐᓂᑕᓴᖅᓯᒪ ᐅᒥᖅᓯᒪᑕᑦ ᐊᒥᓱᓕᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐃᓱᒫᓗᐊᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐊᒻᒪᓗ ᓯᖅᐃᑎᓄᑦ. ᓯᖅᐃᑎᑦᑕ ᐊᖏᕐᓂᒥᐊᖅᑕᐅᓕᖅᑐᑦᑕᐅᖅᑐᑦ ᐃᓐᓇᐅᔨᔪᖅ ᐅᒥᖅᓯᒪᑕᐅᔨᓴᖅᑐᑦ ᐊᒥᓱᓕᕆᓗᐊᖅᑐᑦᑕᐅᖅᑐᑦ ᖃᐅᔨᓴᖅᑐᖅᓂᒃᓄᓪ ᐊᓱᐅᓕᖅᑐᑦᑕᐅᖅᑐᑦ ᑐᖅᑐᓂᓪ, ᐊᒻᒪᓗ ᐃᓂᖏᓪᓗᑕ ᓄᓇᓕᖅᐳᑦ ᖃᐅᔨᓴᖅᑐᖅᓂᓕᕆᓂᖅ. ᐊᕐᑎᑕᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐅᒥᖅᓯᒪᔨᐅᓕᖅᑐᑦᑕᐅᖅᑐᑦ 6%-ᒧᑦ, ᐊᒥᓱᓕᔭᕆᐊᖅᑎᑎᕆᔨᖅ 218-ᒧᑦ 66-ᖅᓯᒪᐅᖅᑐᑦ ᐅᒥᖅᓯᒪᑕᐅᔨᓴᖅᑐᑦ. ᑕᓗᖏᔪᐊᖏᑦ ᐃᒪᔪᑦᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐊᒻᒪᓗ ᓄᓇᓕᖅᖁᐅᑕᐃᑦ ᐃᑲᔪᖅᑐᑦᑕᐅᖅᑐᑦ ᐊᒥᓱᓕᔭᕆᐊᖅᓴᓂᖏᑎᒃ, ᑭᓱᐊᓂ ᑕᑯᔪᒪᔨᑦ ᖏᑯᑎᓄᓂᖅᓴᖅᑎᑦ ᐳᓴᓄᑎᒃ ᒪᖏᓕᐅᑕᐅᔨᓴᖅᑐᑦᑕᐅᖅᑐᑦ 218 ᐅᒥᖅᓯᒪᑕᐅᔨᓴᖅᑐᑦ ᐊᒥᓱᓵᖅᑭᑐᓗᐊᖏᓪᓕᖅᓱᒃ. ᐊᕐᑎᑕᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐃᑲᔪᖅᑐᑦᑕᐅᖅᑐᑦ ᐃᓕᓴᕐᓂᖅᑐᑦᑕᐅᖅᑐᑦ ᐊᒻᒪᓗ ᐱᓕᕆᖏᓐᓂᖅᑐᑦᑕᐅᖅᑐᑦ 275-ᖅᓯᓕᓪᓗᑕ (ᒪᖏᐃᑕᐅᔨᓴᓕᖏᓪᓗᑎᓂᓪ ᐃᓐᓇᐅᑎ 7.5%).

ᐅᓂᒃᑲᖅ ᓇᐃᓐᓴᖅᓯᒪᔪᖅ ᐅᖏᖅᐅᓯᐅᓕᐅᖅᑐᓂᒃ ᐃᒪᔪᑦᕆᔨᐊᓕᒋᑲᑯᓂᓪ ᐊᒻᒪᓗ ᐃᓄᓪᖃᒋᓄᓪ ᑲᑎᒪᓂᖅᑲᖅᑎᑦᓗᑕᑦ.

ᐱᑭᐊᑭᑎ

ᐅᓂᐸᑲᖅ ᑕᑯᐸᐳᑎᑕᑎᐣᖅ ᐊᐢᖏᑎᓕᕆᔨᖅᑯᑦ ᐱᓇᓯᓗᖅᑕᒍᐅᖅᑕᖏᑐᓂᖅ ᓇᒍᐅᑦᑎᐣᓂᖅ ᑐᑭᓯᐅᒪᔪᑎᓗᖅᑕᓂᖅ ᐱᒍᐅᖅᑕᖏᑦ ᐅᖃᖅᑲᖅᑎᖅᑎᓪᓗᑎᑦ ᑲᑎᒪᓂᖅᓂᑦ ᐅᒪᔨᓕᕆᔨᐊᒋᑲᖅᑯᓂᑦ ᐊᒻᒪᓗ ᐃᓄᓗᖅᑕᓂᖅ.

ᐃᓯᒪᑎᕐᐅᐊᓪᓚᑦ ᐅᖃᖅᐅᓯᐅᒍᐅᖅᑐᑦ ᑕᑯᓂᓂ ᑕᑯᐸᐳᑎᑕᑎᓐᖏᑦᑐᑦ ᐊᐢᖏᑎᓕᕆᔨᖅᑯᓂᑦ, ᓄᓇᕗᒻᒥᑦ ᐅᕐᓂᓗᓐᓂᑦ ᓄᓇᕗᒻᒥ ᓴᓂᒪᖅᑯᓂᑦ.

ᐃᓗᓕᖅᑭᑦ

ᓇᐃᔭᖅᓯᒪᔨᖅ i

ᐱᑎᐊᕈᑎ ii

ᐃᓗᓕᖅᑭᑦ iii

1.0 ᐅᓂᒃᑲᐸ ᐱᓵᕎᑎᖓ ᐊᓚᒪᓗ ᐊᐢᑭᑭᐅᒪᓂᖓ 4

2.0 ᐱᓵᕎᑎᖓ ᐅᖏᖅᑐᑎᖏᖅᑕᑦ 4

2.1 ᖃᓄᐃᖓᓂᓚᕆᖓ ᑲᑎᒪᔨᑦ 4

3.0 ᐅᒪᔨᓕᓚᕕᒋᑦᓂᓕᑦ ᐊᓚᒪᓗ ᐃᓄᓗᑦᑕᓂᑲ ᐅᖏᖅᑐᑎᖏᖅᓂᕐᒧᑦ ᓇᐃᔭᖅᓯᒪᔨᖅ 4

3.1.1 ᑕᓗᕐᕕᐊᖅ ᐅᒪᔨᓕᓚᕕᒋᑦᓂᓕᑲ ᐅᖏᖅᑐᑎᖏᖅᓂᕐᒧᑦ ᓇᐃᔭᖅᓯᒪᔨᖅ - 2019 5

3.1.2 ᑕᓗᕐᕕᐊᕆᑦ ᐃᓄᓗᑦᑕᓂᑲ ᐅᖏᖅᑐᑎᖏᖅᓂᕐᒧᑦ ᓇᐃᔭᖅᓯᒪᔨᖅ - 2019 6

4.0 ᐃᓱᓕᓂᖅ - ᐱᓕᓚᐊᖓᑎᐊᕈᒪᓴᖅᑐᑦ 7

5.0 ᐅᒥᖓᓂᑲ ᐊᐅᓕᕆᑎᓂᖅ MX-08-ᒥᑦ ᓇᓕᐅᑦᑕᖅᑕᐅᔨᑦ ᖃᓄᐃᖓᓂᑲ ᐊᓚᒪᓗ ᐊᐅᓕᕆᑎᓂᕐᒧᑦ ᐃᓚᓇᔭᖅᑐᒃᑐᑕᐅᔨᑦ ᑲᑎᒪᔨᕐᑎᑦ 8

ᐅᒪᔨᓕᓚᕕᒋᑦᓂᓕᑦ ᐅᖏᖅᑐᑎᖏᖅᓂᓂᖅ 8

ᐃᓄᓗᑦᑕᓂᑲ ᐅᖏᖅᑐᑎᖏᖅᓂᓂᖅ 9

ᑎᑎᖅᑲᒧᑦ ᐃᓕᑭᐅᐅᔨᖅ 1 13

ᑕᑯᕐᓇᖅᑎᑕᑎᓂᖅ 2019 13

ᑎᑎᖅᑲᒧᑦ ᐃᓕᑭᐅᐅᔨᖅ 2 14

ᖃᓚᑕᐅᐅᖅᑯᑦ ᐃᑲᔪᖅᑕᐅᓂᕐᒧᑦ ᑕᓗᕐᕕᐊᕆᑦ ᐅᒪᔨᓕᓚᕕᒋᑦᓂᓕᑦ 14

1.0 ᐅᓂᐸᑦᑲᐸ ᐱᓴᒋᑎᖏᑲ ᐊᓚᒪᓗ ᐊᐧᖏᑭᐅᒪᓂᖏᑲ

ᐅᓂᐸᑦᑲᖅ ᓇᐃᓰᖏᑎᓭᓴᖅ ᐅᖏᑲᐅᓯᐅᓪᓕᐅᖏᑲᑐᓂᒃ, ᐊᐱᖏᑲᔨᑕᐅᓪᓕᐅᖏᑲᑐᓂᒃ, ᐃᓯᐠᒪᓵᑎᓂᒃ, ᐊᓚᒪᓗ ᐱᖏᑲᐅᔪᐅᓂᒃ ᐅᖏᑲᖏᑎᖏᑲᖏᑎᓪᓗᑎᑦ ᑕᓗᖏᔪᐊᖏᑦ ᐅᒪᔪᓕᓴᐊᓰᑲᔨᓂᒃ ᐊᓚᒪᓗ ᐃᓄᓗᑦᑕᓂᒃ 23, 2019-ᒥᑦ ᐱᓴᒋᑎᑦᓗᑎᑦ ᖏᓄᐃᓐᓴᓂᐅᔪᑦ ᐅᒥᓐᒪᐃᑦ ᓇᒧᓐᑲᖏᑕᖏᓂᕐᑎᓄᑦ ᐊᓚᒪᓗ ᐱᑕᖏᐅᖏᓂᕐᑎᓄᑦ ᖏᐅᐅᓴᐃᓂᖏᑲᒃ ᐅᒥᓐᒪᓕᓂᑦ ᐊᐅᑕᕐᑎᓂᖅ MX-08 ᒥᐦᓴᓄᑦ, ᐊᓚᒪᓗ ᐊᐅᑕᕐᑎᓂᖏᒧᑦ ᐃᒪᓐᓇᖅᑐᑕᐅᔪᑦ. ᓇᐃᓰᖅᓯᒪᔪᖅ ᐊᓚᒪᓗ ᑎᑎᖅᑲᐃᑦ ᑕᙳᓐᓂ ᑕᑯᓂᖅ ᑕᑯᓂᕐᑎᑎᔪᑦ ᑲᒪᕐᓴᐅᑎᓂᓂᒃ ᐱᓴᐅᑕᐅᖏᓐᓄᑦᓂᒃ.

2.0 ᐱᓴᒋᑎᓐᓄ ᐅᖏᑲᖏᑎᖏᑲᓂᕐᓂ

ᐱᓴᒋᑕᐅᓗᐊᖏᑲᑐᖅ ᑲᑎᒪᑕᕐᑎᓂᖏᓄᑦ ᐊᖏᑲᑭᓯᖏᑲᑕᐅᔪᖅ ᑲᒪᑕᑯᓂᕐᑎᓴᑲᒋᓐᓄᑦ ᐃᑕᐅᑎᓐᓂᐊᕐᓗᑎᑦ ᐅᓪᓚᕐᓗᒍᑦ ᐅᓇᓱᐊᕐᓂᖅᑐᑦ ᐊᐅᑕᕐᑎᓂᖏᒧᑦ ᐅᒥᓐᒪᓐᓂᖅ ᐊᐅᑕᕐᑎᓂᖏᒧᑦ ᐅᓂᐸᑦᑎᒥᑦ, ᑲᑎᖏᓯᐃᑲᓐᓂᖏᓂᕐᓗᓂᓂ ᓄᓇᓕᒥᑦ ᖏᑲᐅᔪᒪᓂᕐᑕᐅᔪᓂᑦ, ᐅᖏᑲᐅᓯᓚᓗᑎᑦᓗ ᓯᐅᓂᖏᓴᒥᑦ ᐊᐅᑕᕐᑎᓂᖏᒧᑦ ᐃᒪᓐᓇᖅᑐᑕᐅᔪᑦ. ᖏᓄᐃᓐᓂᑦ 2017-ᒥᑦ ᖏᑦᓯᐅᓂᖅᑎᓄᓂᒃ ᖏᑲᐅᔪᓴᐃᓂᖏᒥᑦ ᖏᑲᐅᔪᑎᑎᓕᐅᖏᑲᑐᓂᒃ ᑲᑎᒪᑕᖏᓂᒃ.

2.1 ᖏᓄᐃᓐᓂᐅᓂᕐᔨᖅᓐᓄ ᑲᑎᒪᓂᕐᓂ

ᑲᑎᒪᖏᑦᑕᓕᐅᖏᑲᑐᓂᒃ ᐅᓐᓄᑲᑦᓂ ᐊᓚᒪᓗ ᐊᑕᐅᓯᖏᒥᑦ ᒪᖏᕐᓄᓂᒃ ᐃᑲᖏᓵᓂᖏᑦᑕᖏᓐᓄᑦ. ᑲᑎᒪᑎᑎᓕᐅᖏᑲᑐᓂᒃ ᓄᓇᕗᒪᒥᑦ ᐊᐃᓴᑲᖏᓯᓚᕐᓗᒥᑦ ᐅᒪᔪᓕᓚᐅᓕ ᐊᓚᒪᓗ ᐊᐃᓴᑲᖏᓯᓚᕐᓗᒥᑦ ᓂᖏᔨᑦᑕ ᒥᐦᓴᓄᑦ ᐊᐅᑕᕐᑎᓕᓕ, ᑕᐧᑦᑯᐊᑲ ᑕᑯᓂᖏᑦᑕᕐᑎᓂᖏᑲᑐᓂᒃ ᐱᓴᒋᑕᐅᓗᐊᖏᑲᑐᓂᒃ ᐊᐱᖏᑯᓂᒃ ᐊᐱᖏᑲᑎᓂᒃ ᐃᓄᐃᑦᑐᓂᒃ ᐃᓕᓴᐅᔪᓂᒃ ᐃᓕᓴᐅᔪᓂᒃ ᐊᓚᒪᓗ ᐃᒪᓐᓇᖅᑐᑕᐅᔪᑦ ᐃᓕᓴᐅᔪᓂᒃ ᐊᐱᖏᑯᓂᒃ ᐃᓕᓴᐅᔪᓂᒃ ᐊᐱᖏᑲᑎᓂᒃ ᐃᓕᓴᐅᔪᓂᒃ ᐊᓚᒪᓗ ᐃᒪᓐᓇᖅᑐᑕᐅᔪᑦ ᐊᐱᖏᑲᑎᓂᒃ ᐊᐱᖏᑲᑎᓂᒃ ᐊᐱᖏᑲᑎᓂᒃ

3.1.1 ᑕᓃᕐᐊᕐᖏᑦ ᐅᒪᔨᓕᓗᐊᓪᓚᖅᑯᓐᓂᒃ ᐅᖏᖅᑎᑎᖅᓂᕐᒧᑦ ᓇᐃᓐᓇᖅᓯᒪᔨᖏᑦ - 2019

<u>ᐱᓵᓕᐅᔨᑦ:</u> ᐅᒥᖏᒪᐃᑦ ᓇᒧᖏᑐᐅᖅᑎᖅᑕᕐᓂᖏᑦ ᐊᓪᒪᓗ ᐱᑕᖅᑎᐅᕐᓂᖏᓐᓄᑦ ᐊᐅᓕᕐᑎᓂᖅ MX-08, Boothia ᓄᑲᐊᓂᑦ, ᐊᑎᓯ 2017. ᖃᖅᔪᐅᓂᖏᑦ ᐅᒥᖏᒪᐃᑦ ᐊᐅᓕᕐᑎᓂᕐᒥᑦ ᐊᒥᓯᖏᓐᓄᑦᑕᓕᐅᑎᐊᖅᑐᑦ.

<u>ᐅᑦᓗᖅ:</u> ᐃ>ᓗ 23, 2019

<u>ᑭᖑᓪᓗᐊᔨᑦ:</u>
ᐊᖓᑎᓕᓪᓚᕆᔨᑦ: ᓕᓴᐃᓗ ᓕᵇᓗᵇ ᐊᓪᒪᓗ ᑭᐊᐃᐅᵒ ᒥᐊᑐᐃᵒ, ᑕᐃᐃᖅ ᐊᓀᐃᓗᵇ

ᐅᒪᔨᓕᓪᓚᐊᓕᑦ: ᓱ ᐊᓯᖏᵇ (ᐃᵇᓯᖏᐅᑕᖅ), ᔨᐊᔭᐅ ᐊᵇᓕ, ᑯᖏᓕᵇ ᑯᑐᵇ, ᑕᒥ ᐊᐃᔭᐅᑦ, ᐅᑯ ᓂᐃᐊᑦᓯᐊᖅ, >ᑐᔪ ᑐᑲᓕᵇ, ᓴᵒ ᓕᐃᔨᑦ.

<u>ᓇᐃᓐᓇᖅᓯᒪᔨᖅ ᐅᖅᑲᑦᑐᓗᓴᕐᑎᐅᑎᐅᔨᑦ ᒥᖅᓴᓗᓂᑦ:</u>
ᑕᑯᕐᓐᓇᖅᑎᑎᓂᒧᑦ (ᑎᑎᖅᑲᒧᑦ ᐃᓕᒋᔭᐅᔨᖅ 1), ᐅᒪᔨᓕᓪᓚᐊᓕᑦ ᓂᔨᓕᖅᑕᕈᓕᐅᖅᑐᑦ ᐊᒪᓗᓐ ᐊᐱᖅᓯᐃᔨᓐᑐᒃ ᑐᑭᓯᐊᕐᑎᐊᕈᒪᔨᓐᑐᒃ ᖃᐅᔨᔭᕐᓂᒧᑦ ᐊᑐᖅᑕᐅᔨᑦ ᐊᒪᓗᓐ ᖃᓄᐃᓕᓴᓂᐅᔨᑦ. ᐅᒪᔨᓕᓪᓚᐊᓕᓕᑦ ᓇᒪᓴᕆᓕᐅᖅᑐᑦ ᖃᐅᔨᔭᕐᓂᒧᑦ ᖃᓄᐃᓕᓴᓂᐅᔨᓂᒃ ᐊᔨᔪᑎᒪᒋᑦ ᖃᐅᔨᔭᕐᓂᓕᕆᓂᖅᓂᒃ ᓄᑲᐅᑦ.

ᐅᒪᔨᓕᓪᓚᐊᓕᓕᑦ ᑐᑭᓯᐅᒪᔨᐊᑎᒃᑲᓐᓂᕐᓂᒃ ᑐᓂᓯᓚᐅᖅᑐᑦ ᓇᑭᓐᓄᓚᐊᕆᔭᕐᓂᓕᒃ ᐅᒥᖏᒪᐃᑦ. ᑕᐃᒪᓐᓇᓂᒃ 1980-ᖏᓐᓂᒃ, ᐅᒥᖏᒪᑕᕐᑎᖅᑲᓚᐅᖅᓂᑦᑐᑦ Boothia ᓄᑲᐊᓂᑦ, ᐅᖅᑎᓚᐅᖅᑐᑦ ᐅᒥᖏᒪᐃᑦ ᑕᐅᑭᓐᓇᓚᐅᖅᑎᖅᓂᐅᓐᓄᑦ Prince of Wales-ᒥᑦ ᐊᒪᓗᓐ Somerset ᖃᑭᖅᑕᒥᑦ. ᑭᓯᐊᓂ, ᐅᔪᓗᐅᓕᖅᑐᖅᑎᖅ, ᐊᓯᐊᓂᓐᓇᕐᓂᖅᑐᖅᑎᖅ, ᓯᑦᓗ ᐱᖏᓐᓇᖏᓂᒃ ᐊᒪᓗᓐ ᓂᒋᐊᓂᒃ ᑕᓚᐅᔪ. ᐅᖅᑲᐅᓯᖅᑕᓚᐅᖅᒥᔪᒃ Boothia ᓄᑲᐊᓱᓂᑐᑦ ᐅᒥᖏᒪᐃᑦ ᓄᓇᖅᑲᓯᓪᓚᕆᔭᐅᓂᖏᓐᓄᑦ ᐅᒥᖏᓖᓂᒃ ᓄᓐᓚᐅᖅᑲᑎᑎᖅᑕᓚᐅᕐᓂᖏᓐᓄᑦ.

ᐅᒪᔨᓕᓪᓚᐊᓕᓴᓄᑦ, ᐊᒥᓯᓇᓚᐊᖅᓂᐅᔨᑦ ᐅᒥᖏᒪᐃᑦ ᐃᓯᓲᓴᑎᓗᓚᐅᔨᖅ, ᐃᒪᓐᓇᐅᑲᓗᖏᓚᑦ, ᐊᒥᓯᓇᓚᐊᖅᑲᑕ ᐅᒥᖏᒪᐃᑦ ᐊᒥᓯᔪᖏᓂᖅᑎᑎᓂᓂᐊᖅᑐᖅ ᑐᓂᑐᖅ. ᓄᓇᓕᒥᑦ ᖃᐅᔨᒪᓂᖅᑐᑦ ᐅᒥᖏᒪᐃᑦ ᐊᒥᓯᓇᒧᐊᖅᑲᑕ ᐃᓇᓐᓂᖏᓐᓄᑦ ᑐᓂᑐᖅ. ᐅᒪᔨᓕᓪᓚᐊᓕᓕᑦ ᑕᐅᑐᖅᑕᕐᓂᖏᓐᓄᑦ ᖃᐅᔨᒪᓂᖅᑐᑦ ᐊᒥᓯᓇᓚᐅᖅᑎᖅᓂᐅᓐᓄᑦ ᓄᓇᖅᑲᑎᑎᖅᑕᓚᐅᕐᓂᖏᓐᓄᑦ ᐊᒪᓗᓐ ᐊᐅᓕᕐᑎᓂᖅᓂᐅᓐᓄᑦ ᐊᒥᓯᓇᓚᐅᖅᑎᖅᓂᐅᓐᓄᑦ. ᐊᒪᓗᓐ, ᑐᖅᑐᓂᒃ ᖃᖅᑲᐅᔨᓕᖅᑐᖅᑎᖅᑕᐅᓕᕐᑎᑦ Boothia ᓄᑲᔪᑦ ᑐᖅᑐᖏᓂᑦ ᖃᖅᔪᐅᓂᖏᓂᑦ ᐃᓕᒪᖅᓗᐊᖅᓯᒪᓕᒃᑲᒪᑦ, ᐊᓯᖏᓂᑦ ᑐᖅᑐᐃᑦ ᖃᖅᑲᐅᔨᓕᐅᖅᓯᒪᓕᑕ ᐊᒥᓯᔪᖏᓂᖅᑐᖅᑎᖅᓂᐅᓐᓄᑦ ᐱᖏᓐᓇᖅᑲᓯᔪᖏᓕᕆᓂᖅᓂᐅᓐᓄᑦ ᖃᑭᑎᒥᐅᑦ.

ᐱᔨᔪᑎᑭᓕᓗᑯ ᐅᒥᖏᒪᑕᐅᖅᑎᖅᑕᕐᓂᖏᓐᓄᑦ ᐃᓕᓴᕐᑎᓂ, ᐅᒥᖏᒪᑕᐅᔨᔪᓇᐅᑦᑐᑦ Ḷᓐᓇᐅᔨᖅ ᐱᓕᐅᔨᖅ ᐱᑕᖅᑎᐅᔨᑦ ᐅᑭᐅᖅᑕᕐᓂᑦ, ᐊᒪᓗᓐ ᓄᓇᑕᐅᔨᑕᓇᐅᔨᖅᑐᑦ ᐅᒥᖏᒪᑕᐅᔨᔪᓇᐅᑎᓂᒃ ᐱᓕᐅᖅᑐᑦ. ᐅᑭᐅᖅᓯᖅ, ᐅᒥᖏᒪᑕᐅᔨᔪᓇᐅᑎᓗᑕᓐᓂ ᐃᓐᓄᑕᐅᔨᖅᑕᕐᓂᖏᓐᓄᑦ, ᐊᒪᓗᓐ ᓄᓇᑕᒥᑎᐅᑦ ᐅᒥᖏᒪᑕᐅᔨᔪᓇᐅᑎᓂᒃ ᐱᓕᐅᖅᑐᑦ. ᐅᒪᔨᓕᓪᓚᐊᓕᓕᑦ ᐊᖏᑎᓕᑎᓪᓚᐊᓕᓕᑦ ᐅᒥᖏᒪᑕᐅᔨᔪᓇᖅᑕᕐᓂᖏᓐᓄᑦ, ᖃᓄᑎᖏᓂᓕᓗᑯᓗᒃ ᐊᒥᓯᔪᖏᓂᖅᑐᖅᑎᖅ ᐃᓐᓄᑕᐅᔨᖅᑕᕐᓂᖏᓐᓄᑦ (ᐅᒥᖏᒪᑕᐅᔨᔪᓇᖅᑕᕐᓂᖏᓐᓄᑦ 218-ᖏᓴᓗᓂᑦ). ᐅᒪᔨᓕᓪᓚᐊᓕᓕᑦ ᐊᑐᐃᓇᐅᓪᓚᕆᔭᑦ ᐅᒥᖏᒪᔪᖏᓇᖅᑎᑎᓂᐅᓐᓄᑦ ᐊᒥᓯᔪᖏᓂᖅᑐᖅᑎᖅᑕᕐᓂᖏᓐᓄᑦ ᐊᑦᑎᑭᐅᖅᓯᒪᓕᕐᓂᖅ, ᑭᓯᐊᓂ ᐅᒥᖏᒪᑕᐅᔨᔪᓇᖅᑕᕐᓂᖏᓐᓄᑦ 7-ᒥᑦ 8%-ᖏᓗᓂᓂ.

<u>ᐃᒪᓐᓇᖅᑐᑎᑕᐅᔨᖅ ᓄᓇᕗᒻᒥ ᑲᕈᒪᓕᐊᓐᓄᑦ:</u>
ᓄᓇᕗᒻᒥ ᑲᕈᒪᓕᐊᓐᓄᑦ ᐊᒪᓗᓐ ᑕᓃᕐᐊᕐᖏᑦ ᐅᒪᔨᓕᓪᓚᐊᓕᓕᑦ ᐊᖏᑎᓕᑎᓪᓚᐊᓕᓕᑦ ᐅᒥᖏᒪᑕᐅᔨᔪᓇᖅᑕᕐᓂᖏᓐᓄᑦ 7-8%-ᒧᑦ ᐊᐅᓕᕐᑎᐊᓂᖅᓴᐅᓂᐊᓂᕐᓂᖏᓐᓄᑦ ᐊᒥᓯᔪᖏᓂᖅᑐᖅᑎᖅᑕᕐᓂᖏᓐᓄᑦ ᐅᒥᖏᒪᐃᑦ Boothia ᓄᑲᐊᓂᑦ.

3.1.2 ᑕᓃᖅᓯᐊᖏᑦ ᐃᓄᓃᒃᑎᑐᑦ ᐅᖃᖅᑎᖅᑎᓂᖅᓗᑦ ᓇᐃᔭᖅᓂᒪᔪᖅ - 2019

<u>ᐱᓱᑦᑕᐅᔪᑦ:</u> ᐅᒥᖏᒪᐃᑦ ᓇᒧᖏᑐᐅᖅᑎᑦᑕᓂᖏᑦ ᐊᓪᒪᓃ ᐱᑕᖅᑐᐅᖅᑎᓂᖏᓐᓄᑦ ᐊᐅᓪᓚᑎᓂᖅ MX-08, Boothia ᓄᕐᐊᓂᑦ, ᐊᑎᓗ 2017. ᖃᐅᔨᐅᓂᖏᑦ ᐅᒥᖏᒪᐃᑦ ᐊᐅᓪᓚᑎᓂᖅᑎᑦ ᐊᒥᓗᖓᖅᓯᓪᓕᐊᑎᖅᑐᑦ.

<u>ᐅᕙᓃᖅ:</u> ᐃ>ᓇ 23, 2019

<u>ᑭᖑᓪᓕᐃᑦᑐᑦ:</u>
ᐊᖏᑎᓪᓚᓴᖅᑯᑦ: ᓀᓴ-ᒪᐃᓇ ᓪᓵᓄᖅ ᐊᓕᒪᓗ ᑭᐊᐃᐃᓐ ᒥᐊᑐᐃᓐ, ᑕᐃᐃᑦ ᐊᓇᐃᓗᓄᖅ

ᐅᒪᔨᓪᓚᓴᐊᓪᓕᑦ: ᓱᐧ ᐊᓯᖏᖅ

<u>ᓇᐃᔭᖅᓯᒪᔪᖅ ᐅᖃᓄᑦᓗᓇᖅᑕᑎᐅᑎᐅᔪᑦ ᒥᐅᓴᓄᑦ:</u>
ᑕᑯᕐᓴᖅᑎᑎᑐᓂᒥᒥᑦ (ᑎᑎᖅᑲᒧᑦ ᐃᓕᑭᕆᐅᔪᖅ 1), ᐃᓄᐃᑦ ᓂᔨᓪᓂᖅᑕᑕᓕᐅᖅᑐᑦ ᐊᓕᒪᓗ ᐊᐱᖅᓯᐃᔨᓗᑎᒃ ᑐᑭᓯᐊᑎᐊᑭᒪᔨᓗᑎᒃ ᖃᐅᐃᓕᓴᐃᓂᒧᑦ ᐊᑐᖅᑕᐅᔪᑦ ᐊᓕᒪᓗ ᖃᓄᐃᙱᓚᑦᑐᓂᖅᑦ. ᓄᓇᓪᒥᐅᑦ ᓴᓖᒪᓴᓪᓕᐅᖅᑐᑦ ᖃᓄᐃᓕᓴᐃᓂᒧᑦ ᖃᓄᐃᙱᓚᑦᑐᓂᖅᑦᑐᓂᒃ ᐊᓕᐅᑎᓪᓚᕆᐊᓂᑦ ᖃᓄᐃᙱᓚᑦᑐᓂᖅᑦᑐᓂᒃ ᓄᓇᒥᑦ.

ᒪᖁᑉᐳᒃ ᓄᓇᓪᒥᐅᑦ ᑕᑯᐊᖃᓐᓂᕐᒪᔪᑦ ᐅᐊᑯᐃᖏᓚᐊᖅᑯᑦ ᒪᖁᐃᑕᐅᔪᓐᓇᖅᓂᒪᓐᓄᑦ ᐅᒥᖏᒪᐃᑦ, 90%-ᖏᑦ ᐅᕐᓀᔭᓐᓂᑦ ᐅᒥᖏᒪᖃᐅᔪᓐᓇᖅᑐᑦ ᑭᓚᓕᖁᖏᑦᓗᓕᑦ. 218 ᐅᒥᖏᒪᖃᐅᔪᓐᓇᖅᑐᑦ ᐊᑦᑭᑭᖃᐅᔪᒪᔪᖅ ᐃᓚᒪᖏᑦᑐᕆᓚᖏᑦ. ᐊᒥᓱᓂᖅᓴᓂᒃ ᐅᒥᖏᒪᔭᓐᓇᖅᑎᑕᐅᔭᕆᐊᖅᑎᖅᑐᑦ ᐅᒥᖏᒪᐃᑦ ᐊᒥᓱᓕᐊᖅᓯᓚᐅᓴᖅᑎᖅᑐᓂᒃ, ᑐᖅᑐᐃᑦ ᓄᓇᒋᑦᑐᓐᓇᖅᓂᐊᖏᒪᒃ. ᐅᖅᑎᐅᓱᓚᐅᑦᒥᔪᖅ ᑭᓚᓕᖅᐳᐅᔪᑦ ᐅᒥᖏᒪᔭᓐᓇᐅᑎᓄᑦ, ᐊᓕᒪᓗ ᐅᒥᖏᒪᔭᓐᓇᐅᑎᑕᑦᓂᖅᓂ ᐊᔪᓐᓇᖅᓱᖅᑕᑦᑐᓐᓄᑦ. ᐊᔪᖅᓇᖓᑕᐅᖅᑎᖅᑕᑦᑐᖅᑐᑦ ᒪᖅᑎᐃᑎᒃᑕ ᓴᓐᑎᑦᑐᓂᒃ ᐊᓕᒪᓗ ᒥᒃᔪᓂᒃ ᖃᖅᔪᖅᓂᑦ, ᐊᓕᒪᓗ ᖃᑭᐃᖅᑎᑕᖅᓂᖅᓂᒃ ᓴᑦᓯᐊᖅᑎᑦᑐᓂᒃ ᐃᓄᐃᑦ ᓄᓇᒥᑦ, ᖃᑭᑭᖅᑑᓐᓇᒥᒥᑦ ᐃᓄᐃᑦᑐᑦᑐᓂᒃ ᑐᖅᑐᐃᑦᑕᐃᓇᓂᖅᑦ. ᐃᓄᐃᑦᑐᑦᑐᖅᑐᑦ ᒪᖅᑎᐃᑎᒃᑕ ᐊᑐᖅᑲᖓᒪᑦ ᑲᓐᓂᑦᑐᓐᓂᒃ ᑐᖅᑐᐃᑦᑕᐃᓇᓂᖅᑦ. ᐃᓄᐃᑦᑐᑦᑐᖅᑐᑦ ᓄᓇᒥᑦ, ᖃᑭᑭᖅᑑᓐᓇᖅᑐᑦ ᑐᓐᓄᓪᓕᐊᑦ ᑐᖅᑐᑭᑦᒪᖅᑕᑦᑕᒪᓕᑕ. ᑕᒪᒃᑯᐊ ᐃᓱᒪᓕᑐᐅᔪᑦ ᐅᒥᖏᒪᔨᐅᓕᖅᑐᑦ ᑕᐃᒪᓐᓇᐃᓕᐅᑎᓂᐅᔪᑦ ᓂᕿᑎᓂᖅ ᓴᖅᑭᑐᐃᓐᓇᓕᖅᓂᖅ. ᐱᓕᕆᓂᐊᓕᒪᓗᒃᑐᑦ ᐅᒥᖏᒪᔨᐅᓕᖅᑐᑦ ᑕᐃᒪᓐᓇᐃᓕᐅᑎᓂᐅᔪᑦ ᐊᑦᑭᓇᖅᓯᖏᑦᑕᖅᑐᓂᒃ ᐃᓚᒋᔭᐅᓂᖅᑐᑦ ᐃᓄᐃᑦᑐᑦᑐᖅᑐᑦ ᑐᓂᓯᓂᖅᑐᑦ ᐃᓄᐃᑦᑐᑦᑐᖅᑐᑦ ᐊᑦᑭᓇᖅᓯᖏᑦᑕᖅᑐᓂᒃ ᐃᓚᒋᔭᐅᓂᖅᑦ ᐊᑦᑭᓇᖅᓯᖏᑦᑕᖅᑐᓂᒃ ᐃᓄᐃᑦᑐᑦᑐᖅᑐ

4.0 ᐃᓱᓕᓐᓂᖅ - ᐱᓕᕆᐊᖃᑎᒌᒍᑎᓕᖅᑐᑦ

ᐊᐢᑎᑎᓕᕆᔨᕼᑦ ᑭᖓᔪᓕᖅᓵᕼᑯᖓᓕᖅᑎᓂᐅᐊᖅᑕᖓᓕ ᐅᒥᖓᒪᐃᑦ ᒥᐦᓴᓄᑦ ᐅᓂᑦᑲᖅ ᐅᖅᑲᓱᐅᐳᑦ ᑲᑎᒪᑎᑎᓂᐅᓂᖅᒥᑦ ᐃᓕᐅᑎᑦᓗᑎᑦ ᐊᒻᒪᓗ ᑐᓂᐅᖅᑲᐅᑎᑦᓗᒍ ᐊᐅᓕᑦᑎᔨᐅᖅᑲᑎᑎᑲᖓᑦᓂᓄᑦ. ᑕᐃᒪᓐᓇᓂᑦ 2017-ᒥᑦ, ᖃᑦᕿᓱᐅᓂᖓᓂᓄᑦ ᖃᑲᐅᔨᓴᐃᓂᖅᓯᑦ ᖃᑲᓄᐃᖏᓂᐅᐸᑦ ᐅᖅᑲᐅᓱᐅᓱᒪᓕᖅᑐᑐᑦ ᖃᑭᑎᖅᒥᐅᑦ ᐅᒪᐸᓕᕆᔨᕼᑯᑦ ᐊᖅᑭᒍᑕᒻᒧᖅ ᑲᑎᒪᑎᑎᓂᐅᓂᖓᓂᑦ 2018-ᒥᑦ ᐊᒻᒪᓗ ᐃᓐᓇᓗ ᐅᒪᐸᓕᕆᔨᐊᓪᓚᕼᑯᓄᑦ, ᐃᓄᓗᑦᑕᓄᑦ, ᐊᓱᖓᓐᓄᑦᔭ ᐊᐅᓕᑎᑎᖅᑲᐅᑎᑦᓄᑦ. ᑐᑭᓱᐅᒪᔨᓴᑦᑐᑦ ᐅᒥᖓᒪᐃᑦ ᐊᒥᓱᖏᓕᕆᐊᖅᑎᖓᓂᓄᑦ ᐅᒥᖓᒪᐃᑦ ᒥᐦᓴᓄᑦ ᐊᐅᓕᑎᑎᓂᐅᒥᑦ MX-08-ᒥᑦ ᑐᕆᖅᑎᑎᑎᕐᓕᐅᓕᐅᖅᑐᖅ. ᐊᒻᒪᓗ, ᓄᓇᕗᒻᒥ ᐅᒪᐸᓕᕆᔨᓴᕐᐊᕼᑯᓄᑦ ᐊᒥᓱᖏᓕᕆᐊᖅᔭᓂᑦ ᒪᓂᓇᓕᐅᕐᖅ ᐅᒥᖓᒪᔪᖏᓇᐅᑎᐅᕐᓂᑦ. ᑕᐃᒪᓂᓇᐃᓕᐅᑐᕈᒪᓂᖅ ᐃᑲᕐᖅᑐᑐᕐᕕᐅᕐᖅ ᑕᓗᕐᐊᐊᖅᒥᑦ ᐅᒪᐸᓕᕆᔨᐊᓪᓚᕼᑯᓄᑦ (ᑎᑎᖅᑲᖅᓗᑦ ᐃᓕᑎᕆᐅᕐᖅ 2).

5.0 ᐅᒥᖏᓂᒃ ᐊᐅᓪᓗᓂᒃ MX-08-ᒥᑦ ᓇᓪᐅᑦᑕᖅᑕᐅᔪᑦ ᖃᓄᐃᖏᑎᒋᑦ ᐊᓕᓐᓄ ᐊᐅᓪᓗᓂᔭᓕᑦ ᐃᓚᖓᔾᔪᑎᑕᐅᔪᑦ ᑲᑎᒪᔨᑦᑐᑦ

ᐅᓚᔪᓪᓕᐊᓯᒃᑯᓂᑦ ᐅᖃᖅᑲᑎᖃᖅᓂᒃ

ᐅᑦᓗᖅ: ᐃᐳᓕ 23, 2019 7-ᒥᑦ 9-ᓗᑦ ᐅᓂᓄᒃᑯᑯᑦ ᖃᑎᖅᒥᐅᓂᑦ

ᐱᔪᑕᐅᔪᖅ: ᐅᓚᔪᓪᓕᐊᓯᒃᑯᓂᑦ ᐅᒥᖏᓚᐃᑦ ᒥᒃᓴᓄᑦ ᑲᑎᒪᓂᖅ, MX-08

ᐅᓚᔪᓪᓕᐊᓯᑦ ᐅᖃᐅᔨᖏᑦ:

- ᐊHᐃᐊᖅ ᑐᐅᑐᐃᑦ ᐅᑭᐅᒃᑯᑦ ᑕᓗᖅᔪᐊᑦ ᖃᓂᖏᓱᓂᖃᑦᑕᕐᓱᖅᑐᑦ
- ᐃᒃᓯᕈᐅᑕᖅ: ᑲᑎᒪᓂᖅᔪᐊᖅᑭᒪᔪᔨᑦ ᑐᐅᑐᐃᑦ ᐱᔪᑎᑎᓗᑎᑦ
- ᓇᓗᓕᖅᓯᒪᓂᖃᖅᑐᑦ ᐊHᐃᐊᖅ ᐊᓕᒪᓗ Boothia ᑐᐅᑐᕈᑕ ᒥᒃᓴᓄᑦ
- ᐃᒃᓯᕈᐅᑕᖅ: ᐊᒥᓯᖏᖅᔪᐊᖅᓯᒪᓂᐅᔪᑦ ᐃᓕᓴᖅᑯᓯᖏᓐᓄᑦ ᑕᐃᒪᓕᐃᓐᓇᕐᓂᖅᓂᑦ
- ᐃᓯᐊᓗᑎᑐᑎᓂᑦ ᐅᒥᖏᓚᐃᑦ ᐊᒥᓯᖏᓪᓗᑎᒃ ᐊᓕᒪᓗ ᑐᐅᑐᐃᑦ ᐊᒥᓱᔭᓐᓂᖅᓗᑎᒃ
- ᐅᒥᖏᓚᐃᑦ ᑕᒪᐅᖏᓪᓕᐊᑕᖅᑐᑦ ᓇᑭᓂᖅᑕᖅᑐᑦ
- ᑕᓗᖅᔪᐊᖅᑎᑦ ᓯᑎᐱᓗᐅᓕᖅᑐᑦ ᑐᐅᑐᐃᑦ ᒥᒃᓴᓄᑦ ᖃᐅᔨᐱᒃᓴᕐᒪᐊᐅᓚᖅ
- ᑐᐅᑕᖅᓯᑎᑦ ᐅᖃᓪᓚᐅᑦ Boothia ᓇᐅᐊᓂᑦ
- ᓇᖅᓗᐅᖅᓵᓚᑦ ᓇᓂᓗᒃᑯᖅᑐᑦ Boothia ᓇᐅᐊᓂᑦ
- ᐃᒃᓯᕈᐅᑕᖅ: ᑐᐅᑐᐃᑦ ᓄᖅᑕᑕᓂᖅᑎᑦ ᐊᓯᐊᖏᓱᓕᓕᖅᑐᑦ ᖃᖏᓱᒃᑯᓂᖅᑎᑦ, ᓇᒧᖏᓱᐅᖅᑕᑕᓂᖅᑎᑦ
- ᐃᓯᐊᓗᑎᑎᑎᐊᖅᑕᐅᔨᑦ ᓯᐅᓂᒃᓱᒥᑦ ᑐᐅᑐᐃᑦ ᐊᒥᓯᖏᓪᓗᐊᓂᓂᖅᑎᓪᓗᑦ
- ᐃᒃᓯᕈᐅᑕᖅ: ᐊᒥᓯᓄᖅᓱᓂᒃ 6%-ᒥᐅᖏᓂᑦ ᓇᓪᐅᑦᑕᖅᑕᐅᔪᓂᒃ ᐱᔪᓚᔪᑦ (ᐅᒥᖏᓚᒍᖏᓇᐅᑎᑦ)
- ᓄᓇᓕᖏᒥᑦ ᓂᖅᑭᓕᓐᖃᖅᑕᐅᒪᔪᑦ ᐅᒥᖏᓚᖓᓄᑦ
- ᐊᒥᓯᐃᑦ ᑐᐅᑐᐃᑦ ᓂᖅᑭᖏᑦ ᓯᕆᖅᑎᑕᖅᑐᑦ; ᓴᓪᑐᓂᓂᑦ ᑐᓂᓯᖏᓚᔪᓂᑦ ᖃᑭᒪᐃᖅᑎᑕᖅᑐᑦ ᓄᓇᓗᑦ, ᑐᓴᓪᓕᖏᓐᓄᒃ ᐱᖅᑎᑕᖅᑐᑦ ᓂᐅᕐᖃᐃᓐᓇᓕᒫᓂᑦ ᖃᑭᖅᑲᓪᓚᐅᓄᑦ

ᐃᓄᓗᖅᑕᓂᖅ ᐅᖃᖅᑲᑎᖃᖅᓂᖅ

ᐅᕙᓗᖅ: ᐃᐱᓇ 23, 2019 7-ᒥᑦ 9-ᒧᑦ ᐅᵒᓄᐦᓴᐦᑯᑦ ᖃᑎᖅᒥᐅᓂᑦ

ᐱᓵᖅᑕᐅᔪᖅ: ᐅᒪᔪᓕᓚᓵᐊᑦᓴᐦᑯᵒᓂᑦ ᐅᒥᖏᒪᐃᑦ ᒥᐦᓴᓄᑦ ᑲᑎᒪᓂᖅ, MX-08

ᐊᐅᑎᓕᓚᓵᐦᑯᵒᓂᑦ:

- ᐱᓕᓚᖃᑎᑎᕐᓗᑎᑦ ᓄᓇᓕᖏᐅᑦ, ᖃᐅᐅᓴᐱᐃᒋᐊᓕᐅᖅᑐᑦ Boothia ᓄᑭᐊᓂᑦ ᖃᐅᐅᓂᐊᖅᓗᑕ ᖃᐅᔨᓂᖅ ᐅᒥᖏᒪᑕᖃᖅᓯᒪᖏᑦ. MX/08, Boothia ᓄᑭᐊ, ᑭᖓᑦᓕᖅᐸᒥᑦ ᓇᓕᐅᑦᑕᖃᐅᓕᐅᖅᑐᖅ 2006-ᒥᑦ, 1100-ᖏᔪᓂᑦ ᐃᓴᐃᓕᐅᖅᑐᑦ 158 ᐃᵒᓇᐅᔪᑦ ᐅᒥᖏᒪᐃᑦ ᑲᑎᔪᐱᑎᓂᒋᑦᓴᓗᑎᒃ. ᖃᐅᐅᓴᒪᑦᓴ, ᑕᓗᖅᔪᐊᖏᐅᑦ ᐃᓱᒪᓕᑎᑎᖃᖅᓯᒪᔪᒪᓕ ᐅᒥᖏᒪᐃᑦᓗ/ᑐᐅᑐᐃᑦᓗ ᑲᑎᖃᑦᑕᐅᑎᓂᖏᵒᓂᒃ ᐊᒪᒪᓗ ᐊᒥᓱᐃᑦ ᑕᑯᓱᒪᑦᓴᑎᑎᒃ, ᓱᓇᐅᕐᓂ, ᑕᐃᒪᖏᓂᑦ 2006-ᒥᑦ, ᐅᒥᖏᒪᐃᑦ ᐊᒥᓱᖏᓪᓕᐊᓱᒪᑕᖅᓯᒪᒪᑕ. ᑕᐃᒪᵒᓇᒧᑦ ᐃᓴᐃᑎᓕᐅᖅᑐᑦ ᐊᑭᓱ 5-ᒥᑦ 13-ᒧᑦ, 2017-ᒥᑦ. ᐃᓴᐃᓂᖅᒥᑦ, ᐃᓴᓕᐅᖅᑐᒍᑦ 702 ᐃᵒᓇᖅᓂᒃ ᑲᑎᔪᐱᑎᓂᖅᒥᑦ. 41%-ᖏᔪᑦ ᑲᑎᒪᔪᓂᑦ ᑕᑯᔫᑎᖏᓂᑦ, ᖃᑭᓂᖅᑐᑯᓗᕐᔪᑦ, ᒥᑭᔫᑦᓴᑎᒃ ᑲᑎᒪᓕᐅᖅᑐᑦ ᒪᖅᑮᓴᑎᒃ 7-ᓄᑦ ᑲᑎᒪᖅᑕᖅᓯᑎᒃ ᐊᒪᒪᓗ ᐊᒥᓱᓂᖅᓴᐃᑦ 19-ᖏᔫᓴᑎᒃ. ᑕᐃᒪᵒᓇᒧᑦ ᖃᐅᐅᓕᔪᒍᑦ ᐅᒥᖏᒪᐃᑦ ᑲᑎᒪᔪᑦ ᐊᒥᓱᓂᖅᓴᖅᑕᓕᐅᔪᵒᓇᖅᑐᑦ ᐅᑭᐅᖏᑎᓕᑐᒍ, ᑲᑎᖅᑕᖅᑐᖅᑐᖅᓴᖅᓯᒪᑕ ᐅᑭᐅᖅᑯᑦ. ᐊᒥᓱᓕᕆᓂᖅᓴᖅᑐᑦ ᐃᓕᓴᐅᑎᖅᑕᖅᑐᖅᓴᖅᑐᑦ. ᐊᓂᖅᖅᑎᑦ ᐃᵒᓇᓂᑦ 14%-ᖏᔪᓕᐅᖅᑐᖅ, ᑕᐃᒪᵒᓇ ᐱᐅᔪᓚᓗᒃ ᐅᒥᖏᓐᓄᑦ, ᐃᵒᓇ ᖃᓄᐃᖏᕐᑐᑦ. ᐃᓕᖏᑎᑦ ᐊᖏᕐᓯᒪᓂᖏᓐᓇ 2006-ᒥᑦ ᐃᓴᐃᓂᒥᑦ ᐊᓚᒪᓗ 2017-ᒥᑦ ᐃᓴᐃᓂᒥᑦ ᖃᓄᐃᑎᑎᑦ ᑎᑭᐅᑎᓂᖏᓐᓇ ᓂᕈᐅᑎᓂᖏᓐᓇ Boothia ᓄᑭᐸ. ᐃᓴᐃᓂᖅᓂ ᐃᐅᖏᓗ ᓄᖅᑲᓕᐅᖅᑐᖅ. ᐱᑕᖃᖅᓯᒪᔪᓂᖏᓐᓇᕐᓂᑐᖅ ᐅᒥᖏᓐᓂᑦ ᐃᓂ ᐃᓴᐃᖏᐅᖏᓐᓂᑐᒥᑦ. ᑕᐃᑲᓂ ᑕᐃᒪᐅᔪᓴᐅᔪᐅᓕᑦ, ᐃᓴᐃᖏᐃᑎᓴᒪᓂᑐᒍᐃᓕᐅᖅᑐᑦ ᐃᓂᓂᖅᓴᖅ. ᐃᓴᐃᖏᐃᐅᓕᔪᖅ ᑭᐸᓵᑐᑐᖅ MX/08-ᒥᑦ. ᖃᖅᑲᓱᐅᓂᖏᓐᓇ ᐅᒥᖏᒪᐃᑦ 3649-ᖏᔪᓕᐅᖅᑐᑦ. ᐊᒥᓱᓯᖅᐸᑦᓕᑎᐊᖅᑐᑦ ᑭᖓᑦᓕᖅᐸᒥᑦ ᐃᓴᖅᑕᐅᓕᐅᖅᒪᑕᒥᑦ. ᑎᓯᒥᓱᓴᑦ ᐃᓴᐃᓵᑐᑦ, ᐃᓴᖅᑲᑎᑦᑕᐅᑦ ᐅᒥᖏᒪᐃᑦ ᐊᒪᒪᓗ ᑎᑎᕋᖅᓗᑐᑦ, ᖃᐅᓂᖅᑕᑎᒥᔪᒍᑦ ᐅᒥᖏᓐᓂᖅᓴᖅ ᓴᓕᐅᓂᖅᑐᖅᓴᖅᑐᑦ ᐊᒪᒪᓗ ᑐᐅᑐᐃᑦ ᐊᐅᐸᓗᖅᑐᖅ ᐱᖓᓱᓂᖅ ᓴᓂᕋᓕᖅ, ᓱᑎᔪᖅ ᐅᔨᕋᖅᑎᑐᖅᑐᖅ, ᓇᓄᐃᑦ ᐊᒪᒪᓗ ᑐᓴᕐᐅᔪᖅ ᐊᐅᐸᓗᖅᑐᑦ ᐊᒪᒋᐃᑦ ᑕᑯᓕᐅᖅᑐᑦ.

- ᒪᑕᓕᖅᑲᑎᑎᑦᓗᑎᑦ ᓄᓇᓕᖏᐅᑦ, ᐃᓂᓵᓴᐃᑎᐊᓕᐅᖅᑐᑦ ᐅᒥᖏᒪᑕᖃᖅᓯᒪᒪᑕ ᑭᓱᐊᓂᑦ ᑕᐃᒪᑕᖅᑲᖅᑕᓕᐅᖅᑐᖅᑐᑦ 6%-ᒋᑎᓪᓗᒍ. ᐊᓱᐊᖏᓪᓕᖅᑎᒃ ᑕᑯᓴᐅᓇᔨᖅᑐᑦ ᐅᒥᖏᒪᑕᐅᔪᓇᖅᑐᓂᑦ, 66-ᓂᑦ 218-ᓄᑦ. ᑕᐃᒪᵒᓇᐃᓕᓪᓕᖅᑐᐃᑦᓴᐃᓇᖅᑕᐅᓂᑦ ᑭᓱᐊᓂᑦ ᐱᓴᑎᓪᓗᐊᑎᓇᓂᑦ ᑕᓐᓴᓂ ᑐᓴᖅᓯᐅᓗᓱ ᐱᖏᓚᐅᖏᖓᒪᑦ ᐊᒥᓱᓄᑦ ᐊᓱᖏᓪᓕᐅᔪᓄᑦ ᐅᒥᖏᒪᑕᐅᔪᓇᖅᑐᓄᑦ.

ᐊᐱᖅᓯᐅᑦ ᐊᒪᓗ ᑭᐅᔪᓯᐅᑦ:

ᓄᓇᓕᖏᐅᑦ:

- Boothia ᓄᑭᐊ ᐅᒥᖏᒪᑕᕐᑲᖅᐸᓪᓕᐅᖅᓱᒪᖏᑎᑐᖅ ᑭᓱᐊᓂ ᑕᐃᒪᑕᕐᑲᖅᑕᑕᕐᑲᖅᑐᖅ Prince of Wales ᕿᑭᖅᑕᒥᑦ. ᑕᐃᑲᓂ ᐅᒥᖏᒪᑕᕐᑲᖅᐸᓱᓕᕐᑲᐃᐅ?

ᐊᒡᑎᓕᓚᓵᐦᑯᑦ:

- ᐅᒥᖏᒪᐃᑦ ᑕᑯᔭᑦ ᒪᵒᓂ (Boothia ᓄᑭᐊᓂᑦ) Prince of Wales ᕿᑭᖅᑕᒥᖏᓪᓗᑲᖅᑐᖅᑐᑦ ᐊᒪᒪᓗ Somerset ᕿᑭᖅᑕᒥᑦ. ᐅᒥᖏᒪᑕᕐᑲᖅᑐᖅ ᓱᓕ Prince of Wales ᕿᑭᖅᑕᒥᑦ ᐊᒪᒪᓗ Somerset ᕿᑭᖅᑕᒥᑦ, ᐃᓕᓐᓂᑦ Boothia ᓄᑭᐊᓄᐊᓕᐅᖅᑐᖅᑐᖅᓴᖅᑐᑦ ᐊᒥᓱᓯᖅᐸᑦᓕᐊᓕᖅᓴᓗᑎᒃᓗ ᑕᐃᒪᓂ.

ᓄᓇᓕᖏᐅᑦ:

- ᐃᓕᓴᐅᖅᑲᓱᐅᖅᑐᒃ ᐅᒥᖏᒪᐃᑦ Prince of Wales ᕿᑭᖅᑕᒥᑦ?

ᐊᕐᑎᑕᓗᓴᖅᑯᑦ:

- 2016-2017-ᒥᑦ ᐅᓴᖅᑕᐅᓕᐅᖅᑐᑦ. ᐅᓴᐃᓕᐅᖅᑐᒍᑦ Prince of Wales ᐊᖏᓚᓗ Somerset ᖁᑭᖅᑕᓂᑦ 2016-2017-ᒥᑦ.

ᓄᓇᓕᖓᒥᐅᑦ:

- ᐅᓯᐅᖏᓚᓗ ᐊᒥᓯᔪᖓᓂᖅᑐᑦ ᑐᐊᒃᑐᑦ.

ᐊᕐᑎᑕᓗᓴᖅᑯᑦ:

- ᐊᑲᒃᑲ, ᑐᒃᑐᓂᒃ ᑕᑯᓕᐅᖅᑐᒍᑦ ᑲᑎᕆᐊᑎᓂᖅᒥᑦ ᑭᓯᐊᓂ ᑐᒃᑐᓂᒃ ᐅᓴᐃᖅᑭᑎᓪᓗᕐᒪ, ᖁᑲᐅᓚᖅᑭᑐᒍᑦ ᖁᖃᖏᑯᑯᑦ ᐱᐅᓯᖏᓪᓕᖏᑎᑦ ᐅᓴᓯᓂᐅᖅᑎᓪᓗᒍ ᐅᑎᕆᓂᖅᑐᐊᓕᓯᑯᑯᑦ ᓄᓇᓕᖓᔪᓂᓗ ᐅᑎᕆᓂᖅᑐᓗᒍ ᖁᑲᐅᓚᖅᑭᑐᒍᑦ ᐅᑎᕆᓂᖅᑐᓗᒍ ᖁᑲᐅᓚᖅᑭᑐᒍᑦ ᑐᒃᑐᓂᒃ ᖁᓕᖓᔪᖓᓂᖅᑐᑦ ᑭᑲᒪᒋᔭᖅ ᑭᒃᑲᓐᖏᕐᕋ ᑭᑲᒪᒋᔭᖅ ᑭᑲᒪᒋᔭᖅᑐᑦ.

ᓄᓇᓕᖓᒥᐅᑦ:

- ᒪᖁᑲᐃᑎᐅᔨᓗᖓᓂ, ᐅᑲᐱᕈᓯᑲᖓᓂ, ᐅᓴᖏᓪᑲᓚᐱᕐᓯᐅᑦ, ᒪᖁᑲᐃᑕᑐᒪᔪᖓᓂ. ᒪᑕᒃᓴᑎᑦ ᐅᓐᓂᖏᑦ, ᐊᒥᓯᒍᖁᑲᕐᓴᒃᓴᑎᑐᒃ ᐊᖏᓚᓗ ᐊᒥᓯᙳᖏᓐᓂᖁᑲᐅᑕᖁᑲᕐᓴᒃᓴᑎᑐᒃ ᐊᖏᓚᓗ ᒪᖁᑲᐃᑎᐅᔨᓗᖓᓂ ᖁᑲᐅᓚᖅᑭᑐᒍᑦ ᐅᓐᓂᖏᑦ.
- ᑕᐃᕐᓯᒪᓂ, ᐃᓐᓇᓗᐳᐅᔪᓂᑦ ᑐᓴᖁᑲᑕᒪᐅᖅᑐᒍᖓᓂ ᐃᓕᓴᖅᑎᑦ ᓂᕆᔨᑦ ᖁᐃᖁᑲᑕᖁᑲᐅᑦ ᐊᖏᓚᓗ ᐃᓕᓴᖅᑎᓐᓂᓕᓂ ᐅᑭᐅᓂᑦ ᐱᑕᖁᑲᓇᖅᑐᒃᑐᓂᑦ ᐃᓕᓴᖅᑎᓐᓂᓕᓂ ᐅᑭᐅᓂᑦ ᐊᒥᓯᙳᓗᒃᑐᒃ. ᐅᖁᑲᐅᓯᑐᖁᑲᖅ (ᐃᓇᐃᑦ ᖁᑲᐅᓚᒪᔭᑐᖁᑲᓕᖅᑎᑦ).
- ᐅᒪᔪᒥᓗᐊᓯᒃᑯᓐᓂᑦ ᐊᐅᓕᑎᑎᐅᑦ ᐱᔭᕆᓕᖅᑎᑦ, ᐅᑦᓗᓗᓴᒋᑦ ᐊᑐᖅᑕᐅᕐᓂᖅ? ᒪᖁᑲᐃᖁᑕᑎᓂᕆ ᐊᖏᓚᓗ ᓇᓕᒥᓂᖅ ᑕᐅᑐᒃᑕᐊᑦ ᑕᒪᑐᐱᓪᓕᑦ, ᑐᑭᓯᓂᖅᑎᑦ 218 ᐅᒥᓪᒪᔨᑎᒃᓕᑦ ᐅᒃᑯᐃᓪᓕᔪᑦ ᒧᐧᓇᓗ ᐅᑭᐅᓗᓴᐸᑎᓪᓗᒍ?

ᐊᕐᑎᑕᓗᓴᖅᑯᑦ:

- ᑕᐃᒪᑐᖅ ᐱᑎᐊᖁᓂᐊᕐᓂᖅ ᐊᓪᒪᐃᑦ, ᓯᑲᓪᓕᖁᑦ ᐅᖁᑲᐅᓯᓗᓴᑲᑦ ᖁᑲᐸᓯᐅᓂᖅᑎᑦ, ᐊᒥᓯᓴᓗᐊᖁᓂᐊᖁᑐᒍᑦ? 218-ᒧᐊᖁᑯᑲᑦ, ᐊᒥᓯᓂᓂᖅᓴᐅᑐᐃᓐᓇᓗᐊᖁᑲᖅᑐᑦ. ᐊᓪᒪᐃᑎᑐᖅ ᑕᐃᒪ ᐊᒥᓯᓴᓗᐊᖁᓯᐅᐅᓂᐊᖁᑲᓓᓂᖅ.

ᓄᓇᓕᖓᒥᐅᑦ:

- 218 ᐅᑭᒪᓄᑦ ᐃᔪᖁᓕᐊᖁᑐᖅ. ᐊᒥᓯᓄᑦ ᐅᑭᐅᓂᑦ ᐊᐱᖁᓯᐃᓐᓇᖅᓴᓂᒃ, ᓄᓇᓕᖓᒥᐅᑦ ᑕᖁᑲᓂᖅᑐᑦ. ᐅᖁᖅᓴᓯᓗ 1200-ᓗᐊᖓᔪᑦ ᐅᒥᖓᒪᐃᑦ ᐊᒪᓗᓗ ᐃᓐᓇᓗᐳᐅᑦ ᐅᒥᖓᒪᐃᑦᓗ ᐅᒥᖓᒪᐃᑦᓗ ᑐᒃᑐᐃᑦᓗ ᑲᑎᕆᐊᖁᑕᐅᑎᓂᐅᓂᖓᓂᑲ ᐅᑭᑦᓂᒃᑯᓕᖅᓯᓯ. ᖁᑲᓯᐊᖁᔪᐃᓐᓇᖁᓂᑲ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᖁᑲᖅᑐᖁᑲ ᐃᓐᓇ ᖁᑲᐅᓚᖅᑭᑐᒍᑦ 90%-ᓂᑲ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᑲ ᒪᓂᒪᓕᖁᑎᑯᔪᓯ. ᖁᑲᓯᐊᖁᔪᐃᓐᓇᖁᓂᑲ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᑕᖁᑲᖅᑐᖁᑲ ᐃᓐᓇ ᖁᑲᐅᓚᖅᑭᑐᒍᑦᓗᒍ ᓄᓇᓕᖓᔪᓕᒫᓂ ᓯᑲᓪᓕᖁᑦᐸᐅᑲᖁᑲᖅᑐᒍᑦ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᓇᖅᑐᑦ ᐃᓐᓇ ᖁᑲᐅᓚᖅᑭᑐᒍᑦ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᑲ ᐊᐱᖁᓯᐃᓐᓇᖅᓴᓂᒃ ᐃᓕᓴᖅᑎᑦ ᑎᑎᕋᖁᔭᓂᒃ ᐊᖏᓚᓗ ᓄᓐᔪᓯᒪᓕᖁᑦᐸᐅᑲᖅᑐᒃᑐᓂᒃ. ᐊᓕᓯᓂᑐᐃᓐᓇᖅᓴᓂᒃ ᐃᓪᓕᐃᓗᑕᖁᑦ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᑲ ᐅᒥᖓᒪᒐᕆᐊᖁᑲᖅᑐᖁᑲ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᑲ ᐃᓐᓇ ᖁᑲᐅᓚᖅᑭᑐᒍᑦᓗᒍ ᐊᓯᐊᓂᑕᐅᖅᑐᖁᑲᖅ ᑐᓴᐅᑐᒃᑕᐊᑦ ᐅᒥᖓᒪᔪᖓᓇᐅᑎᓂᑲ ᐃᓐᓇ ᖁᑲᐅᓚᖅᑭᑐᒍᑦᓗᒍ ᐊᓯᐊᓂᑕᐅᖅᑐᖁᑲᖅᑐᖁ

ᑐᒃᑕᕐᒥᓂᐅᕐᖏᓐᓄᑦ. ᑐᒃᑐᐃᓐᓇᐅᓐᖓᑕᑦ, ᐃᑭᓪᓗᕐᓯᒃᑯᕐᒥᑦ ᕿᓄᐃᓪᑐᑐᐃᓐᓇᓂᖅ ᓂᕆᔨᑎᓂᒃ ᐊᓚᒪᓗ ᕿᓄᑭᐅᑎᑕᕿᖓᓕᖅᑐᑦ ᐃᓕᐃᓗᑕᕿ 2-ᖏᓱᐊᖅ. ᐃᓕᐃᓗᑕᕿ 3, ᐃᓄᐃᑦ ᕿᑭᖅᑕᑉᓗᓕᑦ ᑐᓴᓄᑕᓂᒃ ᑐᒃᑐᕆᕿᑎᑕᕐᒥᑦ ᐊᓚᒪᓗ ᒪᕿᑲᐃᓂᓂᖅ ᐊᐅᖓᓕᑦ ᑐᒃᑐᓯᐅᕿᑎᑕᕐᒥᑦ ᑐᓴᓄᑕᓂᒃ ᑐᒃᑐᑕᑎᒃ ᕿᑭᒪᓗ ᒪᕿᑲᐃᓂᓂᖅ ᐊᐅᖓᓕᑦ ᐃᓕᐃᓗᑕᕿ. ᑕᒪᒃᑯᐊ ᐱᑕᕿᑭᖓᓂᕕᐊᕿᑎᖅᑕᕐᒪᑦ ᕿᓄᐃᓪᑐᐃᓐᓇᓂᑕᕐᒪᑦᑎᓪᓗᑕ. ᕿᑎᖅᓴᓕᒃ.

ᐊᐅᑎᓕᓚᖅᑯᑦ:

- ᐅᕿᑲᒪᒃᖏᓐ ᓯᖏᑦᓕᕐᕿᒥᒃ ᐱᓪᒪᓚᐅᑦᓗᓂᓗ ᐊᖅᓯᐊᓗᖅ ᕿᑯᖅᓴᒥᒃ ᐃᓯᓪᓗᓂᑎᑦᑕᖅᖏᖅᖏᓐᓄᑦ ᐅᕿᑲᐅᑎᑦᕿᑎᔪᑦ, ᐃᓯᓪᓗᓂᐅᑕᐅᓯᒪᓂᖅᑐᑦ ᐱᓪᒪᓚᐅᑦᑕᑎᐊᕿᑲᐅᑦᓗ ᐱᓕᕐᓗᓂᐅᑕᐅᑎᑦᑕᕿᑎᑦᓯᓂᕐᒥᒃ ᐊᑐᕿᑕᐅᔪᒪᓂᖅ ᐊᓚᒪᓗ ᐃᓄᐃᑦ ᓯᕿᐃᑐᐃᓴᖅᑕᕐᒥᑦ ᓗ ᓂᕿᑲᐃᓂᓂᖅᑐᑦ ᐊᑐᕿᑕᐅᔪᒪᓂᖅᑐᑦ ᓯᕿᐃᑐᐃᓐᓇᓂᑕᕐᒥᑦ ᐊᓚᒪᓗ ᐃᓄᐃᑦ ᐊᑐᕿᑎᑦᑕᕐᒥᑦ ᓯᕿᐃᓕᓴᕐᓂᖅ ᕿᑎᕿᓂᖅ. ᑕᒪᓐᓇ ᐱᐅᓚᐱᓐᓇᖅᑕᕐᒥᑦ ᐊᑐᕿᑎᑦᑕᕿᑎᑦᓯᓂᕐᒥᖅ ᐊᓚᒪᓗ ᐃᓄᐃᑦ ᐊᑐᕿᑎᑦᑕᕿᑎᑦᓯᓂ

ᐅᓇᑦᑕᐅᖅ ᐃᓱᒪᑎᓕᐊᕆᒃᑕᐅᖅᑐᑦ, ᓂᖅᐱᖅᑯᑎᐅᖅᑐᑦ ᖃᓄᐃᖕᑎᑕᕆᓕᐊᕆᒃᖃᑕᐅᖅᑐᑦ. ᐊᒧᑦᑲᕆᖃᑉᕋᓱᒃ, ᖃᓄᐃᑎᖅᑕᕐᕋᑦ, ᐅᑎᖓᒪᐃᑦ ᐊᒥᓱᖏᓪᓕᐊᑦᑐᑦᑐᑦ ᑐᓱᔪᕐᖓᓇᖅᑐᒍᑦᓗ ᐊᒧᑦᑲᕆᖃᑦᑕᒃᓴᓕᓗᑕ ᑭᓱᐊᓂ ᐊᒧᑦᑲᕆᖃᑦᑕᖃᖅᓗᑦᓂᓂ ᐱᐅᓂᕆᑲᓕᐊᕆᒃᓯᖃᑐᖅ, ᖃᐅᐅᓚᓂᐊᕆᓗᑦᓂᑕ ᖃᓴᓱᓂᒃ ᐅᒥᐅᖖᓕᐅᖖᓚᖅᑐᑦ ᐅᑭᐅᖅᑎᒥ.

- ᐅᖏᖅᑕᓂ 270-ᓂᒃ ᐅᒥᖖᒪᔪᖖᓇᐅᑎᓂᒃ ᐱᖅᑯᓲᖖᖅ.

ᐊᖏᑎᓕᓵᕆᔨᑦ:

- ᑕᒪᕐᓱᓗᒃᑑᖅ ᖃᐅᔨᖓᒋᑦᑕᒪᕐᓱ ᖃᐅᐃᖅᑲᐅᑦᓱ ᐅᖖᓄᒃ ᐊᓚᒪᓗ ᓂᕐᓕᕐᖖᓇᖅᑲᐅᑦᓱ ᑐᓴᖅᑎᖅᓯᒪᔨᑦ. ᐊᑐᑎᖅᑎᑕᐊᖅᑐᖅ ᐊᐅᓕᑕᑎᓂᖅᓴᑦ ᐱᖏᓕᓕᐊᓂᖅᑭᒃ, ᐊᒃᓱᐊᓗᓗᒃ ᖃᐅᔨᖖᓄᒃᑭᕐᕕᒃ. ᐱᓕᕆᖅᑎᑎᓂᐊᖅᕐᕋᑦ ᐅᒪᔨᓕᕆᓚᓪᓕᐊᓯᓱ ᑕᐃᒪᑐᓴᓗ ᐅᕐᓗᒥᓐᖅᑐᖅᑐᑦᓗ ᐊᑐᐃᓐᓇᖅᑐᑦ ᑐᓂᓗᓂᒃᑲᐅᑦᓱ ᐱᓕᕆᖅᑎᑎᓂᐊᖅᑐᖅᐅᓪᓗᑎ ᐃᓐᓇᑦᑐᓂᒃ ᑐᓂᓗᓂᒃᑲᐅᑦᓱ ᑐᓴᖅᑎᖅᓯᒪᔨᑦ ᐊᓚᒪᓗ ᐅᑎᖅᑐᓂᒃᑲᐅᑦᓱ ᑕᒪᓐᓇᖅᑐᑦ, ᐅᑎᖅᑐᓂᒃᑲᐅᑦᓱ ᐊᒃᓱᐊᓗᓗᒃ ᐊᓚᒪᓗ ᓴᖅᑭᑐᖅᑐᑦᓗ

ᑎᑎᖅᑲᒧᑦ ᐃᓕᑭᐳᐅᔪᖅ 1

ᑕᑯᕐᓇᖅᑎᑕᑎᓂᖅ 2019

ᐊᑕᓂᓕᖅ ᑕᑯᕐᓇᖅᑐᐅᔪᓐᓇᖅᑐᖅ:

Muskox distribution and abundance
Muskox Management Unit, MX08,
Boothia Peninsula

Taloyoak HTO

Lisa-Marie Leclerc
Government of Nunavut
Department of Environment

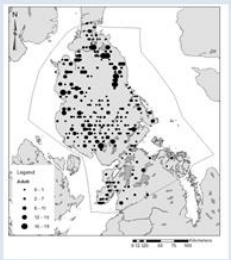
June, 2019

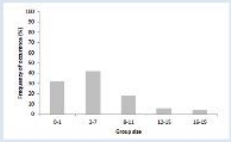
Survey Area

- MX-08, last estimate in 2006 was 1,100 muskoxen, with 158 adults on transect.
- Concerns, muskox-caribou interaction.

Monitoring

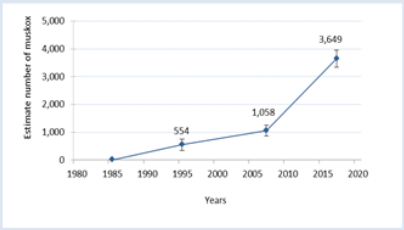
- Survey completed in August 05 to August 13, 2017
- 702 adults counted on transect.
- 41% of the group were small groups of 2 to 7 adults.
- The calves to adults ratio was 14%.



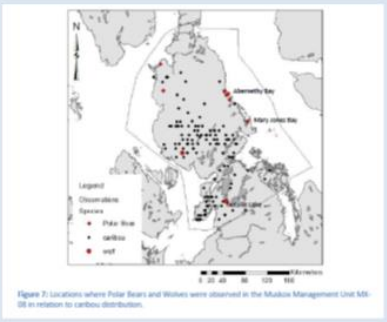


MX08- population estimate

There is an increase
in muskox number with an estimate of
3,649 ± 316 SE



Predators



Managements- Discussion

What is the HTO's management objective?

Are all tags being filled?

What is the community need?

ᑎᑎᖅᑲᒧᑦ ᐃᓕᑭᕆᐅᔪᖅ 2

ᖃᕆᓭᑕᐅᔾᔪᑎᑦ ᐃᑲᔪᖅᑕᐅᓂᖏᓪᒧᑦ ᑕᑐᖏᔪᐊᖏᑎᑦ ᐅᒪᔪᓕᓗᐊᓪᓚᑦᑯᓂᑦ

Smith, Caryn

From: Jimmy Oleekatalik <taloyoak@krwb.ca>
Sent: April 29, 2019 6:50 PM
To: Leclerc, Lisa-Marie
Cc: England, Kate; Methuen, Kevin; Smith, Caryn
Subject: RE: MX08 consultation- accomodation

Hi Lisa,

The Board is in agreement with the 275 for musk-ox tags. If you need more information please let me know.

Best regards,
Jimmy Oleekatalik,
Manager, SBHTA

-----Original Message-----
From: Leclerc, Lisa-Marie <LLeclerc@GOV.NU.CA>
Sent: Monday, April 29, 2019 1:59 PM
To: Jimmy Oleekatalik <taloyoak@krwb.ca>
Cc: England, Kate <KEngland@GOV.NU.CA>; Methuen, Kevin <kmethuen@GOV.NU.CA>; Smith, Caryn <CSmith@GOV.NU.CA>
Subject: MX08 consultation- accomodation

Hi Jimmy,

The Government of Nunavut, Department of Environment, discussed on the possibility to accommodate the recommendation of the HTO in having an increase of the harvesting rate of the muskox population MX08 of more than 6%.

With the comments of the HTO and the community members, we can support an additional increase with a recommendation of 7.5% / 275 animals as this seems reasonable by way of being: aligned with the HTO proposition, community comments, population increase, meeting increased demand for tags, while also reducing risk of meat wastage/wounding animals which is a current problem with the caribou.

Please let me know if the board support this recommendation.

All the best,

Lisa

ᑐᓂᔭᐅᔪᓐᓇᖅ

ᓄᓇᕗᑦ ᐅᒪᔨᓕᕆᔨᒃᑯᑦ ᑲᑎᒪᔨᖏᑐᓂᒃ

ᐱᔭᐅᕐᓯᓗᓕᒋᑦ

ᑐᓴᐅᒪᔨᐸᑎᓯᒪᔪᑦ: ᐃᓕᓴᕐᕕᒃ: ᐃᓱᒪᓕᐅᕈᑎᑦ: X

ᐊᕐᕈᖅᑕᐅᔭᕆᐊᓕᖅᑲ: 2017 ᐅᒥᖓᒪᐃᑦ (OVIBOS MOSCHATUS) ᓇᒃᓂᓂᖏᑦ ᐊᖓᒪᓗ ᐅᓄᕐᓂᖏᑦ, ᐅᒥᖓᒪᐃᑦ ᐊᐅᓚᕋᓂᕐᒃ ᐃᓗᓚᖏᑦ MX-09, ᕿᑎᓗᕐᒃᑐᑦ ᑯᖓᑕ ᐅᐊᓚᓂᐊᓂ

ᐅᕐᒃᐅᓱᕐᒃᓱᒪᓂᖏᑦ

- ᐅᒥᖓᒪᐃᑦ ᕿᑎᓗᕐᒃᑐᒃ ᐅᐊᓚᓂᐊᓂ ᐊᐅᓚᕋᓂᕐᒃ ᐃᓗᓚᖏᑦ (MX-09) ᐅᐊᓚᓂᕿᓂᕐᒃᓴᖓᓕᒫᑦ ᓄᓇᕐᒃᑲᒃᓱᒪᓕᒫᑦ ᐅᒥᖓᒪᐃᑦ ᓂᒋᕐᒃ ᐊᒥᐊᓚᑲᒥ.
- 80-ᓄᑦ ᐊᕐᕿᒍᓄᑦ ᖃᐅᔨᓴᖅᑎᓄᑦ ᖃᑎᖅᑭᓂᖓᒥ ᖃᐅᔨᓴᖅᑕᐅᓂᔪᑕᕐᒃᖅᑐᒃ MX-09 ᐅᒥᖓᒪᐃᑦ ᐅᓄᕐᓂᖏᑐᓄ, ᐃᓄᒋᐊᖏᓂᑕ ᐊᔪᓂᐅᓴᖅ ᖃᐅᔨᓴᖅᓴᓕᒫᓂᒃ ᑕᑭᓂᖅᓴᖓᑦᓗᓂ ᓄᓇᕗᓴᒥ.
- ᕿᐃ-ᓚᑦᓭᓴᐅ ᑯᖓᑕ ᑎᒋᓂ ᕿᑎᓗᕐᒃᑐᒃ ᐅᐊᓚᓂᐊᑕ ᐃᓂᐊᓂ ᑲᑎᖏᓐᓂᒃ ᓇᓚᐅᑦᑕᕐᒃᑕᐅᓕᒫᓂᒃ ᕿᑎᕋᖓᓂᖅᓴᖓᔪᓚᐅᕐᒃᑐᒃ 1988-ᒥ ᑲᑎᑦᓗᓂᑦ 1,805 ± 289 (ᑕᒪᓕᕐᓂᖓ (S.E.)) ᐅᒥᖓᒪᐃᑦ.
- 1994-ᒥ, ᐅᒥᖓᒪᐃᑦ ᐅᓄᕐᓂᖏᑦ ᑕᕐᖓᓂ ᐃᓂᐅᓚᒥᒃ ᐊᕋᑎᒃᓭᓚᕋᕐᓕᒫᑦ 540 ± 139 (ᑕᒪᓕᕐᓂᖓ (S.E.)) ᓂᕐᕕᑎᓂᒃ, ᓭᕿᓕᓗᐊᖏᑲᓇᕐᓴᓂ ᖃᐅᔨᓴᖅᓂᕐᒃᑲᒃᑲᓂᕐᓂᑎᑦᓗᒍ 2007-ᒥ (ᓇᓚᐅᑦᑕᕐᒃᑕᐅᓕᒫᑦᖅ 589 ± 121 (ᑕᒪᓕᕐᓂᖓ (S.E.)).
- 1990-ᓂ, ᐅᒥᖓᒪᐃᑦ ᐅᓄᕐᓄᓐᓂᖏᑦ ᑕᕐᖓᓂ ᐃᓂᐅᓚᒃ ᐱᔨᑎᖃᑦᓗᐊᑕᓚᐅᕐᒃᑐᒃ >ᖓᖏᑦᓂ ᑯᓚᖓᓂ (*Umingmakstrongylus pallikuukensis*) ᐊᓂᓂᐊᒥ ᓇᓂᓕᓂᒥ ᐊᓚᓗᓂ ᓂᕿᑎᖃᑦᑕᕐᒃᑐᑦ ᐊᒃᓕᓄᑦ ᐊᒃᑯᕐᑕᐅᓂᔪᓐᓇᑦ ᐅᓄᕐᓄᓕᓕᒫᓂᑕᐅᓂᖓᓂᓂ (*Ursus arctos horribilis*).
- ᐃᓚᓇᐅᔨᖅ ᐱᔭᐅᔨᓇᖅᑎᑕᐅᔨᑦ ᐊᖏᔪᓕᓚᑕᐅᔨᓇᖅᑎᑕᐅᓗᑎᒃ (TAH) MX-09-ᒧᑦ ᐊᕐᕈᖅᑕᐅᓕᐅᕐᒃᑐᒃ 20-ᒧᑦ ᐅᓕᐃ 1-ᒥᑦ, 2014 ᐅᒥᖓᒪᐃᑦ ᐱᔭᐅᔨᓇᖅᑎᑕᐅᔨᑦ ᐊᖏᔪᓕᓚᑕᐅᔨᓇᖅᑎᑕᐅᓗᑎᐅᑉ ᐱᓕᕆᔭᐅᓚᐊᖅᑐᑦ ᐊᑕᓐᓂ, ᐊᓚᓗ ᑕᓐᓇ ᐊᖏᕐᕋᔪᑕᒫᑦ ᑭᓴᓕᖓ ᑕᐃᒪᐃᓕᖏᓐᓂᒃᑯᓕᓕᒫᓂ ᐊᐅᓚᑦ 1994-ᒥ ᐱᑎᐊᖅᓯᔪᒍ.
- ᐊᖓᑎᓕᓚᓴᒃᑯᑦ (DOE) ᐱᓕᓚᖅᑎᖅᑲᖅᓯᒪᔨᑦ ᕿᑎᓗᕐᒃᑐᒃ ᓄᓇᕗᖓᓂ ᐊᓚᓗ ᐊᐅᓚᕋᓂᕐᓴᐅᖅᑎᑎᓄᑦ ᐃᓕᑎᐅᔨᓄᑦ (ᓄᓇᕗᑦ ᑐᙵᓚᐃᓕᒃᑯ, ᐊᖏᔪᓕᓚᑦ ᒥᑭᑎᐊᓚᑦᑐᓗᖅᑐᓚᓂᑦᓗᓂ ᖃᑐᓂᒥᐅᑦ ᐊᐃᓕᖅᑐᖅᓯᓄᑦ ᐅᒪᔨᓕᕆᔨᓄᑦ (ᓄᓇᕗᑦ ᑐᙵᓚᐃᓕᒃᑯ, ᐊᖏᔪᓕᓚᑦ ᒥᑭᑎᐊᓚᑦᑐᓗᖅᑐᓚᓂᑦᓗᓂ ᖃᑐᓂᒥᐅᑦ ᐊᐃᓕᖅᑐᖅᓯᓄᑦ ᐅᒪᔨᓕᕆᔨᓄᑦ) ᐊᖓᑎᓕᓚᓴᒃᑯᑦ (DOE) ᓴᓇᖅᑎᖅᑲᐃᓐᓇᕐᓂᒃᑯᑦ ᑕᒪᓐᓇᖅᑐᒥ ᐊᐅᓚᕋᓂᕐᓴᐅᖅᑎᑎᓄᑦ ᐃᓕᑎᐅᔨᓄᑦ ᐊᐅᓚᕋᓂᕐᒃᑐᒃ ᐅᓄᕐᓄᓕᓕᒫᑦ ᐅᒪᔨᓕᕆᔨᓄᑦ (MX-09) ᓴᐳᒪᓂᕐᓂᑎᑦᓗᒍ ᐱᐅᔪᓐᓇᖅᑐᓂ, ᐊᐅᓚᕋᓂᕐᒃᑐᒃ, ᐊᓚᓗ ᓂᕿᑭᖅᑎᐊᖓᓂᒃ ᐅᒪᔨᓕᕆᔨᓄᑦ ᐃᓕᓂᖅᑭᐃᓕᐅᕐᒥ ᐃᓅᓂᖅᓴᐅᓚᐅᕐᓂᒃᑯᑦ ᐊᓚᓗ ᓄᓇᕗᐅᕐᒥᒍ.

ᐃᓚᓇᐅᔨᖅ ᖃᐅᔨᓴᕐᓚᓂᖓᓚ

- ᐊᖓᑎᓕᓚᓴᒃᑯᑦ (DOE) ᐃᓱᒪᓕᐅᔪᓐᓇᖅᑐᒃ ᐅᒥᖓᒪᐃᑦ ᐃᓄᐊᓂᖓᓂᒋᑦ ᖃᐅᔨᓴᕐᓚᓂᖓᓂᓇᓂ ᕿᑎᕋᖓᓂᒃ ᐅᐊᓚᓂᐊᓂ ᑲᑎᖏᓐᓂᒃ, MX-09, ᑭᖓᔪᓂᕐᒃᑲᒃ ᖃᐅᔨᓴᖅᑕᐅᓂᐅᕐᓚᑕ ᐊᖓᑎᒍᓂᒃ ᕿᑎᕋᓂᒃ ᖃᐅᔨᓴᖅᑐᒍᓂᒃ.

- ᑐᓴᐅᑦ ᐱᑕᖃᑎᐊᖓᓂᓇᓂᒃ ᕿᑎᓗᕐᒃᑐᒃ ᖃᐅᔨᓴᖅᑐᒃ, ᐊᖓᔪᓂᕐᒃᑲᒃ ᖃᓄᓂᒃᑐᑦ ᓂᕿᑦᓚᒋᓚᐅᕐᓴᐅᑉᑲᒃ ᐊᓯᖏᓐᓂᒃ, ᓴᓇᓗ ᐅᒥᖓᒪᐃᑦ. MX-09 ᐅᒥᖓᒪᐃᑦ ᐃᓄᐊᓂᖓᓂᒃ ᑐᓂᔭᐅᔪᓐᓇᖅᑐᒥ ᖃᐅᔨᓴᖅᑕᐅᓂᐅᕐᓚᑕ ᐃᑲᖅᑲᐊᓂᓇᓂᒃᑯᓕᓕᒫᓂᒃ ᐊᐅᓚᕋᓐᓂᓇᓂᖓᓂᒃ ᐊᓚᓗ, ᑕᐃᒪᐃᓚᒪ, ᓂᕿᑭᖅᑲᓂᓇᐅᓇᓂᓚᐅᕐᒃᑐᒃ ᑭᓴᓂᓂᒃ ᐃᓄᐊᓂᖓᓂᒃᑯᓕᓕᒫᓂᒃ.

- ᐃᓄᒋᐊᖏᓂᓐᓂ ᖃᐅᔨᓴᖅᓂ ᐱᓕᕆᔨᐅᓕᕐᐳᖅᑐᖅ, ᖃᓄᐃᖅᑐᖅᓄᓇᓕᖏᓐᓄᑦ ᑐᑦᕿᕕᖃᖅᖢᓂᑦ, ᐊᖏᓗᓂ 25–ᒥ ᓂᑎᐱᓚ 2–ᔪᑦ, 2017 ᖃᐅᔨᓇᓂᖏᓂᑲᑎᑦ ᐅᓄᖅᓂᖏᑦ ᐊᓚᓐᓇ ᓇᒧᖏᓂᖏᑦ ᐅᒥᖓᒪᐃᑦ MX–09–ᒥ. ᖃᐅᔨᓴᖅᑎᑦᓯᓂᑦ, 87 ᐃᖏᓇᐃ ᐅᒥᖓᒪᐃᑦ ᑎᑎᕆᖅᑕᐅᓕᐅᖅᑐᑦ ᐊᕕᐸᓂᒪᔪᒥᑦ ᐊᓂᐃᓪᓚᒃ ᓇᓕᐅᑦᑕᖅᓂᓯᓯᓂᑎᑎᑦ ᐅᓄᖅᓂᖏᖏᓂ 539 ± 150 (ᑕᓚᕆᓂᖓᑕ (S.E.)) ᓂᕆᐊᑎᓂᒃ MX–09–ᒥ. ᐅᒥᖓᒪᐃᑦ ᐃᓄᒋᐊᖏᓂᖏᑦ MX–09–ᒥ ᓂᕆᑯᓂᓗᐊᖅᓂᒪᖏᑦᑐᖅ ᑕᐃᒪᖏᓂ 1994–ᒥ.

- ᐱᖃᐅᔪᖏᓇᖅᑎᑕᐅᔪᑦ ᐊᖓᔪᓇᓂᑕᐅᔪᖏᓇᖅᑎᑕᐅᓗᑎᒃ (TAH) 20–ᒋᓂᒪᔪᖅ ᑕᐃᒪᖏᓂ 1994–ᒥ. ᑭᓂᐊᓂ, ᐊᕆᕽᔪᓄ ᖃᑯᓂᓄ ᖃᐄᖏᑦᖢᓄ, ᒪᖏᓇᐅᔪᖅ ᑯᑕᐃᑦ ᓄᖓᑕᐅᓂᒪᖏᑦᑐᑦ, ᐊᒪᓚᓗ ᐅᓴᐅᑎᑦ ᖃᑎᖅᑑᓂᑦ 15–ᓂ ᐅᒥᖓᒪᐃᑦ ᓂᕆᑭᐅᓴᑕᐅᓂᒪᔪᑦ ᐊᕆᕽᔪᑕᒪᖅ.

ᑐᐱᐅᒪᑎᕆᑎᖏᑕᐃᖏᓇᕆᓂᖅ:

- ᑕᕽᓂᒪ ᖃᐅᔨᓴᖅᓂᐅᑦ ᐱᔭᖏᑦᓂᐊᓂᖓᑐ ᐊᒪᓚᓗ ᖃᑭᑎᒥᐅᓂ ᐅᒥᖓᒪᐃᑦ ᐊᐅᑕᕆᑎᓂᕐᔪ ᕿᖏᐅᐅᑎᒃ ᑲᑎᖓᓚᓗᑎᒃ ᑲᑎᒪᓂᖓᑎ ᐊᒪᓚᓗ ᐊᖓᔪᓂᖓᓕᒃᑐᖅ ᒥᑭᐅᐊᖏᓂᓗᐊᓂᓗᐊᖅᑐᓂᑦ ᑲᑕᐃᔨᖃᑎᒃᓴᓂᖏᓐᓂᑦ (HTO)–ᑲᑯᓄᓇ ᑐᓂᔨᖃᑎᒌᖓᔪᓂᒃ ᑲᑎᒪᓂᖓᑎ ᖃᑭᑎᒥᐅᓂ ᐊᕐᕿᑲᖅᓯᒪᔪᓄᓇᓕᖏᓐᓂᑦ ᐃᓕᓴᕆᓚᐅᖅᑯᓄᓇᓕᖏᓐᓂᑦ ᑲᑎᒪᓂᖓᑎ ᖃᑭᑎᒥᐅᓂ ᑲᑎᒪᓂᖓᑎ ᐊᕐᕿᑲᖅᓯᒪᔪᓄᓇᓕᖏᓐᓂᑦ ᐃᓕᓴᕆᓚᐅᖅᑯᓄᓇᓕᖏᓐᓂᑦ (KRWB AGM) ᓂᑎᐱᓚ 2017–ᒥ ᐊᒪᓚᓗ ᖃᑭᑎᒥᐅᓂ ᐊᐃᑐᖅᓯᒪᔪᓄᓇᓕᖏᓐᓂᑦ ᐃᓕᓴᕆᓚᐅᖅᑯᓄᓇᓕᖏᓐᓂᑦ ᑲᑎᒪᓂᖓᑎ ᖃᑭᑎᒥᐅᓂ ᑲᑎᒪᓂᖓᑎ ᐊᕐᕿᑲᖅᓯᒪᔪᓄᓇᓕᖏᓐᓂᑦ (KRWB AGM) ᓂᑎᐱᓚ 2018–ᒥ.

- ᐊᐃᕆᓚᓕ 2–ᒥ, 2019, ᑭᖓᓪᓕᕐᕿᖅ ᐅᓂᒃᑲᖅ ᐊᒪᓚᓗ ᐊᐅᓕᕆᓂᕆᔪᓂ ᐱᓇᓴᕐᓯᓕᐅᖅᑐᓂ ᐃᓄᒋᐊᖏᓂᓗᐊᖅᑐᖅ ᐊᖓᔪᓂᖓᓕᒃᑐᖅᑎᑦᓯᓂᑦ ᑕᑭᐅᐊᖏᓂᓗᐊᖅᑎᑦᓯᓂᑦ ᑲᑕᐃᔨᖃᑎᒌᖓᔪᓂᒃ ᑲᑎᒪᓂᖓᑎ (HTO)–ᑲᑯᓂ ᑲᑎᒪᓂᖓᑎᒥ. ᐊᖓᔪᓂᖓᓕᒃᑐᖅᑎᑦᓯᓂᑦ ᑕᑭᐅᐊᖏᓂᓗᐊᖅᑎᑦᓯᓂᑦ ᑲᑕᐃᔨᖃᑎᒌᖓᔪᓂᒃ ᑲᑎᒪᓂᖓᑎ (HTO)–ᑲᑯᓂ ᐱᓕᕆᓂᑯᓚᓕᐱᓕᕆᓂᖅ ᐱᓇᓴᕐᓯᓕᐅᖅᑐᓂᑦ ᐊᖓᔪᓂᖓᓕᒃᑐᖅᑎᑦᓯᓂᑦ (TAH)–ᒥ 20–ᒋᓪᓗᓂ ᐅᒥᖓᒪᐃᑦ ᐃᓄᒋᐊᖏᓂᓗᐊᖅᑐᖅ MX–09–ᒥ.

ᐃᓕᓴᐅᑎᓴᐅᓕᖅ:

- ᑐᕽᓂᒪᔪᓂᖓᑎᑦᓯᓂᕆᔨᒪᔭᖅᑐᓂᒃ, ᐅᓂᒃᑲᖅ ᐊᓯᔪᓴᖅᑕᐅᓕᐅᖅᑐᓂ ᐃᓕᓴᖅᖢᓂᑲᑎᒃ ᓄᓇᓕᖏᓂ ᐊᒪᓚᓗ ᐃᓄᐃᑦ ᖃᐅᔨᓴᖅᓯᒪᔪᖏᓂᒃ, ᓄᓇᓗ
    - ᐅᖃᐅᔨᓯᖅᑲᓚᖏᓐᓂᑦ ᑐᕽᓂᒪᔪᓂᖓᑎᑦ ᐃᓄᐃᑦ ᓇᓂᓯᓂᖏᓐᓂᑦ >ᐸᖓᓂᓐᓂ ᑯᒪᖓᓂ 1970–ᓂ.
    - ᑐᑭᓕᐅᓚᓯᑦ ᓄᓇᓕᖏᓂᒃᑲᖅᑐᓂᒃ ᐃᓕᓴᖅᖢᓂᑲᑎᒃ ᐃᓯᒪᒋᔭᐅᖏᓐᓄᓂ ᖃᖅᖢᓂᒋᐊᕐᕿᖅ ᐅᒥᖓᒪᐃᑦ ᓇᕐᓂᖅᑐᑎᒃ ᒥᖅᓴᓂ ᓇᒧᖏᓂᖏᓐᓂᑦ; ᐃᓕᓴᖅᑎᓯᒋᔭᐅᓯᑦ ᑯᓐᓂᒃᑎᖅᑕᐃᒋᑦᑕᐃᓕᓂᓂᕆᒥ (>ᐸᖓᓂᓐᓂ ᑯᒪᖓᓂ) ᓇᕐᓂᖅᑐᖅᑎᖅᑐᑦ ᐃᑭᓐᓂᖅᑐᓂᓐᓂ.
- ᐊᑐᓕᓯᑯᐅᓯᑦ ᓯᓂᔪᖅᑐᖅ ᐊᖓᔪᓂᖓᑎ ᒥᑭᐅᐊᖏᓂᓗᐊᖅᑎᑦᓯᓂᑦ ᑕᑭᐅᐊᖏᓂᓗᐊᖅᑎᑦᓯᓂᑦ (HTO)–ᑲᑯᖓᓐᓂ ᐊᓯᔪᖅᑕᐅᓕᐅᖅᑐᖅᑐᖅ ᐊᑐᓕᓯᑯᐅᓯᑦ ᖃᐅᔨᓴᖅᓯᒪᔪᓄᓇᓕᖏᓐᓄᑦ ᐊᖓᔪᓂᖓᓕᒃᑐᖅᑎᑦᓯᓂᑦ ᐊᐸᑎᓕᓂᔪᓂᑲᑯᑭᓕᐃᓕᓂᓂᕆᒥ.

ᐊᑐᓕᓯᑯᐅᓯᑦ

- *ᐊᐸᑎᓕᓂᔪᓂᑲᑯᑭᓕ ᐊᑐᓕᓯᑯᐅᓯᑦ ᓄᓇᖓᓕ ᐅᓕᔪᓕᓂᔪᓂᔪᐊᑲᑯᑭᓕ ᑲᑎᒪᓂᖓᑎᒥ ᐱᓇᓴᕐᓯᓕᐅᖅᑐᓂᑦ ᐊᖓᔪᓂᖓᓕᒃᑐᖅᑎᑦᓯᓂᑦ (TAH) 20–ᒋᓐᓂᖓᓕᓂᔪᓂᒃ ᐅᒥᖓᒪᐃᑦ ᐊᐅᓕᕆᓂᖅ ᐃᓗᓕᕆᓂᖅ MX–09–ᔪᑦ.*

ᐊᖏᓇᓂᖅᑎᑦ ᒥᑭᒋᐊᕐᓂᐊᖅᑎᑦᓗ ᑲᑐᔾᕕᖅᑎᑭᒍᓐᓂᐊᓂᑦ (HTO) ᑐᓴᐅᒪᑎᑦᑎᕕᑦᑕᐃᓇᕐᓂᖅ ᐅᒥᖏᒪᐃᑦ ᓇᒡᓘᓂᖅ ᐊᓪᓚᓗ ᐅᓇᕐᓂᖏᑦ, ᐅᒥᖏᒪᐃᑦ ᐊᐅᓚᑎᓂᖅ ᐃᓗᓚᖏᑦ MX–09

ᐊᐃᕐᓇᓚ 2019

ᐊᕐᕌᑎᓚᓵᖅᑯᑦ, ᓄᓇᕗᒻᒥ ᒪᕐᕉᒪᖅᑯᑦ ᐃᕐᑲᓗᐃᑦ, ᓄᓇᕗᑦ

ᐊᐅᓪᓗᓂᔭᕐᓂᖅ ᓇᐃᓐᓇᕐᓂᕐᓂᖅ

ᑭᖑᓂᖅᑐᐃᓕᐅᕐᓂ ᓄᓇᕗᑦ ᑲᑎᒪᔨᑦᓂᓪ, ᐊᖏᑎᓕᕆᓚᐅᑦᓂ (DOE) ᑐᓴᐅᒪᑎᑕᑎᓂᖅᓱᑕᐃᓇᖅᑎᑦᑎᒥ ᐱᓕᕆᓕᐅᕐᓂᖅᑐᑦ ᖃᐅᔨᓂᖅᑐᑦ ᐊᖓᔪᓇᓕᕐᓂᐊᖅᑎᑦ ᒥᑭᐅᑎᐊᖅᑎᑦᓴ ᐅᑐᐱᓂᖅᑎᑎᖅᓯᓂᑦ (HTO)–ᑲᑦᖁᓕᓪ ᐊᐃᖅᕆᓕ 2–ᒥ, 2019. ᑐᓴᐅᒪᑎᑕᑎᓂᖅᑎᑦᑕᐃᓇᖅᑎᐅᕐᓂ ᐱᔨᔪᑎᑎᓕᐅᕐᓂᖅᑕᖓᓕ ᑐᓴᓂᖅᑎᑎᐃᓇᓗᑐᑦ ᓇᓂᐅᓕᖅᑎᐅᕐᓂ ᓇᐅᓕᕐᑕᐃᓂᖅᑕᐅᓗᒪᔨᓂᖅ 2017–ᒥ ᐅᓄᖅᓯᒪᓂᓂᑦ ᐊᓕᐅᕐᑕᐃᓂᖅᑕᐅᓗᒪᔨᓂᖅ MX–09–ᒥᑲ ᐊᓚᒪᓗ ᐅᓂᖅᐅᓗᖅᓱᓗᑎᑲ ᐊᐅᓕᕐᓂᑎᓂᖅᒥᑦ ᐱᓕᕆᖏᑦᑐᐅᓗᒪᔨᓂᖅ ᖃᐅᔨᓂᖅᑐᖅ ᐊᖓᔪᓇᓕᕐᓂᐊᖅᑎᑦᓴ ᑲᑐᔨᖅᑎᑎᐃᓇᖅᑎᑦᓂ (HTO)–ᑲᐅᑦ ᐱᖓᓱᖅᑎᑕᐅᓂᖅᖏᑦᑐᖓᓂᖅ. ᑐᓴᐅᒪᑎᑕᑎᓂᖅᑎᑦᑕᐃᓇᖅᑎᐅᕐᓂᖅ ᐃᓇᐅᑎᖅᑕᐅᓂᖅ ᐅᑎᖅᑕᐃᓂᖅᑐᐅᓗᒪᔨᓂᖅ ᐊᐅᓚᑦᓯᐅᑎᖅᑕᐅᓗᒪᔨᓂᖅ ᐃᓚᐅᕐ

ᐅᖏᐅᓯᖏᓐᓇᐊᑕᑭᑎᖅ

ᐅᓇ ᐅᓂᒃᑲᖅ ᑭᓱᖏᖅᑐᐃᔪᖅ ᐊᐢᕈᑎᑕᓐᓇᓕᒃᑯᑦ ᐊᒃᓴᑲᖅᓂᖅᓱ ᐱᐅᓪᓚᑕᒻᒥᓗᓂᒃ ᐅᓂᒃᑲᐅᓯᖏᓐᓄᑦᑐᓂᒃ ᐅᖏᒃᑲᐅᓯᐅᓪᓚᑦᑐᖅᑐᒃ ᑐᑎᓪᓚᖅᑕᐅᖏᓐᓄᑦᑐᑦᓯᐊᓂᑦ ᐅᑭᒪᓐᓇᖅᑐᑦ ᐊᖏᔪᓇᓕᒃᑯᑦ ᒥᑭᑎᐊᖏᓂᐊᖏᒃᑯᑦᓗ ᑲᑐᔪᖏᓂᑎᒃᓴᐃᓂᖅ.

ᐃᓕᓴᕆᐅᓕᐊᑦ ᑕᕐᕈᓂ ᐊᐢᕈᑎᑕᓐᓇᓕᒃᑯᓂᖅᓱᓂᖏᓐᓄᑦᑐᓕᕆᓂᖅᑐᑦ, ᓄᓇᕗᒻᒥ ᐅᕐᕈᑐᓂᓂ ᓄᓇᕗᑦ ᓴᕈᓚᖏᓂ.

ᖃᓂᖅᓴᖅ ᐃᓗᓕᖅᑎᓂᓂᖅ

ᐊᐅᓚᖅᑎᓂᖅᓯᒋᐊᕐᓂ ᓇᐃᓐᓇᕆᔨᒪᔨᕐᓂ i
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1.0 ᐅᓂᒃᑲᓄᑦ ᐱᓱᑎᐅᔾᔪᑦ ᐊᓪᓚ ᐊᑦᑭᒃᓯᒪᓂᖅ

ᐅᓇ ᐅᓂᒋᖅ ᐱᓱᑎᖃᖅᑐᖅ ᑲᑎᑎᒧᓗᓂ ᐊᓪᓚᓗ ᓇᐃᓐᓇᓕᖅᑎᒧᓗᓂ ᐅᖃᐅᓯᐅᔾᔪᑎᓂ, ᐊᐱᖅᑯᑕᐅᔾᔪᑎᓂ, ᐃᓯᒻᒋᑕᐅᔾᔪᑎᓂ ᐊᐅᓚᓂ ᑎᑎᒧᓗᓂ ᐅᖃᐅᓯᐅᔾᔪᑎᓂ ᐊᓪᓚᓗ ᑎᑎᒧᓗᓂ ᐱᓕᕆᓂᐅᔾᔪᑎᓂ ᐅᖃᐅᓯᐅᔾᔪᓕᕆᓂᓗ ᑐᐱᐅᓚᑎᑦᑎᓂᖅᑐᓂᓗ ᑐᑭᓯᐅᑦᑎᓂᒃᑯᑦ ᐊᖏᖅᑯᓯᒪᔪᓂᒃ (HTO)–ᑲᑯᖏᓐᓂ ᐊᐃᕐᓇᓂ 2–ᒥᑦ, 2019 ᖃᐅᔨᓴᓯᒪᓕᕆᓂ ᐅᒥᖓᓚᐃᑦ ᐊᐅᓚᕋᑎᓂᖅ ᐃᓗᓕᕆᓂ MX–09 ᓇᒡᓘᓂᖏᓐ ᐊᓪᓚᓗ ᐅᓄᕐᓂᖏᓐ ᖃᐅᔨᓴᕐᓂᐅᓚᐅᖅᑐᑦ ᒥᖅᓴᓄᑦ ᐊᓪᓚᓗ ᐊᐅᓚᕋᑎᓂᕐᒧᑦ ᐱᓚᕆᔨᖃᐅᔨᓚᕆᓂᖓ. ᓇᐃᓐᓇᖅᓯᐊᑦ ᐊᓪᓚᓗ ᑎᑎᖅᑲᐃᑦ ᑕᑯᑎᕋᑦ ᐅᖃᐅᓯᐅᓚᐅᖅᑐᓂᒃ ᖃᑎᒌᓄᑦ.

2.0 ᑐᐸᐅᒪᑎᕋᖅᑕᐃᓐᓇᕐᓂᒧᑦ ᐱᓱᑎᑎᑲᐅᔾᔪᑦ

ᖃᑎᒌᓂ ᐁᖅᑭᖅᓯᖅᑕᐅᔾᔪᑦ ᐊᓪᓚᓗ ᐊᐅᓚᑕᐅᓂᖅᓱᑕ ᐊᐋᑎᓕᓚᓵᖅᑯᑦ (DOE)–ᖅᑯᐟᓄᑦ ᐱᓱᑎᑎᑕᐅᖅᑕᖅᓱᒐ ᐊᖏᓇᓯᐅᔾᔪᑎᓂ ᒥᑭᐅᓯᐅᔾᔪᑎᓂᒃᑯᑦ ᐅᖃᐅᓯᐅᔾᔪᑎᓂᒃᑐᓗ ᖃᐃᓐᓇᕐᓂᖅᑎᒍᑎᓂ (HTO)–ᖅᑯᐟᓂ ᐱᕕᓇᖅᑎᖅᑎᓗᓂ ᑐᐸᐅᓚᖅᑎᑦᑎᓂᖅᑐᓂᓗᐃᓐᓇᕐᓂᒧᑦ ᖃᐅᔨᓴᕐᓂᐅᓚᐅᖅᑐᑦ, ᖃᐅᔨᓴᕐᓂᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᐅᖃᐅᓯᐅᔾᔪᓕᕆᓂᓗ ᐊᐅᓚᑕᐅᓂᖅᓱᒐ ᐱᕕᓇᖅᑎᖅᑎᓗᓂ. ᖃᐅᔨᓴᕐᓂᐅᓚᓕᕆᓂ 2017–ᒥ ᐃᓄᐅᐊᖓᓂᖏᓐᓂ ᖃᐅᔨᓴᕐᓂᖅᑎᒥ ᐅᖃᐅᓯᐅᓚᐅᖅᑐᓂ ᑲᑎᒪᓂᓗᐅᑦ. ᐊᖏᓇᓯᐅᔾᔪᑎᓂ ᒥᑭᐅᓯᐅᔾᔪᑎᓂᒃᑯᑦ ᑐᐸᐅᓚᖅᑎᑦᑎᓂᖅᑐᓂᓗᐃᓐᓇᕐᓂᒧᑦ (HTO)–ᖅᑯᓂ ᐊᐅᑐᖅᑕᐅᓂᖅᑐᖅᑐᓂ ᑐᓴᐅᒪᓂᖅᑐᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᑎᓴᒪᓂᖅᑐᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᐱᓕᕆᔨᒃᑯᓂᖅᑐᓂᒃ ᐊᐃᓚᖅᑐᑦ ᐊᐅᓚᑕᐅᓂᖅᑐᓂᓗ ᐃᓕᓴᕐᓇᓗᒍ, ᑕᐅᓇᓗ ᐅᓂᒃᑲᓂᐅᔾᔨᓗᐊᑕᐅᓚᐅᖅᓱᒐᓂ. ᐊᐃᓐᓇᖅᑐᐅᓚᐅᖅᑎᓗᓂ ᐃᓕᓴᕐᓇᓗᒍ ᐊᐅᓚᑕᐅᓂᖅᑐᓂᓗ ᐱᓕᕆᔨᒃᑯᓂᖅᑐᓂᒃ ᐊᐃᓐᓇᖅᑐᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᐱᓕᕆᔨᒃᑯᓂᖅᑐᓂᒃ ᐅᕙᓂ

3.1.1 ᖃᐅᔨᓴᖅᑐᖅ ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᓴ ᑲᑐᔪᖅᑎᑭᖏᓐᓇᐅᑦ (HTO) ᑐᓴᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓯᒪᔨᑦ ᓇᐃᐅᖅᓯᒪᔨᑦ-2019

ᐃᓕᓴᕐᕕᐅᔨᑦ: ᐅᒥᖏᒪᐃᑦ ᓇᒋᓐᓂᑎᑦ ᐊᒻᒪᓴ ᐅᓇᖅᓂᖏᑦ, ᐅᒥᖏᒪᐃᑦ ᐊᐅᓪᓚᑎᓂᖅ ᐃᓴᓪᓂᖏᑦ MX–09, ᖃᐅᔨᓴᖅᑐᑦ ᑰᖏᑕ ᐅᐊᓪᓂᐊᓂ, ᐋᓱᓕᓕ 2017. ᐅᒥᖏᒪᐃᑦ ᐅᓇᖅᓂᖏᑦ ᑕᕐᓀᓂ ᐊᐅᓪᓚᑎᓂᖅ ᐃᓴᓪᓂᖏᑦ ᓕᖅᑰᓕᕐᓂᑐᖅ.

ᑐᑭᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅᓗᑦ ᐱᔾᔪᑎᑎᑭᐅᔨᑦ:
ᑐᑭᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅ ᐋᑦᖁᕻᑕᐅᓪᓚᐅᔨᖅᑐᖅ ᖃᐅᔨᔨᖅᑐᑎᒥ ᐊᐃᖅᑏᓪᓕ 2–ᒥᑦ, 2019. ᑲᑎᒪᓂᐅᔨᑦ ᐱᔾᔪᑎᑎᓕᕆᔨᖅᑕᕐᓂᖏᑦ ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᓴ ᑲᑐᔪᖅᑎᑭᖏᓐᓇᐅᑦ (HTO)–ᖁᓕᓂ ᐱᓕᕆᖅᑲᑎᑭᖏᓐᓇᖅᓂᖅᓗᒥ ᑐᑭᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᑐᓴᓂᖅᑐᒃᓴᓂ, ᐅᓂᒃᑲᑦ ᖃᑎᖃᕐᓗᓂᑦ, ᐊᒻᒪᓴ ᐅᖃᖅᐅᓂᖅᑐᓂᓗᓂ ᓯᕗᓂᕆᓂᖅ ᐊᐅᓪᓚᑎᓂᖅᓗᑦ ᐱᓕᕆᓂᖅᑐᓂᓗᑦᑐᓂᖅ. ᑲᑎᒪᓂᖅ ᐱᓇᖏᓐᓄᐅᓕᕆᔨᖅᑕᐃᓐᓇᖅ ᐃᓕᑕᐊᖅᓯᒪᔨᓴᓂ ᑐᓴᖅᑎᑎᓗᓂᖅ ᐊᖏᓐᓂᓚᐅᔭᖅᑎᑦ (DOE)–ᖁᑦ ᐱᓕᕆᖁᓕᓯᒪᓂᑦ ᐱᓴᐅᔨᓇᖅᑎᑕᐅᔨᑦ ᐊᖏᓐᓇᓕᕆᔨᑦ ᐊᐅᓪᓚᑎᓂᖅᑐᓗᑎᒃ (TAH) 20–ᓂᒃ ᑐᓴᐅᒪᔨᐸᑎᓴᑦ ᐊᑐᐃᓐᓇᐅᐅᓂᖅᑐᑦ ᒪᓕᒃᓴᑦ, ᑭᓯᐊᓂ ᒧᓇᐅᔨᑦ ᓂᕿᑭᖅᑎᓇᓂᑲᓐᓂᒃ ᐊᔨᓇᓇᖅᑐᖅ.

ᐅᓪᓗᓐᓄ: ᐊᐃᓴᓚᓂ 2–ᒥᑦ, 2019

ᑭᖑᓂᖅᑐᐃᓐᓇᐅᔨᑦ:
ᐊᖏᓐᓂᓚᐅᔭᖅᑎᑦ (DOE): ᑎᐅᓚ ᒥᐅᑦᓂ, ᓀᓴ–ᒪᓴ ᓕᑯᑐᒃ ᐊᒻᒪᓴ ᑲᖏᓂ ᒪᑎᐅᓂ, ᖃᓯ ᐊᖏᑭᐊᖅᓴᖅ,
ᐋᓕᓂ ᓂᕆᑕᓇᑎᐊᖅ

ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅᓗᑦ ᑲᑐᔪᖅᑎᑭᖏᓐᓇᐅᑦ (HTO): ᕿᐱ ᐊᓇᐃᓕᑐᒃ, ᑲᖏᓂ ᑭᓪᓕᑭᓂᒃ, ᓴᐃᑯ ᕿᑐᔪᓇᒃ, ᒪᐃᔪᓛ ᐱᑐᓴᓂ, ᐊᐧᐋᓇᓂ, ᓴᓐᓂᑦ ᑲᐅᕿᑐᑐᒃ, ᓴᐃᑯ ᕿᑐᖏᓐᓄᒃ, ᓂᐅᓚ ᐊᓴᓂ (ᐃᒃᓯᖏᐅᑕᖅ).

ᑲᓴᒥ ᖃᑲᐅᓕᓂᖅ ᑲᓇᑕᒥ: ᓂᓐᑕ ᐅᐊᖅᒥᓂ

ᒪᑕᒪᓂᐅᖅᑎᓄᑦ (MLA) ᐃᓕᑎᓴᐅᔨᖅ: ᒥᓪ ᐊᖅᑭᐊᑐᖅ

ᐅᖅᑲᐅᓯᐅᔨᑦ ᓇᓗᓇᐃᓴᖅᓯᒪᓂᖏᑦ:
ᑕᑯᐸᖅᐅᑎᑎᑎᓪᓗᑦ, ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅᓗᑦ (HTO)–ᖁᑯᓂ ᐃᓕᑎᓴᐅᔨᑦ ᐅᖅᑲᓪᓕᓕᐅᖅᑐᑦ ᐊᒻᒪᓴ ᐊᐱᖅᑯᓂᓗᑎᒃ ᐊᐱᖅᑯᑎᑦᑎᓗᒃ ᖃᐅᔨᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᑦᑎᓗᒃᑐᓂᖅᑐᒃᓴᓂ ᐊᒻᒪᓴ ᖃᐅᔨᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᑦᑎᓗᒃᑐᓂᖅᑐᒃᓴᓂ ᐅᖅᑲᓪᓕᓕᐅᖅᑐᑦ ᖃᐅᔨᓴᖅᑐᑦ ᐊᐱᖅᑯᑎᑦᑎᓗᒃᑐᓂᖅᑐᒃᓴᓂ ᐅᖅᑲᓪᓕᓕᐅ

ᐅᒥᖏᓚᐃᑦ ᐅᓄᕐᓂᖏᑎ ᓯᕐᔪᓯᖅᓭᒪᖏᓐᓂᖏᓐᓂᒧᑦ. ᐃᓕᖏᑎᑦ ᐅᒥᖏᓚᐃᑦ ᓄᑦᑐᐴᕆᐅᔪᕐᐃᓄᒃ ᓇᓵᖅᑐᓕᐅᑦ, ᓄᓐᓂᖏᓪᓚ ᐃᓱᒪᑎᐅᑦᓗᓂᓂ ᐃᓂᐅᔪᖅ ᐳᓴᓂᖏᓐᓂᒧᑦ. ᖃᐅᔨᑎᐊᕐᖅ ᓴᓗᓇᐃᕐ ᓇᓵᖅᑐᖃᖅᑐᑦ ᐃᑭᓐᓂᖃᕆᐅᕐᓂᓪᓚᓂ, ᐅᒥᖏᓚᐃᑦ ᓇᒧᓐᓂᖅᑐᐴᕆᐅᔪᕐᐃᓄᒃ ᐳᔪᓂᒋᐃᓕᓂᓐᓂᒧᑦ ᐃᓂᐅᔪᓂᑦ ᐃᑭᓱᓕᕐᓂᓪᓂ ᐃᓂᐅᔪᓂᒃ ᐅᓪᓗᓕᕐᓂᖅ.

ᐅᒥᖏᓚᐃᑦ ᐃᖏᓕᕆᓂᖏᓐᓂᑦ ᑐᑭᓯᐅᓚᕐᑐᐊᖏᓐᓂᑐᑦ. ᐅᓚᔪᓕᓗᓐᑦ ᐅᒥᖏᓚᐃᑦ ᓯᐅᑎᖏᓐᓂᒧᑦ ᓇᑐᓇᐃᒃᑯᑎᖅᓯᓚᐅᖅᓯᓚᔪᓕᑦ ᓇᐸᑕᓕᒃ ᑕᓯᐊᓂ 1990 ᐱᑎᐊᓕᓴᖅᑎᓕᑎᑦ, ᐊᓖᓚᓐ ᐅᒥᖏᓚᒃ ᓇᑐᓇᐃᒃᑯᑎᖅᓱᖅᑐᖅ ᑕᑯᐅᐅᒃᓄᓪᓕᖅᑐᓂ 2017–ᒥᑦ ᑲᓐᑐᔭᑐ ᑕᓯᐊᑕ ᖃᐅᓂᕐᖏᓚᓂ. ᐅᒥᖏᓚᐃᑦ ᑲᐸᐸᑕ ᐸᖏᓚᔭᓐᓇᖅᑐᑦ ᑎᕋᒪᓂᒃ ᑕᓕᒪᓂᑦ ᒪᐃᔪᓂᑦ ᐅᖏᓪᓯᑎᑎᒧᒧᑦ. ᐅᒥᖏᓚᑦᑲᖅᑐᖅ ᓇᓵᖅᑐᑕᐅᐸᑦ ᐳᓴᓂᐊᓂ ᐊᐱᒪᓐ ᐱᐅ ᑕᓯᐊᑕ ᖃᐅᓂᕐᖏᓚᓂ. ᐊᐳᓇᓇᐅᐸᑦ ᒥᐅᓴᓂ ᐃᖏᓕᕆᓂᖏᓐᓂᑦ ᓄᓇᕐᔪᐊ ᓯᖏᓚᐅᓂᖏᓐᓂᒧᑦ (GNWT).

ᓄᓇᕐᔪᐊᒥ ᐅᒥᖏᓚᑕᕐᓂᖏᓪᓚ ᐱᔨᔪᑎᑕᓚᑦᓯᐅᑦ, ᐊᐱᕆᑯᑎᑕᖃᓚᐅᕆᖅᑐᓗ ᑲᑎᕐᕕᓂᑦ ᐅᒥᖏᓚᐃᑦ ᓄᓇᕐᔪᐊᒥ ᐃᓚᒥᐅᒥᑦ ᐅᒥᖏᓚᐃᑦ MX–09 ᐃᓄᒋᐊᖏᓂᖏᓐᓂᒧᑦ. ᑕᐃᒪᐃᓚᒪᑦ, ᑲᑎᒪᖃᑎᑎᖃᒪᔪᑦ ᐊᓚᓗᓐ ᓴᓇᖃᑎᑎᕐᕕᓂᑐᐃᓐᑲ ᓄᓇᕐᔪᐊ ᓯᖏᓚᐅᓂᖏᓐᓂᒧᑦ (GNWT).

ᓂᕆᐸᓴᓂᑲᖅᑐᑦ ᓯᐸᑲᓂᖏᓪᓚ ᐃᓱᒪᑎᓕᓗᓐᑦ, ᑯᑕᐃᑦ ᐊᓪᓚᕐᓇᖅᑐᖅ ᑕᒪᐸᕆᖅᑕᐅᖏᓐᓂᑕᑦ. ᐊᖏᕐᕋᒧᑎᓴᑎᓂᒧᑦ, ᑯᑕᑕᒪᑦ ᑐᓂᔨᓱᒧᑦ, ᐱᔨᑕᒥ 7–ᖃᐅᔪᑦ ᐅᓂᐸᐅᔨᓱᐅᔪᑦ ᓂᕆᐸᓴᓂᐸᑕᐅᑦᑕᓗᑎᐸ 20–ᓂᓐᓄᖃᕐᑐᓂᓂ. ᓯᓕᐸᐅᓂᖅ, ᐊᖏᕐᕋᒧᓂᐸ ᖃᑯᓂᓂᐸ ᖃᐄᖓᓴᖅᑐᓂᐸ ᑯᑕᐃᑦ ᓄᖐᒋᐅᔨᓴᖅᑎᓴᖏᓐᓂᑐᑦ, 15–ᖃᐅᔪᑦ ᐊᑐᖅᑕᐅᔨᓯᒪᔪᑦ 20–ᐅᑦ ᖃᑯᓯᔨᓂᖅᓵᖅᓄᖅ, ᐊᖏᓚᓐ ᐅᓂᐸᐅᔨᓯᑕᖃᓄᖏᓐᓂᖓᓪᓗᓂ ᓂᕆᐸᓴᓂᐸᑎᐅᔪᓂᐸ ᐃᓱᐊᓗᑎᐸᓴᐃᑦ ᒥᐸᓚᓄᑦ. ᐊᖏᕐᕋᒧᓂᐸ ᒥᑭᑎᐊᖏᐊᖅᑲᓐᓗ ᑲᑐᔭᓴᐸᑎᐅᔪᓂᑦ (HTO)–ᐸᑯᑦ ᐃᓄᖏᓴᐸᑦ ᐊᑐᖅᑕᐅᓂᖅᓴᓂ ᑕᐃᒪᐃᖏᓐᓂᓐᓇᖅᑐᑦ MX–09–ᒧᑦ 20–ᓂᐸ ᐱᔨᐅᔪᓐᓇᖅᑎᑕᐅᔪᑦ ᐊᖏᕐᕋᒧᓂᐸᑕᐅᔪᓐᓇᖅᑎᑕᐅᓗᑎᐸ (TAH)–ᓂᐸ. ᐳᓴᐊᓂ, ᐊᖏᕐᕋᒧᓂᐸ ᒥᑭᑎᐊᖏᐊᖅᑲᓐᓗ ᑲᑐᔭᓴᐸᑎᐅᔪᓂᑦ (HTO)–ᐸᑯᑦ ᐃᓱᒪᓴᖅᓯᓴᑎᑕᐅᔪᑦ ᒪᓕᑐᖏᐊᖏᓐᓂᒧᑦ ᐊᑐᕈᓐᓇᓗᑎᐸ ᐱᖏᔪᓴᑎᑦᑎᓇᓯᓗᓂ ᓄᓇᕗᓂᓂ–ᐱᓕᕆᐊᖏᔪᓂᐸ ᐊᐅᓚᕐᑎᓂᓕᒻᒥᑦ ᐱᓕᕆᑎᐊᕈᑎᓂᐸ, ᒥᑭᑐᒥᑦ ᐊᐸᔪᓴᑎᑦᑎᓚᐅᖅᓴᓕᒻᒥᑦ 1–ᒧᑦ ᐅᖏᓪᓗᓐᓂᑦ 2–ᓄᑦ ᐅᒥᖏᓚᐃᑦ ᐅᓄᕐᓂᖏᓐᓂᑦ ᐃᑲᔪᕆᓂᖅᑎᓂᑦᑎᓚᐅᖅᓴᓂᑦ, ᓯᓂᔨᖏᑐᓄᖏᓐᓂᖓᓐᓂ, ᐱᕈᖅᑦᓯᓚᐅᖏᓐᓂᖓᓪᓗᑎᐸ, ᐊᔨᓂᐊᓚᐅᖅ.

<u>ᐱᓚᓇᐃᓴᐅᓕᔪᓂᑦ ᓄᓇᕗᑦ ᑲᖏᒪᐃᑦ (GN)–ᐃᔨᓐᓄ:</u>

ᖃᐃᓪᓗᖅᑐᖅ ᐊᖏᕐᕋᒧᓂᐸ ᒥᑭᑎᐊᖏᐊᖅᑲᓐᓗ ᑲᑐᔭᓴᐸᑎᐅᔪᓂᑦ (HTO)–ᐸᔨᖏᓐᓄ ᐱᓚᓇᐃᓴᐅᓕᔪᑦ *ᐊᑐᖅᑕᐅᓂᖏᓪᓗᓂ* ᑕᐃᒪᐃᖏᓐᓄᓇᐃᓕᑦ MX–09–ᒧᑦ 20–ᓂᐸ ᐱᔨᐅᔪᓐᓇᖅᑎᑕᐅᔪᑦ ᐊᖏᕐᕋᒧᓂᐸᑕᐅᔪᓐᓇᖅᑎᑕᐅᓗᑎᐸ (TAH)–ᓂᐸ. ᓂᕆᐸᓴᓂᐸᑎᐅᔪᑦ ᐊᓚᓗᓐ ᐊᑐᖅᑲᓐᓂᑦ ᐅᓂᒃᑳᑕᐅᖅᑕᑕᑐ ᐊᔨᔪᓴᖅᑐᐃᓐᑲ ᑕᐃᒪᒥᓂᐸ ᐃᓄᐅᑕᐅᔪᓂᑦ ᐅᖏᓪᓗᓐᓂᑦ 2–ᓄᑦ ᐅᒥᖏᓚᐃᑦ ᐃᓄᐅᑕᐅᔪᓂᑦ, ᐃᓄᐅᑕᖃᓄᐃᓕᓂᑦ ᐱᓚᓇᐃᓴᐅᓕᔪᑦ 2023 ᐊᖏᓂᓂ.

4.0 ᑭᖓᓯᓪᓗ ᐅᖃᐅᓯᕐᓂᖅ – ᐱᖃᑎᒋ ᐱᓕᕆᔨᖓᕆᓯᓪᓗ

ᐊᕐᑎᑎᓕᕆᓴᒃᑯᑦ (DOE)–ᒃᑯᑦ ᐁᖃᑭᒃᓯᕗᑎᐊᖅᓂᐊᖅᑐᑦ ᐅᒥᖓᒪᐃᑦ ᐅᓂᒃᑲᖏᑐᓂ ᑲᑎᒪᑎᓯᓗᓄᑦ ᐅᖃᐅᓯᓯᐅᔪᓂᒃ ᐃᓕᖅᖢᓗᑎᒃ ᐊᒻᒪᓗ ᑐᓂᐅᖅᖢᖅᑕᐅᓗᑎᒃ ᐊᐅᓕᕆᑦᐅᖅᑲᑎᑎᒋᓐᓂᑦ ᐃᓕᑕᑯᐅᔪᓂᑦ. 2019–ᒥᑦ ᐅᒥᖓᒪᓂᑦ ᖃᐅᐅᔪᖅᓂᖅᖢᖅ ᐸᖅᓇᖏᐅᑎᓯᓗᒍ ᐃᓗᐃᓯᓕᖅᖢᒥ, ᐊᕐᑎᑎᓕᕆᓴᒃᑯᑦ (DOE)–ᒃᑯᑦ ᐃᓱᒪᑕᑯᖅᖢᖅᑎᐊᖅᑐᑦ ᐅᖃᐅᓯᓯᐅᔪᓂᒃ ᐊᒻᒪᓗ ᑎᓕᐅᕐᔪᐊᑎᓂᒃ ᐊᖖᓇᓴᐅᓕᖅᑐᑦ ᐊᒻᒪᓗ ᑎᑭᑕᐊᖅᖢᖅᑎᐊᖅᑐᑦ ᐅᖃᐅᓯᓯᐅᔪᓂᒃ (HTO)–ᒃᑯᑦ ᑐᓴᐅᖅᑎᑎᒋᔭᐅᓂᒃᑯᑦᑕᐃᓐᓇᖅᓂᖅᑎᓯᓗᓂ, ᖃᓄᐃᓕᐅᕐᓂᖅᑎᑦᓗᓂ ᐃᓕᓴᕐᓂᐊᕐᓗᒍ.

ᑕᐃᒪᓐᓇᓂ 2017–ᒥ, ᐅᓇᖅᓂᕐᓂᓂᒃ ᖃᐅᔨᓴᖅᓂᖅᖢᑦ ᖃᐅᔨᓴᐅᓕᕆᓪᓚᑦ ᐅᖃᐅᓯᐅᓕᕆᓪᓚᑦ ᕿᑎᖅᑕᐅᑦ ᐊᐃᔨᖅᑐᑦ ᐅᑭᐅᖅᑕᕐᒪᑦ ᐱᒪᓪᓚᑦᑎᓪᓗᒍ ᑲᓇᑕᒥ ᐃᓚᖓᑕᐅᓪᓚᑦᑎᓪᓗᒍ ᑲᑎᒪᓂᕐᓂᖅ 2018–ᒥ ᐊᒻᒪᓗ ᐊᖓᔪᖅᑲᐃᑦ ᒥᑭᐅᐊᓂᖅᑲᑎᑎᒋᓐᓂᒃ ᑲᑎᒪᔨᖃᑎᒌᖓᓂᑦ (HTO)–ᒃᑯᓂᒃ ᐊᒻᒪᓗ ᐊᓯᖏᓐᓂᒃ ᐊᐅᓚᑦᑎᐊᕐᑲᖃᑎᒌᖓᓂᒃ ᐃᓕᑕᑯᐅᔪᓂᑦ. ᑐᓴᐅᒪᔪᑎᑦᑎᓂᓕᕆᓂ ᓴᖅᑯᓴᖅᓯᒪᔪᓂᓕᕆᓂᒃ ᐅᒥᖓᒪᐃᑦ ᐅᓇᖅᓂᕐᓂᓂᒃ ᐅᒥᖓᒪᐃᑦ ᐊᐅᓚᑦᑎᓂᓕᒃ ᐃᓗᓕᕆᓂᖅ MX–09–ᐅᐸ ᒥᖁᓇᓂᑦ ᐅᖃᐅᓯᐅᓯᒪᔪᓂᑦ. ᓴᓚᖁᓂᓕᒃ, ᐱᓕᕆᖅᑯᓂᖅᑎᐅᓯᒪᔪᓂᓪᓚᑦ ᓄᓇᕗᑦᒥ ᐅᒪᓐᓇᕐᓂᖅᓚᑦ (NWMB)–ᒃᑯᓂᓄᑦ ᐱᓇᔨᖃᖅᓂᐊᕐᑐᖅᑐᒃ ᐱᓴᐅᔨᓇᖅᑎᑕᐅᔨᓇᖅᑎᑕᐅᓗᑎᒃ (TAH) ᖃᓄᐃᓕᐅᕐᔪᓂᒃᑯᑦ ᑕᐃᒪᐃᓕᖅᓇᓴᓗᑎᑦ, ᑭᓇᓇ ᑕᐃᒪᐃᓂᐅᔪᓂᖅᖢᖅ ᒪᓕᒃᔪᐃᒍ, ᖃᓄᐃᓗᖅᑐᑦ ᐊᖓᔪᖅᑲᐃᑦ ᒥᑭᐅᐊᓂᖅᑲᑎᑎᒋᓐᓂᒃ ᑲᑎᒪᔨᖃᑎᒌᖓᓂᑦ (HTO)–ᒃᑯᕐᓂᓂᒃ ᐃᑲᔨᖅᓯᖅᑕᐅᓕᓐᓂᓗᒃ. ᑐᓴᐅᒪᑦᑎᓂᖅᑕᐃᓇᓂᐅᕐᓕ ᐱᓕᕆᔨᖓᓪᓗ ᐊᐅᓚᑦᑎᐊᖅᑐᑦ ᑭᓇᓯᒥᖓᓪᓗ ᐅᓂᒃᑲᒥᑦ.

ᐅΔᒍᖅ I

ᖃᕆᓴᐅᔭᒃᑯᑦ ᑕᑯᓴᐅᑎᕐᑎᓴᓪᓚᐊᑎ 2019

ᐅᒥᓪᒪᐃᑦ ᓇᒋᔭᓂᖏᑦ ᐊᖏᒪᓗ ᐅᓄᕐᓂᖏᑦ ᐅᒥᓪᒪᐃᑦ ᐊᐅᑕᕐᑎᓂᓐᓄᖅ ᐃᓗᓕᖏᑦ, MX09–Γ, ᖃᐅᓪᓗᖅᑐᑦ ᑯᓴᒪᑦ ᐅᐊᓕᓂᐊᓂ

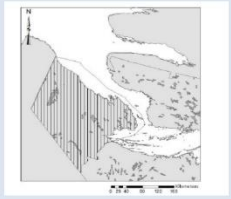
ᖃᐅᓪᓗᖅᑐᖅ ᐊᖏᒃᓇᓕᕆᑦ ᒥᑭᑎᐊᓂᐊᕆᑦᑐᓗ ᑲᑎᕐᓴᖃᑎᒌᒃᓴᓕᔪᓂᑦ (HTO)–ᑲᑯᑦ

ᓵᓴ–ᒪᓴ ᑕᑭᓗᑲᒃ
ᓄᓇᕗᑦ ᒐᕐᒪᖃᑯᑦ
ᐊᐅᑎᑕᓴᕐᑯᑦ

ᔪᓂ, 2019

ᖃᐅᔨᓴᖅᑕᓱᑦ ᐃᓂᖏᒐ

- MX–09, ᑭᐃᓕᓐᖅᐸᒋᑦ ᓇᑕᐅᑦᑕᖅᑕᐅᓕᑦ 2007–Γ 598 ᐅᒥᓪᒪᐃᑦ, 159 Δᵊᓇᐅᔨᓗᑎᒃ ᐊᑲᓴᓯᒪᔪᒋᑦ.
- Δᓯᒪᓗᑎᑦ, ᐊᖏᓇᓯᑎᑦ ᓂᕆᐅᔭᓂᖅᑕᖅᑯᑦᓴᑕᒪᓴᖅᑐᒃ ᑐᓴᑐᓂᓐ ᐃᓇᓐᑎᑕᐅᔨᓪᓗᓂ ᑕᵊᕐᓂ ᐃᓂᐅᔪᒋᑦ.



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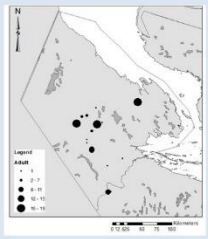

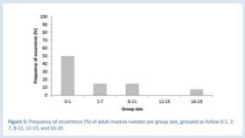

MX09, – 2017 Δᓄᑎᐊᖅᓂᕐᑎᑦ ᓇᑕᐅᑦᑕᖅᑕᐅᔨᑦ

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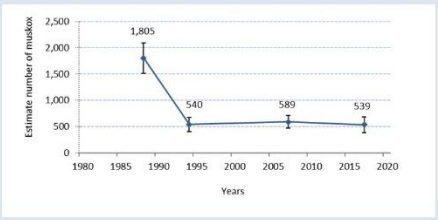

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MUSKOX (*Ovibos moschatus*) DISTRIBUTION AND ABUNDANCE, MUSKOX MANAGEMENT UNITS MX09, WEST OF THE COPPERMINE RIVER, AUGUST 2017.

Lisa-Marie Leclerc[1]

Version: January 2019

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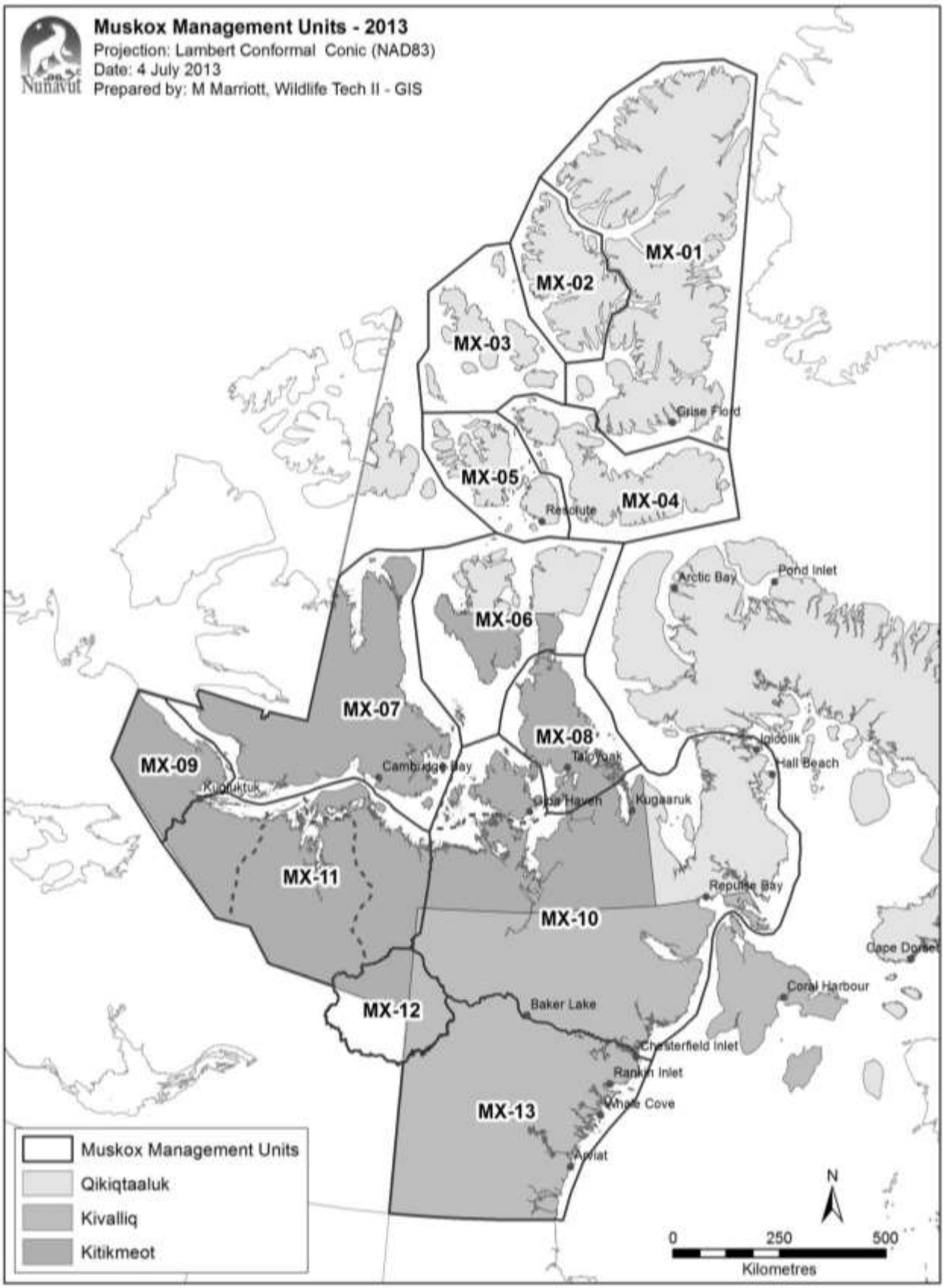
Muskox Management Units - 2013
Projection: Lambert Conformal Conic (NAD83)
Date: 4 July 2013
Prepared by: M Marriott, Wildlife Tech II - GIS
Nunavut
MX-01
MX-02
MX-03
MX-04
MX-05
MX-06
MX-07
MX-08
MX-09
MX-10
MX-11
MX-12
MX-13
Grise Fiord
Resolute
Arctic Bay
Pond Inlet
Cambridge Bay
Kugaaruk
Repulse Bay
Cape Dorset
Coral Harbour
Baker Lake
Chesterfield Inlet
Rankin Inlet
Whale Cove
Arviat
Hall Beach
Muskox Management Units
Qikiqtaaluk
Kivalliq
Kitikmeot
N
0
250
500
Kilometres

Executive Summary

The muskox in the West Kugluktuk management unit MX-09 are the westernmost indigenous muskoxen in North America. A systematic strip transect survey took place in August 2017 to determine the muskox abundance and distribution in this management unit. A total of 8,591 km^2 were flown, representing 16% coverage of the study area of 53,215 km^2. During the survey, 87 adult muskoxen were recorded on transect resulting in an estimated number of 539 ± 150 (S.E.). The population in MX-09 has been mostly stable since 1994, and the survey results are consistent with local observations. The muskox distribution does not appear to have changed from the observed historical distribution. Muskoxen have taken advantage of the wetter and lower-lying areas in the Rae-Richardson River Valley that is within the proximity of uplands, providing them with suitable forage and a refugee from predators. The calf to adult ratio was 38 calves/100 adults and the average number of group, excluding calf, was small, 6.21 ± 6.6. (S.D.). Muskox density in MX-09 was the lowest density within the Kitikmeot region with 0.010 muskox / km^2 since 2013. The next survey of this area is planned to be completed no later than 2023 in order to review the harvest quota.

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Introduction

During the last century, muskox on the upper Rae-Richardson River Valley were able to persist on the landscape while the species was nearing extirpation (Barr, 1991). This small pocket of muskox was known as the population of the North of Great Bear Lake. This population increased in numbers allowing for a small harvest quota to be established for Kugluktuk (6 tags) and Paulatuk (8 tags) (Urquhart, 1980). The muskox population of the North of Great Bear Lake kept growing, giving rise to the muskox population of the Rae-Richardson River watersheds, which was recently renamed the West Kugluktuk muskox management unit, MX09.

The muskoxen of the West Kugluktuk management unit are the westernmost indigenous muskoxen in North American, and represent the muskox population with the longest monitoring effort representing over 80 years of monitoring (Kelsall *et al*., 1971). Based on comparable systematic population surveys, this population peaked in 1988 before it plummeted and remained stable with around 550 animals between 1994 to 2007 (Gunn, 1995; Dumond, 2007). Even if this population is subject to year-round hunting by the community of Kugluktuk, the quota has been limited and enforced to 20 tags. It is known that harvesters do not fill this limit annually, since there is general local knowledge that these muskoxen are “sick” and found in few numbers (Kugluktuk HTO, pers. comm.). In the 1990s, the discovery of a lungworm parasite (*Umingmakstrongylus pallikuukensis*) was found to be an attributable factor to the decline of the population (Gunn, 1995, Kutz, 2000). Although the effect of this parasite on the muskox population dynamics is not fully understood, this lungworm is known to decrease the respiratory capacity of the individual muskox thus making it more vulnerable when being chased by a predator.

The Rae-Richardson Rivers Valley area constitutes a rich habitat for predators, as it sustains wolf (*Canis lupus arctos)*, wolverine (*Gulo gulo*), and grizzly bear (*Ursus arctos horribilis*). Potential increases in grizzly bear could suggest that this predator is exercising a greater pressure in inhibiting the muskox population growth (Gun, 1995). However, a recent study shows that the grizzly bear’s diet contains a large proportion of small tundra herbivores (e.g. ground squirrels). On the other hand, muskox represents an important food source for wolves and wolverines (L’Herault *et al*., 2016). People and predators both depend on muskox and have always co-existed on the landscape. Under Inuit natural laws, all living things should be respected as well as the role they play in the ecosystem. This concept is referred to, by elders and hunters, as *Maligaits* (L’Herault *et al*., 2016).

Since caribou herds in this region of Nunavut are declining in numbers and are becoming harder to harvest, harvesters are seeking alternative sources of meat, such as muskox, to address issues around food security. In addition to the conservation and recovery of this small size population of muskox, the preservation of harvesting rights should be considered a priority for future management and for maintaining food security. This task would be difficult to achieve without re-assessing the muskox population of MX-09, and revisiting the Total Allowable Harvest (TAH) This project aims to provide an update of the muskox population, MX-09 and provide updated TAH recommendations. Consistent with other muskox surveys, the Nunavut wide monitoring approach was used. This scientific information will be provided in balance with Traditional Knowledge to review existing management strategies and promote a sustainable harvest of muskox.

Objectives

This project aims to address the concerns and requests of Inuit hunters, as well as to provide up to date scientific information for management purposes. Therefore, the main objectives of this study are:

1. Determine the estimated number of muskox;
2. Determine muskox distribution and density;
3. Determine calf:adult ratio and group size.

By achieving the objectives of this project, it will be possible to have better information on current muskox abundance and distribution in the muskox management unit MX-09. Information on group structure, calf production, group size and density, are essential to gain insight on the relationship between these variables and population dynamic.

Materials and Methods

Study Area

The study area is the muskox management unit MX-09, also called West Kugluktuk management unit. The boundaries for this area are to the west and south, the Nunavut boundary with the Northwest Territories and, to the north, the coast line of the Dolphin and Union Straight. Muskox management unit MX-09 is separated to the east, from the adjacent muskox management unit MX-11, by the Coppermine River.

This area is part of the Southern Arctic Ecozone, transiting following the latitudinal gradient from the boreal forest around Great Slave Lake to the tundra. In this subdivision, two terrestrial ecoregions are found, the Takijuq Lake Upland and Coronation Hills regions (Environment Canada, 1995). In the south of the subdivision, it includes; the edge of the tree line, eskers,

rocky barrens, with lakes going through the landscape to provide a physical uniqueness. The taiga forest is present, but restricted to a locally warm and dry place with scattered stands of spruce. The taiga changes to northward vegetation covers, which are dominated by sedge meadows and shrubs, such as dwarf birch, willow mixed with various herbs, lichens, and mosses. The entire eastern boundary of the study area is characterized by vegetated rock outcrops that are common on the Canadian Shield (Environment Canada, 1995). To the North, the Coronation Hill region prevails. The relief of the lower Coppermine River valley and coast line is characterized by weather-worn plateaus and south facing hill systems. These topographic features along with the climate, influence the biotic processes differently. Plant cover becomes discontinuous to null at higher elevations, on dry exposed sites, and on the low profile sand dunes boarding the coastline north of the Bluenose Lake.

Survey Area

Prior to survey, no reconnaissance survey was undertaken to maximize the coverage area investigated. Instead, anticipated muskox distribution patterns were obtained from past ground surveys, hunter observations, and Inuit Traditional Knowledge/*Inuit Qaujimajatuqangit* (IQ). Since it was reported that muskox groups are still found in low numbers across MX-09, the whole management unit was surveyed at 16% coverage with not strata of different effort allocation (Figure 1).

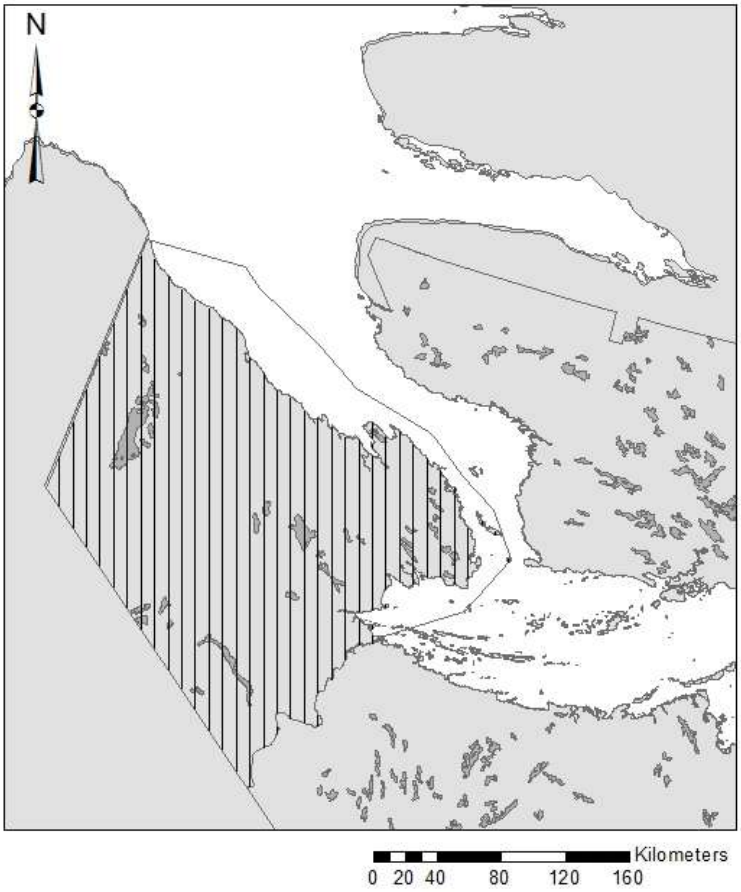

Figure 1: Transect lines of the muskox management units MX-09 representing 16% coverage of the area.

To increase the precision of the survey areas, ESRI'S ArcGIS software with an adapted survey design tool was used to randomly plot the transect lines until desired percentage of coverage was archived. The tool allows the user to determine the precise number of transects and the distance between each transect line required in function of the transect strip width and the total area of the management unit. Orientation of the transect lines within the stratum was determined in function to have the most homogeneous and shorter transect line length under the assumption that muskox groups are randomly and uniformly distributed on the landscape (Figure 1).

Table 1, below, summarizes the total area, the percentage of cover, the total number of kms of transects of different length, the number of lines, the resulting distance between each transect line and the orientation of the transect line. In sum, the management unit, MX-09, of 53,215 km^2 was surveyed with a total of 5,237 km of transect lines, which represented 31 transect lines of different length at a spacing of 8.5 km (Table 1).

Table 1 Characteristic of the study area and the transect lines per stratum in the Management Unit MX-09.

Stratum	Total area (km^2)	Percentages (%)	Total transect lines (km)	Number of lines	Distance between transect line (km)	Orientation
MX-09	53,215	16	5,237	31	8.5	North-South

Aircraft configuration

A systematic transects line survey was flown with a fixed-wing single engine turbine aircraft, a grand caravan. The transect lines were surveyed at a speed of 160 km/hr and the survey altitude of about 121 meters, which was mostly maintained following the relief of the study area using a radar altimeter. The pilot responsibilities were to monitor this air speed and altitude while following the pre-programmed transect on a Geographic positioning system (GPS). The strip transect was 800 meters on each side of the aircraft, for a total transect width of 1.6 kilometers. The pre-determined transect width of 800 meters was set on each wing based on calculation using the formula of Norton-Griffiths (1978) and others (Gunn and Patterson 2000; Howard 2011).

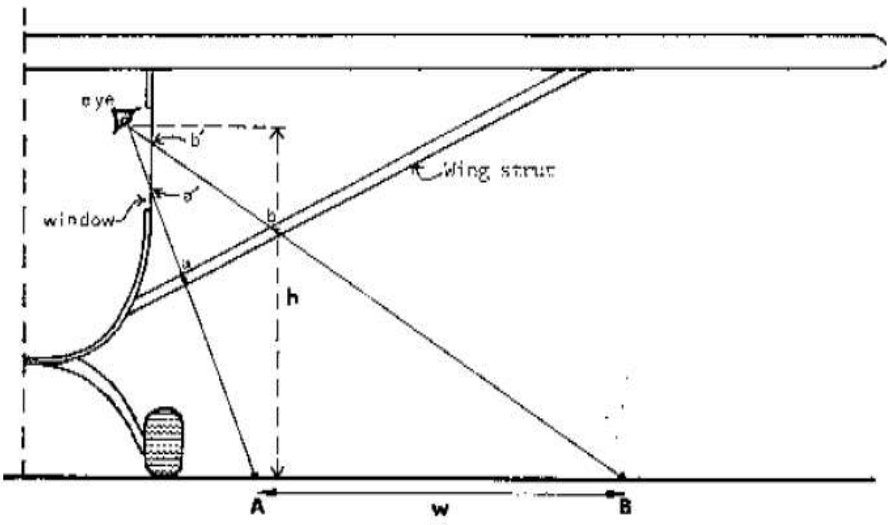

Figure 2: Schematic diagram of aircraft configuration for strip width sampling North-Griffiths (1978). W is marked out on the tarmac, and the two lines of sight a'-a-A and b'-b-B establish, whereas a'- and b' are the window marks.

$$w = W*h/H$$

Where, W= the required strip width; h= the height of the observer's eye from the tarmac; and H= the required flying height.

The entire survey was set up with an observer/recorder crew: two recorders, one left side observer and one right side observer. Each left and right side observers and the recorders were divided into a team. Observers were responsible to continuously search for and count muskox; the number of calves (5-6 months old) were counted when they were conspicuous while on transect. No sex and age classification counts were systematically attempted. The data recorded include the number of muskox and GPS location. Only counts of adults were used in the final population estimate. Even if this survey focused on muskox, additional sightings of other species were also recoded, such as caribou, grizzly bear, wolverine, and wolf.

Analyses

As this survey focused mainly on obtaining an estimated number, only unambiguous classification criteria were used to determine the number of calves and adults. The group was then broken down into adults (female/male) and calves (Howard 2011). The flying height and speed did not allow for accurately distinguishing male from female muskox in a group from horn size. Therefore, the proportion of calves per female cow was not determined, and no information on the recruitment or productivity was generated. The group structure was however described such as calf crop, mean group size, and the number of single lone bull encounter.

To determine the number of muskox in the study area, only adult muskox sightings recorded on transect were analyzed using Jolly's Method 2 for unequal sample sizes (Jolly 1969) using a coefficient limit of 95%. Such methodology is commonly used in calculating the muskox estimate in other management units in Nunavut. The population estimates for fixed-width strip sampling using Jolly's Method 2 for uneven sample sizes (Jolly 1969; summarized in Caughley 1977) are derived from the following equation:

$$\hat{Y} = RZ = Z\frac{\sum_i y_i}{\sum_i z_i}$$

Where $\hat{Y}$ is the estimated number of animals in the population, R is the observed density of animals (sum of animals seen on all transects $\sum_i y_i$ divided by the total area surveyed$\sum_i z_i$), and Z is the total study area. The variance is given by:

$$Var(\hat{Y}) = \frac{N(N-n)}{n}\left(s_y^2 - 2Rs_{zy} + R^2 s_z^2\right)$$

Where N is the total number of transects required to completely cover study area Z, and n is the number of transects sampled in the survey. s_y^2 is the variance in counts, s_z^2 is the variance in areas surveyed on transects, and s_{zy} is the covariance. The estimate $\hat{Y}$ and variance $Var(\hat{Y})$ are calculated for each stratum and summed. The Coefficient of Variation (CV = σ/$\hat{Y}$) was calculated as a measure of precision.

Density, the number of muskox per unit area (muskox/km^2), will be determined using the number of adult muskox seen on transect divided by the total area of the study area. Lakes and stream areas will be not subtracted from the total area calculations used in muskox density (Statistical analysis based on Campbell and Setterington (2001).

The area occupied by muskoxen and the time of the survey within the study area was determined. Thus, the distribution was illustrated by plotting each muskox sighting on transect based on their precise geospatial position captured with GPS. In addition, the number of animals composing each group was highlighted using an increasing size of circles to represent groups of 0-1, 2-7, 8-11, 12-15, and 16-19 animals.

Given the importance of predators, Arctic wolf, wolverine, and grizzly bears, we collected standardized information of predator sightings in the management units using the predator index (Heard, 1992). The predator index reports all predator sighting per species against the total number hours flown, in this case also including the ferry time. It is then possible do have a yearly trend, as the number of predators observed is expressed per 100 hours.

Results

The survey was conducted out of the community of Kugluktuk from August 25th to August 30th, 2017. The management unit was surveyed at 16% coverage, which was surveyed in 51 hours, including on transect and ferry flight from Kugluktuk airport to the start of the transect lines. Low ceiling and fog prohibited the ability to survey continuously from the west side to the east on the management unit. Therefore, some sections of the coastline were left to be completed at a later time, when weather was permitting. The sedentary muskox behavior (Adamczewski *et al*., 1997) reduces the probability that an individual moves a significant distance within the short survey time frame.

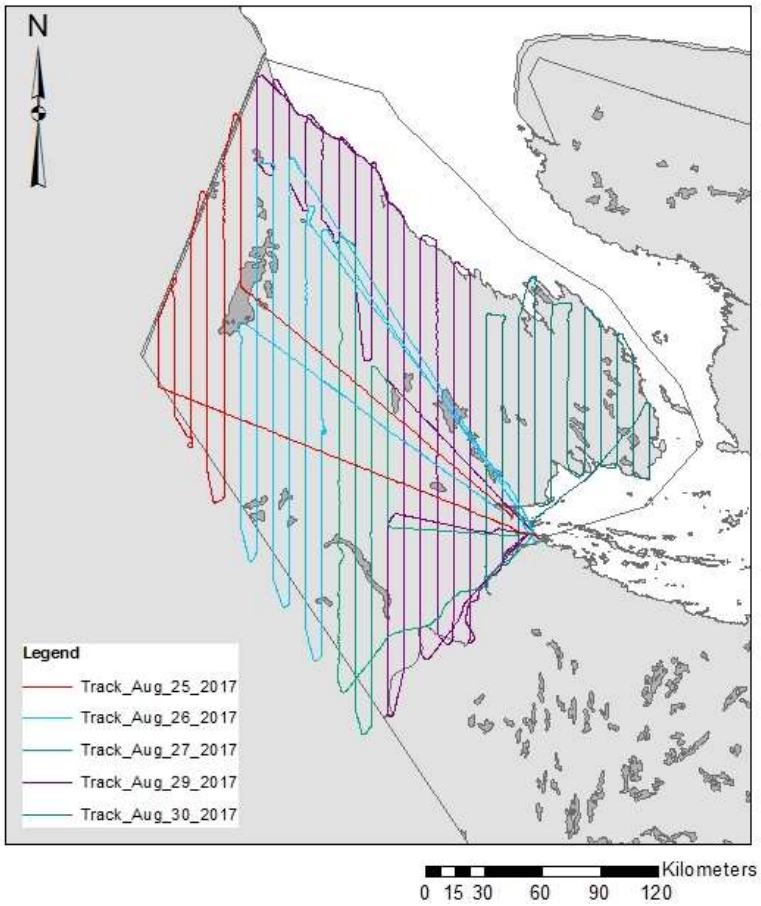

Figure 3: Daily track flown to cover 16% of the muskox management units MX-09 from August 25th to 30th.

Distribution

The adult muskox distribution in the management units is represented in the Figure 5 below. During the survey, 18 groups of muskoxen were seen on and off transect. The large groups of muskoxen, 16-19 adult animals, were distributed mostly north of Dismal Lake, and 25 km from the coast south of Stapylton Bay. Most muskoxen sighted were found between Emagyok Lake and Dismal Lake in the Rae-Richardson River Valley. No muskoxen were seen on the west side of the management unit, or around Bluenose Lake.

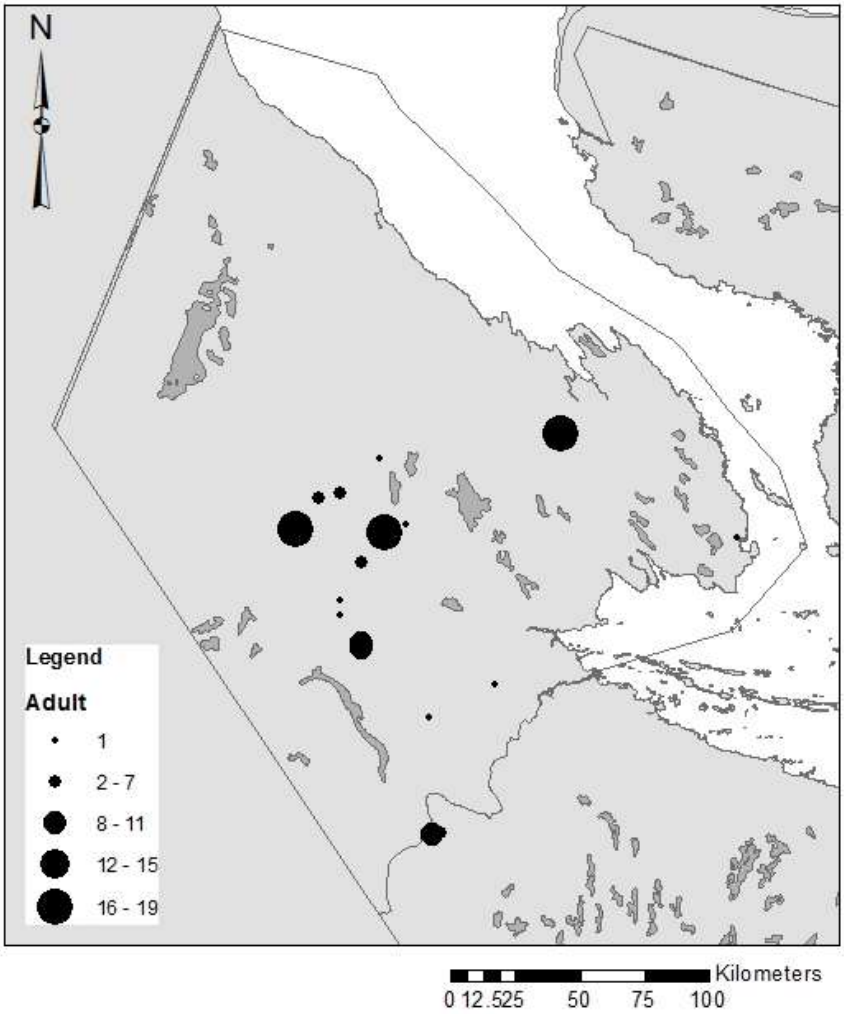

Figure 4: Muskox distribution, on and off transect, in the management unit MX-09 during the survey where the number of animals per group was grouped as 0-1, 2-7, 8-11, 12-15, and 16-19.

Group Characteristic

During the survey, 14 groups of muskoxen were recorded on transect, and seven groups were single muskox, mostly lone bulls. The majority of the groups (58%) were small groups of 2 to 11 adults (Figure 5). The average number of adults (+1 year and older) per group was 6.21 ± 6.6 (S.D.). The highest number of adults counted in one group was 18. Calves were not including in the group size, but in the calf to adult ratio.

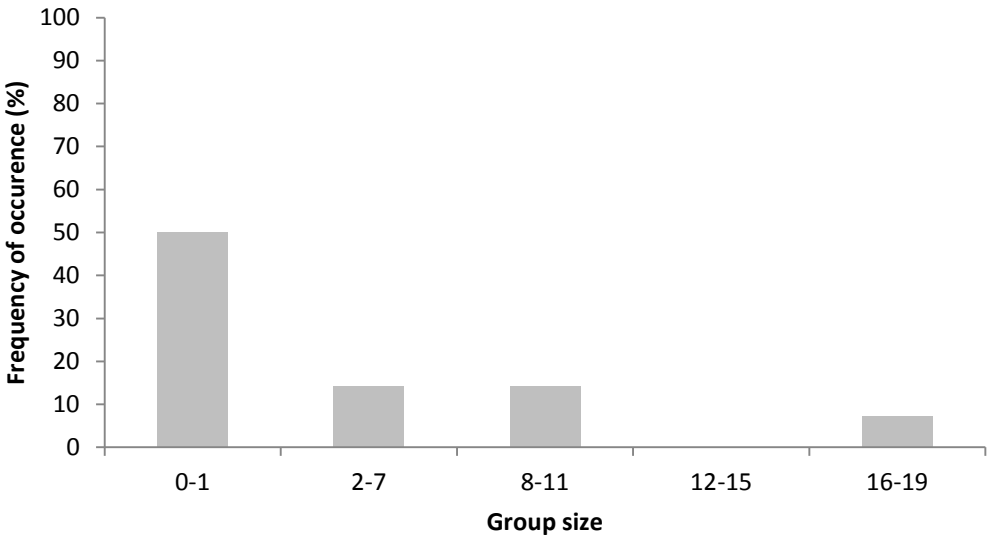

Figure 5: Frequency of occurrence (%) of adult muskox number per group size, grouped as follow 0-1, 2-7, 8-11, 12-15, and 16-19.

The calf to adult ratio was determined for each group of muskox seen on transect. Since the identification was done from a fixed-wing, it was impossible to distinguish with certainty the sex of the adult or the group age based on the horn shape and length. A little more than one third of the groups seen (36%) had calves. For the group that had calves, most of them had at least one calf, but some larger groups had 12 and where distributed mostly in the Rae-Richardson River Valley (Figure 6). Thus, the calf to adult muskox ratio was 38 calves/100 adults.

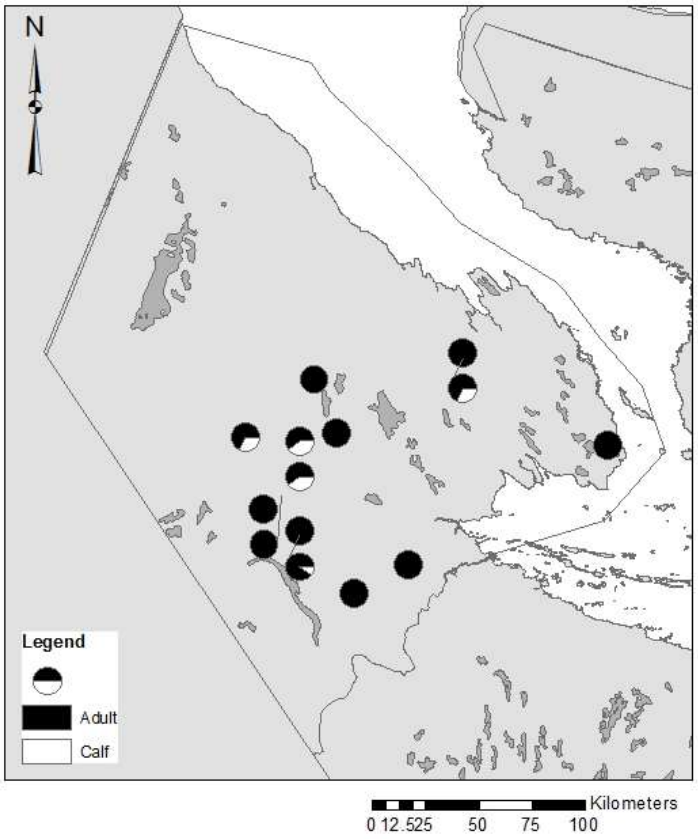

Figure 6: Proportion of calves per adult muskox in each group observed on transect in the Muskox Management Unit MX-09.

Estimate

The percentage of the overall cover of the management unit surveyed with 8,591 km^2 represented 16% of the total study area (53,215 km^2). During the survey, 87 adult muskoxen on transect were recorded. The estimated number of muskox in the management unit MX-09, is 539 ± 150 (S.E.) ($p<0.005$, $t = 1.696$, $N = 132$ and $n = 31$). For this estimate, the total number of transect at 100% coverage was 132 (N) and 31 (n) transect lines were surveyed (Table 2). Overall, the muskox density of the management unit was 0.010 muskox / km^2.

Table 2 Muskox estimate in the Muskox management Unit MX-09

Stratum	Area Survey (km^2)	Total area (km^2)	Muskox on Transect	Estimate	Standard error (S.E.)	95% CL (±)	CV
MX-09	8,591	53,215	87	539	150	255	0.28

* $p<0.005$, $t = 1,696$, $N = 132$ and $n = 21$

Predator sighting (wolf, wolverine, and grizzly bear)

In 2017, during the 51 hours of flying within the management units, 6 wolves, 2 wolverines, and 26 adult grizzly bear sightings were recorded (Figure7). The wolves (blue dots) were found on the southern part of the study area, south of Dismal Lake, meanwhile the wolverines (yellow dots) were along the coast or between Bluenose Lake and Emagyok Lake. Grizzly bears (red squares) were found dispersed in the study area in 24 different locations. Among them, one group was composed of a female and cubs and two other groups were composed of a female and two cubs. Predator sightings, using the predator index, (Heard, 1992) reveled 12 wolves / 100 hours, 4 wolverines / 100 hours, and 51 grizzly bears / 100 hours.

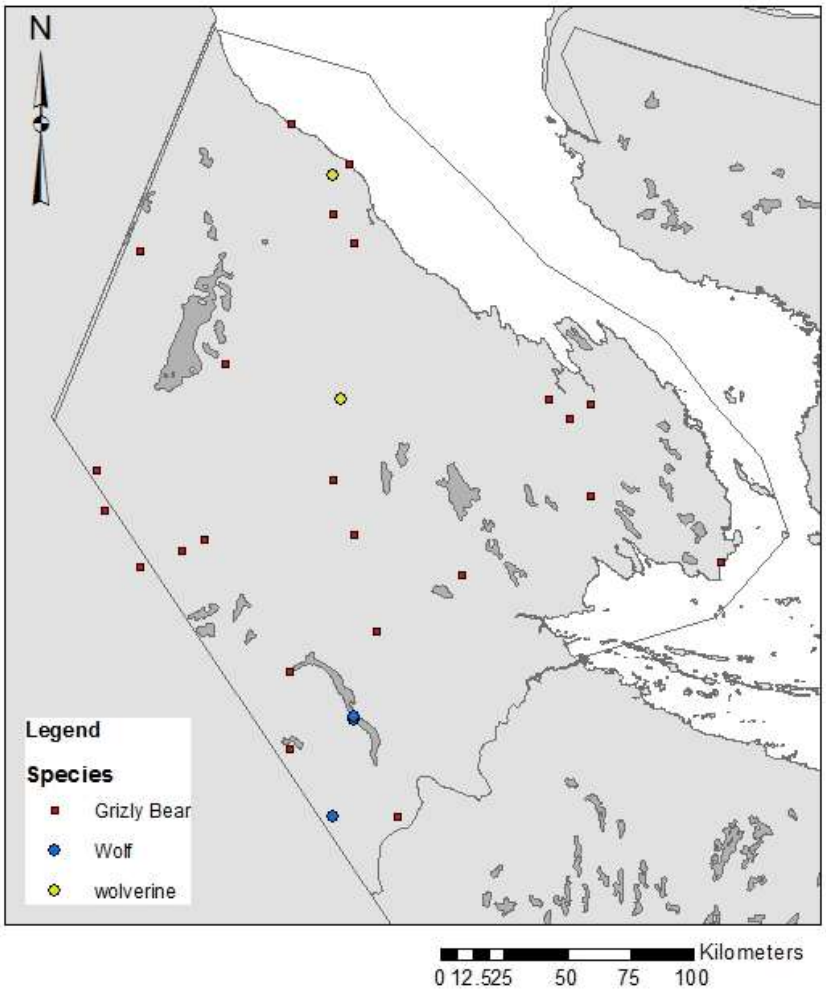

Figure 7: Locations of Wolf, Wolverine, and Grizzly Bear observations in the Muskox Management Unit MX-09.

Discussion

Distribution

Historically muskoxen were abundant in the area of Dismal Lake, and in the 1950s-1960s few muskoxen were seen in the Bluenose Lake area (Kelsall *et al*., 1971). No muskoxen were observed around Bluenose Lake in 2017. In 2007, the majority of the muskox observations on and off transect were north of Emagyok Lake (Dumond, 2007), whereas in 2017 most of the muskoxen were distributed south of the same lake in the Rae-Richardson River Valley, which is typical of muskox suitable summer habitat characterized by a wetter and lower-lying area within the proximity of uplands (Danks and Klein, 2002). Thus, the difference between the 2007 and the 2017 muskox distribution suggests a summer movement away from the coastal plain. Such change in distribution could be explained by the specific time that the survey was completed, June versus late August. Muskox have potentially moved their main distribution slightly to the south and this might be due to a change in the aridity of the soils that influenced the foraging conditions and due to the proximity the upland serves as an escape terrain from predators. The use of summer areas may vary from year to year and within the season.

Group Characteristic

The small group sizes encountered, 6.21 ± 6.6 (S.D.), during the survey is attributable to the timing in which the work was completed. After the peak of the rut, in mid-August, the herds are still small and have not started to merge into the larger wintering herds. They are then easier to count. The small herds are the result of competition within herds to access the rich forage of the Rae-Richardson River Valley. Bulls tend to be solitary, which explains the number of solitary muskox seen during the survey (7 groups of solitary muskox) (Gunn, 1990). During mating season, one strategy suggests that the bulls disperse to ensure less contested breeding opportunities, reducing the cost of combat between adult bulls.

The calf to adult muskox ratio was 38%. This ratio is normally associated with a population that would be increasing, since it has been established that 10.5% of calf to adult ratio is necessary to keep the muskox population stable (Freeman, 1971). Since the calf ratios have been known to vary greatly between years, longer term data is needed to determine a trend (Reynolds, 1998).

Density

The mechanism driving muskox density is not fully understood. Heard (1992) noted that group size is not generally related to muskox density. In this management unit, a vast area to the west of the management unit does not seem to be utilized by muskox that, at low muskox numbers, drive the overall density down- 0.010 muskox / km^2. The 2017 muskox density in the

management unit MX-09 is the lowest muskox density recorded in the Kitikmeot Region since intensive survey work started in 2013.

Abundance Estimate

The extent of this area remained relatively the same even after all the muskox management units were re-delineated across Nunavut in 2015. For this management unit, small changes included extending the delineation to reach the Northwest jurisdiction boarder to the west and south. This relative similarlity between the old and new delineation allows the 2017 survey to be compared to the previous population estimates.

During the muskox moratorium, the small number of muskoxen remaining on the upper Rae-Richardson River Valley were able to increase in number on the north part of Great Bear Lake. In 1980, the muskox population in the area was estimated to be 869±279 (S.E.) and the successive survey, done three years later, recorded a continuous increase in the muskox population of the Rae-Richardson Rivers watersheds with an estimate of 1,295 ± 300 (S.E.) animals (Gunn 1995; Fournier and Gunn, 1998). The peak of the muskox population in the area was recorded in 1988, with 1,805 ± 289 (S.E.). From the three consecutive surveys, in 1980, 1983, and 1988, the number of tags increased respectively from 12, 35, to 50. In 1994, muskox abundance in this area plummeted to 540 ± 139 (S.E.) muskoxen and remained relatively stable based on the subsequent survey conducted in 2007 with a population estimate of 589 ± 121 (S.E.) muskoxen (Gunn 1995; Dumond, 2007). The latest estimate of 2017 with 539 ± 150 muskoxen show that this population has remained stable. This current status is supported by the observations of hunters.

Since 1994, the muskox population has remained relatively stable and the number of tags has also remained consistent with 20. Environmental factors such as predators, forage quality and quantity, diseases, and harvesting might have contributed to the stabilization of the muskox population in this management unit. Ongoing and more frequent population monitoring, carried out every 5 years, might allow for early detection of signs of recovery in the herd through indicators of population growth. Information provided in this report should guide future muskox surveys in allocating the survey effort proportionally with the distribution of muskox. This will help to produce a more precise estimate.

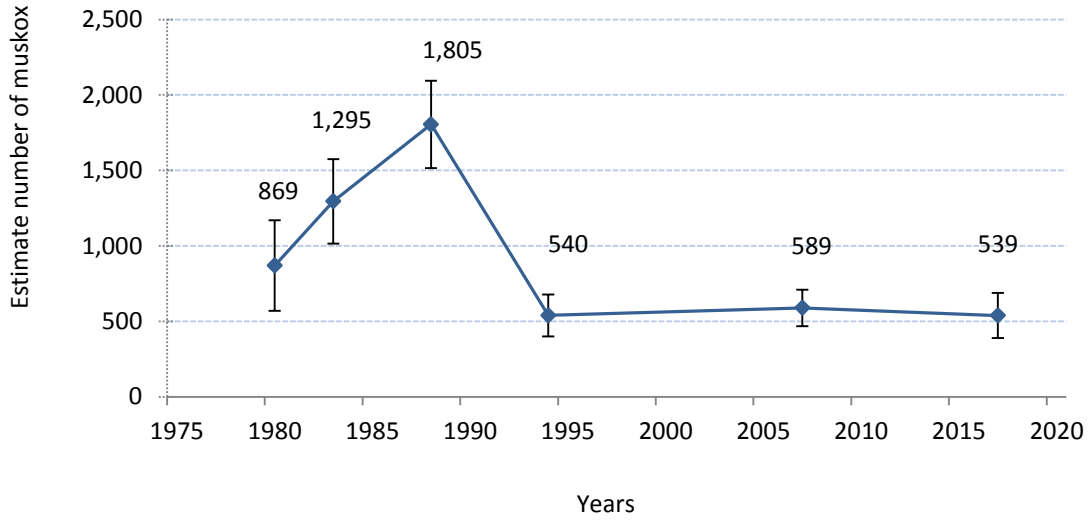

Figure 8: Muskox population estimates for MX-09.

Predator sighting (wolves, wolverine, grizzly bear)

Wolves and wolverines are muskox predators and are found in the study area and pose an additional cause of mortality that might affect the muskox population to a greater extent in the future. Although the number of grizzly bear sightings (51/ 100 hours) is greater than the other two predators (12 wolves / 100 hrs and 4 wolverines/ hours), their impact on influencing the muskox dynamics might be disproportional to their number, since their diet mostly consists of small tundra herbivores (e.i. ground squirrels) (L'Herault *et al*., 2016). However, the grizzly bear has adapted to their tundra environment and have maximized of available food sources. Local community members have observed novel grizzly bear hunting strategies which involve chasing the herd to break the calf-cow pair and then sit there to attract the lost muskox calf, as the muskox calf is attracted by the darker silhouette. The community members are however more concerned about the predation effect on caribou, than the effect of these predators on the muskox.

Acknowledgements

I wish to thank the pilot, Dan Hall, for his dedication and in making the survey possible. A special acknowledgement goes to Darien Evyagotailak for his logistical support and his enthusiastic participation. I am grateful to the Kugluktuk HTO for assistance in providing meaningful insight essential to the planning of this survey and the selection of observers: Aislyn Bolt, Jonathan Niptanatiak, Saveah Ilgok, Albert Anavilok, Amber Adjun, Kevin Ongahak; their involvement needs to be acknowledged as they assisted as observers during the aerial survey. I also want to acknowledge the precious time that Tristan Brewer gave as technical support for the data analysis. This project was funded by the Department of Environment (Government of Nunavut) and the Nunavut Wildlife Research Trust Fund under the Research Permit 2017-13.

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Building Nunavut Together
Nunavuliuqatigiingniq
Bâtir le Nunavut ensemble

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ᔪᓂ, 2019

ᑲᑎᒪᔨ ᓄᓇᕗᑦ ᐱᖃᓕᕆᔨᒃᑯᑦ
Building Nunavut Together
Nunavuliuqatigiingniq
Bâtir le Nunavut ensemble

ᓄᓇᕗᑦ ᑲᖏᒪᒃᑯᑦ ᐃᒪᐃᓕᒪᔪᑦ:

ᓄᓇᕗᑦ ᑲᖏᒪᒃᑯᑦ-ᐊᖏᑎᓕᓱᓴᒃᑯᑦ (DOE) ᑐᓴᖅᑎᕐᑎᓯᒪᔪᑦ ᑲᑎᒪᓴᓂᒃ ᖃᐅᓴᓴᓂᓂᐅᑦ, ᐊᓪᓚᓐ ᖃᒥᖅᑭᓂᐅᑦ, ᐱᖓᓕᓕᐊᓂᕐᓗᒍᓂ ᐅᒥᖓᒪᐃᑦ ᐊᐅᓚᕐᑎᓂᖅ ᐃᓐᓕᓕᖅ, MX09–Ⅾ 2017–Ⅾ, ᐊᓪᓚᓐ ᑐᓂᓯᓐᑎᒃ ᐄᓱᓇᐅᔪᖅ ᐊᐅᓚᕐᑎᓂᕐᒧᑦ ᐊᑐᓕᖁᕆᐅᓯᒪᔪᑦ

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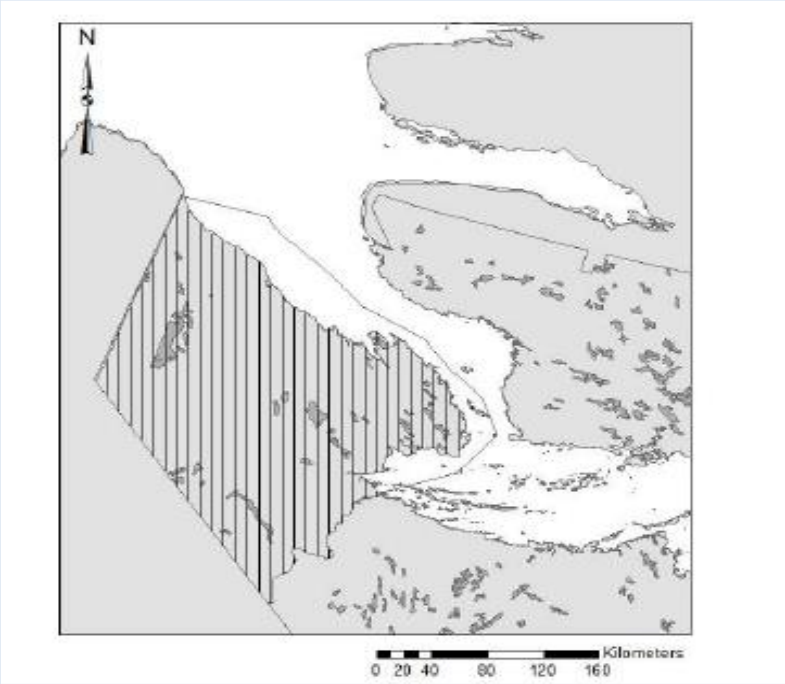
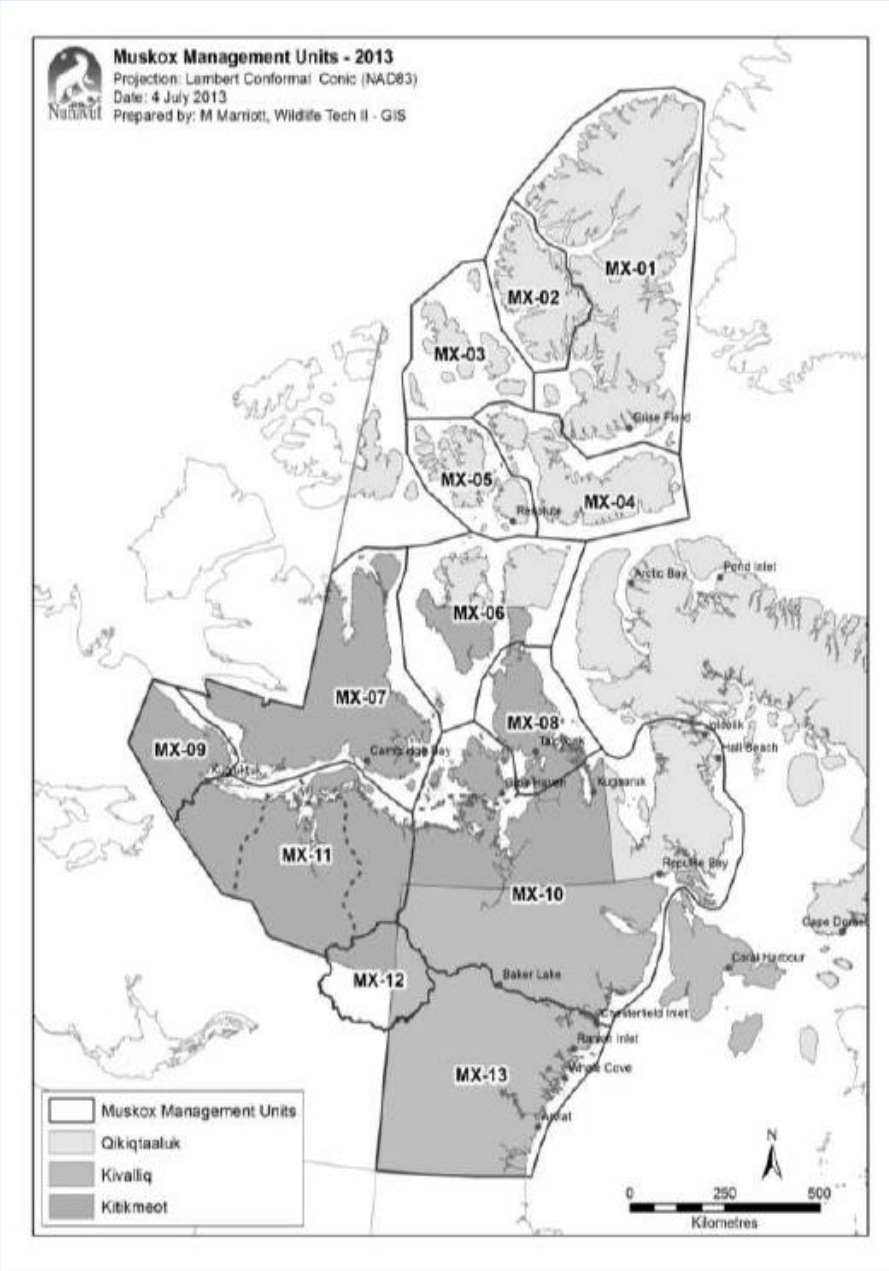
- MX–09, ᑭᖓᑦᓗᖅᓯᒋᑦ ᓇᓗᐅᑦᑕᖃᑦᑕᐅᔪᑦ 2007–ᒥ 598 ᐅᒥᖓᒪᐃᑦ, 159 ᐃᓐᓇᐅᑦᓗᑎᒃ ᐊᐃᕐᓯᒪᔪᒥᑦ.

- ᐃᓱᒪᓕᕆᑦ, ᐊᖓᔪᓇᓴᕆᑦ ᓂᕿᒃᓴᓂᖅᑕᓇᐅᑉᐳᖅᑐᑦ ᑐᓴᑐᓂᒃ ᐃᓇᖓᑎᑕᐅᑦᓗᓂ ᑕᐢᖓᓂ ᐃᓂᐅᔪᒥᒃ.





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- ᖃᐅᔨᓴᖅᓂᖅ ᐱᓕᕆᖃᑕᐅᓕᐅᖅᑐᖅ ᐊᑲᓗᓯ 25–ᒥ ᔪᑎᐱᓕ 02–ᒍᑦ, 2017 (16% ᐃᓂᐅᔪᒥᒃ)
- 103 ᐃᓐᓇᐃᑦ ᐅᓴᖅᑕᐅᓕᐅᖅᑐᑦ ᐊᓚᒐᓱ 87 ᐊᐃᒃᓯᒪᔪᒋᑦᓴᑎᒃ

- 58% ᑲᑎᖏᓚᑦ ᐅᓄᖑᓯᑉᓗᑦᓱᑎᒃ ᑲᑎᖏᓚᓕᐅᖅᑐᑦ, 2–ᒥᑦ 11–ᒍᑦ ᐃᓐᓇᐅᑦᓱᑎᒃ
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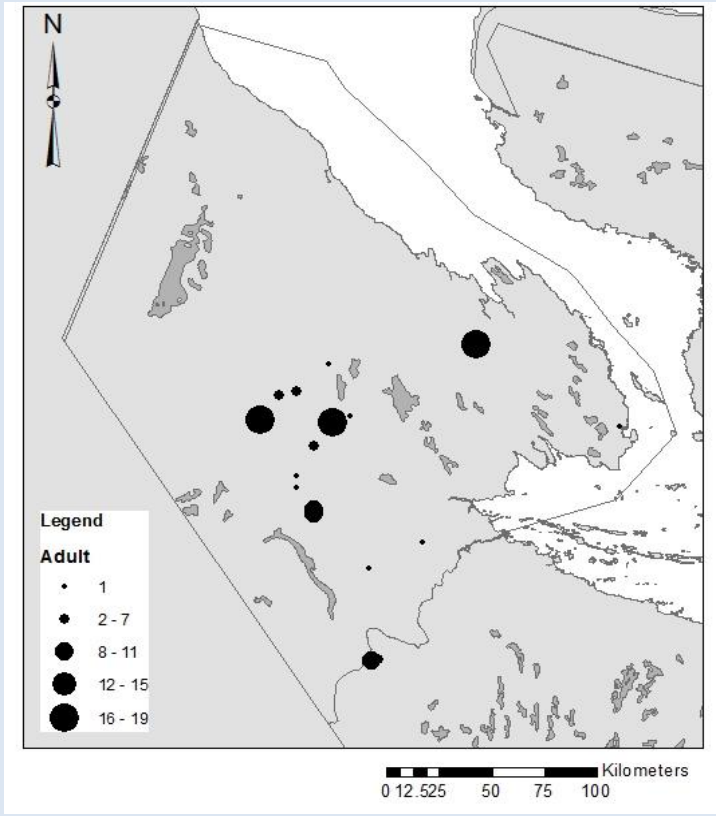

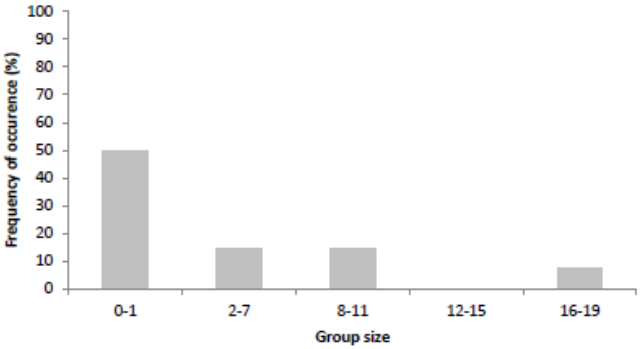

Figure 5: Frequency of occurrence (%) of adult muskox number per group size, grouped as follow 0-1, 2-7, 8-11, 12-15, and 16-19.

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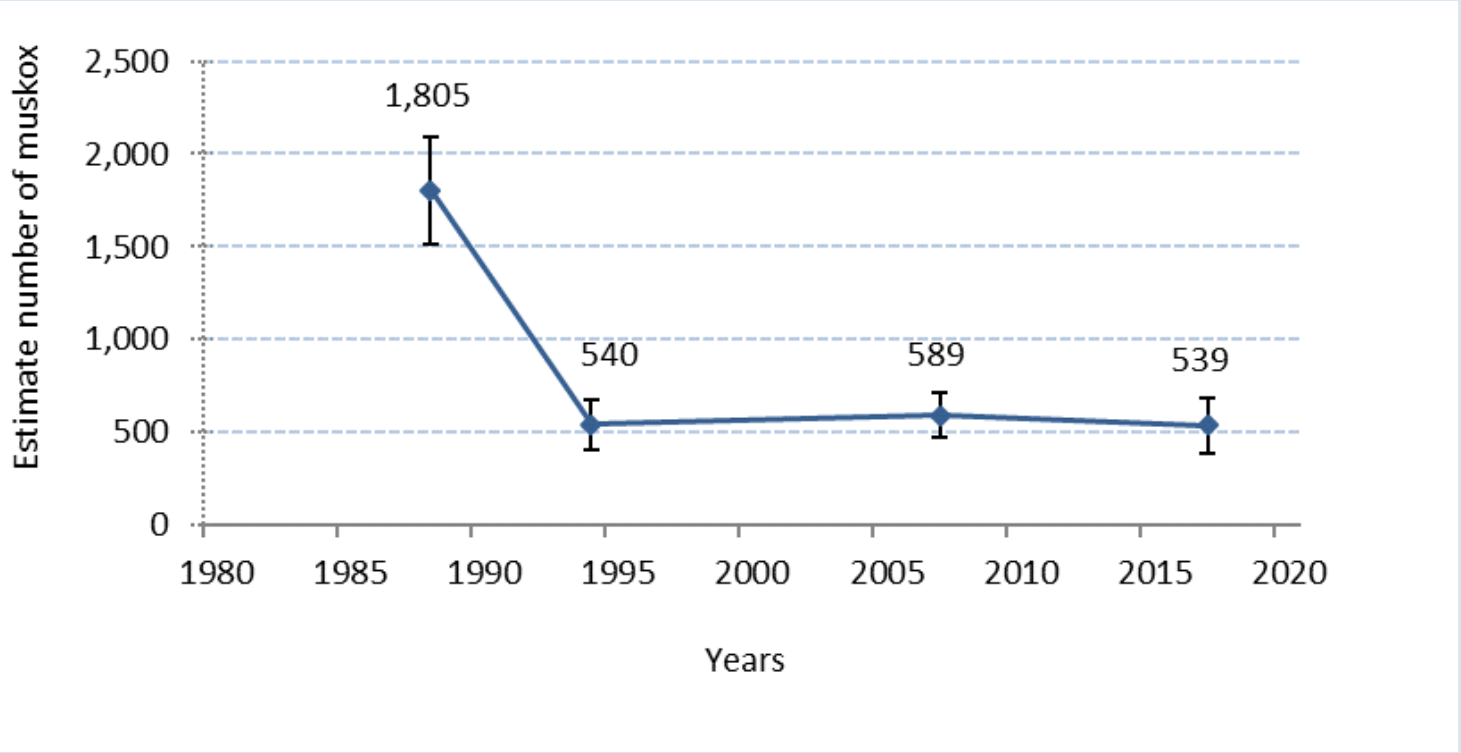
539 ± 150 ᑕᒪᒪᕐᓂᖓ SE



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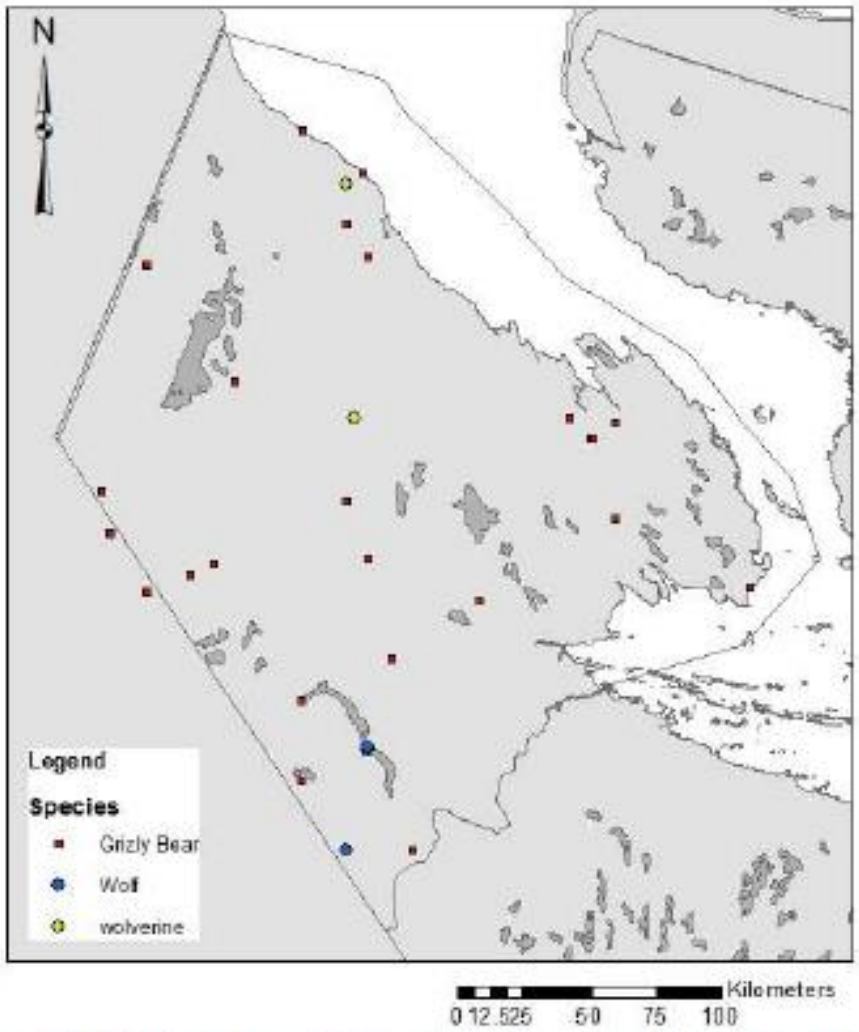

Figure 7: Locations of Wolf, Wolverine, and Grizzly Bear observations in the Muskox Management Unit MX-09.

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ᑭᓯᐊ ᐊᖓᔪᓕᓴᐃᓂᒃ ᒥᑭᒋᐊᖃᓂᐊᖅᐃᓂᓪᓗ ᑲᑐᔨᖃᑎᒋᐊᖅᓯᒪᔭᖓᓂᒃ (HTO)–ᒃᑯᓂ ᐊᐅᓪᓗᑎᓂᕐᒧᑦ ᐱᓕᕆᔨᒃᑯᐅᕐᓂᒃ?

ᑯᒋᐃᓂᒃ ᓄᖓᒋᐅᖃᖅᑕᖅᑭᓂ?

ᓄᓕᓪᓂᒥ ᑭᓯ ᐱᓯᒪᔪᐅᐊᖓ?

ᑲᑎᒪᓗ ᓄᓇᕗᑦ ᐱᓕᕆᔨᒋᐊᕐᓂᒃᑯᑦ
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Bâtir le Nunavut ensemble

ᓄᓇᕗᒻᒥ ᑲᖏᒪᑲᓂᓕ ᐊᑐᓕᕐᓂᐊᕐᓗᒍᓪᓚᕆᓪᓚᕆᓪᓕᕐᓂ

ᖃᐅᔨᓴᖅᓂᖅᓯᒪ ᖃᐅᔨᓴᕆᔨᕐᓗᒍᓪᓚᕆ ᐊᓚᒪᓂ ᓄᓇᓕᖅᓂ ᐊᓚᒪᓂ ᐊᖏᓐᓇᓱᑲᑎᑦ ᒥᑭᒥᐊᖅᓂᐊᖅᑎᑦᓂ ᑲᑐᔨᕿᑎᑎᖅᑐᑦᓯᓂᖅᓂᑦ (HTO)–ᑲᑎᓂ ᑐᓴᐅᒪᑎᑦᑎᖅᑦᑕᐃᓐᓇᖅᓂᒥᒻ ᒪᓕᒃᓯᑎᓪ, ᐅᑎᐊᒐᓴᖅ ᐊᑐᓕᕐᓂᐊᕐᓗᒍᓪᓚᕆᓪᓚᕆᓪᓕᕐᓂ ᐃᖅᑭᒃᑕᐅᔪᓪᓚᕆᓪᓚᕆ ᐅᒥᖏᒪᐃᓪᓚᕆ ᐅᒥᖏᒪᐃᓪᓚᕆ ᐊᐅᓚᓐᓇᓕᔪᓐᓇᑎᑦ ᐃᓗᓕᕐᑎᑎᑦMX–09–Ꮀ:

ᐊᐢᑎᓕᓚᔪᓂᒃᑯᑦ ᐊᑐᓕᕐᓂᐊᕐᓗᒍᓪᓚᕆᓪᓕᕐᓂ ᓄᓇᕗᒻᒥ ᐅᒪᓕᔪᓕᓚᔪᓂᒃᑯᑦ ᑲᑎᒪᓂᔪᓐᓇᓂᓪ ᐱᓗᐅᔪᓐᓇᖅ ᖃᐅᔨᓴᕆᔨᕐᓗᒍᓪᓚᕆᓪᓚᕆᓪᓕᕐᓂ ᐊᖏᓐᓇᓱᑲᑎᑦᓯᓂᖅᓂᑦ ᖃᐅᔨᓴᕆᔨᕐᓗᒍᓪᓚᕆ (TAH) 20–ᒋᑦᓯᓗᓂᒃ ᐅᒥᖏᒪᐃᓪᓚᕆ ᐊᐅᓚᓐᓇᓕᔪᓐᓇᖅ ᐃᓗᓕᕐᑎᑎᑦ MX–09–Ꮀᑦ.

ᐊᖏᓇᓂᖅᑎᑦ ᒥᑭᒋᐊᕐᓂᐊᖅᑎᑦᓗ ᑲᑐᔾᕕᖅᑎᑭᒍᓐᓂᐊᓂᑦ (HTO) ᑐᓴᐅᒪᑎᑦᑎᕕᑦᑕᐃᓇᕐᓂᖅ ᐅᒥᖏᒪᐃᑦ ᓇᒡᓘᓂᖅ ᐊᓪᓚᓗ ᐅᓇᕐᓂᖏᑦ, ᐅᒥᖏᒪᐃᑦ ᐊᐅᓚᑎᓂᖅ ᐃᓗᓚᖏᑦ MX–09

ᐊᐃᕐᓇᓚ 2019

ᐊᕐᕌᑎᓚᓵᖅᑯᑦ, ᓄᓇᕗᒻᒥ ᒪᕐᕉᒪᖅᑯᑦ ᐃᕐᑲᓗᐃᑦ, ᓄᓇᕗᑦ

ᐊᐅᓪᓗᓂᔭᕐᓂᖅ ᓇᐃᓐᓇᕐᓂᕐᓂᖅ

ᑭᖑᓂᖅᑐᐃᓕᐅᕐᓂ ᓄᓇᕗᑦ ᑲᑎᒪᔨᑦᓂᓪ, ᐊᖏᑎᓕᕆᓚᐅᑦᓂ (DOE) ᑐᓴᐅᒪᑎᑕᑎᓂᖅᓱᑕᐃᓇᖅᑎᑦᑎᒥ ᐱᓕᕆᓕᐅᕐᓂᖅᑐᑦ ᖃᐅᔨᓂᖅᑐᑦ ᐊᖓᔪᓇᓕᕐᓂᐊᖅᑎᑦ ᒥᑭᐅᑎᐊᖅᑎᑦᓴ ᐅᑐᐱᓂᖅᑎᑎᖅᓯᓂᑦ (HTO)–ᑲᑦᖁᓕᓪ ᐊᐃᖅᕆᓕ 2–ᒥ, 2019. ᑐᓴᐅᒪᑎᑕᑎᓂᖅᑎᑦᑕᐃᓇᖅᑎᐅᕐᓂ ᐱᔨᔪᑎᑎᓕᐅᕐᓂᖅᑕᖓᓕ ᑐᓴᓂᖅᑎᑎᐃᓇᓗᑐᑦ ᓇᓂᐅᓕᖅᑎᐅᕐᓂ ᓇᐅᓕᕐᑕᐃᓂᖅᑕᐅᓗᒪᔨᓂᖅ 2017–ᒥ ᐅᓄᖅᓯᒪᓂᓂᑦ ᐊᓕᐅᕐᑕᐃᓂᖅᑕᐅᓗᒪᔨᓂᖅ MX–09–ᒥᑲ ᐊᓚᒪᓗ ᐅᓂᖅᐅᓗᖅᓱᓗᑎᑲ ᐊᐅᓕᕐᓂᑎᓂᖅᒥᑦ ᐱᓕᕆᖏᑦᑐᐅᓗᒪᔨᓂᖅ ᖃᐅᔨᓂᖅᑐᖅ ᐊᖓᔪᓇᓕᕐᓂᐊᖅᑎᑦᓴ ᑲᑐᔨᖅᑎᑎᐃᓇᖅᑎᑦᓂ (HTO)–ᑲᐅᑦ ᐱᖓᓱᖅᑎᑕᐅᓂᖅᖏᑦᑐᖓᓂᖅ. ᑐᓴᐅᒪᑎᑕᑎᓂᖅᑎᑦᑕᐃᓇᖅᑎᐅᕐᓂᖅ ᐃᓇᐅᑎᖅᑕᐅᓂᖅ ᐅᑎᖅᑕᐃᓂᖅᑐᐅᓗᒪᔨᓂᖅ ᐊᐅᓚᑦᓯᐅᑎᖅᑕᐅᓗᒪᔨᓂᖅ ᐃᓚᐅᕐ

ᐅᖏᐅᓯᖏᓐᓇᐊᑕᑭᑎᖅ

ᐅᓇ ᐅᓂᒃᑲᖅ ᑭᓱᖏᖅᑐᐃᔪᖅ ᐊᐢᕈᑎᑕᓐᓇᓕᒃᑯᑦ ᐊᒃᓴᑲᖅᓂᖅᓱ ᐱᐅᓪᓚᑕᒻᒥᓗᓂᒃ ᐅᓂᒃᑲᐅᓯᖏᓐᓄᑦᑐᓂᒃ ᐅᖏᒃᑲᐅᓯᐅᓪᓚᑦᑐᖅᑐᒃ ᑐᑎᓪᓚᖅᑕᐅᖏᓐᓄᑦᑐᑦᓯᐊᓂᑦ ᐅᑭᒪᓐᓇᖅᑐᑦ ᐊᖏᔪᓇᓕᒃᑯᑦ ᒥᑭᑎᐊᖏᓂᐊᖏᒃᑯᑦᓗ ᑲᑐᔪᖏᓂᑎᒃᓴᐃᓂᖅ.

ᐃᓕᓴᕆᐅᓕᐊᑦ ᑕᕐᕈᓂ ᐊᐢᕈᑎᑕᓐᓇᓕᒃᑯᓂᖅᓱᓂᖏᓐᓄᑦᑐᓕᕆᓂᖅᑐᑦ, ᓄᓇᕗᒻᒥ ᐅᕐᕈᑐᓂᓂ ᓄᓇᕗᑦ ᓴᕈᓚᖏᓂ.

ᖃᓂᖅᓴᖅ ᐃᓗᓕᖅᑎᓂᓂᖅ

ᐊᐅᓚᖅᑎᓂᖅᓯᒋᐊᕐᓂ ᓇᐃᓐᓇᕆᔨᒪᔨᕐᓂ i
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1.0 ᐅᓂᒃᑲᓄᑦ ᐱᓱᑎᐅᔾᔪᑦ ᐊᓪᓚ ᐊᑦᑭᒃᓯᒪᓂᖅ

ᐅᓇ ᐅᓂᒋᖅ ᐱᓱᑎᖃᖅᑐᖅ ᑲᑎᑎᒧᓗᓂ ᐊᓪᓚᓗ ᓇᐃᓐᓇᓕᖅᑎᒧᓗᓂ ᐅᖃᐅᓯᐅᔾᔪᑎᓂ, ᐊᐱᖅᑯᑕᐅᔾᔪᑎᓂ, ᐃᓯᒻᒋᑕᐅᔾᔪᑎᓂ ᐊᐅᓚᓂ ᑎᑎᒧᓗᓂ ᐅᖃᐅᓯᐅᔾᔪᑎᓂ ᐊᓪᓚᓗ ᑎᑎᒧᓗᓂ ᐱᓕᕆᓂᐅᔾᔪᑎᓂ ᐅᖃᐅᓯᐅᔾᔪᓕᕆᓂᓗ ᑐᐱᐅᓚᑎᑦᑎᓂᖅᑐᓂᓗ ᑐᑭᓯᐅᑦᑎᓂᒃᑯᑦ ᐊᖏᖅᑯᓯᒪᔪᓂᒃ (HTO)–ᑲᑯᖏᓐᓂ ᐊᐃᕐᓇᓂ 2–ᒥᑦ, 2019 ᖃᐅᔨᓴᓯᒪᓕᕆᓂ ᐅᒥᖓᓚᐃᑦ ᐊᐅᓚᕋᑎᓂᖅ ᐃᓗᓕᕆᓂ MX–09 ᓇᒡᓘᓂᖏᓐ ᐊᓪᓚᓗ ᐅᓄᕐᓂᖏᓐ ᖃᐅᔨᓴᕐᓂᐅᓚᐅᖅᑐᑦ ᒥᖅᓴᓄᑦ ᐊᓪᓚᓗ ᐊᐅᓚᕋᑎᓂᕐᒧᑦ ᐱᓚᕆᔨᖃᐅᔨᓚᕆᓂᖓ. ᓇᐃᓐᓇᖅᓯᐊᑦ ᐊᓪᓚᓗ ᑎᑎᖅᑲᐃᑦ ᑕᑯᑎᕋᑦ ᐅᖃᐅᓯᐅᓚᐅᖅᑐᓂᒃ ᖃᑎᒌᓄᑦ.

2.0 ᑐᐸᐅᒪᑎᕋᖅᑕᐃᓐᓇᕐᓂᒧᑦ ᐱᓱᑎᑎᑲᐅᔾᔪᑦ

ᖃᑎᒌᓂ ᐁᖅᑭᖅᓯᖅᑕᐅᔾᔪᑦ ᐊᓪᓚᓗ ᐊᐅᓚᑕᐅᓂᖅᓱᑕ ᐊᐋᑎᓕᓚᓵᖅᑯᑦ (DOE)–ᖅᑯᐟᓄᑦ ᐱᓱᑎᑎᑕᐅᖅᑕᖅᓱᒐ ᐊᖏᓇᓯᐅᔾᔪᑎᓂ ᒥᑭᐅᓯᐅᔾᔪᑎᓂᒃᑯᑦ ᐅᖃᐅᓯᐅᔾᔪᑎᓂᒃᑐᓗ ᖃᐃᓐᓇᕐᓂᖅᑎᒍᑎᓂ (HTO)–ᖅᑯᐟᓂ ᐱᕕᓇᖅᑎᖅᑎᓗᓂ ᑐᐸᐅᓚᖅᑎᑦᑎᓂᖅᑐᓂᓗᐃᓐᓇᕐᓂᒧᑦ ᖃᐅᔨᓴᕐᓂᐅᓚᐅᖅᑐᑦ, ᖃᐅᔨᓴᕐᓂᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᐅᖃᐅᓯᐅᔾᔪᓕᕆᓂᓗ ᐊᐅᓚᑕᐅᓂᖅᓱᒐ ᐱᕕᓇᖅᑎᖅᑎᓗᓂ. ᖃᐅᔨᓴᕐᓂᐅᓚᓕᕆᓂ 2017–ᒥ ᐃᓄᐅᐊᖓᓂᖏᓐᓂ ᖃᐅᔨᓴᕐᓂᖅᑎᒥ ᐅᖃᐅᓯᐅᓚᐅᖅᑐᓂ ᑲᑎᒪᓂᓗᐅᑦ. ᐊᖏᓇᓯᐅᔾᔪᑎᓂ ᒥᑭᐅᓯᐅᔾᔪᑎᓂᒃᑯᑦ ᑐᐸᐅᓚᖅᑎᑦᑎᓂᖅᑐᓂᓗᐃᓐᓇᕐᓂᒧᑦ (HTO)–ᖅᑯᓂ ᐊᐅᑐᖅᑕᐅᓂᖅᑐᖅᑐᓂ ᑐᓴᐅᒪᓂᖅᑐᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᑎᓴᒪᓂᖅᑐᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᐱᓕᕆᔨᒃᑯᓂᖅᑐᓂᒃ ᐊᐃᓚᖅᑐᑦ ᐊᐅᓚᑕᐅᓂᖅᑐᓂᓗ ᐃᓕᓴᕐᓇᓗᒍ, ᑕᐅᓇᓗ ᐅᓂᒃᑲᓂᐅᔾᔨᓗᐊᑕᐅᓚᐅᖅᓱᒐᓂ. ᐊᐃᓐᓇᖅᑐᐅᓚᐅᖅᑎᓗᓂ ᐃᓕᓴᕐᓇᓗᒍ ᐊᐅᓚᑕᐅᓂᖅᑐᓂᓗ ᐱᓕᕆᔨᒃᑯᓂᖅᑐᓂᒃ ᐊᐃᓐᓇᖅᑐᐅᓚᐅᖅᑐᓂᒃᑯᑦ ᐱᓕᕆᔨᒃᑯᓂᖅᑐᓂᒃ ᐅᕙᓂ

3.1.1 ᖃᐅᔨᓴᖅᑐᖅ ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᓴ ᑲᑐᔪᖅᑎᑭᖏᓐᓇᐅᑦ (HTO) ᑐᓴᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓯᒪᔨᑦ ᓇᐃᐅᖅᓯᒪᔨᑦ-2019

ᐃᓕᓴᕐᕕᐅᔨᑦ: ᐅᒥᖏᒪᐃᑦ ᓇᒋᓐᓂᑎᑦ ᐊᒻᒪᓴ ᐅᓇᖅᓂᖏᑦ, ᐅᒥᖏᒪᐃᑦ ᐊᐅᓪᓚᑎᓂᖅ ᐃᓴᓪᓂᖏᑦ MX–09, ᖃᐅᔨᓴᖅᑐᑦ ᑰᖏᑕ ᐅᐊᓪᓂᐊᓂ, ᐋᓱᓕᓕ 2017. ᐅᒥᖏᒪᐃᑦ ᐅᓇᖅᓂᖏᑦ ᑕᕐᓀᓂ ᐊᐅᓪᓚᑎᓂᖅ ᐃᓴᓪᓂᖏᑦ ᓕᖅᑰᓕᕐᓂᑐᖅ.

ᑐᑭᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅᓗᑦ ᐱᔾᔪᑎᑎᑭᐅᔨᑦ:
ᑐᑭᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅ ᐋᑦᖁᕻᑕᐅᓪᓚᐅᔨᖅᑐᖅ ᖃᐅᔨᔨᖅᑐᑎᒥ ᐊᐃᖅᑏᓪᓕ 2–ᒥᑦ, 2019. ᑲᑎᒪᓂᐅᔨᑦ ᐱᔾᔪᑎᑎᓕᕆᔨᖅᑕᕐᓂᖏᑦ ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᓴ ᑲᑐᔪᖅᑎᑭᖏᓐᓇᐅᑦ (HTO)–ᖁᓕᓂ ᐱᓕᕆᖅᑲᑎᑭᖏᓐᓇᖅᓂᖅᓗᒥ ᑐᑭᐅᒪᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᑐᓴᓂᖅᑐᒃᓴᓂ, ᐅᓂᒃᑲᑦ ᖃᑎᖃᕐᓗᓂᑦ, ᐊᒻᒪᓴ ᐅᖃᖅᐅᓂᖅᑐᓂᓗᓂ ᓯᕗᓂᕆᓂᖅ ᐊᐅᓪᓚᑎᓂᖅᓗᑦ ᐱᓕᕆᓂᖅᑐᓂᓗᑦᑐᓂᖅ. ᑲᑎᒪᓂᖅ ᐱᓇᖏᓐᓄᐅᓕᕆᔨᖅᑕᐃᓐᓇᖅ ᐃᓕᑕᐊᖅᓯᒪᔨᓴᓂ ᑐᓴᖅᑎᑎᓗᓂᖅ ᐊᖏᓐᓂᓚᐅᔭᖅᑎᑦ (DOE)–ᖁᑦ ᐱᓕᕆᖁᓕᓯᒪᓂᑦ ᐱᓴᐅᔨᓇᖅᑎᑕᐅᔨᑦ ᐊᖏᓐᓇᓕᕆᔨᑦ ᐊᐅᓪᓚᑎᓂᖅᑐᓗᑎᒃ (TAH) 20–ᓂᒃ ᑐᓴᐅᒪᔨᐸᑎᓴᑦ ᐊᑐᐃᓐᓇᐅᐅᓂᖅᑐᑦ ᒪᓕᒃᓴᑦ, ᑭᓯᐊᓂ ᒧᓇᐅᔨᑦ ᓂᕿᑭᖅᑎᓇᓂᑐᖅ ᐊᔨᓇᓇᖅᑐᖅ.

ᐅᓪᓗᓐᓄ: ᐊᐃᓴᓚᓕ 2–ᒥᑦ, 2019

ᑭᖑᓂᖅᑐᐃᓕᐅᔨᑦ:
ᐊᖏᓐᓇᓕᕆᔨᖅᑕᐃᓐᓇᖅᓂᖅ (DOE): ᑎᐅᓚ ᒥᐅᑦᓂ, ᓀᓴ–ᒪᓴ ᓕᔨᑉᓴᒃ ᐊᒻᒪᓴ ᑲᐅᑦ ᒪᑎᐅᓂ, ᓴᓯ ᐊᖏᑭᐊᖅᓂᖅ,
ᐋᓕᓂ ᓂᕐᕆᓇᕎᑎᐊᖅᓂᖅ

ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅᓗᑦ ᑲᑐᔪᖅᑎᑭᖏᓐᓇᐅᑦ (HTO): ᕿᐱ ᐊᓇᐃᓴᓂᖅ, ᑲᑦᓂ ᑭᓪᓕᑭᖅᓂᖅ, ᓴᐃᔨ ᕿᓴᔪᓇᖅ, ᒪᐃᔭᓂ ᐊᐅᓪᓚᑎᓂᖅ, ᓴᐃᔨ ᕿᓴᓐᓴᑦᖅ, ᓴᐃᔨ ᕿᓴᓐᓴᑦᖅ (ᐃᖅᓯᐊᐅᑕᖅ).

ᑲᔪᓂᒥ ᖃᐅᔨᓴᒪᓂᖅ ᑲᓇᑕᒥ: ᓕᓐᑕ ᐅᐊᖏᑐᓐ

ᒪᕕᑦᑕᐅᖅᑎᓄᑦ (MLA) ᐃᓕᑎᕆᔨᕐᓂᖅ: ᒥᓪ ᐊᖏᑭᐊᖅᓂ

ᐅᖃᐅᔨᔨᐅᔨᑦ ᓇᓗᓇᐃᔪᖅᓯᒪᓂᕐᓂᑦ:
ᑕᑯᓴᐅᑎᑎᑎᑦᓗᓂᑦ, ᐊᖏᓐᓇᓕᕆᔨᑦ ᒥᑭᐊᖅᓯᐊᖅᑎᑦᑎᐸᖅᑕᐃᓐᓇᖅᓂᖅᓗᑦ (HTO)–ᖁᑯᓂ ᐃᓕᑎᕆᔨᔨᑦ ᐅᖃᖅᑕᓕᕆᔨᖅᑕᐅᑦ ᐊᒻᒪᓴ ᐊᐱᖅᑯᑎᓗᒃ ᐊᐱᖅᑯᑎᑎᓂᖅ ᖃᐅᔨᓴᖅᐅᓕᕆᔨᖅᑕᐅᑦ ᐃᓕᑎᕆᔨᑦ ᐊᐅᓪᓚᑎᓂᖅᓗᑦ ᐊᐱᖅᑯᑎᑎᓂᖅ ᐊᐱᖅᑯᑎᑎᓂᖅᓗᑦ ᐊᐅᓪᓚᑎᓂᖅᑐᓴᓂᖅ ᐃᓕᑎᕆᔨᑦ ᐊᐅᓪᓚᑎᓂᖅᑐᓴᓂᖅ (ᑐᑭᐊᖅᓯᐊᐅᑕᖅ) ᐃᓕᑎᕆᔨᑦ ᐃᓕᑎᕆᔨᑦ ᐊᐅᓪᓚᑎᓂᖅᑐᓴᓂᖅ

ᐅᒥᖏᓚᐃᑦ ᐅᓄᕐᓂᖏᑎ ᓯᕐᔪᓯᖅᓭᒪᖏᓐᓂᖏᓐᓂᒧᑦ. ᐃᓕᖏᑎᑦ ᐅᒥᖏᓚᐃᑦ ᓄᑦᑐᐴᕆᐅᔪᕐᐃᓄᒃ ᓇᓵᖅᑐᓕᐅᑦ, ᓄᓐᓂᖏᓪᓚ ᐃᓱᒪᑎᐅᑦᓗᓂᓂ ᐃᓂᐅᔪᖅ ᐳᓴᓂᖏᓐᓂᒧᑦ. ᖃᐅᔨᑎᐊᕐᖅ ᓴᓗᓇᐃᕐ ᓇᓵᖅᑐᖃᖅᑐᑦ ᐃᑭᓐᓂᖃᕆᐅᕐᓂᓪᓚᓂ, ᐅᒥᖏᓚᐃᑦ ᓇᒧᓐᓂᖅᑐᐴᕆᐅᔪᕐᐃᓄᒃ ᐳᔪᓂᒋᐃᓕᓂᓐᓂᒧᑦ ᐃᓂᐅᔪᓂᑦ ᐃᑭᓱᓕᕐᓂᓪᓂ ᐃᓂᐅᔪᓂᒃ ᐅᓪᓗᓕᕐᓂᖅ.

ᐅᒥᖏᓚᐃᑦ ᐃᖏᓕᕆᓂᖏᓐᓂᑦ ᑐᑭᓯᐅᓚᕐᑐᐊᖏᓐᓂᑐᑦ. ᐅᓚᔪᓕᓗᓐᑦ ᐅᒥᖏᓚᐃᑦ ᓯᐅᑎᖏᓐᓂᒧᑦ ᓇᑐᓇᐃᒃᑯᑎᖅᓯᓚᐅᖅᓯᓚᔪᓕᑦ ᓇᐸᑕᓕᒃ ᑕᓯᐊᓂ 1990 ᐱᑎᐊᓕᓴᖅᑎᓕᑎᑦ, ᐊᓖᓚᓐ ᐅᒥᖏᓚᒃ ᓇᑐᓇᐃᒃᑯᑎᖅᓱᖅᑐᖅ ᑕᑯᐅᐅᒃᓄᓪᓕᖅᑐᓂ 2017–ᒥᑦ ᑲᓐᑐᔭᑐ ᑕᓯᐊᑕ ᖃᐅᓂᕐᖏᓚᓂ. ᐅᒥᖏᓚᐃᑦ ᑲᐸᐸᑕ ᐸᖏᓚᔭᓐᓇᖅᑐᑦ ᑎᕋᒪᓂᒃ ᑕᓕᒪᓂᑦ ᒪᐃᔪᓂᑦ ᐅᖏᓪᓯᑎᑎᒧᒧᑦ. ᐅᒥᖏᓚᑦᑲᖅᑐᖅ ᓇᓵᖅᑐᑕᐅᐸᑦ ᐳᓴᓂᐊᓂ ᐊᐱᒪᓐ ᐱᐅ ᑕᓯᐊᑕ ᖃᐅᓂᕐᖏᓚᓂ. ᐊᐳᓇᓇᐅᐸᑦ ᒥᐅᓴᓂ ᐃᖏᓕᕆᓂᖏᓐᓂᑦ ᓄᓇᕐᔪᐊ ᓯᖏᓚᐅᓂᖏᓐᓂᒧᑦ (GNWT).

ᓄᓇᕐᔪᐊᒥ ᐅᒥᖏᓚᑕᕐᓂᖏᓪᓚ ᐱᔨᔪᑎᑕᓚᑦᓯᐅᑦ, ᐊᐱᕆᑯᑎᑕᖃᓚᐅᕆᖅᑐᓗ ᑲᑎᕐᕕᓂᑦ ᐅᒥᖏᓚᐃᑦ ᓄᓇᕐᔪᐊᒥ ᐃᓚᒥᐅᒥᑦ ᐅᒥᖏᓚᐃᑦ MX–09 ᐃᓄᒋᐊᖏᓂᖏᓐᓂᒧᑦ. ᑕᐃᒪᐃᓚᒪᑦ, ᑲᑎᒪᖃᑎᑎᖃᒪᔪᑦ ᐊᓚᓗᓐ ᓴᓇᖃᑎᑎᕐᕕᓂᑐᐃᓐᑲ ᓄᓇᕐᔪᐊ ᓯᖏᓚᐅᓂᖏᓐᓂᒧᑦ (GNWT).

ᓂᕆᐸᓴᓂᑲᖅᑐᑦ ᓯᐸᑲᓂᖏᓪᓚ ᐃᓱᒪᑎᓕᓗᓐᑦ, ᑯᑕᐃᑦ ᐊᓪᓚᕐᓇᖅᑐᖅ ᑕᒪᐸᕆᖅᑕᐅᖏᓐᓂᑕᑦ. ᐊᖏᕐᕋᒧᑎᓴᑎᓂᒧᑦ, ᑯᑕᑕᒪᑦ ᑐᓂᔨᓱᒧᑦ, ᐱᔨᑕᒥ 7–ᖃᐅᔪᑦ ᐅᓂᐸᐅᔨᓱᐅᔪᑦ ᓂᕆᐸᓴᓂᐸᑕᐅᑦᑕᓗᑎᐸ 20–ᓂᓐᓄᖃᕐᑐᓂᓂ. ᓯᓕᐸᐅᓂᖅ, ᐊᖏᕐᕋᒧᓂᐸ ᖃᑯᓂᓂᐸ ᖃᐄᖓᓴᖅᑐᓂᐸ ᑯᑕᐃᑦ ᓄᖐᒋᐅᔨᓴᖅᑎᓴᖏᓐᓂᑐᑦ, 15–ᖃᐅᔪᑦ ᐊᑐᖅᑕᐅᔨᓯᒪᔪᑦ 20–ᐅᑦ ᖃᑯᓯᔨᓂᖅᓵᖅᓄᖅ, ᐊᖏᓚᓐ ᐅᓂᐸᐅᔨᓯᑕᖃᓄᖏᓐᓂᖓᓪᓗᓂ ᓂᕆᐸᓴᓂᐸᑎᐅᔪᓂᐸ ᐃᓱᐊᓗᑎᐸᓴᐃᑦ ᒥᐸᓚᓄᑦ. ᐊᖏᕐᕋᒧᓂᐸ ᒥᑭᑎᐊᖏᐊᖅᑲᓐᓗ ᑲᑐᔭᓴᐸᑎᐅᔪᓂᑦ (HTO)–ᐸᑯᑦ ᐃᓄᖏᓴᐸᑦ ᐊᑐᖅᑕᐅᓂᖅᓴᓂ ᑕᐃᒪᐃᖏᓐᓄᓇᖃᓕᔪᑦ MX–09–ᒧᑦ 20–ᓂᐸ ᐱᔨᔪᑦᓇᖅᑎᑕᐅᔪᑦ ᐊᖏᕐᕋᒧᓂᐸᐅᔪᑦᓇᖅᑎᑕᐅᓗᑎᐸ (TAH)–ᓂᐸ. ᐳᓴᐊᓂ, ᐊᖏᕐᕋᒧᓂᐸ ᒥᑭᑎᐊᖏᐊᖅᑲᓐᓗ ᑲᑐᔭᓴᐸᑎᐅᔪᓂᑦ (HTO)–ᐸᑯᑦ ᐃᓱᒪᓴᖅᓯᓴᑎᑕᐅᔪᑦ ᒪᓕᑐᖏᐊᖏᓐᓂᒧᑦ ᐊᑐᔨᓇᓕᓗᑎᐸ ᐱᖏᔭᓴᑎᓕᑎᓇᓱᓗᓂ ᓄᓇᓕᖅᓂᒧᑦ–ᐱᓕᕆᐊᖏᔪᓂᐸ ᐊᐅᓪᓕᖅᑐᓴᒪᓂᑦ ᐱᓕᕆᑎᐊᔨᑎᓂᐸ, ᒥᑭᑎᑦ ᐊᔪᓴᓴᑎᓂᑦ 1–ᒧᑦ ᐅᖏᓪᓗᓂᓂ 2–ᓄᑦ ᐅᒥᖏᓚᐃᑦ ᐅᓄᕐᓂᖏᓐᓂᑦ ᐃᑲᔪᓴᓱᓓᓕᑎᑦ, ᓄᓇᕐᔪᐊᖏᓇᓱᓓᓕᑎᑦ, ᓱᓴᓱᖏᓐᓄᓇᕐᓂᖏᓐᓂᑦ, ᐱᔨᖅᑎᑦᓕᐊᖏᓐᓂᖓᓪᓗᓂᐸ, ᐊᑯᓂᐊᓕᐸ.

<u>ᐱᓕᕆᓴᐅᔨᔪᓂᑦ ᓄᓇᕗᑦ ᑲᖏᒪᑯᑦ (GN)–ᐸᑯᓄᑦ:</u>

ᓴᓇᓗᖅᑐᖅ ᐊᖏᕐᕋᒧᓂᐸ ᒥᑭᑎᐊᖏᐊᖅᑲᓐᓗ ᑲᑐᔭᓴᐸᑎᐅᔪᓂᑦ (HTO)–ᐸᑯᖏᓐᓄ ᐱᓕᕆᓴᐅᔨᔪᑦ *ᐊᑐᖅᑕᐅᓂᖏᓪᓚᓂ* ᑕᐃᒪᐃᖏᓐᓄᓇᓴᐅᔪᑦ MX–09–ᒧᑦ 20–ᓂᐸ ᐱᔨᔪᑦᓇᖅᑎᑕᐅᔪᑦ ᐊᖏᕐᕋᒧᓂᐸᐅᔪᑦᓇᖅᑎᑕᐅᓗᑎᐸ (TAH)–ᓂᐸ. ᓂᕆᐸᓴᓂᐸᑎᐅᔪᑦ ᐊᓚᓗᓐ ᐊᑐᖅᑲᓐᓂᑦ ᐅᓂᒃᑳᑕᐅᖅᑕᑕᒃ ᐊᔪᓴᓴᑎᓂᐸᐅᔪᑦ ᑕᓯᐸᒥᓂᐸ ᐃᓄᒋᐊᖏᓂᐸᐅᔪᑦ ᓴᓇᓗᓂᖏᓐᓂᒧᑦ ᐅᖏᓪᓗᓂᓂ 2–ᓄᑦ, ᐃᓄᒋᐊᖏᓂᓴᒧᑦ ᓴᐸᐅᖅᓯᓂᖏᓐᓂᒧᑦ, ᐃᓄᒋᐊᖏᓂᓴᒧᑦ 2023 ᑐᖏᓚᓂ.

4.0 ᑭᖓᓯᓪᓗ ᐅᖃᐅᓯᕐᓂᖅ – ᐱᖃᑎᒋ ᐱᓕᕆᔨᖓᕆᓯᓪᓗ

ᐊᕐᑎᑎᓕᕆᓴᒃᑯᑦ (DOE)–ᒃᑯᑦ ᐁᖃᑭᒃᓯᕗᑎᐊᖅᓂᐊᖅᑐᑦ ᐅᒥᖓᒪᐃᑦ ᐅᓂᒃᑲᖏᑐᓂ ᑲᑎᒪᑎᓯᓗᓄᑦ ᐅᖃᐅᓯᓯᐅᔪᓂᒃ ᐃᓕᖅᖢᓗᑎᒃ ᐊᒻᒪᓗ ᑐᓂᐅᖅᖢᖅᑕᐅᓗᑎᒃ ᐊᐅᓕᕆᑦᐅᖅᑲᑎᑎᒋᓐᓂᑦ ᐃᓕᑕᑯᐅᔪᓂᑦ. 2019–ᒥᑦ ᐅᒥᖓᒪᓂᑦ ᖃᐅᐅᔪᖅᓂᖅᖢᖅ ᐸᖅᓇᖏᐅᑎᓯᓗᒍ ᐃᓗᐃᓯᓕᖅᖢᒥ, ᐊᕐᑎᑎᓕᕆᓴᒃᑯᑦ (DOE)–ᒃᑯᑦ ᐃᓱᒪᑕᑯᖅᖢᖅᑎᐊᖅᑐᑦ ᐅᖃᐅᓯᓯᐅᔪᓂᒃ ᐊᒻᒪᓗ ᑎᓕᐅᕐᔪᐊᑎᓂᒃ ᐊᖖᓇᓴᐅᓕᖅᑐᑦ ᐊᒻᒪᓗ ᑎᑭᑕᐊᖅᖢᖅᑎᐊᖅᑐᑦ ᐅᖃᐅᓯᓯᐅᔪᓂᒃ (HTO)–ᒃᑯᑦ ᑐᓴᐅᖅᑎᑎᒋᔭᐅᓂᒃᑯᑦᑕᐃᓐᓇᖅᓂᖅᑎᓯᓗᓂ, ᖃᓄᐃᓕᐅᕐᓂᖅᑎᑦᓗᓂ ᐃᓕᓴᕐᓂᐊᕐᓗᒍ.

ᑕᐃᒪᓐᓇᓂ 2017–ᒥ, ᐅᓇᖅᓂᕐᓂᓂᒃ ᖃᐅᔨᓴᖅᓂᖅᖢᑦ ᖃᐅᔨᓴᐅᓕᕆᓪᓚᑦ ᐅᖃᐅᓯᐅᓕᕆᓪᓚᑦ ᕿᑎᖅᑕᐅᑦ ᐊᐃᔨᖅᑐᑦ ᐅᑭᐅᖅᑕᕐᒪᑦ ᐱᒪᓪᓚᑦᑎᓪᓗᒍ ᑲᓇᑕᒥ ᐃᓚᖓᑕᐅᓪᓚᑦᑎᓪᓗᒍ ᑲᑎᒪᓂᕐᓂᖅ 2018–ᒥ ᐊᒻᒪᓗ ᐊᖓᔪᖅᑲᐃᑦ ᒥᑭᐅᐊᓂᖅᑲᑎᑎᒋᓐᓂᒃ ᑲᑎᒪᔨᖃᑎᒌᖓᓂᑦ (HTO)–ᒃᑯᓂᒃ ᐊᒻᒪᓗ ᐊᓯᖏᓐᓂᒃ ᐊᐅᓚᑦᑎᐊᕐᑲᖃᑎᒌᖓᓂᒃ ᐃᓕᑕᑯᐅᔪᓂᑦ. ᑐᓴᐅᒪᔪᑎᑦᑎᓂᓕᕆᓂ ᓴᖅᑯᓴᖅᓯᒪᔪᓂᓕᕆᓂᒃ ᐅᒥᖓᒪᐃᑦ ᐅᓇᖅᓂᕐᓂᓂᒃ ᐅᒥᖓᒪᐃᑦ ᐊᐅᓚᑦᑎᓂᓕᒃ ᐃᓗᓕᕆᓂᖅ MX–09–ᐅᐸ ᒥᖁᓇᓂᑦ ᐅᖃᐅᓯᐅᓯᒪᔪᓂᑦ. ᓴᓚᖁᓂᓕᒃ, ᐱᓕᕆᖅᑯᓂᖅᑎᐅᓯᒪᔪᓂᓪᓚᑦ ᓄᓇᕗᑦᒥ ᐅᒪᓐᓇᕐᓂᖅᓚᑦ (NWMB)–ᒃᑯᓂᓄᑦ ᐱᓇᔨᖃᖅᓂᐊᕐᑐᖅᑐᒃ ᐱᓴᐅᔨᓇᖅᑎᑕᐅᔨᓇᖅᑎᑕᐅᓗᑎᒃ (TAH) ᖃᓄᐃᓕᐅᕐᔪᓂᒃᑯᑦ ᑕᐃᒪᐃᓕᖅᓇᓴᓗᑎᑦ, ᑭᓇᓇ ᑕᐃᒪᐃᓂᐅᔪᓂᖅᖢᖅ ᒪᓕᒃᔪᐃᒍ, ᖃᓄᐃᓗᖅᑐᑦ ᐊᖓᔪᖅᑲᐃᑦ ᒥᑭᐅᐊᓂᖅᑲᑎᑎᒋᓐᓂᒃ ᑲᑎᒪᔨᖃᑎᒌᖓᓂᑦ (HTO)–ᒃᑯᕐᓂᓂᒃ ᐃᑲᔨᖅᓯᖅᑕᐅᓕᓐᓂᓗᒃ. ᑐᓴᐅᒪᑦᑎᓂᖅᑕᐃᓇᓂᐅᕐᓕ ᐱᓕᕆᔨᖓᓪᓗ ᐊᐅᓚᑦᑎᐊᖅᑐᑦ ᑭᓇᓯᒥᖓᓪᓗ ᐅᓂᒃᑲᒥᑦ.

ᐅΔᒍᖅ I

ᖃᕆᓴᐅᔭᒃᑯᑦ ᑕᑯᓴᐅᑎᕐᑎᓴᓪᓚᐊᑎ 2019

ᐅᒥᓪᒪᐃᑦ ᓇᒋᔭᓂᖏᑦ ᐊᖏᒪᓗ ᐅᓄᕐᓂᖏᑦ ᐅᒥᓪᒪᐃᑦ ᐊᐅᑕᕐᑎᓂᓐᓄᖅ ᐃᓗᓕᖏᑦ, MX09–Γ, ᖃᐅᓪᓗᖅᑐᑦ ᑯᓴᒪᑦ ᐅᐊᓕᓂᐊᓂ

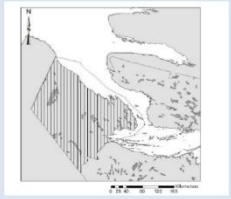
ᖃᐅᓪᓗᖅᑐᖅ ᐊᖏᒃᓇᓕᕆᑦ ᒥᑭᑎᐊᓂᐊᕆᑦᑐᓗ ᑲᑎᕐᓴᖃᑎᒌᒃᓴᓕᔪᓂᑦ (HTO)–ᑲᑯᑦ

ᓵᓴ–ᒪᓴ ᑕᑭᓗᑲᒃ
ᓄᓇᕗᑦ ᒐᕐᒪᖃᑯᑦ
ᐊᐅᑎᑕᓴᕐᑯᑦ

ᔪᓂ, 2019

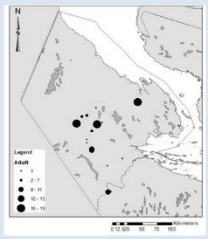
ᖃᐅᔨᓴᖅᑕᓱᑦ ᐃᓂᖏᒐ

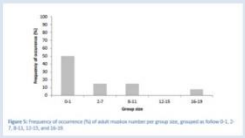
- MX–09, ᑭᐃᓕᓐᖅᐸᒋᑦ ᓇᑕᐅᑦᑕᖅᑕᐅᓕᑦ 2007–Γ 598 ᐅᒥᓪᒪᐃᑦ, 159 Δᵊᓇᐅᔨᓗᑎᒃ ᐊᑲᓴᓯᒪᔪᒋᑦ.
- Δᓯᒪᓗᑎᑦ, ᐊᖏᓇᓯᑎᑦ ᓂᕆᐅᔭᓂᖅᑕᖅᑯᑦᓴᑕᒪᓴᖅᑐᒃ ᑐᓴᑐᓂᓐ ᐃᓇᓐᑎᑕᐅᔨᓪᓗᓂ ᑕᵊᕐᓂ ᐃᓂᐅᔪᒋᑦ.



ᖃᓄᐅᑦᑐᓂᖅ

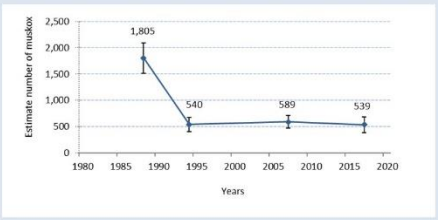
- ᖃᐅᔨᔭᕆᓂᖅ ᐱᔭᕇᖅᑕᐅᓕᐅᖅᑐᖅ ᐊᑎᓕ 25–Γ ᓯᑎᐱᓗ 02–ᒧᑦ, 2017 (16% Δᓂᐅᔪᒋᑦ)
- 103 Δᵊᓇᐃᑦ ᐃᓕᖅᑕᐅᓕᐅᖅᑐᑦ ᐊᑎᓕᔭ 87 ᐊᑲᓴᓯᒪᔪᓕᖅᔭᓄᑦ
- 58% ᑲᑎᕐᓗᓂᑦ ᐅᓄᕐᓂᑲᓴᓕᔭᓄᑦ ᑲᑎᕐᓗᓕᐅᖅᑐᑦ, 2–Γᑦ 11–ᒧᑦ Δᵊᓇᐅᔨᓴᓕᔭᓄᑦ
- ᓄᕐᕆᐃᑦᐃᵊᓇᐃᑦ ᐊᑲᓴᖅᓯᒪᓂᕐᓂᑦ ᐃᓚᐃᓕᓐᓗᓕᐅᖅᑐᖅ 38 ᓄᕐᕆᐃᑦ/100 Δᵊᓇᐃ





MX09, – 2017 Δᓄᑎᐊᖅᓂᕐᑎᑦ ᓇᑕᐅᑦᑕᖅᑕᐅᔨᑦ

ᐅᒥᓪᒪᐃᑦ ᐅᓄᓂᖅᑎᑦ ᓯᕆᔨᖅᑯᑲᒪᔨᕐᓂᕐᑎᑦ, ᑕᐃᒪᓪᒪᓂᓐ 1995–Γ, ᓇᑕᐅᑦᑕᖅᑕᐅᓴᓪᓚᐊᑎᒃ 539 ± 150 ᑕᓪᒪᖅᓂᖅ SE



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NWMB Regular Meeting 002-2019

Please accept this document, Kugluktuk Angoniatit Association Bluenose East Community Caribou Management Plan for decision/approval.

Kugluktuk Angoniatit Association representatives will be present at the Regular Meeting in Ottawa on June 19, 2019.

Thank you,

Amanda Dumond

Manager

KUGLUKTUK ANGONIATIT KATIMAYIT

AHIAKMIUTANIK KIVATANI

NUNALINGNILU

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MUNAKGIYAUYUKHANIK
PANGNATTAUYUKHANIK

TITIKGAKHIMAYUNIK HAVAKHAUTIT TALVANI ILIDJUHIKGIYAINIK HAVAKHAUTIKHANIK

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Tahapkununa havakhautit munakgiyauyukhanik pangnattauyukhanik havakhikhimayunik tahapkunanga Kugluktuk Angoniatit Katimayit, Kavamat talvani Nunavut uvanilu Nunavut Tunngavik Katimayit.

KAUHIMAYAUVAKTUNIK HULI

Tahapkuat Kugluktuk Angoniatit Katimayit pidjutiplumaplugu kuyakgiyumaplugu tamna Cheryl Wray, Nunavut Tunngavik Katimayit tamnalu Lisa-Marie Leclerc, Kavamat talvani Nunavut, pidjutivakhimayuk ikayukhutik mikharut aullaktikgutikhanik tahapkununa pangnattauyukhanik.

Kitikmeot Inuit Katimayit pivakhimayut ilauvakhimayut aiyuukiktuihimavaktunik pilihalliktilugit havagiyaulikgungullikgamik aullaktikhimayunik mikharut pangnattauyukhanik.

TITIKGAKHUGU NAUNAITKUTARGIPLUGU ILITUKGIDJUUTAUTIT

ACCWM	Tutkikhaiyit Katimayit mikharut Havakatigiknikmun Anguutinut Havakhautikhanik Munakgiyauyukhaniklu
BNECH	Ahiakmiutanik Kivatani Ammihuakyunik
COSEWIC	Katimayit mikharut Naunaitkutauvaktunik Nallautakguttauvaktunik tahapkuninga Nirgitaikpalliayunik Pikpaliayunik Anguutinik talvani Kanatami
DOE	Havakvit talvani Nunalikiyit, Kavamat talvani Nunavut
ENR	Nunalikiyit talvanilu Anguhikiyit, Kavamat talvani Tunungani Nunatsiami
GN	Kavamat talvani Nunavut
GNWT	Kavamat talvani Tunungani Nunatsiami
HTO	Anguniaktit imalu Nanirgiaktukpaktunik Katimayit
KIA	Kitikmeot Inuit Katimayit
KRWB	Kitikmeot Nunatuttukanik Anguhikiyit Katimayit
NA	Nunavut Angiutauvaktunik
NTI	Nunavut Tunngavik Katimayit.
NQL	PIKANGITUNIK-Kufiutilanginik Naunaitkutak Nungumanginik Kangilimaitunik
RWO	Nunatuttukanik Anguhikiyit Katimayit
TAH	Tamakpianganut Pidjutautaaktunik Angunnahuaknikmun
TK	Ingilgangnitanik Pitkuhiktukpaktunik Kauhimayuttukanik

TUKLIA-HAVAKHAUTIT MUNAKGIYAUYUKHANIK PANAARIT IMALU HAVAGIYAUYUKHANIKLU

KUGLUKTUK ANGONIATIT KATIMAYIT tahapkuat Kugluktuk HTO piyukhauyunik pilutik havagiyauyukhanik tutkikhainikmun mikharut TAH; katitikgiyukhanik mikharut TK; havakhautikhanik munakgiyauyukhanik; malikgakhaniklu tahapkunani katimakatigiktukhanik uvunalu havagiyauvaktukhaniklu ungniguutinakpaktukhanik havakhautikhanik mikharutlu malikgakhanik talvaniktunik Nunavutmi Angiutauvaktunik; utiutifaakpaktunik uvunalu nuttanguktiklugit kakungukgaikpat pangnattauyukhanik pidjutikariakaktunik havagiyauhainaktukhanik uutakiukgutinaitumiklu.

NUNAVUT TUNNGAVIK KATIMAYIT piyukhauyunik pilutik havagiyauyukhanik tutkikhainikmun mikharut havakhikariakaktunik tamakpianginik havagiyauvaktunik itkumannahuaklutiklu havatiaklutik talvani Nunavutmi Angiutauvaktunik. NTI pilutik havagilugit ikayuktauyukhanik ihaagiahutigiyauyuniklu.

KITIKMEOT NUNALINGNILU ANGUHIKIYIT KATIMAYIT pilutik ikayutauyukhanik tahapkununga HTO.

KAVAMAT TALVANI NUNAVUT, HAVAKVIT TALVANI NUNALIKIYIT pilutik havagiyauyukhanik mikharut nalukhanginamik havaktit naunaiyainikmun hungnautilanginik naunaitkutakhanik havakhautikhanik munakgiyauyukhanik; ataniktutaulutik ihiviukhinikmun naunaiyainikmun hungnautilanginik kinikhianikmun; ikayutauyukhaniklui uniklutikhanik havakhikariakaktunik; ikayutauyukhanik tukihigiarutinik uvunalu ikayutauyukhanik havakhikariakaktunik.

1. PANGNATTAUYUKHANIK AULLAKTIKGUTIKHANIK

Tahapkununa kauhimadjutiplugit mikharut pangnattauyukhanik piyukhanik tahapkunani Kugluktuk Angoniatit Katimayit piyukhanik havagiyauyukhanik angunnahuaktauvaktunik, havagilugit nunalingni nallautakgutauvaktunik, pihimayunik havakhikhimayunik talvanitunik Nunavutmi Angiutauvaktunik talvunakukhimayunik Uiwuak 5.7.3 "Tahapkunanilu Alungukuttakpaktuniklu talvanitunik Hakugikhukhaaknikmun Naapkiyaulimaitutlu HTOs pilutik ilaulugit mikharut malikgakhanik: (aah) havagilugit malikgakhanik angunnahuaknikmun havagiyauvaktunik pitkuhiktukpaktunik imalu nalukhanginamik havagivaktamingnik tahapkunani inukatigikpaktunik, ilauyunik atuktauvaktunik tahapkunani pikangitunik angunnahuaknikmun nirgitinut nungumanginik kangilimaituniklu; (b) tahapkununalu tutkikhailugit pilugit havakhiktaulutik nunalingni ihaagiahutigiyauvaktunik nallautakgutauvaktunitlu imalu tutkikhakhikhimayunik ihaagiahutigiyauvaktunik tahapkunani inukattigikpaktunik; (c) tahapkunani malikgakhainik pikangitunik naunaitkutaitunik nunataaknikmun aviktuinikmun inukatigiknikmunlu, pikaktuniklunin naliak ipiknakpiaktunik ihumadjutigilugit imalu ihumadjutigilugitlu, nalikmik avaakukhimayunik nunalingni kauhimayauvaktunik ihaagiahutigiyauvaktunik nallautakgutauvaktunik imalu tutkikhakhimayunik kauhimayauvaktunik ihaagiahutigiyauvaktunik nallautakgutauvaktunik; uvunalu (d) tamatkiumannahuakhugit, mikharut havakhautikhanik munakgiyauyukhanik angunnahuaktauvaktunik ingmingnut inukatigikpaktuniklu."

2. NIRGIITIT HURAAT TUKIHIGIARUTINIK

Nunamiutanik tuktunik (Taryukmiutanik Tuktuktutainik Tunungani *Rangifer tarandus groenlandicus) inngniktukpargainik talvanitunik Nunatukanginik Nunavutmi talvanilu Tununganl Nunatsiami. Tahapkuninga tuktutainik ilihimayauvaktunik taakuyauvaktunik aah avaatingnut nungumanginik ungahiktummut pidjutivaktunik ammihuakyuvaktunik.* Pilakivaktunik hila malikhugu ingilgayanginik ivaavit atukpagainik imalu ilidjuhikgivakgainik alangukpaliangmut hila nirgititlu tahapkunani ilikuaktunik ammihuakyuinik, alangayuniklu ammihuakyunik ilitukgiyauvaktunik havagiyauvaktunik munakgiyauvaktuniklu. Una Thomas uvani (1969) mi hivuliuvakhimayuk hivulikpakpianguyunik atikhautikhanik tamna atia " Ahiamiutanik ammihuakyunik Ahiakmiutanik ammihuakyunik" tahapkuninga tuktutainik uukivaktunik tununganl waatanivyak tahapkunani tuktuinik uukivaktunik tununganivyak wuanivyaklu talvani Great Bear Tahik ingayangit ivaavit talvani nunani imalu ingniukvit talvani Bluenose Tahik. Uvani kitikuktunik ukiunginik-1990s, havakhikhimayunik tuniyauvakhimayunik naunaitkutanik ihiviuktauvaktunik nalvakhiuktauvaktunik ihiviugakhanik havagiyauvaktunik havakhikhimavaktunik akungainik uvani havakhikhimayunik tuniyauvaktunik naunaikutakhanik pihimayunik tingmitukhutik naunaiyainikmun naunaitkutakhanik havakhikhimayunik akungainik uvani 1966 talvanilu 1993, pivakhimayunik havakhikhimayunik ilangunik itkumayunik havagiyauvaktunik pingahunik ammihuakyunik tuktunik pilutiklunin paaikpalialutik talvani avaatingnut nungumanginiklu tahapkuninga tuktunik tuktuikpalialuni kakugu talvani nunani avaatingnut nungumanginiklu "Ahiakmiutanik ammihuakyunik" (Nagy *et al.*, 2005). Tamna atigiyanik Ahiakmiutanik Kivatani ammihuakyuinik pihimayunik havakhikhimayunik mikharut tuktutainik nunamiutanik kufiutilanginik kanuklu gulluvait ilitukgiyauvaktunik hivituyumik hakugiktuklu atuktauvaktuk huli aipaagungukgangut nunatuttukanginik uutifaakpagait ivaavit ingniukvit kivatani talvani Bluenose Tahik (ACCWM, 2014; Nagy *et al.*, 2011).

Tahapkununa kufiutilanginik nunamiutanik tuktutainik kufiutilangit ilihimayauvaktunik malikhugu aah ammigaitunik ukiunik 60 nik atuktauvaktunik pitkuhiktukpaktuk nallautakgutauvaktunik cycle. Pihimangmataah hivitunianik mikharut atuktauvaktunik pitkuhiktukpaktuk nallautakgutauvaktunik, ingilgangnitanik pitkuhiktukpaktunik kauhimayauyukhauyut huli ipiknakpiakmataalu ikayutauyukhanik tukihigiarutinik mikharut kaakugu atuktauyukhanik naunaitkutakhanik kufiutilangnik nallautakgutauvaktunik pihimayunik pinnalimaitunik havakhikhimayunit tahapkunanga upalungaikhimayunik pilgumikhimayunik tingmitikut havakhikhimayunik ihiviukhinikmun naunaitkutakhanik. Mikharut ilidjuhikgiyanik, tamna havaktit ihiviukhinikmun hunauyakhainik avaatainik naunaitkutakhanik havalikhimayuk munakgilikhugit tahapkuninga Ahiakmiutanik Kivatani tuktutainik uvani havaktit umayulikiyit nalukhanginamik avakhautikhanik havalikhimayunik munakgilikhugit Ahiakmiutanik Kivatani tuktunik uvani 2000 pivakhimayunik ammihuakyunik nallautakgutauvakhimayunik tamakpianginik uvuna 120,000 tuktunik. Tahapkuat nungumaagivakganik mikharut kufiutilanginik havakhikhimavaktunik naunaitkutanitunik adjiliukhimayunik 122,697 nirgitinik 2010 hivuagut aah ikiikliyunik ilitukgiyauhimayut uvani 2013, kanukgitilugit talvuna ammihuaktyuit nallautakgutauvaktunik kufiutilanginik 68,295 tuktunik. Uvani 2015, tahapkuat kufiutilanginik pihimayunik nallautakgutauvaktunik kufiutilanginik 38,592 niritiniklu (Boulanger *et al.*, 2016).

Tahapkununa kilaminuak ikiiklihimayunik Ahiakmiutanik-Kivatani ammihuakyuinik aktikuttilanginik kuugluktitaaihimayut aah ilidjuhikginialugit ihumadjutigiyauyukhanik, imalu kilamiukgutaulutik aullaktikgutikhanik havakhautikhanik munakgiyauyukhanik upidjutaulutik tahapkunani atuktauganginakpaktunik kingnauyunitlu. Ammigaitunik nunalingni: Wrigley, Norman Wells, Tulit'a, Deline, Whati, Gameti, Bechoko, Paulatuk talvanitunik Tunungani Nunatsiami talvanilu Kugluktuk uvanituk Nunavutmi pidjutiyunik nahugidjutivakgaait nikikhut ammihuakyuit tahapkuninga nikikhakgiplugitnaahugidjutivakgaait (ACCWM, 2014). Kihimi, tuktuit inngniktuivaktut inuttiaknaktut ammigaitunik ihaagianakninik paaigutinaitumik nikikhaigutinaitumiklu. Inuitlu talvanikmiutanik nunatuttukanginiklu imalu tuktuit atadjutikgiktut pihimayut inngniktuivaktut inuttiaknaktut pitkuhiktukpaktuk atuktauvaktuk huli tadja, humikvikhaituk tuktunik pihimayut iluumi atadjutauyunik Inungni ingilgangnitanik, pitkuhiktukpaktunik puigugakhaungitunik ipiknakpiaktunik, imalu atadjutigiknikmun inukatigiknikmun pitiakatigiknikmun talvanitunik nunalingnu. Tuktu nikigiplugu ilihimayauvaktuk nakukguttauyuk ikayutaugami ilingaingnaktuk Inungni pitkuhik atukhugu.

2.1. Tuktunik Ilingaitakhainik

Aniakgutilingniklu, nauvaktunik kaakuvikgangukpaktunik imagidjakhutik alauyuniklu kupilgungukpaktunik, ilihimayauvaktutlu pidjutivaktunik ilingakpaktunik angniakgungnikpaktunik, inuvaktuniklu huli, atukluaktauvaktuk huli inungni inuumakgingnaktuk pikuhiktukpaktunik, kanukvikhaituklu mikharut hivunikgiliktainik tadja aiyuuknakhivaliayuk tuuktukainalimainakpaliagamilu ikiiklivaliaplutik taakungnakluakgungnaikhutik nallautakgutauvaktunik. Kanuklu hivuniktuiliklutik ihuigutivalialutik tuktuikpaliagumi nallautakgutauvaktunik. Inuitlu ihaagiahutigiyauvaktunik pinaktakhamingnik ungunnahuakginakgialingnik ilingaitunik nirgitinik pidjutiyanginik havagilugit havagivaktamingnik ungniguutinakpaktukhanik pikuhiktukpaktunik inukatiggiknikmun-pittiakutigiknikmunlu imalu nikigilluakpaktamingnik nikitianik ilingaitunik (Carlsson *et al.*, 2016; Jenkins *et al.*, 2015). Taimaitumik, aullaktikgutikhanik aah ilingaitukhanik havakhikhimayunik pilutik havagiyauyukhanik munakgiyauyukhanik havakhautikhanik pihimayut ipiknakpiaktunik mikharut uingaiktauyukhaniklu ihuinakgutivaktunik nauvaktunik nirgiitinik aniagutingukpaktunik imalu kufiutilanginik nungumanginiklu mikharutlu ilingaikhimayukhanik nirgitinik tahapkunani tunungani nunalingni.

Aniagutit taakuyauvaktunik aah kayaknaitut tahapkununga Ahiakmiutanik-Kivatani ammihuakyuit pihimayunik tutkikgutivaktunik timainik kanukgitakhainiklu, hingaivaktunik kufiutilanginik, imalu inuuvaktuniklu. Tahapkuat Ahiakmiutanik-Kivatani ihiviuktauvaktunik naunaitkutakhanik katitikhimayunik talvuna nunalingni havakhikhimayunik havakhautikhanik munakgiyauyukhanik havakhautikhanik angunnahuaknikmun talvuna 2004 talvunga 2014 pihimayait atukhugit naunaitkutakhanik havakhikhimayunik pidjutitkiyumikhimayunik naunaitunik aniakgutilingnik ilihimayauvaktunik tutkaaikgutivaktunik tuktunik. Tahapkuat Ahiamiutanik-Kivatani tuktuitainik pikaktunik aniakgutinik Aniakgut Aukmi Alphaherpesvirus (Aniakgut Aukmi), Aniakgut Kilaalakivaktunik avaatainik Kuupilgunguktaktuniklu Aukmi (Aniakgut Kilaalakivaktunik), uvunalu Aniagut Kaalakni Nuuvaknitunik kanukgitunik 3 (PI3), Aniakgut Puuvakni Aukmungakpaktuklu 4 (Puuvak) uvanilu Aniakgut Aukmi taryumiutanik tuktutainik *Toxoplasma gondii* (Aniakgut Aukmi). Tahapkununa pivikhaitut aniakgutingukpaktunik aukmi Niakungnungakpaktuniklu

NUNAVUT INUIT ANGOTIGHALIGIYIIT
ELIKKOOKTUK EKAYOOTIGHATIGOT ANGGIKATIGIIGOT

Ona anggikatigiigot eniktiktaohimayok ajiikotigiplogo, holi atokhimmaaktuk obloanit sainiktaolgaaktillogo katojikatigiinit.

OKOAK Kugluktukmi Angoniatit Katimayiinit

OKOALO Nunavut Inuit Angotihaligiyikyoaginnit
Kagikhinigmit, NU

Ehomaghakhiotigiplogit atoktaghat, himmaohiotigilogit okagiiktaohimayut, angikatigiigotaohimayollo, piyaghalioktaohimayollo titigakhimapllotik omaaniittut, naliak koak anggikkatigiigotaohimayut imaa;

1. Havaagiyaghait Atanggoyaoyut

1.1 Kugluktuk Angoniatit Katimayiit havaagiyaghait ekayoghimmaaklogit Kugluktukmi Angonahoaktoligiyiit holiyaghaitigot havaagiyaghaitigollo malikalogit okakhimayot anggikatigiigotaoyomi.

1.2 Atanggoyap havaagiyaghait havakvigiyani ekayyyoooktoklogit okoa Nunavut Inuit Anggotighaligiyikyoanggit tamaita ehoaghaktaghat nakatkikhimavagiaginni nakoatollo ehoaknighaoyotigot piyagiikhimavagiaginni.

1.3 Nunavut Inuit Anggotighaligiyikyoanggita havaagiyaghaginggitait hapkoa hilataaniitpata elaliotihimaittollo Anggikatigiigonmot kihiani Kugluktuktup Angoniatit Katimayiita naamagikpajok anggikhimakpatjokloniit titigakhimayomit.

2. Havaaghat Enikhimayaoghaoyot

2.1 Nunavut Inuit Wildlife Anggotighaligiyikyoanggit tigohiyoghaoyut havaagotkoyaoyonik ayoghaotigihimaitomik pittiaklogillo imaa, titikanik totkiotiakhimalotik manilikotiniklo okoa ekayoklogit Kugluktuktup Angoniatit Katimayiit hapkoningga immakaa: :

- Ekayokpaklotik ehoaghailigaikpata okiokot maniknik atoktaghainik Kugluktuktup Angoniatit Katimayiita;
- Ekayokpaklotik titigakviani havaagoyoghat nakoatot aolapkaklogit maniktigot ekayotighanik tonihimnahoagaikpata okakhimayok maliklogo anggikatigiigotigihimayainni Kugluktuktup Angoniatit Katimayiita

miitiigahoaligaikpatalo ehoaghaotivaklogit katimajotigiyaghainik maligaghallo atoktaghallo nalliomanahoaklogik okagiikhimayut anggigotigihmayaini.

- Eniktigivakloni manighaagiyaghainik havaktit katimayillo manighaaghainik kmiitiktaagaikpata maniktigollo himmaotighanik kojitigiyaghaniklo itkaptaaksiligiyit toghiktogaikpata maniktigot atoktaohimayonik;
- Titigakhikmavaklogit tamaita akiligaghailo okoa Kugluktuk Angoniatit Katimayiita elaokpata honatigot aolajotigivagaitigot Kugluktuk Angoniatit Katimayiita
- Tonivaklogit akiligaghat akiliktaagaikpata okononggа togaangmata maniknik toniyaohimayonik Kitikmeoni Anggotighaligiyit Katimayiinota manikakvianot anggiktaoyagaipatalo okonangga Kugluktuktup Angoniatit Katimayiinit;
- Ekayokpalogillo ehoaghailigaikpata maniknik manikakviani okoa okonangga Kugluktuktup Angoniatit Katimayiini;
- Takopkaktitivaklotiklo okiop elanggagot manilikotigiyainnik takoogiakakmatigik okonangga Kugluktuktup Angoniatit Katimayiini hapkoa tamaita manighaaghallo akiligaghallo alaniklo naonaiktaghanik eniktigivakloni;
- Ekayokpalogitllo okonangga Kugluktuktup Angoniatit Katimayiini katimayit anggommatihimakoplogit honatigot piyaghanik Malitgaitigot pitkoyaohimayonik;;
- Ekayokpaklogolo Kitikmeoni Anggonahoaktoligiyiita titigaktiat totkighaiyi ehoaghailigaikpat manilikotainik okonangga Kugluktuktup Angoniatit Katimayiini , ovalo
- Alallo Manilikotit titikaniklo totkoktoinik pikahiotihimakpata okiokot maniknik atoktaghanik enigiiyaghimayonot maniit pihimayot Nunavunmi Anggotighaligiyikyoatkot Katimayiit Nunavutlo Kavamainit kakoknggogaikpat takoyaoyomakatakpakniakmata.

3. Hivitonighaa

3.1 Ekayoktoijotit pipkakhimayait Nunavut Inuit Anggotighaligiyikyoaginnit April 1-mi, 2019 March 31-mot, 2020. Ona anggikatigiigot atogoikniaktuk nonggotpat March 31, 2020-gonkkat kihimi havaotighatigot anggikatigiigot oigoyaokpat titikatigot pihimaloni.

4. Akiligaghat Ekayoktoijotaoyotigot ovalo Aoaltitijotaoyotigollo, akiligaghanik Akiligiikhimayonotlo

4.1 Okakhimayut maliklogit atoagaghat malikayaghallo, ehomaghakhiotaohimayollo kanoktot hivitoyomik havaagovaktot eliktaovakmagaagita okonangga omani anggikatigiigotaoyomi Nunavut Inuit Anggotighaligiyikyoaginit t, talva Kugluktuktup Angoniatit Katimayiit akilikpakniaktait okoa anggikatigiigotaoyomi Nunavut Inuit Anggotighaligiyikyoagit okiok elainnagagot pihimalotik imaa $3,125-taalamik ovaloniit okiokot $12,500-taalamik.

5. Nunavut Inuit Angotighaligiyiita Titigakviat pihimayut elikkookhotik kaantolaaktioyut

5.1 Ona Anggikatigiigut anggikatigigotaoyoghak nakoatot havaagoyoghat aolatiagotighaitigot, ovalo Nunavut Inuit Angotighaligiyiit havakviat atoktaghagitayit maliklogit okakhimayut anggikatigiigotaohimayuk imaa

ONA ANGGIKATIGIIGUT sainiktaoyuk atolikhonilo okongnangga Nunavut Inuit Anggotighaligiyiinit okoalo Kugluktuktup Angoniatit Katimayiinit anggiktaohimalikmat tamagiknit.

Okonongga Nunavut Inuit Anggotighaligiyiinit:

Jason Mikki	Oblok
Haklovioknaihiyoghak	Oblok

Okonongga Kugluktuk Angoniatit Katimayiini:t

Eghivaotalik	Oblok
Atanggoyak	Oblok

(Niakumungakpaktunik Aniakgut), Kikturgiaanit Aniakgut Hakugiktuk (WNV), uvunalu Aniagut Puuvakni Imakpaktut Naamuniayaakhutik Puuvak Bovine respiratory syntical virus (BRSV) (Ilidjuhit 1) (Carlsson *et al.*, uvunalu havakhiktauyunik makpiraagakhanik; Kutz *et al.*, 2001).

Ilidjuhinik 1: Aukmi aniakgutingukpaktunik pihimayunik ihiviuktauhimayunik aniakgutinik ingniktikhimayunik Ahiakmiutanik Kivatani tuktunik.

Aniakgutit	Ungnalluk Ihiviukhimayunik aukmi aniakgutingukpaktunik (%)	Unguhulluk Ihiviukhimayunik aukmi aniakgutingukpaktunik (%)
Herp	56	46
Pesti	35	33
PI3	31	13
Neo	--	--
Bru	2	0
Toxo	0	14
WNV	--	--
BRSV	--	--

Ahiakmiutanik-Kivatani ammihuakyuinik havakhikhimayunik nalvaktauvaktunik pihimayunik pikaktunik ilitukgiyauvaktunik *Puuvakni Aniakgut* Aukmungakpakhunilu 4 uvunalu *taryukmiutanik tututainik* tahapkuatlu aniakgutingukpaktunik ilihimayauyunik pidjutitaktunik atadjutiniklu inungni. Tamna Puuvakni Aniakgut, kanukgilidjutivaktuk pilakivaktuk imaktitpakhugu niki, imalu tadja ihuigutivaktunik aah ikitiyunik nallautakgutauvaktunik ilikuaktunik talvanitunik ammihuakyuinik. Angunnahuakpaktunik ilitukgittaktuniklu aniakgutikakut tuktu pigumik talvani kuniakhugu avaatainik puuvingayakhainik naliak tuhiiatkumik ilihimangnakhigumik. Tahapkununa taakungnakhigumik avaatainik pidjutiyuutut aniaktuk tuktu naunaiktuk aukmi aniakgutilik kaahakgumik tuktunik nikanik pikakgungnakhiyuk aniakgutimik aukmi karitangmungaktaktunik (Kutz *et al.*, 2001). Uvuna pidjutigumik tamangnut kayaknaktunik, nikitianik nikinik upalungaiktauvaktunik nikikhanik aviktukhugit panikhikhugit atugakhanik piyukhauyunik ilauyunik kuniaklugit nikiitanik uutpiaklugitlu imalu unguyaugangutaalu hangnakittiaklugit kilaamik nikiikhanik nirgitinik pilutik kayakgitiaklutiklu havagiyautiak pilutiklutik atugakhainiklu.

Aulladjutiplugit havagiyauyunik Ahiakmiutanik-Kivatani tingmitikut havagiyaugumik pihunguyut ikayutauyukhanik ilitukgiyauyukhanik kayaknaktunik angutinik nirgitinik kufiutilanginik mikharutlu tunungani nunalingni. Kihimi, ilingaitunik pidjutikiyumiktauvaliktunik nakyunik nirgitinik pikangitunik aniakgutaitunik uvunalu taakunganktunik ilingaitunik ilidjuhikgiyainik (Boorse, 1977). Tahapkununa naunaitunik aah ilingaitunik tuktunik ammihuakyunik pidjutitaktunik ungigliyauhimayunik naapkiyukhanik mikharut kangikgianginik tahapkuninga nirguakgutinik kufiutilanginik ilidjuhikgiyakhainik imalu unngniguutinakpaktukhanik, nirgitinik ilingaitakhainik, uvunalu talvanituniklu kuugluktitainikmunlu (Macbeth unalu Kutz. taiguakgakhanik.). Uvunalu hivuniktuiyunik utiutifaakgutilugit pulakfaaklugit uvaptingni tadja upidjutauyunik mikharut uingaikgutauyukhanik tahapkuninga Ahiakmiutanik Kivatani ilingaitunik pidjutitkiyumiklugit nakutkiyauyunik upidjutikhanik havagiyauyukhanik.

2.2 Ingilgayangit Atukhugit Ivaayaktukpaktunik Ingniukviit, Hila Malikhugu Avaatingnut Nunguumanginiklu, Nirgiitinik Huraat Atukpagainik Huli

Nunamiutanik tuktutainik ilihimayauvaktunik pidjutivakgainik ingilgayakgivakgainik avaatingnut-nungumakgiktunik ivaavit ingniukvit imalu atukpagainik alaakinik nunatuttukanik nunani namukgivakgainik hila malikhugu atukpagainiklu. Nunauyami nayugainik pidjutivaktunik naunaikgutivaktunik humitakhainik tuktuit ingilgavaktunik, ikattunik naliak naapktunit aputikaknit kikumaanitlu kuugait kuugalliagakaknit, pidjutivaktunik hila taimailitkangut nakugiyaulluakpargait ingilgayangit ivaavit ingniukvit. Tahapkuatauk ilitukgiyaulikpatut ahiagut ingilgayangukhutik apkutait ayuikgutauvakhimayut gullavaknik nurgaitunik tuktunik malikpaktunik hivuliuktiuyunik gullavaknik nirgitinik tahapkuatauk pidjutivaktunik ingilgayangukhutik atukhugit pitkuhiktukpaktunik apkutait (Nicholson *et al.*, 2016). Tamakpiangani ukiumi, tahapkunani hila malikhugu avaatingnut nungumanginiklu havakhikhimayunik aviktukhimayunik tahapkunanga havaktit umayulikiyit nunalikiyit pihimayut havakhikhimayunik ammigaikhugit havakviuyukhanik: ivaavit ingniukvit/nunaini ingniukvit, auyaalihami, kitikuktumi auyami, auyaakhaktumi uukiakhami/nuuliaktut, uukiakhami/hivuagut nuuliaktut, uukiumi, uupingami ivaavit ingniukvit imalu hivuagut-ivaavit ingniukvit (Nagy *et al.*, 2005). Malikhugu hila kanukgitakhanganiklu (niglaumatilangaah, hilakiumannianik, apuut naaptunianik, nunakgikgiakhanik ilingaitakhanik naunianik) naliaktalvanitunik kayaknaktunik nirgitinik huraanik, tuktuitlu atuktaktaitlu nirgiavigiyainik aahikuklutik ilihimadjutigiplugit ingilgaayakgivakgainik. Uvunau ilidjuhikgiyainik, aah kauhimadjutiplugit ihuinakgutit nakungikgutivaktunik mikharut naunianik nuttanik nunaap naunia pilakittaktutlu aaihikpaktitaktuklu upingami (Nicholson *et al.*, 2016).

Karitauyakukgutinik atukhutik ilitukgiyauyukhanik humitakhanik naunaitkuakhanik havakhikhimayunik tahapkunani unguhulluknik Ahiakmiutanik Kivatani talvuna 1996 talvunga 2004 havakhikhimayunik naunaiyakhimayunik tutkikhaiyanginik ammigaitunik nalikmik hila malikhugu avaatingnut nungumanginiklu (Ilidjuhinik 2) (Nagy *et al.*, 2005). Tamna ivaavit ingniukvit talvanitunik Ahiakmiutanik Kivatani tuktunik pargutihimayunik talvanga kivatanit Bluenose Tahik talvunga nunalingni Kugluktukmit. Taimailitkangut ammihuakyunikangut, Kugluktukmiut ilihimadjutigivakgat itkauyaktunik unguhulluknik tuktunik talvanitunik nunalingni. Talvuna auhaalihami, tahapkuat gulluvait imalu nurgaitlu kangniglivaktunik hingnaanut Taryukgalukmi, tahapkuat gulluvait imalu nurgaitlu ingilgavaktunik c hivuagut ingilgatinagit hivumut hivurganut hiamitutik talvanga taryumi talvunga tununganut hivurani talvani Great Bear Tahik Auyaaktumi. Tahapkuat nirginiakpaktunik talvani nunaini talvanga atulihami Nikilikivia talvunga Hikutilikvia (auyakhaaktumi). Taimailitkangut uukiakhukgangut/nuuliaknakhigangut, tahapkuat ingilgaalvit gulluvait tamatkiutivagut tununganì talvanilu kivatani talvani Great Bear Tahik, talvanilu Mctavish Ikuuhianik, tikkitpagatlu talvunga Dehcho Nunatuttukanik. Talvunalu uukiakhami/hivuagut nuuliakpaktunik pidjutauvakmiyuklu talvuna kanuk gulluvait tuktunik ingilgayakgivakgainik apkutigivakgainik akungainilu hivugani Great Bear Tahik talvanilu Dehcho Nunatuttukanik uvunalu pidjutivaktunik uukivaktunik talvani nayuganik. Talvuna uukiumi avaatingnut nungumanginiklu tahapkuninga Ahiakmiutanik Kivatani tuktunik ilihimayauvaktunik tamatkiumayumik pargutivaktunik tahapkuninga ingilgayuniklu Kingaumiutanik tuktunik ammihuakyunik. Tikitainakhugu ungniguutivaktunik, taimailitumi hila upingakhaktuk ivaavit ingniukvit nayulikhugit, tahapkuat tuktuit ingilgalikpaktunik tungnumut ingilgayanginik apkutainik mikharut ivaavit

ingniukvit nayugainut imalu kangniglivaktunik talvunga Kugluktuk. Kihimi, tadja havaktit umayulikiyit uingaikgutivakgainik tahapkunanga Ahiakmiutanik Kivatani ammihuakyunik nirgitinik atuktauvaktunik naunaiyainikmun havakhikhimayunik nungumalingnik kangilimaitunik, pihimagumik havaktituaggivakgamigit kihimi mikharut mikitkiyanik kufiutilanginik gulluvaktunik tuktunik havakhikhimayunik kunguhikmialingnik naunaitkutaliukpaktunik.

Ilidjuhinik 2: Ahiakmiutanik-Kivatani Hila Malikhugu havakhiktauyunik kilamiukhutik havakhikhimayunik uvuna Nagy et al., (2005).

Hila Malikhugu avaatingnut nungumangniklu	Talvuna
Ivaavit ingniukvit/ hivuagut-ivaavit ingniukvit	Imakguktikvia 1 talvunga 25
Auyaalihami	Imakguktikvia 26 talvunga Tarhilikvia 15
Kitikuktumi Auyami	Tarhilikvia 16 talvunga Nikilikivia 7
Auyakhaliktumi	Nikilikivia 8 talvunga Hikutilikvia 7
Uukiakhami/nuuliaktut	Hikutilikvia 8 talvunga 31
Uukiakhami/ hivuagut-upalungaiyaktunik nuuliakniaikhimayunik	Tatiaknakhivia 1 talvunga 31
Uukiumi	Ubluikvia 1 talvunga Kitikauyakvia 31
Upingami, upingami ingilgayangit ivaavit ingniukvit, hivuagut-ivaavit ingniukvit	Kitikautiyuk 1 talvunga Imakguktikvia 31

Naahugiyauyunik mikharut nallautakgutauvaktunik tahapkuninga Ahiakmiutanik Kivatani hila malikhugu avaatingnut nungumanginiklu ipiknakpiaktuklu malikgakhanik ilidjuhikgiyakhainik nirgitainik havagiyaulutik ilitukgiyaulutik ipiknakpiaktunik nunamiutanik naunianik. Kakugu ipinakpiaktukhanik havakhautit naliak pirgumiyauyunik nayugainik piluni takungnakluni nallautakgutauvaktunik nallikmiktunik piyukhaniklu ilidjuhikgiyainik kanukgilidjutivaktuniklu naamukgiyauyungnaikpaktunik nunainik kanukgilidjutigangumik pidjutitaktuniklu naunaikluni ingilgayanginik aaihikuklutiklunin ammihuakyuinik avaatingnut nungumanginiklu (Nicholson *et al.*, 2016). Kanukgituniklu nalikmutlu pilutik ungniguutiyukhanik havakhikhimayunik atadjutigiktunik havakhautit upidjutaulutik, humi nakutkiyauyunik havakhikufuuluktunik havakhautit havakhikariakaktunik pihimayunik haatitikhimayunik kakungiakgangut mikharut kiudjutikhanik alanguktikpaktunik talvanitunik atugakhanik.

3. KAYAKNAKTUNIK

Ammigaitunik naunaitpaktunik aah kayaknaktunik ilihimayauvaktunik, kihimi havakhikhimayunik tahapkununani atukpaganik katimayinit mikharut Naunaitkutauyunik likiklvaliayunik Paaikpaliayunik Nirgitinik Huraanik talvani Kanatami (COSEWIC), kayaknaktunik havakhikhimayut nallautakgutauvaktunik nuutkangayuitunik naliak pivaktunik talvanikpiaklu biplutik nakuyungyaikguttivaktunik ihuigutiplutik ilikuaktunik kufiutilanginik. Kayaknaktunik nirgitinik huraanik paagutigangumik (nakungikgutivaktunik ihuinakpaktunik nirgitinik huraanik), angunnahuaknikmun aaiyuknakhunilu havaginahuakhugu, hila alangukpaliayuk hilaaklukyuakginalikhuni tadjaguinakpalikhunilu kayaknakhiyuni, nirgitinik umayunik alangukpaliayunik, kuupilgungnikhutik imalu aniakgutinikpaktunik, uvunalu ikummutitukpaktunik imalu alguyaktuktunik uvunalu uyagakhiukvit pivaktutlu pinnaktaktunik talvanikpiak kayaknaktuniklu. Kayaknaktuniklu pidjutitautaktut upalungaikhimagumik imalu nallautakgutauvaktunik havakhikhimayunik tahapkununa IUCN-CMP (Hilaryuami Ilidjuhikginnahuakniaklugit Katimayit-Ilidjuhikginnahuakniaklugit Nallautakgutauvaktunik Panaarit) ilihimayaungitunik kayaknaktunik naunaitkutat atugakhanik (Hivitunikhak *et al.*, 2009). Una naunaitkutak piyauyukhak iniktiktauyukhak mikharut Ahiakmiutanik Kivatani ammihuakyunik imalu kanukgiliukgutauyukhanik havakhautikhanik pilutik havgiyauyukhanik tahapkunani nunalingni.

3.1 Inungnit Kuugluktitainikmun

Atugakhaliuknikmun uyagakhiuktit imalu aullaktikgutikhanik, ulapkiviuvaktunik hulilukakviuvaktunik pidjutitaktutlu ihuikgutilutik talvanitunik tuktunik ilidjuhikgiyainik imalu ihuikgutilutiklu tuktunik nakungitumik. Talvunikpiaklu ihuinakgutinaktuk tahapkunanga inungnit hulilukaktunik pidjutitaktuniklu pilakipkailutik alanguktiklutik pinaktukhaniklu, kanukgitakhainik imalu kufiutilanginik nungumanginik naliak ilanganik talvani nunaap naunianiklu kanukgilidjutitaktuklu nayugainik nakugiyauyungnaikluni tahapkunani tuktunik imalu pilakilunilu aaihikpaktittaktuklu apkuutainik naliak aaihikpaklutik (Nicholson *et al.*, 2016). Kanukgilidjutigangumik ikikligangumik ammihuakyunik, pivaktut hugilluayuitait inuit kuugluktitaigangumik imalu pidjutivaktutlu hugiyuitait inuit pidjutivangmiyuk ammigaikgutitaktuklu tuktunik.

Talvanitunik Ahiakmiutanik Kivatani avaatingnut nungumanginik, pikaktunik uktutikhanik uyagakhiuktinik havakhautikhanik. Talvani Tunungani Nunatsiami, tahapkuat Mackenzie Valleymi pilik ilingaitunik ukhuuklukaktuk imalu ammigaitunik nalvakhiukpaktunik talvani nayuganik havagiyauyunik. Uktutikhanik kanukgituniklu Mackenzie Gasilik Havakhautit (MGP) ilidjuhikgiyanik gasilik tuukhualgit imalu Mackenzie Valleymi Apkuut ilaugumi pidjutihunguyuklu tahapkunani Ahiakmiutanik Kivatani tuktunik (ACCWM, 2014). Talvani Tungani Nunatsiami uvanilu Nunavutmi pikaktunik aah ilingaitunik uyaganitunik kiplagikpiaktunik imalu alauyunik ipiknakpiaktunik uyaganitunik pargutivaktunik talvani avaatingnut nungumanginiklu. Tamangnik nalvakhiuknikmun imalu uyagakhiuktunik havakhautit pidjutitaktut ihuikgutilutik tahapkuninga Ahiakmiutanik

Kivatani tuktunik kanukgitakhainik mikharut nayugainik imalu ilingaitakhainiklu. Kalikgiktikpaktunik ihuikgutivaktunik pilugit ihumadjutigiyauyukhanik.

3.2 Kayaknaktunik nirgitinik huranik

Amaguk (*Canis lupus arctos*), Akhaak (*Ursus arctos*), Kalvik (*Gulo gulo*) unalu Kupanuakpak (*Kupanuakpak*) pihimayut ilihimayauvaktunik kayaknaktunik nirgitit huraat. Tuktuit ilihimayauyut nikigivakgait amakuut, akhaitlu, imalu kalvitlu, kanukvitunik amaguut nikigilluakpargait tuktuit nikigiplugit (90%) (L'Herault *et al.*, 2016). Pivakulluaktut kanukvikhaituk, amakuut ilihimayauyut kauhimayauyunik kayaknaktut nirgitit huraat nunamiutanik tuktunik, tahapkununa hivituyumik mikharut kanuktun amakuut ihuigutivaktunik ikiiklitivaktunik imalu naufaakpaktunik tuktunik ammihuakyuinik ilitukgiyauyuitunik huli, talvunalu ahuutaklu amagut nakungikgutivaliayunik ikiklitivaliayunik imalu naufaakpakgiakhainik tuktunik ammihuakyuknik ilihimayaungitut. Talvunalu nirgitinik nikauvaktunik atadjutivaktunik ihumadjutigiyauvaktunik aah kakungukgaikpat havakhiktaulutik kanukgiliukgutauvaktunik amaguanik nuutiktauvaktunik ikiklivaktunik tuktunik pinnaktukhanik (Klaczek *et al.*, 2016). Pilakivaktunik mikharut tuliikgutauvaktunik tahapkununa nirgitinik kayaknaktunik kufiutilanginik kiudjutauvaktunik tahapkununa tuktuikpaliayumik, talvunalu kayaknaktunik nirgitinik pilutiklunin ilanganik angiglihimalutik talvuna (ACCWM, 2014). Uvunalu pidjutitaktuk uingaikgutaulutik huuk angunnahuayuktunik imalu nunalingni inukatigikpaktunik pidjutivakgainik ihumadjutigiplugit amagunik ammigaikpalaktunik uvunalu pidjutitkiyumikpaktunik nalukhanginamik havagivaktamingnik pidjutitaktuniklu naudjutilutik kayaknaktunik nirgitinik munakgiyauyukhanik havakhautikhanik ikiklitiyukhanik kayaknaktunik nirgitinik ayuknakhitivaktunik.

Tadja ubluanganut, umayulikiyit havagivakgait huli kiudjutikhanik mikharut havagiyauyukhanik amagunik ammigaikpalakmataalu havagiyauyukhauyut umalyulikiyit huli kiudjutikhanik mikharut havagiyauyukhanik amagunik ikiklitinnahuaktukhanik havagiyauyukhanik. Pilgumikhimayunik kayaknaktunik amagunik ammigaikpalakmataalu havagiyauyukhauyut, tahapkunanilu Finlayson talvanilu Aishihik nayugainik talvanitunik Yukonmi, pivakhimayuk takungnakhunilu nakuyumik havakhikufuuluktunik nakukgutauyunik havagiyautiakpaktuk taimaliukgangumik nakukgutauvaktunik pidjutaugakpakhutik mikharut kulinik ukiunik. Kihimi, tamna havagiyauvaktuk akhuukgutauplunilu imalu nutkaktauhainakgangutaah, tahapkuat amagut nauhainakpaktutlu, (Hayes, 2010).

Malikhugit tahapkuninga Pitkuhiktukpaktunik Kauhimayutukanik, tahapkuninga akhaat talvanitunik Ahiakmiutanik Kivatani pidjutiyunik ammigaiktunik ihumagiyauvaktunik, pivaktunik ammigaikhutik takungnakhivaktunik kayaknaktunik nirgitinik talvani ivaavit ingniukvinik imalu hivuagut talvungauvaktunik ivaavitnut avaatingnut nungumanginiklu (Kugluktuk HTO, pers. comm.). Akhait pinniakhimavaktunik nurganik, ungnilihaniklu nurganuanik, kihimi pivaktuniklu unguhullukniklu. Akhait pidjutiliktunik ayuikgutiplugit alauyunik atugakhanik tuukutiyanginik

tuktunik: kimalatiplutik tuktunik, kimalatiplutik ingilgayunik tuktunik, uutakiukhugit tuktunik ingilgayanginik apkuutainik ikakvinik kugakni, naaimaplugit nurgainik (Reynolds & Garner, 1987). Pivakhimayunik kuniakpakhimayunik tahapkuninga Akhaanik tukutiyunik tuktunik adjiliuktauvakhimayunik, tahapkuat Akhaatlu tuktunik atadjutigikgiakhainik ilihimannaitut (Skoog 1968; Murie 1981).

Kuupanuakpak nikigivakgainik mikauyunik nirgitinik, kanukgituniklu sika jaapanikgalik tuktu (*Jaapanikgalik tuktu*) unalu Dall's imngnaik (*Imngnaik*) (Kerly, 2013; Nette *et al.*, 1984). Uvani Finlandmi, Kuupanuakpak imalu ilihimayauyuk nakukgutiggivakganik nikigigangumiuk tamna tukutauyuk tuktu (Tututainik) nurgakniki (Nybakk *et al.*, 1999). Kuupanuakpait ubluliukpaktunik talvani hivugani hungmiplugit Kungnahik kugak kuutikguani, pivaktunik ammigivakgainik Ahiakmiutanik ivaavit ingniukvit. Tavunalu pinaktaktunik tahapkuninga kuupanuakpak tukuniakhugu ungnilihak Ahiakmiutanik Kivatani nurgakmik nakukgutikhaminik tukulugu. Talvuna tukutiyunik pilutik nikikhakhiuktunik pidjutivakgainik nirgitinik, hikhiknik imalu mikiyunik nirgitinik, pidjutiyuniklu ikhiuktunik uupipkakhugit ahiit huttumiaktukhanik alauyunik nirgitinik (Nette *et al.*, 1984; Nybakk *et al.*, 1999). Tadja havakhautikhanik havagikyauyunik naahugiyauyukhanik kuupanuakpanik tuktunik atadjutiggikpagiakhainik, imalu naunaitkutakhanik tahapkuninga ihiviuktauvaktunik pilugit upalungaikhimayukhanik hivunikgiyakhanik nirgitinik ayuuknakhivaktunik Ahiakmiutanik Kivatani tuktunik.

4. ANGUNAHUAKTAUVAKTUNUIK NALLAUTAKGUTAUVAKTUNIK HAVAGIYAUVAKTUNIKLU

4.1 Nunalingni pivaktunik Angunnahuakpaktunik Ahiakmiutanik Kivatani Tuktunik Nunavut

Kugluktuk kihimi nunalingni talvanitunik Nunavutmi pivaktunik angunnahuakhugu tuktu talvanga Ahiakmiutanik Kivatani ammihuakyunik. Pikaktunik alauyunik nunalingni pivaktunik angunnahuakhugu talvanga ammihuakyuinik NWT.

4.2 Atukhugit talvani Kufiutilanginik imalu Ingilgangnitanik Angunnahuaknikmun Havakhautit

Tahapkuat Ahiakmiutanik Kivatani Ammihuakyunik (BNE) pivatunik ipiknakpiaktunik inudjutigivakgaat tahapkunani Inungni nikigivakgaat/inudjutigiplugu. Tuktut pidjutauvaktut angunnahuaktauplutik mikharut kakungukgangut ukiunganik pidjutauyukhanik alangayunik atugakhanik. Tuktu atuktauvaktuk nikigiplugu, anurgak, iglugiplugu, atugakhaniklu. Inuit angunnahuakpaktut kihimi ihaagiahuktamingnik. Taimani pidjutiyuitut ikaakuklutik nirgitinit.

Ublumi, angunnahuakpaktunik atukpaktunik ingnikutilingnik imalu hikuttinik angunnahuaknikmun tuktunik. Imalu kaayumitkiyaulikmut imalu ayuuknaituni nalvakhiuknikmun nirgitiniklu.

Hila malikhugu:

- *Talvuna ingilgayanginik apkuutait atukhugit, hugihuiklugit hivuliuktit, angunnahuaknikmun/kuugluktitainikmun pilakilugitlunin aihikpaktilugit talvanga ingilgayanginik apkuutait atukhugit them may alter migration.*
- *Ubluktuhivia talvunga Kitikauyakvia – pidjutitkiyumikpaktunik ungnalluknik*
- *Kikaiyakvia talvunga Imakguktikvia – pungnik*
- *Tarhilikvia talvunga Nikilikivik – alaakinik kihimi mikiyumik angunnahuaknikmun*
- *Apitilikvia talvunga Ukiakhuk – Pungnik*
- *Tatiaknakhivia talvunga Ubluikvia – unguhulluknik imalu unguhulluknik*

Uvani 2007, tahapkuat Kugluktuk HTO pivakhimayut havakhikhimayunik nirguakgutinik nutkaktitaiyukhanik angunnahuaknikmun imalu ungunnahuaknikmun niuvgutikhanik tamakpianginik ammihuakyunik talvanitunik Kugluktum nayugainik. Una pihimayuk ikayutauyukhanik tahapkunanga nunalingni nanminilingnik piutigiyainik havagivakgainik talvuna

5. HAVAKHAUTIT/AULLAKTIKGUTIKHANIK/Munakgiyauyukhanik

5.1 Havakhautit Havagiyauvaktukhanik

Tahapkununa atugakhanik talvanitunik ICCMP ilauyunik:

1) Havakhikhimayunik aah nungumagiyakhanik kangilimaitunik KHTO katimayinik angunnahuaknikmun munakgiyauvaktunik tahapkunanga KHTO;
2) Ataniktutauyunik BNECH angunnahuaknikmun uniklutinik tahapkununga KHTO by katimayinik;
3) Havakhiktauyunik aah 'Piyukhaungituk talvani nayuganik angunnahuaknikmun Tuktunik Avaatingnut nungumanginik' piyukhanik ikiklitiyukhanik BNECH angunnahuaknikmun nayuganik pivaktunik ayuuknaitumik pinaktukhanik tahapkununa apkutainik imalu ingnikutilingnik akhaalutinik;
4) Tahapkununa havakhikhimayunik mikharut KHTO munakgiyauyunik havakhiktukhauyunik atugakhanik kauhimayauyunik BNECH angunnahuaknikmun nallautakgutauvaktunik imalu avaatingnut nungumanginiklu, ataniktutauyunik uniklutinik, uvunsalu angunnahuaknikmun havagiyauvaktunik;
5) Naudjutikhanik aah havakhautikhanik imalu kinikhiayunik mikharut kinikhiayunik piyukhanik naudjutikhanik aah kayaknaktunik munakgiyauyukhanik havakhautit piyukhanik ikiklitiyukhanik nirgitinik tuktivaktunik ayuuknakhivaktunik BNECH;
6) Auladjutigiplugit imalu nakuhitkiyumiktukhanik ilihaknikmunikmun KHTO katimayinik tahapkuninga tuktunik, pitiaknikmun angunnahuaknikmun havagiyauvaktunik, uvunalu alauyuniklu nirgitinik angunnahuaknikmun; unalu
7) Havakhiklutik pinnahuaknikmun talvuna ungigliyumiktukhanik naamukgiyauvaktunik naunaitkutak angunnahuaknikmun nallautakgutauvaktunik tahapkunani KHTO kauhimayauyunik omingmaknik talvanitunik Kugluktuk angunnahuaknikmun nayugainik piyukhanik ungnigutiyukhanik angunnahuaknikmun ayuuknakhivaktunik tahapkunani BNECH.

5.2 Pangnatauyunik Aullaktikgutikhanik

Tahapkunani KHTO Angiutauvaktunik Atadjutigiknikmun Nunaligni Tuktunik Munakgiyauyukhanik Pangnattauyunik.

<u>Havakhikhimayunik aah nungumagiyakhanik kangilimaitunik KHTO katimayinik angunnahuaknikmun munakgiyauvaktunik tahapkunanga KHTO</u>

Tahapkuat KHTO nallautakguttauvaktunik talvuna Imakguktikvia 2015 talvunga Imakguktikvia 2016 kanuk 190 BNECH pihimayunik angunnahuaktauvaktunik uvani ukiungani talvanitunik. Tahapkuat KHTO nallautakgutauvaktunik hila malikhugu aviktukhikhimayunik nunguyumik ukiunganik BNECH angunnahuaktauvaktunik mikharut: 20 nik uvani auyami; 20 nik uvani ukiakhami; 100 nik ukiumi, imalu 50 nik uvani Kitikakvia. Talvani 100 BNECH ungunahuaktauvaktunik ukiumi, tahapkuat angunnahuaktauvaktunik talvani Napaktolik, kanukgitilanganik ungahiktilanganik 200Km

hivigiyanivyak- kivatanivyaklu talvanga Kugluktuk humi pidjutivakhimayunik takuyauvaktunik ammihuakyuknik tuktunik ukivaktunik uyagakakniknini imngnaani nayugainik. Pidjutauvakhimayunik ukkautigiplugit kanukgitakhainik tahapkunani BNECH, kihimi itkumagiyauyunik tahapkuninga KHTO.

Pigulluaktilugit tahapkuat KHTO angunnahuaknikmun pihimayunik nallautaktiakhimayunik pihimayunik 75%, tamakpianganik ungunnahuaknikmun pihimayunik ikitkiyauyunik tahapkunanga 340 TAH. Pidjutivakhimayunik ilanginik ukiunginik angunnahuaknikmun pidjutitkiyumikhimayunik imalu atadjutiyunik tahapkununga BNECH kufiutilanginik pihimayunik ammigaikyumihimayunik, naliak tahapkuat BNECH pihimayunik kangniglivaktunik nunalingni talvuna pivakhimayut.

Talvuna kauhimayauvaktunik mikharut ikikliyumiktunik angunnahuaknikmun uvani Imakguktikvia 2015-Imakguktikvia 2016 talvanitunik ammigaitunik, uvunalu ilauyuniklu:

a. ammigaitunik Inuit pikangitunik atugakhanik, naliak ayuikgutauvaktunik talvuna angunnahuaknikmun tuktunik;
b. tahapkuat BNECH kufiutilanginik pihimayunik ikiiklivaliayunik imalu uktutikhanik pidjutiyukhanik imalu angunnahuaknikmun BNECH imalu ikikliyumiktunik naunaikhutik (aah ilidjuhikgiyanik kiudjutinik alunguktikattakpaktunik);
c. pikangituk ukiumi apkuutaituk, naliak tingmitinik atuktauvaktunik angunnahuaknikmun tuktunik avaatainik Kugluktuk, pidjutivaktunik naluvyaktunik angunnahuaknikmun, uvunalu tahapkuninga pikangitunik atugakhanik aulakgiami ungahiktumut talvanga nunalingni pidjutilimaitut aulaklutik angunnahuaklutik BNECH.

Kanukgilidjutivaktunik, tahapkuat KHTO pilutik havagilugit havakhiktauyukhanik piutigiyakhainik nungumakgiyakhainik avaatkulimaitunik 340 tuktunik angunnahuaktauvaktunik uvani 2016-17 atukhutik havakhikhimayunik tahapkunani KHTO piyakgaihimayunik havakhikhimayunik. Tahapkuat KHTO pilutik aulaktikgutikhanik atugakhanik atugakhanik pidjutauyukhanik tuniyaulutik huppimutaitunik piyukhanik KHTO inukatigikpaktunik ungunnahuaknikmun tahapkuninga 340 nirgitinik tahapkunanga BNECH. Pihimayunik amigakhuuktauyunik kilamiukgutaugamik tahapkunani NWMB pidjutauyukhanik pidjutitkiyumiklugit iniktikhimayunik havakhiktauyukhanik naunaitkutanik tahapkununa KHTO havakhiktauyukhanik pangnatauyunik, kihimi tahapkuat KHTO imalu pidjutitaktunik havakhiktaktuniklu.

Tahapkuat HTO havakhikhimayunik aah nirguakgutauvaktunik pidjutitaktukhanik angunnahuakpaktunik pihimaluktuukhanuik naunaitkutanik angunnahuaknikmun nugudjutilingnik ubluinik atuknianut hivuagut piyukhanik tunifaaklugit tahapkununga HTO. Tahapkuat HTO havakhikhimayunik aah nirguakgutauvaktunik pidjutiyukhaungitunik angunnahuaknikmun aah atauhikmik tatkikhiutinik talvuna Imakguktikvia 1-Tarhilikvia 1st. Talvunalu Angutinik Malikgak havakhikhimayunik pidjutiyukhaungitunik kimailutik niknik tuktunit ikaktaulutik uvunalu kanukgilidjutivaktunik ihuinakgutauvaktunik mikharut Angutinik Malikgak. Karyuk Malikgak aah .223 uvunalu ungitkiyak tahapkununga ungiyunik ungutinik.

5.3 Munakgiyauyukhanik havakhautit imalu Uniklutikhanik

Uvuna ilauyunik mikharut ataniktutauyukhanik uniklutikhanik, tahapkuat KHTO talvanilu GN havakatigiktunik atautimi piyukhanik tunihiyukhanik havaktit umayulikiyit ihiviuktauyukhanik hivunikgiktukhanik kakugu munakgiyauyukhanik imalu havakhautikhanik ilingaitunik ammihuakyuinik. Imalu ilauyukhaniklu KHTO havakhiktauyukhanik nungumakhainik kangilimaitunik inukni 340 BNECH piyukhanik angunnahuaktautyukhanik, tahapkuat KHTO pilutik aulaktikgutikhanik atugakhanik munakgiyauyukhanik angunnahuaknikmun BNECH. Uvunalu pilutik ikayutauyukhanik kiudjutikhanik kanukgilidjutivaktunik ICCMP pihimayunik havakhikhimayunik ungniguutivaktunik havagiyauyukhanik. Pigumik angunnahuaknikmun naunaitkutanik pigumik kangikgumik tahapkuat KHTO havakhikhimayunik nungumanginik kangilimaitunik, tahapkuat KHTO atugakhanik hungiutiyauyukhanik munakgiyauyukhanik mikharut nallautakgutaulutik tahapkuat ICCMP mikharut tikilugit talvuna nungumanginik kangilimaitunik. Imalu ilanganik uvuna angunnahuaknikmun uniklutikhanik, pilugit kiudjutikhanik ukallautigiyauhimayunik tahapkunanga angunnahuakpaktunik mikharut ilingaitakhainik imalu kufiutilangink tahapkunani BNECH talvuna atuktaulutik havakhikhimayunik pinnaktukhanik naunaitkutakhanik tahapkunani BNECH. Talvuna ilidjuhikgiyainik, tahapkunani kufiutilanginik nurganiklu, tahapkunani hingaiyunik gulluvaknik imalu alauyunik aulayuitunik tukihigiarutinik pihimayunik iniktikhimayunik havakhikhimayunik adjiliukhimayunik ammigaitunik kufiutilanginik omingmaknik talvani nunainakmi kivatani Kugluktuk kanuk naunaikhimayunik ammigaikpaliayunik TAH mikharut nirgitinik. Tahapkuay KHTO pihimayunik havakhiktukhanik piyukhanik atugakhanik upalungaikhimayukhanik mikharut havakhiktukhanik uniklutikhanik tahapkununa BNECH angunnahuaknikmun.

KRWB talvanilu KHTO piyunik havakutigiktunik talvani aullaktikgutikhanik aah ingilgangnitanik kauhimayutukanik havakhautikhanik pilutik Apkutit naunaiyaklugit piyukhanik adjiliuktauyukhanik takuvaktunik, angunnahuaknikmun, hila alangukpaliayuk havakhiklutik naunaitunik havakhikhimayunik.

5.4 Pidjutinaituk Hamani Nunani Angunnahuaknaituk Avaatingnut Nungumanginiklu

1) <u>Havakhikhimayunik aah 'Pidjutauyukhaungituk Talvani Nunani Angunnahuaknikmun Tuktunik Avaatingnut Nungumanginiklu' talvuna ikiklitiyukhanik BNECH angunnahuaknikmun nayugainik pikaktunik pinnaktaktunik apkuutainik.</u>

Tahapkuat BNECH kakungukgangut kangniglivaktunik talvunga Kugluktuk talvuna upingami uvunalu uukiakhami ingayanginik ingilgavaktunik apkutainik. Talvuna upingami imalu uukiakhak ingayanginik ingilgavaktunik apkutainik. Pinnaktaktunik talvunga hivuganivyak wuatanivyaklu talvani Kugluktuk pivaktuniklu ungahivyaktuk kanukgitilanganik 5-10 nik ungahiktilanganik havakhikhimayuk pilakivakmut kugak naliak kugalialakivaktunik ingilgayunik havakhikhimayunik apkuutiliukhimayunik. Uvunalu ikayutauvaktunik kilamik pinnaktukhanik angunnahuayuktunik talvani avaatingnut nungumanginiklu atugakhaniklu naliak ayuikgutauvaktunik nalukhakgungnaikhugit mikaknik

nayugainik kangnituanganik Kugluktuk mikharutlu BNECH kakungukgangut apkuutigivakgatlu apkuhakhutik. Aah ammigaitunik BNECH angunnahuaknikmun talvani BNECH apkuhakhutik talvunakukhutik nayuganiklu.

Tahapkuat KHTO havakhikhimayunik Makpiraliukhimayunik 5.7.3 pikangitunik-naunaitkutak nirgitinik atugakhanik nungumanginik kangilimaitunik talvani Nunguyumik Ukiunganik Tamakpianganut Katimadjutauyunik uvani Imakguktikvia 12, 2017 angunnahuaknikmun piyukhaungituniklu tuktunik angunnahuaknikmun talvani nunani, pihimayunik takuyautaktunik uvani Ilidjuhinik 1 imalu Ilidjuhinik 2. Una nayuganik avaatingnut 300 nungumangniklu talvanga pidjutitkiyumikpaktunik pinnaktaktunik angunnahuaknikmun talvanga Nunatuttukanik Kugluktuk. Pikaktunik hugyaitunik nunainik nukuttanik nayuganik pidjutautaktunik atuktaulutik havakhiktukhanik pidjutauyukhaungituk talvani nunani angunnahuaknikmun tuktunik avaatingnut nungumanginiklu ilauhimayunik ungiyunik kugaknik imalu imngnakniklu. Talvuna havakhikhimayunik uvuna avaatingnut nungumanginiklupilutik naunaiklutik hivituyunik ikiklitilutik tuktunik angunnahuaknikmun hivunikgiyainik ukiunik imalu hila maliklugu tahapkuat BNECH ingilgayanginik apkutainik atukhutik kangnitanginik nunalingni talvanitunik nayuganik, Uungahiktunik, ammigaitunik angunnahuakpaktunik pihimayut nalukhuvyakpaktunik angunnahuaknikmun atugakhainik naliak ilihaktauyuinamik mikharut angunnahuaknikmun tuktunik pitiakgutilugit, imalu ammigaikpalakpaktunik angunnahuakpaktunik nayuganik talvanitunik pidjutitaktuniklu kayaknakhitiluni inukmik imalu angunnahuakpaktunik kayaknaitumik ilihaktauyukhanik kauhimauyauyukhanik. Una pidjutauyukhaungituk talvani nunani angunnahuaknikmun tuktunik avaatingnut nungumanginiklu havakhikhimayunik atugakhanik IQ (ilauyunik atianik talvani titikgakhimayunik). Ilauyunik ubluliukhimayunik kakugu nunai nayugainik havakhiktaulutik uumiktaulunilu mikharut angunnahuaknikun.

Ilidjuhinik 1. Tahapkuat KHTO "Pidjutauyukhaungitunik talvani nunani Angunnahuaknikmun Avaatingnut Nungumanginiklu" naunaiyakhimayuk titikgauyakhimayuk aupayaktumik. Una avaatingnut nungumanginik atadjutauyunik talvanga Kugluk (Kugluktuk) uvanilu Kuungnahik (Kungnakhik) Kugak, tamna taryuk tunungani, uvunalu atautimiuyunik imngnaknik talvanitunik Hatongaat nayuganik talvunga hivugangnivyak wuatangnivyak talvanilu Gurling Nuvukmi Imngnait talvunga tununganivyak kivatani.

Ilidjuhianik 2. Talvuna Angunnahuaknikmun Avaatingnut Nungumanginiklu" naunaiyakhimayuk titikgauyakhimayuk aupayaktumik adjikiktunik Uktutikhanik KHTO, kihimi nunatuttukanik havakhikhimayunik.

KANUKGILIDJUTIVAKTUNIK: KUFIUTILANGINIK UVUNA HIVITUYUNIK (120,000 +)

HAVAGILUGU: MUNAKGIYAUYUKHANIK HEALTHY POPULATION

PILUGIT ATUGAKHANIK

- Ilihaknikmun kanuktun ilidjuhikgiyakhainik angunnahuaknikmun
- Kayaknaktunik Nirgitinik Huraat Havakhautikhanik
- Anguyauvaktunik Havakhautikhanik
- Ingilgangnitanik Pitkuhiktukpaktunik katitiklugit
- Piyukhaungituk Hamani Nunani Angunnahuaknikmun Avaatingnut Nungumanginiklu Havakhikhimayunik
- Ihiviuktauvaktunik Naunaitkutakhanik katitiktauvaktunik – atugakhaniklu

KANUKGILIDJUTIVAKTUNIK: KUFIUTILANGINIK UVUNA AKUNGAINILU IKIIKLIVALIAYUNIK (60,000 +)

HAVAGILUGU: HAVAGILUGIT AMMIHUAKYUIT NAUYUKHANIK

PILUGIT ATUGAKHANIK

- Havagiyumiklugit havakhautikhanik munakgiyauyukhanik
- Nungumalingnik angunnahuaknikmun
- Kayaknakpaktunik Nirgitinik Havagiyauyukhanik
- Ataniktutaulutik Uniklutikhanik
- Ataniktutaulutik Ihiviukhinikmun Naunaitkutakhanik katitiklugit
- Ingilgangnitanik Pitkuhiktukpaktunik katiklugit
- Ilihaknikmun kanuktun
- Havakhautikhanik Munakgiyauyukhanik katitiklugit
- Havagilugit Kakungukgaikpat Katimadjutikhanik pilingnik Tulia-havakhautikhanik panaarit, kavamait, katimayit, nunalingni
- Ivaavit Ingniukvit/Nunaap Naunia Pirgumiyauyuk
- Angunnahuaknaituk Avaatingnut Nungumanginiklu
- Angunnahuaknaituk nanminiliknik piutigiyainik
- Aiyukiktuinikmun tahapkuninga pangniiknik angunahuaknikmun
- Aiyukiktuinikmun alauyunik nirgitinik angunahuaknikmun
- Aiyukiktuinikmun tingmitikut/kufiutilangi nik/gulluvak-nurgak nallautakgutauvaktunik ;nurgak ungnakpaktunik nallautakgutauvaktunik ; ingniktikhimayunik havakhiklutik naunaitkutakhanik ukiukgaikpat

KANUKGILIDJUTIVAKTUNIK: KUFIUTILANGINIK UVUNA IKIIKLIYUNIK (20,000 -)

HAVAGILUGU: HAVAGILUGIT AMMIHUAKYUIT NAUYUKHANIK

PILUGIT ATUGAKHANIK

- Mikiyunik/Piyukhaungituk Angunnahuakluni
- Pangnik Angunnahuaktautaktuk
- Havakhikhimayunik aah Hila Maliklugu
- Ataniktutauyukhanik Unikluktukhanik
- Havagiyumiklugit havakhautit munakgiyauyukhanik
- Kayaknaktunik Nirgitinik Munakgiyauyukhanik
- Atanitutauyukhanik katitikgiyukhanik
- itkuhiktuktunik Kauhimayauvaktunik katitikgilutik
- Ilihaknikmun ilidjuhikgiyakhainik angunahuaknikmun
- Ivaavit Ingniukvit/Nunaap Naunia Pirgumiyauyukhanik
- Angunahuaknaituk talvani Aullaktikgutikhanik
- Angunnahuaknaituk talvani nanminilingnik piutigiyainik angunahuaknikmun

panaarit, katimayit, nunalingni

- Narhugiyauvaktunik Havakhautit alauyunik alauyunik angunnahuaknikmun
- Aiyukiktuinikmun tingmitikut/kufiutilangi nik/gullavuk-nirgak kufiutilanginik; nirgak inuvaktunik nallautakgutauvaktunik ; nugyuinik kataakpaktunik ingmingnik ihiviuktauvaktunik naunaitkutakhanik aipagungukgangut havagiplugit
- Havagiyumiklugit Havakhautit Munakgiyauyukhanik

6. HAVAKHIKHAINAKTUKHANIK

1) Talvuna havakhikhimayunik aullaktikgutikhanik mikharut establishment KHTO atsaniktuktauyumik havakhikhainaktukhanik atugakhanik kauhimayauyunik controlled enforcement system regarding BNECH anguuyauvaktunik nallautakgutauvaktunik imalu avaatingnut nungumanginiklu, ataniktutauyukhaniklu uniklutikhanik, uvunalu angunnahuaknikmun havagivakgainiklu angunnahuaknikmun nallautakgutauvaktunik uvunalu avaatingnut nungumanginiklu, havakhautikhanik uniklutikhanik, imalu angunnahuaknikmun havagivakgainiklu.

Tahapkuat KHTO ihumadjutigiyainik mikharut pitiakgutaulutik tahapkunani nunalingni, uvunalu akhukgutaulutik havakhiktauyukhanik malikgakhanik mikharut uktutikhanik ICCMP. Tahapkununa havakhikhimayunik kilamiuktauvalakgamik tahapkunani NWMB pidjutiyukhanik havakhiktukhanik kauhimavaktunik havakhiktauyukhanik malikgakhanik ilauyukhanik uktutikhaniklu, kihimi pilugit ayuuknaitunik iniktauyukhanik. Havakhiktauyukhanik malikgakhanik ilauyunik nungumalingnik hivunikgiyakhainik uktutikhanik atadjutiyunik uvuna: BNECH angunnahuaknikmun uktutikhanik, pinnaktukhanik talvuna KHTO tutkikhakhimayunik havakhikhimayunik nakukgutauvaktunik; KHTO akiliktuktauhimayunik alauyunit aaituktuivaktunik nunalingni angunnahuaktukhaniklu, uvunalu KHTO panaarinik tahapkuatlu kavamat havakvit nalikmiklu ungutinik uvunalu nunami hulilukakgutauvaktunik. Tamakpaliayunik atukgungnaikpaliayunik tahapkuninga atuktauvaktunik pidjutauyuk ayuuknaktuk mikharut angunnahuaknikmun uktutikhanik imalu pikgutivaktunik havagiyauvaktukhanik inungni kanukgilidjutivaktuniklu malikgakhanik hugiyuitainiklu, imalu ilauyunik kangunaktunik ilauyunik ihuiyakhugit nakungiyakhugit nunalingni havakhiktauyukhanik malikgakhanik uvunalu ipiknakpiaktunik kanukgiliuktukhaungituniklu.

Kanuk pihimagamik tahapkuat KHTO havakhikhimayunik malikgakhanik, naliak niplikhimayunik malikgakhaniklunin taimani pihimayut pitiaktauvakhimayut. Pivaktunik kauhimayauvaktunik malikgakhanik havakhikhimayunik pidjutitkiyumikpaktunik nunalingni nakukiyauvaktuniklu tuhaayauyukhanik pivaktutlu pitiaknakmut malikgakhanik havakhikhimayunik angiutauvaktunik taakuyauvaktunik alauyunik pidjutiplutik, naliak atadjutiggingittunik tahapkunani nunalingni.

Aah ilidjuhikgiyaatut, tahapkuat KHTO ungiutauvaktunik naunaitkutalingnik malikgaliukhimayunik ammigaiktunik uukiunik mikharut piyukhaungitunik ilukpallilutik talvani Kugakmi Coppermine River tunungani Kugluk, pihimangmataah ihumadjutigiplugit ammigaikpalakgumik iglukpaknik pillakihungungmataah kuugluktitaihungungmataah huraanik nirgitinik, ilauyunik tuktunik atukpaganik nunatuutukanginiklu. Una ilaagiyanik kuugak havakhikhimayunik ilauhimayunik nunataaknikmun aviktuinikmun pilingnik 'Piyukhaungitut Hamani Nunami Angunnahuaknaituk avaatingnut nungumanginiklu' naunaitkutalingnik uvani kulani. Pihimaitkulluakhutik tahapkuat KHTO pikangitunik malikgalingnik angiutauvaktunik havakhiktauyukhanik nakukgutauyukhanik (pihimangmataah nunaginginaamitku), tahapkuat nunalingni, pivaktunik ungiyuniklu, pivakhimayut pittiakgutiplugit KHTO nakukgutauyukhanik imalu ikittunik naliak pitunik nuttanik iglukpaknik pidjutaungitunik talvani Kugluktumi Kugani. Uvunalu ilidjuhikgiyanik kanuktun tamangnik Inuit imalu Inungitunik talvani Kugluktukmi pitiakgutivaktunik tapkuninga KHTO.

Hivunikgiyakhainik, havakhiktauyunik aah ICCMP pidjutitkiyumiklugit ikayutauyukhanik pilutik hakugiktunik kauhimayauvaktunik ihumadjutigiyauvaktunik mikharut KHTO talvanilu Kugluktuk nunalingni inungni malikgakhainik atuklugit atugakhanik. Pigumi inuk uunikhiukgumagumi tapfumunga ataatiaminut anaanutiaminut, ungayukaminut, ungangminut, ataakminut naliak ilaanakmut huuk ihuinaktuk tamna, unalu pidjutitkiyumikpaktunik haakugituklu kauhimayauvakhuni ilidjuhikgiyainik taimaitunik imalu haakugitkiyauyuk kauhimayauvaktuk pilugit malikgakhanik mikharut pihimaitumik ihumalukhimaitumiklu talvuna kavamat malikgakhanik piluni ilitukgiyaulunilu ungahikluni imalu kauhimayaungitunik. Tahapkunani havakhikhimayunik nuttanik mikharut pitiktaknikmun angunnahuaknikmunlu.

GN pilutik havakhiktauyukhanik talvanitunik 340 talvanilu HTO pilutik havagiyauyukhanik ungnaluk/unguhuluk nallautakgutauvaktunik NQL pilutik havagiyauyukhanik pidjutinaitunik talvani nunani angunnahuaknikmun talvuna ivaayunik ingniukvit hila malikhugu.

7. HIVUMUT HIVUMUKNIKMUN

1) Naudjutauhimayuk mikharut havakhautikhanik imalu kinikhiayunik panaarikhanik piyukhanik naudjutiyukhanik aah kayaknaktunik nirgitinik munakgiyauyukhanik havakhautikhanik piyukhanik ikikliniakhimayukhanik nirginiakhimayunik nirgitinik aiyuknakhivaktunik mikharut BNECH.

Mikharut ukiunik ammigaitunik, tahapkuat KHTO pivakhimayut ukaakhugit tahapkuat GN-DOE mikharut takuyauvaktunik ammigaiktunik kayaknaktunik nirgitinik, kanukgitunik amaguknik akhaaknik talvanitunik avaatingnut nungumanginik BNECH. Pitkuhiktukpaktunik, Inuit pivaktunik angunnahuaknikmun amagunik imalu akhaknik talvanitunik nayuganik pivakgainik ubluktamat inuhinik nunamitutik aaihikpani. Tadja tamakpiavyaitlu Inungni talvanitunik aulakgungnaikpaliaplutiklu aihikpaanut aaihimainalikhutik inugiangnikmi, pidjutiyunik tadja ingnungni ikitkiyaullikhutik Inuit aaihikpanitunik aulaknit uvunalu pidjiyungnaikhutik uktutikhanik angunnahuaknikmun tahapkuninga nirgitiniklu.

Inuit pikaktut angunnahuaktungnikmun nalukhanginamik tamakpianginik, imalu pikaktunik pidjutivakhimayunik aah hivituyunik ingilgangnitanik kauhimayuutukauyunik angunnahuayuktunik amargukhiuyuktunik talvanitunik pikaktunik hakugiktunik aklingnaktunik angunnahuaknikmun nalukhanginamik tamainik, imalu pikaktunik ammigaitunik ingilgangnitanik amagukhiuyuktunik. Uvunalu atautimilu ammigaitunik ingilgangnitanik pitkuhiktukpaktunik kauhimayutukanik tahapkuninga amagunik akhaaknik pipkaivaktut Inungni mikharut pinnahuaktungniktunik havaktungniktunik atanguyaunikmunlu. Uvunalu ilidjuhikgiyainik, Inuit ilihimayunik humi ammigaitunik hitinik nayugainik amagunik talvanitunik BNECH avaatingnut nungumangnik talvani Nunavut.

Inuit auladjutivaktunik angunnahuaknikmun amagunik imalu akhaknik talvanitunik. Ayuuknakhunilu, tahapkununa akikhautainik pidjutauyunik akikhikpiakhutik akhuut mikharut ammigaitunik ukiunik pidjutiyunik akikhikpiakhutik adjikgiakhainiklu mikharut akittukpaktuk angunnahuaknikmun. Tahapkuat GN pidjutiyunik uktutikhanik havakhautit ikayutauyukhanik tahapkunani KHTO pilingnik kauhimayauyunik tahapkuninga kayaknaktunik nirgitinik ikiklingangnahuakhugit. Uvunalu ilauyunik, tahapkuat KIA pikaktunik manikhakhautikhanik ikayutauyukhanik tahapkunani KHTO pidjutauyukhanik havakhautikhanik mikharut kayaknaktunik nirgitinik munakgiyauyukhanik. Tahapkuat KHTO piyunik auladjutiyunik havakutigikhutik tahapkunani GN unalu KIA mikharut manikhakhautikhanik ikayutauyukhanik kayaknaktunik nirgitinik munakgiyauyukhanik pilugit havakhiktauyakhainiklu ilingaitunik imalu ilidjuhikgiyainiklu tahapkunani BNE Tuktunik Ammihuakyunik.

Taimaitulluamik havakhikhimayunik aah kayaknaktunik nirgitinik munakgiyauyukhanik pangnattauyunik. Kaipkaklugit atautimut havakatigiktukhanik aullaktikgutikhanik pangnattauyukhanik. Amagut ilihimayauyunik nikigivagainik tuktunik. Talvunalu ihumagiyaungingnaktunik aah amagut nikigivakgainik kanukmi 25 nik huna nallautakgutauvaktunik atauhikmik. Aah havakhktiakhimayunik iliktikhimayunik amagunik munakgiyauyukhanik

havakhautikhanik atukpaktunik pitkuhiktukpaktunik ayuiknikmun Inungni pidjutitaktunik aklikgungnakluni havaktungniktunik ikayutauyukhanik munakgiyauyukhanik tahapkunani BNECH, uvunalu pilutik havagilugit puiguktukhaungitunik Inungni angunnahuaknikmun ayuikgutauvaktunik. Pilutit talvuna ilidjuhikgiyanik havakhikhimayunik manikhakhautikhanik ungunnahuaknikmun havakhikhimaunik pidjutiyanginik pinnahuaktungnikhiyanginik Inungni angunnahuaknikmunlu pidjutiyukhanik aihikpani aullakhimalutik angunnahuaknikmunlu nikikhakhiuknikmun imalu angunnahuaknikmun amagunik pidjutiyunik, pivaktunik angunahuaknikmun nikikgiplugit, naunaitkutanit aah anguyauvaktunik 30 nik amagunik pidjutitaktunik nikigilugit 750 nik tuktunik nirgiyaungitunik tahapkunanga amagunit, uvunalu upalungaikhimayunik ikayutauyukhanik pilingnik kufiutilanginik nauvakgiakhainiklu. Pilutiklu naunakhitilutik igluvaktilugitlu hilaa ukiunganik malikhugu angunnahuaknikmun tahapkunanilu KHTO katimayit.

Tahapkununa pidjutauyuniklu pitkuhiktukpaktunik kauhimayauvaktunik akhanik kufiutilanginik ammigailikhutik aklingnakhiplutik talvani Wuatani Kitikmeot Nunatuttukanik talvani Nunavutmi imalu takungnakhivaliayuniklu tunungani talvanilu kivatani. Mikharutlu ilihimayauvaktuk piliktunik ammigailikhutik kufiutilangnik akhanik talvaniliktunik ivaavit ingniukvinik hivuagutlu ingiukgiaktukpaktunik avaatingnut nungumangnik tapkuninga BNECH. Akhanik ilihimayauvaktunik pidjutivaktunik tuktunik nirginiakhimayunik. Tahapkuat GN pihimayunik paaikhugu tamna nauanaitkutauvaktuk angunnahuaknikmun akhanut tahapkunani Inungni imalu naahugiyauvaktunik mikharut angunnahuaknikmun nallautakgutauvaktunik akhanik nayuganik kangnituanganik Kugluktuk piluni ungigliyumiklutik naunaitkutakhanik. Ammigaiyumiyunik nallautakgutauvaktunik angunnahuaknikmun piyakgaailiktunik takungnakhiyunik.

Tahapkuat Kitikmeot Inuit Katimayit piyakgaailiktunik havagiplugit kayaknaktunik nirgitinik munakgiyauyukhanik nallautakgutauvaktunik imalu pilutik ikayuklutik pilingnik pirgumiyauyunik manikhakhautikhanik mikharut kanukgitunik havakhautikhanik. Talvani Tunungani Nunatsiami (NWT), angunnahuakpaktunik pidjutitaktunik ammigaivyaktunik manikhakhautikhanik mikharut ikayutauyukhanik ungunahuaknikmun manikhakhautikhanik pidjutilutik tuukutiklugit amagunik pidjutitaktunik naapkitaktunik akikhautikhanik niritinik amiinik, pigumitku amaguk tuukutaugumi talvuna taatkikhiutinik taimailitkangut kanukgilidjutivaktunik ukiungani kanukgitunik amiinik nirgitinik ilingaitut talvuna. Tahapkuat Kavamat Nunavutmi, uvani nunauyaliukhimayunik nayugainik humi ilidjuhikginiakhugit ihumadjutigiyauvaktunik tahapkunani tuktunik, pidjutaulutik adjiliuktuutut tahapkunani GNWT havakhautikhanik tahapkunun ikayutauyukhanik angunnahuaknikmun manikhakhautikhanik pilugit havakhautikhanik amagunik. Tahapkununa ikayutauyukhanik angunnahuaknikmun manikhakhautikhanik pilugit pilugit tuniyauvaktukhanik tamanut ukiunganik pidjutiyanginik angunnahuaknikmun talvani hitini amaguanik takungnakgumiklu.

Tahapkuat KHTO havagiyauyukhanik mikharut aullaktikgutikhanik havakhautikhanik pidjutiyukhanik nuttiklugit amaguanik talvuna auyami, kanuk akhuut ikiklitiyumikpaktuk nikikhanik nikikariakaktunik tahapkunani amagunik. Uvunalu ilauyunik nakuyunik iliktikhimayunik ikayutauyukhanik angunnahuaknikmun manikhakhautikhanik atugakhanik angunnahuaknikmun amagunik pilutik atuklutik aklingnaktunik amagukhiuknikmun ayuikgutauvaktunik Inungni Inuit piyukhanik

munakgiyauyukhanik ammihuakyuknik. Hivunikgiyakhainik, alakangitunik alauyunik nunataknikmun katimayinik talvanitunik NWT, Inuit pikangitunik pitkuhiktukpaktunik ikumagiyauvaktunik ukuahiknik atadjutivaktunik tahapkunani angunnahuaknikmun naliak amikiplugit amagunik. Amakut puuhitait puuhitakgiyauvaktut atuktauvaktunik ublukpakgulluaktilugu atuktauvaktunik ukiukhiutinik anurani puuhitakgiplugit talvanitunik Kitikmeot Inuit. Tahapkununa ammigaikyumiyunik akhanik naunaitkutanik nungudjutilingnik ubluinik tugakhanik atuknianut pilutik ikayutaulutik tahapkunani. Tahapkuat KHTO pidjutiyunik niuviktauyukhanik mikharut advocating 10 nik akhaknik naunaitkutanik nungudjutilingnik ubluinik tugakhanik atuknianut tahapkunani angunnahuakpaktunik ungiyunik nirgitinik akilikhugu angunnahuaktauyukhanik nirgitinik. Naunaiyainikmun havakhikhimayunik ataniktutauyukhanik uvani auyami mikharut Kuupanuakpanik (kupanuakpak) imalu pivaktunik ikhinaktunik tuktunik nurganut.

Tahapkuat Havakvit Nunalikiyit havakhikhimayunik aah Amagunik Havakhautikhanik naunaitkutakhanik katitikgiyukhanik talvanitunik Kitikmeot Nunatutukanik. Talvuna havagiyauvaktukhanik uvuna havakhautikhanik piyukhanik katitikgiyukhanik katitiktauyukhanik naunaitkutakhanik tahapkuninga amagunik imalu ammigaikpaliayunik kufiutilanginik amagunik angunnahuaktauvaktunik.

2) Auladjutiplugit imalu nakuhitkiyauliktunik ilihaknikmun uvuna KHTO katimayit havagiyauyukhanik tuktunik, pitiaknikmun angunahuaknikmun havagiyauvaktuniklu, uvunalu avaatingnutlu nirgitinik angunnahuaktauyukhanik

Tahapkuat KHTO pidjutauyunik havakhikhimayunik tuktutigut ilihaktauyukhanik havakpauhinik talvani Kugluktukmi. Uvani havakpauhinik ilauyunik havagiyauyukhanik kanuk angunnahuaknikmun havagiyauvaktunik, uvunalu angunnahuaknikmun alakiniklu nirgitinik. Pikaktuklu ammigaitunik alauyunik nirgiyauvaktunik nirgitinik talvani Kugluktukmi nunani ilauyunik omingmaknik, tuktuvak, ikalukpik, natik, imalu tingmiakgiaktuklu pidjutigufuuktauhimayunik atiinik tainiakgitkaa. Una ilihaknikmun havakpauhinik auladjutauyunik pidjutauyuniklu ublukpakgulluaklunilu ayuikgutauvaktunik kauhimayutukanik tahapkunanga Inungni tahkununga inulguminik kingulinik imalu ilauyunik tukihigiarutinik ayuikgutauvaktuniklu pivaktuk atadjutauvaktunik tahapkunani ICCMP. Ilihaknikmun pilugit iniktiktaulutik malgunguklugit wuiwuanganik: ilihaktitqauyukhanik ilihaknikmun kanukgitunik uvani pangnatauyunik (i.e. angunnahuaknikmunlu avaatingnut nungumanginiklu imalu pitutiyukhaungitunik angunnahuakningmun talvani nunani ilitukgiyaulimaituniklu tutkumayunik imalu kauhimayaungitunik ubluliuktauyukhaniklu, tahapkununa naunaitkutakhanik mikharut ammihuakyuknik imalu pitiagutauvaktunik angunnahuaknikmun havagiyauvaktunik tuktunik (avaatainik pilugit atugakhanik, kanukgitunik taimaitunik hikuktuutinik)

3) Ungigliyumiktukhanik pinnahuaknikmun mikharut ungigliyumiktukhanik naamukgiyauvaktunik naunaitkutanik nirgitinut nungudjutilingnik ubluinik atugakhanik atuknianut tahapkuat KHTO kauhimayauyunik omingmaknik talvanitunik Kugluktuk angunnahuaknikmun nunani piyukhanik angnnahuaknikmun ayuuknakhivaktunik tahapkuninga

BNECH.

Aah nuttamik omingmaitigut munakgiyauyukhanik avaatingnut nungumanginiklu havakhikhimayunik tahapkunanga GN is MX-11. Tamna takungnaktuk talvani nunainakmi talvani kivatani Kugluktum Kuganik tavungakaluk kangiklugu Kingauk talvunga wuatani Ellice Rivermi. Tahapkuat KHTO havakhikpaktunik ukiunik titikgakhimayunik tahapkuninga omingmaknik kufiutilanginik mikharut ammigaitunik titikgakhimayunik tahapkuninga omingmaknik kufiutilanginik talvani hivugani kugluktumi pidjutiyunik ungigliyumiktunik imalu havakhikariakaktuniklu tahapkuat GN havagilugit pidjutifaaklugit-tutkikhaklugit tahapkuninga TAH. Uvani 2013 tahapkuat GN havagivakgainik naunaiyaknikmun naunaitkutakhanik talvanitunik wuatani 25% of MX-11 kanuklu kangnitkiyauyunik Kugluktuk. Tahapkuat GN nallautaktauhimayunik pikaktunik aah omingmaknik kufiutilanginik uvuna 6746 +/- 1851 talvanitunik kihimi wuatani talvani MX-11. (GN – DOE Kitikmeot Omingmanik Angunahuaknikmun Munakgiyauyukhanik Pangnatauyunik 207-18). Tamakpianganik Pidjutautaktunik Angunahuaknikmun tahapkuninga MX-11 is 225.

Uvani Apitilikvia 2017, tahapkuat KRWB tuniyauvaktunik naunaitkutanik nirgitinut nungudjutilingnik ubluinik atugakhanik atuknianut maliklugit: Ikaluktutiami HTO – 25; Omingmaktok HTO – 40; Kingauk HTO – 40 talvanilu Kugluktuk HTO – 120.

Uvani Tatiaknakhivia 2017, tahapkuat KHTO tuniyauvaktunik tahapkuninga 120 naunaitkutanik nirgitinut nungudjutilingnik ubluinik atugakhanik atuknianut maliklugit: 60 – aviktuinikmun nunataknikmun; 10 – pikangitunik-naunaitkutanik aviktuinikmun nunataknikmun; 30 – nikitianik nikinik havakhautit; 20 – akiliktuktauhimayunik ungunnahuaknikmun ungunnahuakpaktunik ingnit.

9. NAUNAITKUTAUYUNIK

Tahapkuat Nunavut Angiutauvaktunik, Angiutauvaktunik Akungainik talvanitunik Inuit Nunavutmi Nunatutukanginik Nunani unalu Tamna Atanikpak hivuliukti Queenguyuk talvani talikpiani uma Kanatami imalu tutkikhakhikhimayunik.

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Kugluktuk Angoniaktit Katimayit

"Anguniaktit imalu Nanirgiaktukpaktunik Katimayit pihimayut havakhikhimayunik mikharut kauhimayauyukhanik pirgumiyauyukhanik angutinik nirgitiniklu talvanitunik Nunavutmi Nunalingnilu, imalu havagilugit mikharut inuhikgingnaktumik mikharutlu pitkuhiktukpaktunik ingilgangnitaniklu tahapkunani nunalingni pilutik malikgakhaniklu, havagiyaukyukhanik munakgiyauyukhanik, aullaktikgutikhanik uvunalu havagiyauyukhanik nalikmiklu uiwuanik talvanitunik Nunavutmi Angiutauvaktunik mikharut ikayutauyukhanik tamakpiangnik Inungni hivunikgiktukhanik ukuingnik kakugu." Malikgakhanik uvunalu Hivumuktukhanik talvani HTOs talvanitunik Nunavutmi: Naunaitkutanik Hivuliuyunik Ataniktutauyukhanik Kavamainik, Nunavut Inuit Angoniaktit Havaktiat Tutkikhaiyi Titikgaktiat.

Tadja Katimakatigikpaktunik talvani Ikhivautalingnik Ilanginik Katimayit:

Larry Adjun, Ikhivautalik Hivuliuktuiyi

Kevin Klengenberg, Tuklia-Ikhivautalik Hivuliuktuiyi

Bobby Anavilok, Titikgaktiat Tutkikhaiyi-Manilikiyi

Stanley Carpenter, Ikhivautalik Ilangut Katimayit

Kevin Ongahak, Ikhivautalik Ilangut Katimayit

Jayko Palongayak, Ikhivautalik Ilangut Katimayit

Myles Pedersen, Ikhivautalik Ilangut Katimayit

Amanda Dumond, / Atanik Havaktiat Tutkikhaiyi Titikgaktiat

LARRY ADJUN
CHAIR, KUGLUKTUK HTO

ጋᓂᖅᑕᐃᓄᑦ ᐅძᓇᖅᑐᓄᑦ

ᓇᓐᖏᓗᒥ ᐅᒪᔨᓪᓚᕕᓴᕐᕕᐊᖅᑯᑦ ᑲᑎᒪᔨᖅᑎᓇᓂ (ᐅᒪᔨᓪᓚᕕᓴᕐᕕᐊᖅᑯᑦ)

ᑲᑎᒪᓂᕆᓐᓇᓱᓚᖏᖅᓂᒃ ᑲᑎᒪᓂᖅᑎᓂ 002-2019

ᐅძᓂᖅᑐ ᐱᔨᑎᖏᖅᓂᖓᓕᒃ

ᖃᐅᔨᓴᖏᑎᐊᕈᑕᐅᓂᖓᓕᒃ: ☐ ᐃᓱᒪᓕᐅᕈᑕᐅᓚᕆᐊᖃᖅᓂᖓᓕᒃ: ☒

ᐱᔨᑎᖏᖃᖅᑐᒃᓴᖅ: *ᐊᖏᕈᑎᐊᖅᑕᐅᓂᖅᓯᓂ ᑕᒪᓐᓇᒃ ᑲᑎᓪᓗᓂ ᐅᒪᔨᖅᑕᐅᔨᓐᓇᓂᖅᓯᓂ ᖃᓄᖅᑕᒍᖓᓐᓂ ጋᔭᐅᖅᓯᓂ 2019*

ᖃᓄᖅ ᐱᓕᕆᓚᐊᒃᑎᓂᖅᓯᒪᓂ ᑕᒪᓐᓇ:

ᑕᐃᒪᖓᑦ 2014-ᖓᓯᓚᐅᖅᑐᑎ, ᓯᑲᓕᓕᒥ ᑕᒪᐃᓐᓂᒃ ᐱᔭᐅᔨᓐᓇᓂᕆᖅᑐᑎᑕᐅᓚᐅᖅᑐᑦ ᐃᒪᓐᓇᑎᑎ 0-ᓂᒃ (ᐱᑕᖅᖁᕕᑐᓂᒃ) ᑐᔭᑐᓂᒃ, ᑭᖓᓂᐊᒍᓪᓗ ᑕᒪᐃᓐᓂᒃ ᑲᑎᓪᓗᕕ ᐱᔭᐅᔨᓐᓇᖅᓯᓚᐅᖅᑐᑦ 250-ᐅᓪᓗᑎᒃ ᐊᖓᔪᓪᓗᐃᑦ ᑐᔭᑐᐃᑦ (<ᖁᓕᑦ) ᖃᑭᖅᑕᒋᔪᑦ ᑐᔭᑐᕕᖁᓕᒃ. ᖃᑭᖅᑕᒥ ᐊᖓᔪᓄᑎᑕᒪᓕᓂ ᑲᑎᒪᔨᐃᓕ ᐊᒪᓚᓪᓗ ᐊᖓᔪᓄᑦᑎᓂ ᑕᒪᐃᓐᓇᓂᕆᖅᑐᓂᖅ ᑕᒪᒃᑯᐊᓂ ᐊᐃᕐᓂᖅᑐᓂᖅ ᑕᒪᐃᓐᓇᓂᕆᖅᑐᓂᖅ ᑐᔭᑐᐃᑦ ᐱᑕᖅᖁᕕᑐᓂᖅᓯᒪᓂ ᑕᒪᐃᓐᓇᓂᕆᖅᑐᓂᖅ ᑐᔭᑐᐃᑦ, ᐊᖓᔪᓄᑦᑎᓂᖅᓯᒪᓂ ᑕᒪᐃᓐᓇᓂᕆᖅᑐᓂᖅ ᐊᖓᔪᓄᑦᑎᓂᖅᓯᒪᓂ ᑐᔭᑐᐃᑦ ᐱᑕᖅᖁᕕᑐᓂᖅᓯᒪᓂ ᑐᔭᑐᐃᑦ,

ᐊᖏᑭᐊᖅᑕᐅᔾᑎᐊᑲᓂᓂᕐᓂᓄᑦ ᑐᑦᑐᐅᔾᓇᖅᓂᕐᑎᑦ ᖃᑭᖅᑕᒪᔭᖅ ᑐᑦᑐᐊᓂᖅ.
ᑲᔪᓯᑎᑕᓪᓚᐅᖅᓴᑦ ᐱᖁᔨᔾᔪᑎᑭᒃ ᐊᒥᓯᕆᐊᖅᑕᐅᖁᔨᓂᖅᓴᑦ ᑕᒪᐃᓂᓂ ᐱᓴᐅᔪᓇᖅᓂᕐᓂᓂᓂ
ᑲᑎᒪᔨᑦ ᑐᑦᑐᐅᔪᓇᖅᑐᑦ, ᐊᓘᒪᓗ ᐊᖏᕿᖁᔨᓗᑎᒃ ᐃᓕᓴᕐᓂᓂ ᐊᖁᓪᓗᓂᓂ ᑐᑦᑐᑎᓇᐅᑎᒃᓴᖏᓂᓂ.
2018-ᒥ ᐊᖁᕐᔪᒥ, QWB-ᓂᑦ ᐅᒪᔪᓕᓚᔨᓪᓕᒻᒥᑦ ᑲᑎᒪᔨᑦ ᑐᓴᕐᓯᓚᐅᖅᑐᑦ ᑕᐃᓂᓯᓚᐅᖅᑐᑦ
ᐅᒪᔪᓕᓚᔨᖁᔨᓚᓂᓂᓂ

ᓄᓇᕗᒻᒥ ᐅᒪᔨᓕᓚᓵᕐᔨᐊᕆᔭᕐ ᐃᓕᓴᖅᓯᓴᐊᕐᑲᖅᑐᑦ ᑕᕆᔭᕐᑐᐃᓐᓇᕐᓂ ᐊᑐᕐᔪᐊᑐᑦ QWB-ᔭᑦ ᐊᖏᓚᓯᕆᓕᒫᑦ ᑲᑎᒪᓵᖏᓐᓄᑦ ᐊᐅᓪᓗᕆᓂᓕᕆᓂᖅ ᐊᖏᕐᕕᐊᔪᖏᓐᓇᒪᑦ ᑐᓴᓯᖅᑕᐅᓂᖏᓐᓂ ᐱᐅᓯᐊᕈᑕᐅᔨᓚᓐᓂᑦ ᐊᖏᓴᕐᓗᐃᓚᕐᓂᓕᖅᑐᒃ ᑐᓴᑐᓯᐅᔭᓯᐅᓯᓚᓂᖏᓚᓄᑦ ᓴᐳᖅᑕᒪᔪᒻᒪ. ᐃᓕᓴᖅᖏᑕᒪᓚᓚᓂ ᓴᓇᓄᖅ ᑐᓴᑕᓚᐊᖅᓴᓂᖏᓚᓂᓂᑐᒃ ᒪᓕᖅᑕᐅᓗᑎᒃ ᐃᓇᐃᑦ ᓴᓇᐅᓪᓚᖅᑐᖅᑕᑐᒃ.

ᑐᐳᔭᓂᐊᕈᑕᐅᔭᒪᓂᖏᓐᓂᑦ:

ᒎᓴᔨᐊ ᐊᖏᓚᓯᕆᓕᒫᑦ ᓴᑎᒪᔨᖏᓐᓂᑦ, ᐊᖏᓚᓯᕆᑲᓪᓗ ᑲᑐᔨᓴᓇᑎᑲᖏᓐᓂᑦ ᓇᓇᓐᓪᓗ ᐃᓇᐃᑦ ᓴᕐᐳᑎᑦᓴᖅᑕᕆᔭᒪᔪᑦ ᐃᔭᒪᓗᑎᒋᔨᖏᓐᓂᓂᒃ ᐊᑎᔭᐃᓴᔭᕆᓂᒃ ᐊᑦᓴᔪᓴᓕᕆᕕᐅᓂᖅᑕᖏᓐᓄᓂᒃᑐᓂ, ᑕᐃᓇᑲᓂᑐᐃᓂᓇᐅᔨᖏᑦᑕᓴᖅ ᐊᖏᓚᓯᕆᓕᒫᑦ ᓇᓇᓕᓂᑦ QWB-ᔭᑦ ᐊᑦᓴᔪᑕᒧᖅ ᓇᓇᒪᑦᔨᐊᕆᓚᕈᑕᖅᑕᖏᓐᓄᓂᒃᑐᓂ, ᓇᙵᐱᓚᐅᓕᐅᖅᑐᒥ 2017-ᒥ. ᐊᖏᓚᓯᕆᑲᑦ ᓴᑐᔨᖅᑎᑲᖏᓐᓂᑦ QWB-ᔭᑦᓗ ᐊᖏᓚᓯᕆᓕᒫᑦ ᓇᓇᒪᓂᖏᓐᓂᑦ ᐃᓕᐅᖅᑕᑦᑕᖅᓯᒪᖏᓕᒃᑐᑦ ᐋᖅᑭᒋᐊᖅᑕᐅᓯᒪᓂᖏᓐᓄᓂᒃ ᓇᓇᕐᓂᓂᒃ ᑐᐳᔭᓂᐊᕐᕕᐊᓇᕐᓂᒃᑐᒥ, ᒪᓕᓂᓂᔪᑦ ᓇᓇᕐᓂᑦ ᑲᙵᒪᐃᑦ ᐊᙵᑎᓕᓚᓵᔪᓂᓂᒃᑐᒥ, ᐋᖅᑭᒋᐊᖅᑕᐅᓕᐅᖅᓚᑕ ᑕᒪᐅᓂᒃ ᓴᓇᐋᓚ 2019-ᔪᑦ.

ᐊᑐᕐᔨᔪᖅᓱᓂ:

ᐊᖏᓚᓯᕆᓕᒫᑦ ᓇᓇᒪᓂᖏᓐᓂᑦ ᐊᑐᕐᔨᔭᐅᑦ ᑕᐃᔪᔨᐊ ᑕᒪᕐᒥᔪ ᓇᑎᑦᓂᑎᑦ ᐱᓴᐅᔭᓂᓗᕆᓂᖏᓐᓂᑦ ᑐᔪᑐᐃᑦ ᐊᒋᐅᓂᖅᑐᒥ 2019-ᒥ ᐋᖅᑭᖅᑕᐅᓴᕐᔨᓚᑎᑦ ᐃᒪᓐᓇ, 278-ᓂᕐᓂᓚᑎᑦ ᐊᖅᑎᔪᓕᒫᖅ, ᒎᓴᔨᐊᓪᓚᑎᑦ 45-ᖃᓪᓚᑎᔪ ᐊᖅᓇᑦᔨᑦᓚᑎᔪ. ᑕᐃᔪᔨᐊ 45-ᖃᓪᔭᑦ ᐊᖅᓇᑦᓚᖅᓂᔪ ᑐᔪᑐᑐᓇᐅᑲᓂᓂᕈᑕᖅ ᐊᑐᖅᑕᐅᔭᓯᐅᓗᐊᖅᑲᖅᓯᒪᔪ ᑐᔪᑎᑦᓚᓚᖅᑐᓗᓂᖅ ᐊᖏᓴᕐᓗᓚᖅᓂᖅ

ᑎᑎᕆᖅᑕᖏᓐᓂ ᐅᒪ:

ᒪᐃᔨᑦ ᐅᓴᓴᓐ, ᓴᑭᖅᑕᔨᓐᒥ ᐊᖏᓚᓯᕆᑲᓪᓗ ᑲᑎᒪᓵᖏᓐᓂᑦ, 613-407-1197

ᐅᑦᓗᖅ:

ᒪᐃ 15, 2019

Kitikmeot Regional Wildlife Board
ᖃᑎᒃᒪᓕᐅᑦ ᓄᓇᓕᖕᒥᓄᑦ ᐅᒪᔨᓕᕆᔨᖅᑕᐅᓯᒪᔪᑦ ᑲᑎᒪᔨᖏᑦ
Kitikmeot Nunaliit Avikhimaniani Angutikhaligiyit Katimayit

Kitikmeot
ᖃᑎᒃᒪᓕᐅᑦ
Kitikmeot
Cambridge Bay
ᐊᒻᒪᓗᒋᑦᓴᖅ
Kugaaruk
ᑰᒑᕐᓗᒃ
Kugluktuk
ᑰᒡᓗᒃᑑᒃ
Gjoa Haven
ᒫᓐᓇᐅᔭᖅ
Taloyoak
ᑕᓪᓚᐅᔭᒃ

May 22, 2019

Nunavut Wildlife Management Board
P.O. Box 1379
Iqaluit, NU
Attn: Denis Ndeloh, Terrestrial and Avian Species Biologist

Dear Mr. Ndeloh,

Re: Kitikmeot Regional Wildlife Board Support for EHTA Request for Additional Grizzly Bear Sport Hunts Tags

In response to our conference call on May 21st, 2019, KRWB attempted to followed up with Mr. Drikus Gissing to see if the Government of Nunavut had any concerns regarding the EHTA request for five (5) additional grizzly bear tags. Drikus was out on duty travel at the time and we did not get his response. Further, he was to return from duty travel Monday, May 27th, 2019. After the NWMB deadline to receive briefing notes for their June 2019 meeting.

I'd also like to mention the rationale for EHTA's request for additional tags are needed. EHTA has reduced caribou sport hunts on the Dolphin and Union herd after conservation concerns were raised and that EHTA will stop the sport hunts on the herd after the fall of 2019. This will take away income for our local hunters.

We hereby submit this request for five (5) additional grizzly bear tags for the Kitikmeot region and ask NWMB that GN's comments or concerns be included once they are available. Once again, Mr. Gissing is out on duty travel at the time of our request for their input on this request.

If you have any questions, please do not hesitate to contact me or our staff, Ema Qaggutaq at (867) 769-1007 or via email at krwb@niws.ca

Yours truly,

James Panioyak,
Chairperson

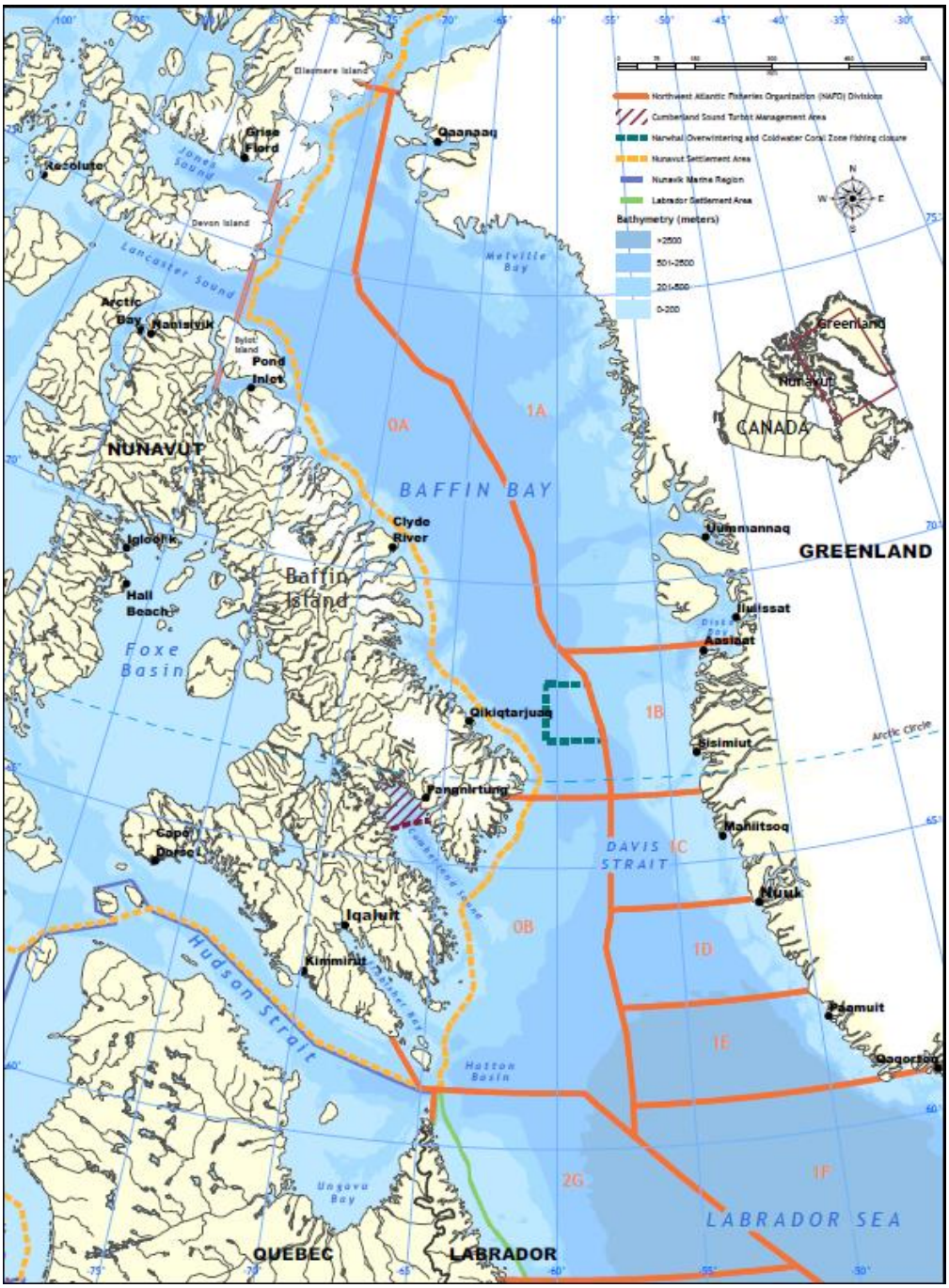
cc. Drikus Gissing, GN
Bert Dean, NTI Wildlife
Robert Greenley, Chair of EHTA

ᑐᓂᓴᐅᔨᓪᓕ ᐅᑭᖏᑐ
ᓄᓇᑭᓗᒥ ᐅᒪᓲᓕᖅᑕᓚᐅᕐᓂᖅ ᑲᑎᒪᔨᖏᑦ

ᑐᑭᓯᕋᑭᐱᖅᓯᒪᓂᖅ: ᐃᓕᓴᕐᓂᖅ

ᐱᓱᕈᖏᖅ: ᐊᖏᑦᓴᑕᐅᓂᖏᖅ 2019 ᐃᓕᓴᐅᑎᓯᒪᔨᓪᓚᑦ ᐃᓴᖃᑐᓕᓂᕐᓗᒍ ᐊᐅᓕᕆᓂᕐᓗᒍ ᐸᖏᓕᐅᔨᐊᑯᑦᑐᑦ ᖃᖅᓕᕆᓂᖏᓐᓄᑦ (ᓇᒡᓘᓇᐃᕐ)

ᓄᓇᖕᒪᐊᓘᖅ: NAFO ᐊᐅᓚᑎᕃᐅᔨᓪᓚᑦ



ↃPᓂᑕᐊPᑎᖏᑦ:

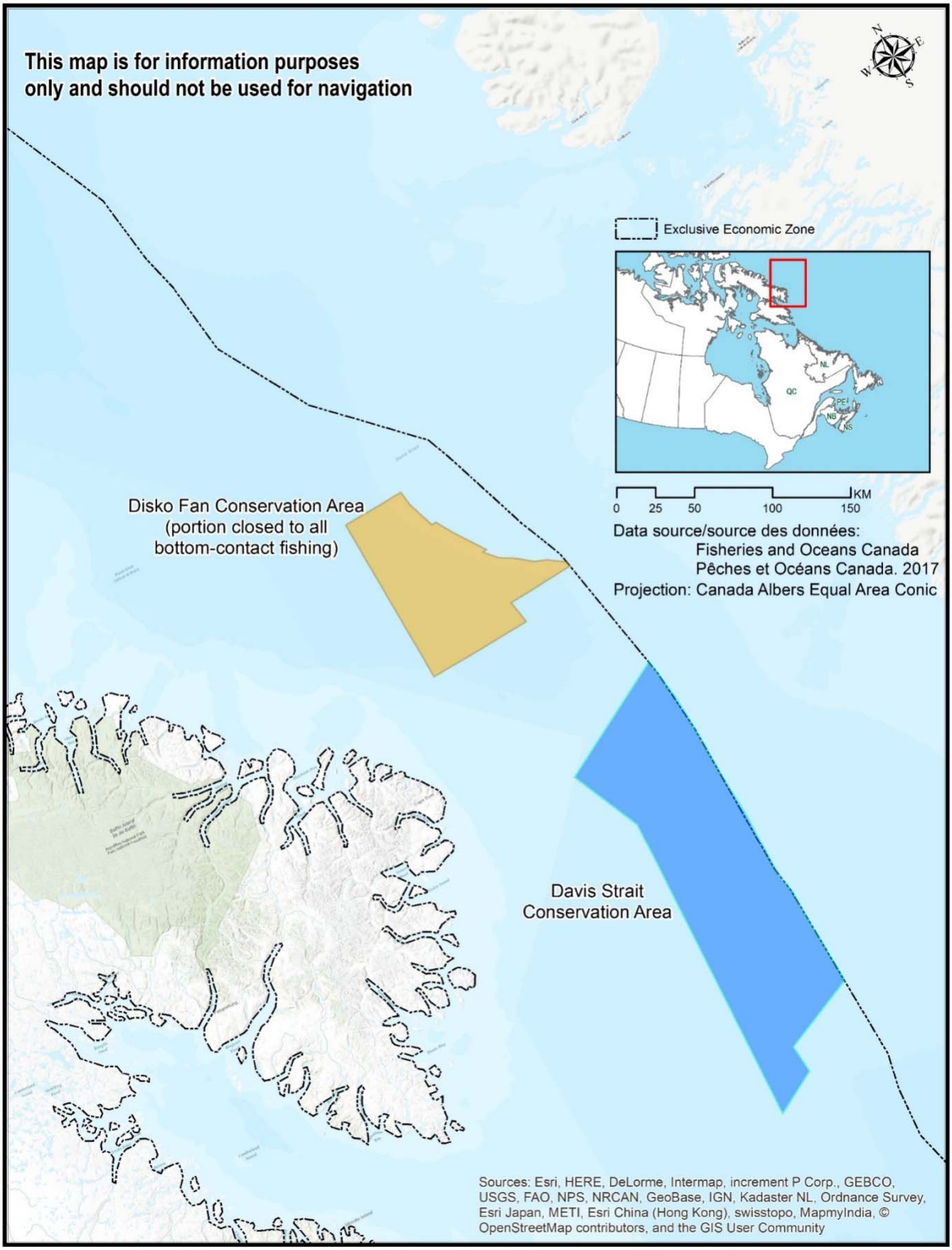
2013 ᐃᒋᓕᐅᑎᓂᒪᓕᕐᓂ ᐃᖃᓗᓕᕆᓂᖅᓯᒪᓂ ᐊᐅᓕᕆᑎᓂᖅᓯᒪᓂ <ᖅᓇᐅᑦ (IFMP) ᐊᑯᑭᑐᑦ ᖃᓕᕋᓕᖏᓐᓂᒃᓂ (ᓇᒡᖅᓇᐃᑦ) ᐅᖅᓂ ᐅᐊᓐᓇᖏᑕ ᐱᖏᓐᓇᖏᓂ ᐊᑦᓯᓐᑎᒃ ᐃᖃᓗᓕᕆᓂᖅᓯᒪᓂ ᑲᑐᔾᔨᖃᑎᒋᑦ (NAFO) ᓇᒥᓐᓂᖏ 0 ᐅᑦᓗᒥᒍᖏᓕᕐᕕᑎᑕᐅᓂᒪᓕᕐᓂᖅ ᐊᓯᒪ ᐋᖅᑭᑎᐊᖅᑕᐅᓂᒪᓕᕐᓂᖅ ᐊᑎᖃᔪᖏᓚᓂ 2019. IFMP ᐱᑕᖅᖃᑎᑎᑎᐊᑦ ᓇᓗᓇᖏᑦᑐᑐᓂ ᐊᓯᒪ ᑕᓯᖅᓂᒪᓕᕐᑐᓂ ᓇᓗᓇᐃᔪᖅᓂᒪᓕᖏᓐᓄᓐᓂ ᐊᑯᑭᑐᑦ ᖃᓕᕋᓕᖏᓐᓄᓐᓂ ᐃᖃᓗᓕᕆᓂᖅᓯᒪᓂ ᖃᓄᐃᓕᔨᖅᑭᑎᑎᓂᖏᓐᓄᓐᓂ, ᐱᖅᓂᐅᑎᑦᓗᑎ ᐊᑐᖅᑕᐅᖅᑭᑕᖅᓂᒪᓕᕐᓂᖅᐃᑦ, ᓇᒡᖅᓂᖏᓐ, ᐃᖃᓗᓕᕕᐊᑭᑎᑦ, ᐃᓕᐅᖅᑕᐅᓕᕐᐊᑦ, ᐊᐅᓕᕆᑎᓂᒪᓕᕐᓂᑦ ᐃᓂᒻᓗᑕᐅᓕᕐᐊᑦ, ᐃᓂᒪᓕᐅᖅᑮᑦᓕᐊᑎᓂᒪᓕᕐᓂᑦ ᐊᓯᒪ ᖃᐅᔨᓴᖅᑕᐅᑎᖏᓐᓄᓐᓂ ᐊᑯᑭᑐᑦ ᖃᓕᕋᓕᖏᓐᓄᓐᓂ (*Reinhardtius hippoglossoides*). IFMP ᓇᓗᓇᐃᔭᐃᔪᖅ ᐊᖏᑎᖅᑕᐅᓂᒪᓕᕐᓂᖅᓱᓪᓗᖅᑐᑎᓂ ᐊᐅᓕᕆᑎᓂᒪᓕᕐᓂᑦ ᐊᑐᖅᑕᐅᐊᓐᓂ, ᒧᑦᓇᐅᐊᓐᓂ, ᐊᓕᐅᖅᑕᐅᐊᓐᓂ, ᐊᐅᓕᕆᑎᓂᒪᓕᕐᓂᑦ ᐃᓂᒪᓗᑕᐅᐊᓐᓂ, ᒪᓕᑐᐃᑦ ᐊᓯᒪ ᐱᑦᓕᓚᐅᐊᓐᓂ ᐊᑕᖏᑎᖅᑐᒍ ᐃᖃᓗᓕᕆᑎᓂᒪᓕᕐᑎᑦ NAFO ᐃᓂᖏᑲ 0 (ᐊᐃᕐᑐᖅᓂᒪᓕᕐᓂ 0A ᐊᓯᒪ 0B). IFMP ᐱᖃᓂᐅᖅᓴᖏᓕᑐᖅ ᑭᓇᓂᖅᑐᖅ ᐃᑭᖏᑕᓂ ᖃᓕᕋᓕᖏᓐᓄᓐᓂ ᐊᐅᓕᕆᑎᓂᒪᓕᕐᓂᑦ ᐃᖃᓗᓕᕆᓂᖅᓯᒪᓂ. ᐃᓕᓴᐃᑲᖅᑐᖅ IFMP ᓴᖅᑭᑎᑕᐅᓂᐊᖅᑐᖅ ᑕᕐᓂᒧᖏᑕ ᐃᖃᓗᓕᕆᓂᖅᐅᐊᒧᑦ.

ↃPᓂᑕᐊPᑐᑦ ᐱᕐᓯᑎᖅᖅᑐᑦ ᓄᓇᕗᓪᓕᒧᑦ ᐃᓕᓐᓄᓐᓂ (NSA) ᐊᓯᒪ ᐊᑐↃᐊᓂᓕᑦ ᐃᒪᐃᑦ ᐊᖏᑦᖃᑕᐅᑦᓴᖅᓂᒪᓕᕐᓂ ᐃᓂᒪᓕᐅᖅᑮᑦᓕᐊᑎᓂᒪᓕᕐᓂᖅᓴᑦ ᑲᑎᒪᔨᓂᑦ ᐊᓯᒪ ᒥᓂᑕᒍᑦ. ᐱᓚᒪᐅᐊᖅᓂ ᓇᓗᓇᐃᔭᓗᒍ ᑕᒃᓴᐃᐊ IFMP ᐃᓗᓕᖅᖃᓗᐊᖅᑎᓪᓗᒍ ᐃᓕᓐᓂᖅᓂᑦ ᓄᓕᑦᓂ ᑐᑭᓯᐊᓂᑦ, ᐱᖅᓂᐅᔨᔪᖏᑦᑐᖅ ᑐᓴᓂᖏᓐᓇᐊᓂᑦ ᐊᓂᔪᖅᑕᐅᐊᓚᐊᓂᑦ ᐊᐅᓕᕆᑎᔨᐅᐊᓂᑦ ᐊᑦᑐᐃᓂᖅᑲᖅᑐᑦ ᐃᓕᐅᖅᑲᖅᑕᐅᓂᖏᓐᓄᓐᓂᓕᒃ ᑎᒍᒥᐊᖅᑎᒍᑦ, ᐃᖃᓗᓕᕆᑎᓂᐊᖅᑎ ᐅᕐᖓᔪᓂᑦ ᑎᒍᒥᐊᖅᑎᐅᖅᑕᐅᓚᐊᖅ ᐃᖃᓗᓕᕆᓂᖅᓯᒥᒥᑦ ᐅᕐᖓᑎ ᐅᕐᖓᔪᓂᑦ ᐊᑦᑐᐃᓐᓇᓂᓕᑦ NSA.A ᓇᓗᓇᐃᔭᖅᓂᒪᓂᑦ ᐊᓂᔪᖅᑐᑦ ᑎᑎᕋᓚᐊᖓᓱᖅᑕᐅᓂᒪᓕᕐᓂᑦ IFMP ᐱᖅᓂᐅᑎᓂᒪᓕᕐᓂ ᐅᖅᓂ ᐊᓗᑐᖅᓂᒪᓕᕐᓂᖅ 2.

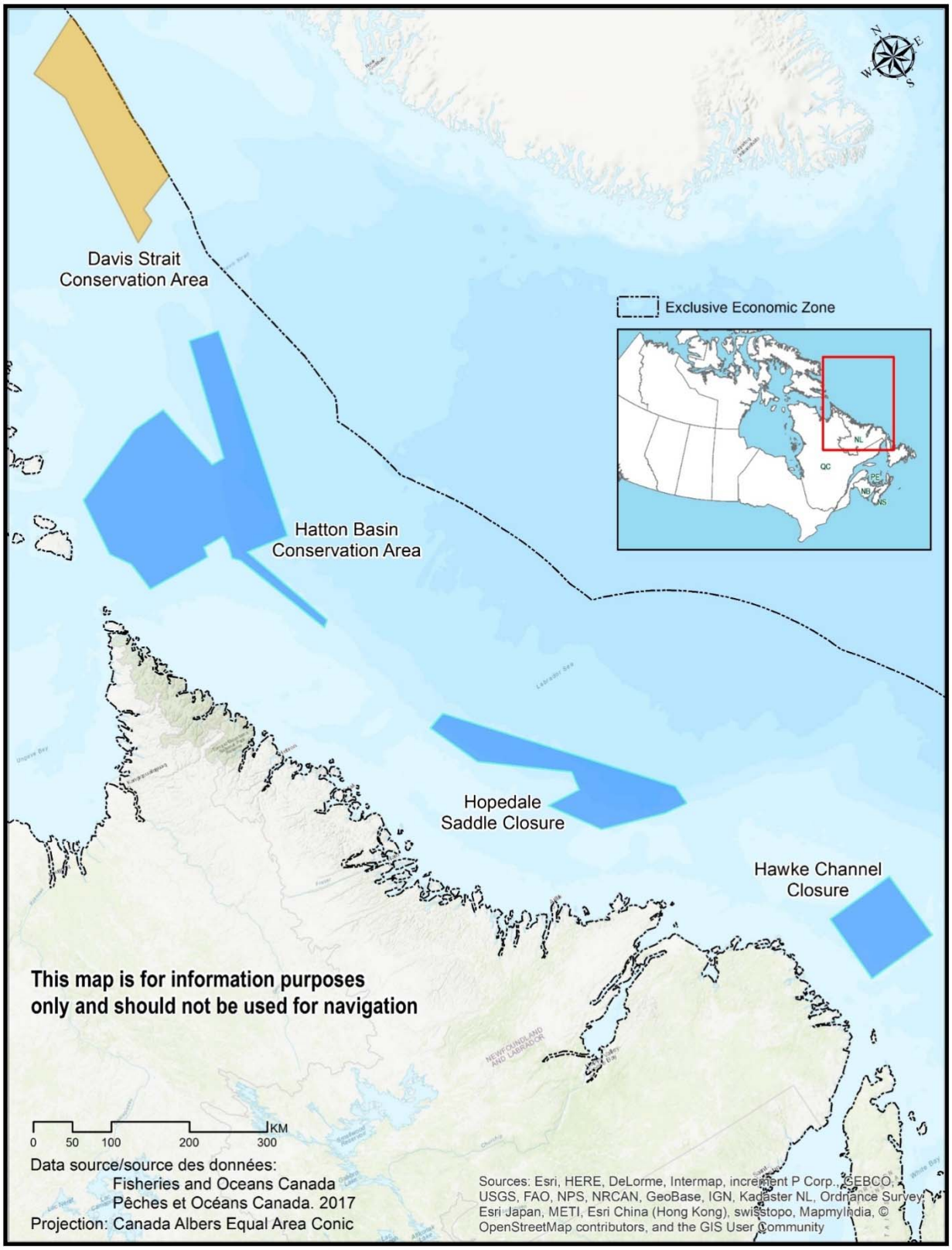
ᓇᓗᓇᐃᖅᑕᐅᓂᖏᓐᓄᓐ, ᓄᑦᓱᖅ ᐃᓕᓐᔪᑦᑲᐅᓂᖅᓯᒪᓕᕐᓂᖅ 2019 IFMP ᓇᓗᓇᐃᔭᖅᓂᒪᓂᑦ ᓄᑦᓱᑦ ᐃᒪᖏᒥ ᐊᑦᑕᓇᖏᑦᑐᑦᑐᑦ ᓴᖅᑭᖅᑎᑕᐅᓕᕐᓂ ᐃᓗᓐᓂᐊᓂ ᓇᒡᖅᓂᖏᓐ 0. ᐊᑦᑕᓇᖏᑦᑐᑦᑐᑦ ᐊᑐᓕᖅᑎᑕᐅᓕᐅᖅᑐᑦ 2018 ᑐᓴᖅᑎᑦᑎᓂᒪᓕᖅᑐᓕᐅᖅᑎᓪᓗᑎᑦ ᐊᓯᒪ ᐱᑕᓚᖅᑎᖅᖅᑐᑎᑦ ᐃᓗᓐᓂᐊᓂᑦ ᐊᑯᑭᑐᑦ ᐱᑕᖅᖃᑎᑎᑎᐊᑦ ᐃᓗᓐᓂᐊᓂᑦ, ᐊᓯᒪ ᐱᑕᖅᖃᑎᑎᑕᐅᓕᐅᖅᑐᑦ ᑲᑎᒪᔨᓄᑦ ᓱᑦᓂᒥᑦ ᐊᓯᒪ ᓕᑎᐱᓚ 2017.

ᓄᓇᖓᐊᑦ ᓄᑦᓱᑦ ᐊᑦᑕᓇᖏᑦᑐᑦᑐᑦ ᐱᖅᓂᐅᑎᓂᒪᓕᕐᓂᑦ ᐊᑕᑦᓂᒥ.

ᓄᓇᕗᑦ: Disko Fan ᐊᒻᒪ ᖃᑭᖅᑕᓗᒍ ᐊᑯᑭᑐᐃᓐᓂ ᓄᓇᖅᑐᑦ ᐃᑭᕋᖅᖢᑎ ᓴᓴᒥᐅᑉᐳᓵᑦ



ᓄᓇᖅᐊᖅ: ᖃᑭᖅᑕᓗᒍ ᐊᑯᑭᑐᐃᓐᓂ ᓄᓇᖏᑦ ᐃᑭᕆᔨᖅ ᐊᒻᒪ Hatton Basin ᓴᓗᒥᐅᕕᒃᓴᑦ



ጋϟᖅበናበᓂᖅ:

ΔLᖏᐅᑕᑕᓱᓴძᑦ ხᓇᑕᒥ ᖃᑎᐊᓇᑦ ᐅᑭᐅᑕᖅጋᒧᑦᓗ ᐱᙆᐊϟᖅᑎᖅᑎᑎᓂᓱᒪᓪᑦ ᑎᒧᒥᐊᖅᑎᐅᖅᑐᖅᑕᐅᓪᓂᑦ ᐊძᑭᑐᑦ ᖅᑕᕆᓕᖅᑭᓂᑦ Δᖅᓗᑕᑕᓱᓂᖏᒧᑦ ᖃᑎᖏᑕᓗᑎ ᐊᒪᒪ ᐅᖅᑲᐅᓂϟᖅᑎᖅᑕᓗᑎ ᑎᑎናᓱᐊᖅᓯᖅᑕᐅᓂᓱᒪᓪᑦ IFMP ᑲᑎᒪᓂᐅᖅᑲᑕᖅᓂᒪᓪᑦ ხᓇᖅᓱᖅᓯ ᐅᑭᐅᑕᖅጋᒥ Δᖅᑲᒥᐅᑕᒥᑦ Δᖅᓗᓴᖅᓂᑦ ᑎᒧᒥᐊᖅᑯᑦ ᐅᖅᑲᐅᓴᕐᐊᖅᑯᑦ ᑲᑎᒪᔨᖏᑦᑐᖅᑎᑦ (EAGSAC). ᒧᖅᓇᐅᔨᒥᑦ ᑎᑎናᓱᐊᖅᓯᖅᑕᐅᓂᓱᒪᓪᑦ ᐅᖅᑲᐅᓂᐅᓪᓚᖅᑕᐅᓂᑎᑦ 2019 EAGSAC ᑲᑎᒪᓂᖅᓯᖅᓂ ᕕϟᐊᓗ 19-20 ᐊᒪᒪ ጋᓂᐅᖅᑲᑕᐅᓪᓚᑎᑐ EAGSAC Δᓪᒋᕆᐅᔨᓄᑦ ᐊᒪᒪ ᐊᓂᖅᑐᓄᑦ ᑎᒧᒥᐊᖅᑎᐅᖅᑲᑕᐅᔨᓄᑦ ᖃᑎናᑦᓚᑐᑎ ᐊᒪᒪ ᐅᖅᑲᐅᓂϟᖅᑎᖅᑕᑦጋᑎᑦ ᒧᑦ 25-ᒥᑦ, 2019. ᓇᓚᓇΔᕐᖅᓂᒪᓂᖅᒋᑦ ጋϟᖅበናበᓂᓱᒪᓪᑦ ᐊᒪᒪ ᑎᒧᒥᐊᖅᑎᐅᖅᑲᑕᐅᔨᑦ ᓇᓂᕆᐅᔪᓇᖅጋᑦ ᐅᖅᓂ ᐊᖏጋᖅᓂᒪᔨᖅ 3.

ᐊጋᓕᖅძᕆᐅᔨᖅ:

ᐱᔨᐁᑎᑦᓕᔪ ᐅᑦᓗᒥᒍᖏᓕᖅᑎᑕᐅᓂᒪᓂᖏᑐ 2019 IFMP:

- ᓇᓚᓇΔᕐΔᔨᖅ ᒧᓇᓇᐅᔨᒥᑦ ᐊᐅᓕᑎᓴᐅᔨᖅ ᑎᒥᐅᔨᖅ ᑦᓴᓂᒥᖏᑐ Δᖅᓗᑕᑕᓱᓂᖏᒧᑦ ᐊᒪᒪ ርძᐊᖅᐅᑎᑎᓄᔨᖅ Δᓂᒪᑕᐅᖅᑕᐅᓂᒪᕆᔨᓱᓰᖅጋᓂᑦ, ᐱᖅᓂᐅᑎᑦᓗᑎ ᐊᖅᑭᖅጋᑦ ᐱᓕᑦᑕᐊᓂᖏᑐᓄᑦ ᖅᑎᒪᓴᑦ ᐊᒪᒪ ᒥᓂᑦር;
- ᖃᖅᓄΔᑕᓂᓂᖅᑎᑎᖏᑦᑕᖅ ᖃᖅᓄΔᑦᑐΔᓇᖏᓂᑦ ᒧᓇᖏᖅᑕᐅᓂᒪᕆᔨᓱᓰᖅጋᓂᑦ ᐱᓕᑦᑕᐊᓂᖏᓐᓂ / ᐊጋᖅᑕᐅᔨᖅ ᓇᒥጋΔᓇᓇᖅ Δᖅᓗᑕᑕᓱᓂᐅᔨᒥᑦ, Δᓗᐊᓂᖅᓗᓗᐊᕈᑎ ᐅᕐᓇᐃᓇᓂᓂᑦ ᐊᑐጋᐊᓂᖅᑎᕆᓗᐊᕈᑎ NSA;
- ᑐᓂᔨᖏᑦᑐᖅ ᐅᕐᓇᐃᓇᓂᓂᑦ ᐱᐅᖅᑎᖅᑎᑎᖏᑦᑐᖅ ᖃᖅᓄΔᑦᑐΔᓇᓇᓂᑦ ᐊᐅᓕᑎᓂᓱᒪᓪᑦ ᐊጋᖅᑕᐅᔨᓇᓂᑦ; ᐊᒪᒪ
- ᐊᑕᖏᕐᕕᓂᑦ ጋϟᖅበናᑎᓂᑦ ᐱᖏᓕᐊᓂᖏᓐᓂ ᐊᑐጋᖅᑕᑕᒧᐊᖏᓂᑦ ᑎᒧᒥᐊᖅᑎᐅᖅᑲᑕᐅᔨᓄᑦ, ᐊᒥᓂᒪᓕᓇᓂᑦ ᐱᙆᐊϟᖅᑎᖅᑐᖅᑕᐅᓂᑦ ᐊᐅᓂጋᖅᓂᒪᔨᖅᑐᖅ.

ᐊጋᓕᖅძᕆᐅᔨᖅ ᑕᕆძᐊ NWMB ᐊᖏᑦᑐᖅᓂᒪᓂᑦ ᐊᖅᑭᑎᑦᐊᖅᓂᒪᔨᑦ 2019 IFMP ᓄᑕᖓᔨᑦ ᑎᑎናᖅᑕᐅᓂᒪᔨᑦ ᑭᖓᖏᕐᐃᓇᑐᖅ 2013 IFMP.

ᑎᑎናᖅጋᖅ:

DFO ᓄᓇᒋᖏᓴᖅጋᓂᑦ ᐊᐅᓕᑎᓂᓱᒪᓪᑦ, ᖃᑎᐊᓇᑦ ᐅᑭᐅᑕᖅጋᒧᑦᓗ ᐊᕐᓇᑐᖅᓂᒪᔨᓄᑦ.

ᐅᑦᓗᖏᓐ:

ᒪΔ 15, 2019

ᐱᖅᓂᐅᑎᓂᒪᔨᑦ (3):

ᐊᖏጋᖅᓂᒪᔨᖅ 1: 2019 IFMP GHL 2019 ᓇᓚᓇΔᕐᖅᓂᒪᓂᖅᑎᑎᓂᒪᔨᑦ (ጋᑭᓂᓇᖅᓂᑎᑕᐅᓂᒪᔨᑦ)

ᐊᖏጋᖅᓂᒪᔨᖅ 2: ᓇᔨᓚᓇΔᕐᖅᓂᒪᓂᖅᑎᑎᓂᒪᔨᑦ 2019 GHL IFMP ᐊᓂᔪᖅጋᑦ (ጋᑭᓂᓇᖅᓂᑎᑕᐅᓂᒪᔨᑦ)

ᐊᖏጋᖅᓂᒪᔨᖅ 3: ጋϟᖅበናበᓂᒪᓪᑦ ᓇᓚᓇΔᕐᖅᓂᒪᓂᖅᑎᑎᓂᒪᔨᑦ (ጋᑭᓂᓇᖅᓂᑎᑕᐅᓂᒪᔨᑦ)

Fisheries and Oceans Canada | Pêches et Océans Canada

Fisheries Management | Gestion des pêches

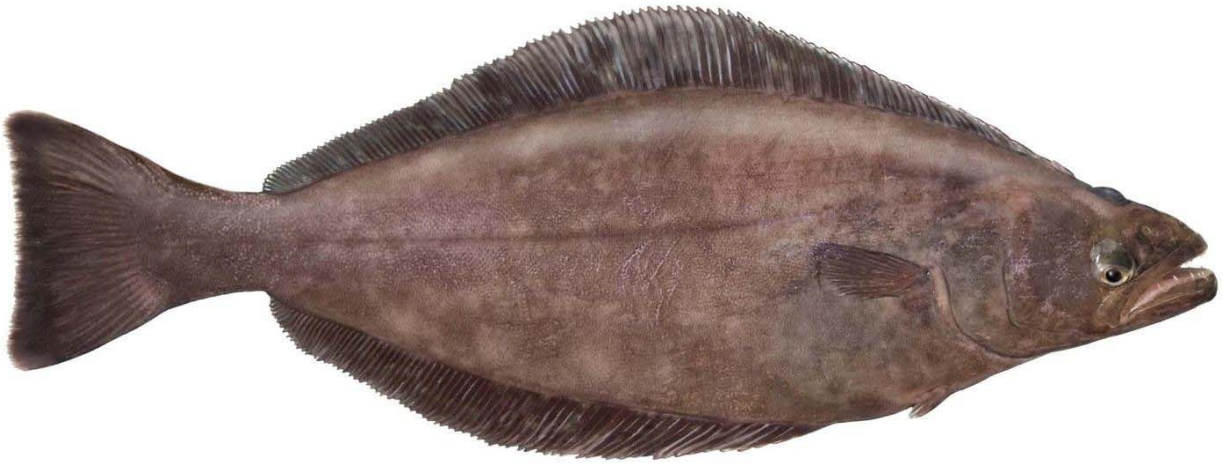
ᐊᑐᓕᖅᑎᑕᐅᓯᒪᓕᕆᓂ ᐃᖃᓗᓕᕕᒋᔭᖏᑦ
ᐊᐅᓚᕐᑎᐸᐊᑎᒃᓴᑦ ᕿᖏᐅᑎᖑᕐᓂ ᓇᐃᓐᓇᖅᓯᒪᔪᓂ

ᖃᓄᕐᕿᓂᕐ

(ᕿᐃᓐᖓᕆᐊᑎᐊᖅ ᖄᐃᕐᓴᓕᔭᕐᓯᐃᓐᕿᓕᖅ)

ᐅᐊᖑᓇᖑᓕᓂ ᑲᓇᑕᕐ ᑲᓇᓐᓇᖑᑕ ᐃᖃᓗᓕᕕᒋᔭᖏᑦ ᑲᑎᒪᔨᖃᑎᒌᖑᓂᕐ ᐊᕐᕌᕐᑐᖃᓯᒪᒃᑯᓂᖑᓪᓗᓂ 0

ᐊᑐᓕᖅᑎᑕᐅᓂᒃᑯᑦ
2019



Canada

ᐱᑭᐊᖏᓂᖅ

ᐱᓴᐃᑎᖅ ᑕᑯᓗᒃ ᐊᑐᓕᖅᑎᑕᐅᓪᓚᐃᔪᓂ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᖅ (IFMP) ᓇᐃᓵᖅᓯᒪᓂᖅᑎᑦ ᓇᓗᓇᐃᔨᐃᓚᑦᑎᑎᑐᑎᒍᑎᐅᔪᖅ ᑐᑭᓯᐊᕐᑭᑎᓂᑦ ᓇᓂᑦᐅᓯᓂᖅ ᐃᓕᓴᕐᓂᖏᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᓂᓂᖅ. ᑕᐃᓐᓇ ᑎᑎᖅᑲᖅ ᑐᓴᖅᑎᕐᑎᔪᕐᑎᑦᑕᐅᑐᖅ ᑐᑭᓯᐊᕐᑭᑎᖅᑎᑦ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᓪᓚᓗ ᐊᐅᑕᕐᑕᐅᓂᖏᑦ ᑕᑯᔨᐃᓂᖅ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᓪᓚᓗ ᐃᒪᒋᑎᑕᕐᓂᓚᓵᔨᐃᔨᑦ ᑲᑎᒪᔨᖅ (DFO) ᐃᖏᓄᓇᐃᔭᖅᑎᖅᑎᓯᓐᓄᑦ, ᐱᓴᔨᔭᖅᑎᒍᑐᖅᑐᑦ ᐊᐅᑕᕐᑎᔪᐅᖅᓄᑎᑦᑐᑦ ᑲᑎᒪᔨᖅᑎᑦ ᐊᓪᓚᓗ ᐊᔨᓯᖅᑎᖅᑎᑦ ᐱᒋᒋᔭᖅᑎᑦ. ᑕᐃᓐᓇ ᐊᑐᓕᖅᑎᑕᐅᓪᓚᐃᔪᓂ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᓂ ᑐᑭᓯᐅᓪᓚᑎᑕᐅᔪᖅ “ᒪᓕᓐᓄᕆᓪᑐᐃᓐᓇᖅᑐᓂᓕᖅ” ᒪᑭᒪᔨᓐᓇᖅᑐᖅᓯᓂ ᐊᐅᑕᓕᐅᔨᐊᑎᖅᓴᖅᑎᓯᓐᓄᑦ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ.

ᑕᐃᓐᓇ ᐊᑐᓕᖅᑎᑕᐅᓪᓚᐃᔪᓂ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᓂ ᐱᓴᔨᔭᖅᑎᒍᑦ ᓚᓕᑕᐅᔨᓴᖅᑎᖅᑐᖅᑐᖅᓴᖅᑐᑦᑕᖅ ᑎᑎᖅᑲᖅ ᑐᖓᓚᐅᑦᔨᐅᑕᑭᓐᓇᖅᑐᐃᓐᓇ ᐃᖏᓄᖅᑐᐃᐃᓚᔨᓂᒍᑦ ᐊᐃᕐᐅᑎᖅᑎᑕᐅᔪᖅᑐᐃᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᓂ ᐋᑦᑭᒋᐊᖅᑕᐅᓇᖅᑐᖅ ᖃᐅᖓᑕᐃᓇᑎᑎᓂᒍᑦ ᐊᓪᓚᓗ ᓄᖅᑲᖓᑎᑎᓂᑎᓯᐊᕐᓂᒪᓂᖅ ᒥᓂᑕᐅᑦ ᐃᓱᒪᖅᓯᓂᐅᓂᖅ ᓴᖅᑭᓂᖓᓂᑦ ᐋᑦᑭᒋᐊᖅᑕᐅᓱᒪᔨᓂᒍᑦ *ᐃᖃᓗᓕᓴᕐᓂᖏᓐᓇᓪᓗᒍ ᐱᖁᔭᖅᑎᑦ*: ᒥᓂᑕᐅᔨᖅᑐᖅ ᐃᒪᐃᔪᓇᖅᑐᖅᑐᖅ, <<ᖏᓐᑎᖅᑎᓇᓯᐊᕐᓂᓂᖅ ᐅᕐᕕᔭᓕᓂᑦ ᐊᓯᐊᑎᔪᑦ ᐱᓴᔨᔭᖅᑎᖅᑐᖅᓴᖅᑐᑦᑕᖅ ᐃᓗᓕᖅᑐᓂᑦ ᐊᑐᓕᖅᑎᑕᐅᓪᓚᐃᔪᓂ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᓂᓂᖅ ᒪᓕᓐᓂᒍᑦ ᑐᓂᔪᖅᑲᖅᑐᖅᑐᑦ ᒪᓕᓐᓂᒍᑦ ᐃᖃᓗᓕᓴᕐᓂᖏᓐᓇᓪᓗᒍ ᐃᖃᓗᓕᓴᕐᓂᖏᓐᓇᓪᓗᒍ.

ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᓪᓚᓗ ᐃᒪᒋᑎᑕᐅᓪᓚᐃᔪᓂ ᑲᒪᕐᓴᐊᖏᖅᑎᓯᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦᑐᓂᖅ ᐱᓴᔨᔭᖅᑎᒍᑦ ᒪᓕᓐᓂᒍᑦ ᓄᓇᕗᖓᓪᓗᒍ ᐊᖏᕐᕈᑎᑦ, ᑕᐃᓐᓇ ᐊᐅᑕᕐᑎᓂᖏᑦ ᐃᖏᓄᔨᓕᓂᖏᓐᓇᓪᓗᒍ ᐊᐅᑕᕐᑎᓂᖏᑦ <ᖏᓐᐅᑎᖅᑎᓂ ᐊᐅᑕᕐᑎᓂᖏᑦᑐᓂᖅ ᒪᓕᓐᓂᒍᑦ ᐱᓴᔨᔭᖅᑎᒍᑦ ᒪᓕᓐᓂᒍᑦ ᓄᓇᕗᖓᓪᓗᒍ ᐊᖏᕐᕈᑎᑦ ᑎᑎᕋᖅᑭᒪᔨᖏᓐᓇᓪᓗᒍ ᓄᓇᕗᖓᓪᓗᒍ ᐊᖏᕐᕈᑎᒋᑦᑐᓂᖅ ᓴᖅᑭᓂᖓᓂᑦᑎᓯᐊᕐᓂᒪᓂᖅᑐᑦᑕᖅᑐᓂᖅ ᐊᔪᔨᖅᑕᖅᑐᓂᖅᑐᑦᑕᖅᑐᓂᖅ.

1. ᓇᓗᓇᐃᔪᖅᑐᓄᖅᓯᒪᓂᖅᐸᑦ ᐃᖅᑲᓇᑕᓐᓄᓚᐅᑦ

ᖃᓄᖅᓯᓐᓂᐅᕕᖏᓐᓄᖅ ᐅᖅᑲᐅᑕᐅᔨᖅ ᑕᕐᓂᓂ ᐊᑐᓕᖅᑎᑕᐅᓂᕐᓂᕆᔨᑐᓂ ᐃᖅᑲᐅᓕᑕᓂᖏᓪᓗᑦ ᐊᐅᓚᕆᑎᓂᖏᓪᓗᑦ ᐅᑭᐅᖅᑕᕐᓂᖅ (IFMP) ᐅᐊᖓᓇᖓᓂᒃᓯᒃ ᑲᓇᑕᑦ ᑲᓇᖓᓇᖓᑕ ᐃᖅᑲᐅᓕᑕᓂᖏᓪᓗᑦ ᑲᑐᔪᓴᖃᑎᒌᖓᓂᖅ (NAFO) ᐊᐃᑐᑐᖅᓯᒪᓂᖅᑲᖓᓂᖅ 0 (ᐊᔪᓐᖓᒍᖅ 1). ᐊᐃᑐᑐᖅᓯᒪᓂᖅᑲᖓᓂᖅ 0 ᐊᐃᑐᑐᖅᓯᒪᔨᖅ ᐅᑭᐅᖅᑕᖅᑐᒥᖓᓂᖅᑕᐅᓕᓕᓂᐅᓂ, ᐊᐃᑐᑐᖅᓯᒪᓂᖅᓯᒪ 0A (ᖃᑭᑲᖅᑕᓕᓛᑦ ᐃᒪᖓᓯ) ᐅᖓᓯᑕᐅᔪᔨᖅ ᐅᐊᖓᓯᑦ 78°10’N ᑎᑭᑐᐃᒍ 66°15’N, ᐊᓚᒪᓗ ᓂᕆᐊᓄᑦ, ᐊᐃᑐᑐᖅᓯᒪᓂᖅᓯᒪ 0B (ᖃᑭᑲᖅᑕᓕᓛᑦ ᐃᑭᕋᓖᓯᓯᒪ) ᐅᖓᓯᑕᐅᔪᐅᕐᖅ ᐃᐊᖓᓯᑦ 66°15’N ᑎᑭᑐᐃᒍ 60°12’N.

ᐊᐃᑐᑐᖅᓯᒪᓂᖅᓯᒪ 0A ᐃᖅᑲᐅᑕᓕᓂᓂᖅ ᑭᓇᐅᔪᑦᑕᕋᓂᔪᐅᕐᖅ ᓴᓂᐊᖅᑕᐅᓂᒪᑦᓯᐅᓄᖅ ᐃᖅᑲᐅᑕᓕᓂᖅ ᑯᑕᖃᖅᑎᑕᐅᓂᒃᓄᖅ ᓄᓇᕐᓂ ᑭᓇᐅᔪᖅᑕᖏᓐᓄᖅᑕᐅᓂᒃᓄᖅ, ᐊᐃᑐᑐᖅᓯᒪᓂᖅᓯᒪ 0B ᑯᑕᖃᑦ ᐅᑦᓗᒥᐅᔨᖅ ᑲᑐᔪᐅᕐᖅ ᐊᔪᐅᖅᓯᒪᓂᖅᓯᒪᓂᖅ ᓴᓂᐊᖅᑕᖅᑐᒥᖓᓂᖅᑕᐅᓂᒃᓄᖅ, ᑭᓇᐅᔪᐅᑦᑕᕋᓂᐊᕐᓂᐅᑎᓂᐅᔪᐅᕐᓂᖅᓯᒪᓂᖅᑲᖓᓂᖅ ᓴᓂᐊᖅᑕᖅᑐᒥᖓᓂᖅᑕᐅᓂᒃᓄᖅ ᐊᐃᑐᑐᖅᓯᒪᓂᖅᓯᒪ 0A ᑯᑕᖃᖅᑎᑕᐅᓂᒃᓄᖅ ᐊᐅᓚᐅᖅᑕᐅᓂᒃᓄᖅ ᐃᖅᑲᐅᑕᓕᓂᖅ ᑯᑕᖃᖅᑎ

ᑲᓇᑕᒥ *ᐃᖃᓗᓕᕆᓂᖏᑦ ᐱᖁᔭᖓ,* ᐊᒻᒪᓗ *ᐃᖃᓗᓕᕆᓂᖏᑦ (ᑕᒪᐃᓐᓄᑦ)* ᐊᑐᐊᒐᕋᓱᒃᑎᓂᒃ ᐊᒻᒪᓗ *ᑲᓇᑕ*ᒃ *ᑲᓇᓐᓄᑦ ᐃᖃᓗᓕᕆᓂᖏᑦ ᐊᑐᐊᒐᕋᓱᒃᑎᓂᒃ*; ᐊᒻᒪᓗᓂᑕᐅᖅ ᓂᐃᒪᖏᑕᐅᓕᕆᓂᖏᑦ ᐱᖁᔭᖓ ᐊᒻᒪᓗ *ᐅᒪᔨᔪᐃᑦ ᖃᕋᓴᐃᓐᓇᓕᒥᓂᖏᑦ ᐱᖁᔭᖓ (SARA)* ᐊᑐᖅᑕᐅᓪᓗᖅᑐᑦ ᑲᑦᓴᕋᒪᓂᖃᖅᑐᖓ 0 ᖃᑦᓯᓚᐅᓂᒃᓴᓂᒃ ᐊᐃᑕᑐᖅᑭᒪᓂᖃᖅᑐᖓ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᓄᓇᕗᒻᒥᐅᑦ. ᓴᖅᑭᓂᓂᑦ ᑐᓂᔭᐅᓪᓗᖅ ᒪᓕᑕᑐᑦ ᑕᖅᔪᐊ ᐱᖁᔭᖓ ᐊᑐᐊᒐᕋᓱᒃᑎᓂᒃ ᐱᓕᕆᓚᐅᑎᖅᑎᑦ ᒥᓂᑕᐅᔨᑦ ᓚᓇᓚᐃᖅᑭᓕᕆᓂᖏᑦ ᓴᖅᑭᓂᑐᐅᓚᐅᑎᖅᑎᑦ ᐅᒥᐊᔪᐊᖅᑐᖓ ᐊᑐᖅᑐᖅᑎᑐᓂᒃ ᐃᓕᓴᖅᑐᖓ, ᐃᓕᓴᖅᑐᖓ ᐱᖁᔭᖓ ᐃᖃᓗᓕᕆᓂᖏᑦ ᐊᐃᑕᑐᖅᑭᒪᓂᖃᖅᑐᖓ ᐊᐃᑕᑐᖅᑭᒪᓂᖃᖅᑐᖓ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐃᓕᓴᖅᑐᖓ, ᐃᓕᓴᖅᑐᖓ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐊᓯᖏᓂ, ᐃᓕᓴᖅᑕᐅᓪᓗᖅᑐᑦ ᐸᓪᓕᑕᐅᕐᓂᖅᓴᓗᒃᑐᓂ, ᐃᓕᓴᖅᑐᖓ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐊᓯᖏᓂ, ᐃᓕᓴᖅᑕᐅᓪᓗᖅᑐᑦ ᐸᓪᓕᑕᐅᕐᓂᖅᓴᓗᒃᑐᓂ, ᑐᒥᐊᔪᐊᖅᑐᖓ ᐅᑐᖅᑎᖅᑎᑦᑐᓂᒃᓴᓂᒃ, ᐊᐃᑕᑐᖅᑭᒪᓂᖃᖅᑐᖓ ᐊᐃᑕᑐᖅᑭᒪᓂᖃᖅᑐᖓ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ ᐊᑐᐊᒐᕋᓱᒃᑎᓂᒃ ᐊᐅᓚᑕᐅᓪᓗᖅᑐᑦ

ᐊᓚᓗ ᐅᐸᐱᓚᓴᐅᐅᖅ ᐃᖃᓗᐃᑦ ᑯᓯᐅᒍᓚᒥᑦ ᑕᐱᐅᕆᒥ, ᓯᕐᓗᐅ ᑲᓇᐅᖓᓗᓄ ᐅᐊᖏᓇᖓᓂ ᑲᓪᓚᓐᓂ ᓄᓇᖏᓂ ᐊᓚᓗ ᓂᒋᐊᓄᑦ ᐱᖓᓱᖓᓗᓄᓕᕐᓂᓗ ᖃᑭᖅᑕᕐᓗᑎᒃ, ᑕᐱᐅᖓᓗᓄᑦ ᐊᓚᓗ ᖃᑭᖅᑕᕐᓗᑎᒃ ᑲᓪᓚᓐᓂ ᓇᓕᑕᓗ ᐃᑭᕐᓴᖖᖓᓗᓄᑦ.

ᐃᓄᐃᑦ ᐊᓚᓗ ᐃᖃᓗᖓᓯᐊᖅᑎᑦ ᐱᖃᑯᓂᖃᖅᑎᒍᒧᑦ ᐊᐅᑎᒥ ᖃᐅᐅᓕᒫᑦ (TEK) ᐱᓚᒪᐅᐅᖅ ᐃᖃᓗᑐᕆᓂᖅᑲᒃ ᐊᐅᑕᓪᓚᓂᒃ ᐊᓚᓗ ᐊᑐᖅᑕᐅᓯᓂᖅ ᖃᓄᓕᒫᓂᖅᖃᑦ ᐃᓕᓴᐅᒪᓂᖅᑕᑦ ᐃᓯᓚᑦᑐᖅᑕᐅᓂᖅᑕᑦ ᐊᐅᑕᓪᓚᓂᒃ

ᐊᕐᑎᑎᒥᖅᓗᒍ ᑲᒪᕆᐊᓂᖅᓗᒍ ᓇᓗᓇᐃᓕᓴᐃᓇᖅᑐᖅᓂᖅ ᐃᖅᑲᓗᓪᓚᓂᖅᑕᖃᒪ ᓄᓇᕕᐊᖅᑐ ᓇᓗᓇᐃᓕᓴᐃᓇᖅᑭᖅᑕᕐᓗᐃᑦᑐᑦ ᑎᒥᐅᔪᓂᑦ, ᐊᐅᓪᓚᐅᑎᓂᖅ ᐊᖏᐅᑕᐅᔾᔨᐊᓕᖅᓯᒪᓕᖅᑐᑦ ᓂᐅᕐᓚᐅᖅᑕᑐᓄᑦ ᓂᕆᐅᓪᓚᐅᖅᑐᖅᓯᓪᓚᕆᐊᖃᖅᑐᑦ ᓄᐃᑦᑐᕐᓂᖅᑐᑦ, ᐃᓕᓴᕐᓂᐊᒍᑎᖅᑲᓪᓚᕆᐊᖃᖅᑐᑦ ᐊᑐᐊᓂᖅᑐᑦ ᐊᑐᐊᓂᖅᑐᑦ

ᐅᓂᖅᐅᓱᖅᑎᓂᖅ – ᐱᓪᓚᑎᖅᐅᖅᓯᒪᓕᒃ ᐱᕎᐊᖅᓯᒪᓂᖕᓂᓐᓂᒃ ᐃᑭᓱᑐᐊᑭᐅᑦ ᐅᓂᖅᐅᐅᑕᐅᓱᑦ ᐃᓴᖃᑕᓚᓗᖅᑯᑦ ᐊᓚᒐᓗ ᐃᓚᖏᑕᐅᑕᓚᓗᖅᑯᑦ ᑲᓇᑕᒥ ᐱᓕᕆᓂᒃᑯᑦ ᐊᓯᖏᓂᒃᑯᑦ, ᐊᓱᖏᑦ ᐅᓚᔪᐃᑦ ᐱᓕᖅᖏᓐᓄᑦ ᐊᖏᓂᒥᑦ, ᓄᓇᖅᑕᐅᔪᑦ, ᐊᓱᖏᑦᓴᓗ. ᖃᖏᕐᓴᑯᑦ ᐅᓂᖅᐅᐅᓱᖅᑎᓂᖅᑎᑦ ᐱᔪᓯᑕᐅᑭᐅᖓᓂᒃ ᐃᓕᖏᓂᒃᓴᓗ. ᑕᒪᓐᓇ ᑐᐱᓱᐊᕐᑎᖓᓐᓄᑦ ᐊᑐᖅᑕᐅᓱᖅᓱᖅ ᐅᐊᑎᑎᓴᐅᔪᑦᓴᓗ ᐊᑐᓚᔭᑎᑦᓴᓗ ᐊᑐᓚᑎᔪᓐᓇᖅᑐᑦ ᐊᑐᓚᑎᔪᓐᓇᖅᑐᑦ ᐊᑐᓚᓕᒪᔪᖅᑕᐅᔪᑦ ᐊᑐᓚᑎᔪᓐᓇᖅᑐ

4.3. ᑕᓗᐅᑦ ᐊᓚᒐ ᓇᐅᑎᓴᐅᐸᑦ ᐃᓱᒪᖏᖅᓱᐅᑭᑎᑭᐅᓂᖏᑦ

ᐃᖏᓄᑕᓘᓴᕆᔨᑦ ᐊᓚᒐ ᐃᒪᖏᐅᑕᑕᓘᓴᕆᔨᑦ ᑲᓇᑕ ᒪᑕᖏᐅᑕᑕᓘᐢᑎᑦ ᓇᓄᑦᑐᑭᓘᐢᑎᑦ ᐊᐟᑎᖏᑦ ᐅᒪᐟᖏᑦᓄ ᑕᓗᐅᖏᑦ (EBSAs) ᑲᓇᑕ ᑕᓗᐅᖏᑦᓂᓂ. ᐊᐟᑎᖏᑦ ᐅᒪᐟᖏᑦᓄ ᐱᓚᓗᐅᐸᑦ ᑕᓗᐅᖏᑦ ᐱᖏᓂᕆᔨᑦᑐᑦ, ᑭᓱᐊᓂᒋ ᐃᓱᒪᑭᓴᐅᓴᓘᐊᓕᒋᑦ ᑕᓗᐅᑦ ᐊᑕᑕᓇᖏᐃᖏᓱᓴᐅᓴᓘᐊᓕᒋᑦ ᐊᐅᒋᑕᐅᐸᐟᐊᖏᑭᑎᒍᑦ ᐸᖏᓇᑕᐅᐸᐟᐊᑎᑭᓄᓪᓗᑦ ᐊᓚᒐ ᐃᓱᒪᑕᐅᑭᑎᑭᓄᓪᓗᑦ.

ᑲᓇᑕᒥ ᑐᖇᒪᑲᑦ ᐱᑕᓘᐊᖏᖅᓂᐊᖏᖅᓂᖏᓱᒪᐟᑦ ᓱᓂ ᓴᐳᓴᓗᑎᑦ 10%-ᖏᐟᓂᑦ 2020–ᑭᑭᓂ. 2017-ᒥ ᐊᓚᒐ 2020–ᒥ ᑎᑭᑕᐅᓇᓱᖏᖅᑐᑦ ᑕᒪᐃᓐᓂ ᑕᐃᔨᐅᓱᒃ ᑲᓇᑕ ᑕᓗᐅᖏᑑᒃ, ᐅᐊᖏᓇᖏᑦᑑᓱᒃᑕᐅᐸᐟ, ᐅᐊᖏᓇᖏᑦᑑᓱᒃᑕᐅᐸᐟ ᐊᐅᑕᑎᐃᔨᐅᐸᐟ ᓇᓄᑎᐊᓂᓂᖏᑦ. ᐊᐃᑐᖏᑦᓂᓂᑑᒃ ᐊᐃᑐᖏᑦᓄᓪᓗᑦ ᐊᓚᒐ ᑎᔪᒥᐊᖏᑎᑭᐅᓴᐅᐸᐟᐊᖏᑦ ᐊᐃᑐᖏᑦᓄᓪᓗᑦ ᐱᖏᓂᕆᔨᑦᑐᑦ, ᐅᐊᖏᓇᖏᑦᑑᓱᒃᑕᐅᐸᐟᐊᖏᑦ ᐊᐃᑐᖏᑦᓄᓪᓗᑦ ᐊᐅᑕᑎᐃᔨᐅᐸᐟᐊᖏᑦ ᐱᖏᓂᕆᔨᑦᑐ

ᐊᑐᓕᖅᑎᑕᐅᓴᖅᑐᑦ ᐊᐃᕐᑐᖅᓯᒪᓂᒃᑲᓐᖓᓂᒃ 0 ᖃᓄᕎᓕᓐᓂᐊᕐᓂᓕᕆᓂᕐᒧᐃᖓᔪᓂᒃ ᐱᖓᓕᓕᐊᔪᑎᖃᕐᓂᑦ ᐊᒻᒪᓗ ᖃᓄᐃᓕᓕᐅᕙᓪᓕᐊᑎᑎᕐᓴᐅᓗᑎᑦ ᐊᒥᓱᓄᑦ ᐊᖏᔪᒍᓐᓇᖅᑐᒍᓄᑦ ᒪᓕᒐᑎᑦ ᓯᖓᓕᓕᐅᑎᖏᓕᕐᑐᑦ ᐃᓴᖅᑲᓕᕐᓂᐊᕐᓂᑦ ᐊᒻᒪᓗ ᐃᒪᓕᕐᒥᐅᓕᕐᓂᐊᕐᓂᑦ ᑲᑎᕐᓂᒃ ᐃᓴᖅᑲᓕᕐᖏᖓᓂᒃ ᐊᒻᒪᓗ ᐊᓯᖏᓐᓂᒃ ᑎᒍᒥᓄᑦᑐᐅᓂᒃ.

ᐅᖃᐅᓕᐅᕆᓚᐊᖃᖅᖢᓴᑦ, ᐊᒥᓱᓚᓂᐅᓴᑦ ᖃᓄᕎᓕᓕᑎᐅᓴᑦ ᑕᕝᖅᓄ ᐃᖃᓗᓯᓚᐊᖃᓕᕆᐅᓴᒃ ᓂᐅᓴᐅᓱᐃᖓᒪᒪᑕ ᐅᖃᓕᓴᐅᓱᐃᖃᖅᑎᖓᓂᖅᓄ ᐊᖃᖅᑭᒍᓱᒪᓕᕐᓂᑦ (ᓱᐃᓗᓴ ᑲᓪᓕᓐᓂ ᓄᓇᖅᓄ). ᑲᓇᑕ ᐃᖃᖅᑲᓇᐃᓴᖃᖅᑎᑕᒃ ᑲᓪᓕᓐᓂ ᓄᓇᖅᓄ ᐃᖃᖅᑲᓇᐃᓴᖃᖅᑎᖓᓂᖅᓄ ᐊᖃᖅᑭᒍᓱᒪᓕᕐᓂᑦ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄᑦ ᑲᑎᕐᓂᒃ ᑕᐅᖃᓱᐃᓕᕐᓂᑦ ᖃᓄᕎᓕᓕᑎᐅᓴᑦ ᓄᓇᖅᓄ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄᑦ ᑲᑎᕐᓂᒃ ᑕᐅᖃᓱᐃᓕᕐᓂᑦ ᓄᓇᖅᓄ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄᑦ ᐊᖃᖅᑭᒍᓱᒪᓕᕐᓂᑦ ᑲᑎᕐᓂᒃ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄᑦ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄᑦ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄᑦ ᐊᖃᖅᑭᒍᓱᒪᓕᕐᓂᑦ ᐃᓕᓴᖅᑕᒥᐅᓂᖅᓄ

	• ᑐᑭᓯᒪᔪᑦᑎᓯ ᓄᐊᑕᐅᓂᖅᓯᑦ ᐱᐅᓯᑎᑕᐅᑎᕐᓯᒪᓂᑦ ᐊᑐᖅᑕᐅᓯᒃ ᐱᐅᓯᖏᓪᓗᒍ ᐱᐅᓯᖏᓪᓗᒍ ᖃᐅᔨᓴᖅᑕᐅᑎᐊᕐᓯᒪᓕᒃᑕᑦ ᑎᓐᓯᐊᓂ ᑕᓗᐅᕐᓂᐅᑕᐃᑦ ᐱᓕᐅᖏᓪᓗᒍ ᐊᒻᒪᓗ ᐊᑐᐃᓚᖅᑭᓂᖅᓯᒃᑯᓐᓂᖅᓯᑦ ᑕᒪᐃᓐᓇᖅᑐᑦ.
ᑲᑐᔨᓗᑎᑦ ᐸᐸᑎᐊᓪᓚ	
ᐱᓕᓐᓄᑦᑲᓂᐱᐊᖏᑲᒡᒍᓗᑎᑦ, ᐃᓕᐅᖏᑲᒡᐅᖏᑲᒡᒍᓗᑎᑦ ᐃᓱᒪᓕᐅᑭᐅᓂᖅᐳᐅᔪᓂᖅ, ᐊᒻᒪᓗ ᑲᑐᔨᓗᑎᑦ ᑲᒪᔨᓴᖅᑭᑦᓯᐅᓂᖅᐳᖅ, ᖃᐅᔨᓴᖅᑕᐅᓯᐅᓂᖅᐳᖅ, ᐊᐅᑕᕐᑐᐊᕐᓯᒪᓂᖅᐳᖅ ᖃᐅᔨᓴᖅᑕᐅᓯᐅᓂᖅᐳᖅ, ᐊᐅᑕᕐᑎᐅᐅᖏᑲᒡᐅᔪᓄᑦ ᑲᑐᔨᓴᖅᑎᑲᒡᒍᓗᑎᑦ ᐊᓱᖕᖓᓐᓂᒃ ᐱᓕᐅᖏᑲᒡᒍᓗᑎᑦ ᐊᓱᖕᖓᓐᓂᒃ ᐱᓕᐅᖏᑲᒪᕐᓚᓐᓂᑦ.	• ᖃᐅᔨᓴᓐᓂᐊᕐᓯᒪᓂᖅᓯᑦ ᑲᑎᒪᑎᕐᑎᖃᐅᔨᓴᖅᑕᐅᓯᑦᓯᒪᓂᑦ ᑎᒍᒥᐊᖃᑎᖅᑐᐅᔨᓴᖅᑕᐅᔪᓄᑦ. • ᐱᓕᐅᖏᑲᒡᒍᓗᑎᑦ ᓇᓗᓇᐃᖃᑕᐅᓂᒪᓚᐅᖏᑲᒡᒍᓗᑎᑦ, ᖃᐅᔨᓴᖅᑭᑲᒡᒍᓗᑎᑦ, ᐱᓰᑦᓯᓂᖅᓴᓂᖅ, ᐊᑐᐅᖏᑲᒡᒍᓗᑎᑦ ᐊᒻᒪᓗ ᑭᖑᓪᓕᐊᓂᖅ (SMART) ᐱᓇᓂᐊᒪᓂᐅᑦᔭᓂᑦ ᐊᒻᒪᓗ ᓇᓗᓇᐃᖃᑕᐅᑎᖅᑐᐅᔨᓴᖅᑕᐅᓯᑦᓯᒪᓂᑦ ᑎᒍᒥᐊᖃᑎᖅᑐᐅᔨᓴᖅᑕᐅᓯᑦᓯᒪᓂᑦ ᐱᓰᑦᓯᓂᖅᓴᓂᖅ. • ᐊᑐᖅᑕᐅᓰᓯᓂᐊᕐᓯᑦ ᑲᑐᔨᓴᖅᑕᐅᔪᓂᑦ ᖃᒪᓕᕆᔨᖅᓴᖅᑎᑲᒡᒍᓗᑎᑦ, ᓴᒃᑕᐅᓂᒪᓚᐅᖏᑲᒡᒍᓗᑎᑦ ᐃᓱᒪᓕᐅᑭᐅᓂᖅᐳᐅᔪᓂᖅ ᑭᓇᐅᔪᓂᖅ ᐊᐅᑕᕐᑎᐅᐅᖏᑲᒡᒍᓗᑎᑦ ᐃᓚᖓᔪᓂᖅ ᐱᓕᐅᖏᑲᒡᒍᓗᑎᑦ ᐃᓚᖓᔪᓂᖅ ᐊᐅᑕᕐᑎᐅᐅᖏᑲᒡᒍᓗᑎᑦ ᐃᓚᖓᔪᓂᖅ ᐱᓕ

6. ᐊᑐᓕᓂᑲᒪᔨᖅᑐᓂᒃ ᐊᓚᓗ ᓴᓂᖅᓭᐃᓂᖅ ᑯᑎᒃᑐᓄᒃ

ᒪᕿᑦᓚᑎᒃ ᐃᓗᓕᒃᑎᑦ ᑐᖓᓚᓕᐊᓕᕆᔨᓂᐊᓂᖅ ᑲᑐᔪᓗᑎᑦ ᑕᓚᐅᖅᑎᓂᖓᓂᒃᑐᓂ ᐊᑐᖅᓴᓂᒃ: ᐊᑐᓕᓂᑲᒪᔨᖅᑐᓂᒃᑎᑦ (ᓂᖅᓗᒃ ᓕᐃᓴᓂᓯᒋᓂᖅ ᐊᓚᒪᓗ ᐅᓗᐃᐃᒃᑯᑦ ᐃᓕᐅᖅᑕᐅᓂᒃᑎᓂᐅᓂᒃ ᐃᖅᑲᓗᓕᕆᓂᖅ) ᐊᓚᒪᓗ ᓴᓂᓯᒋᐃᓂᖅ (ᓂᖅᓗᒃ ᑯᑎᑦᑎᓂ ᓂᑯᑎᐊᔪᖅᑲᖅᑕᐅᓂᒃᑎᑦ). ᒥᓂᑦᑕᐅᔪᖅ ᑕᐃᒪᒍᓐᓇᖅᑐᖅ, ᐱᔪᐊᑎᖅᑎᓗᓂ <<ᕐᕆᓂᐊᖅᓗᓂ ᐅᕙᓪᓕᐊᓂᒃ ᐱᔪᐊᑎᖅᑎᑕᐊᖅᓗᓂ, ᐋᖅᑭᒃᑭᐊᖅᓂᓗᓂ ᐊᑐᖅᑭᖅᐅᓕᕆᓂᖅᓂᒃᑐᓂᒃ, ᓴᓂᓯᖅᑕᑦᐅᓂᖅᓂᖅᓂᒃ ᑯᑎᑦᑐᓄᒃ ᐊᓚᒪᓗ ᑲᑎᒪᔨᖅᑎᑦ ᐋᖅᑭᒃᑭᐊᑎᑕᐊᖅᓂᖅᓂᒃᑐᓂᒃ ᐃᓕᓴᕐᓂᖅᑐᓄᒃ ᐊᐅᑦᑎᓂᒃᑐᓂᒃ ᐊᑐᓕᓂᑲᒪᔨᖅᑎᑦ ᐊᐅᓪᓚᕐᓂᖅᓂᒃ ᐱᓕᕆᔨᒃᑐᓂᒃ ᐊᑐᖅᑭᖅᐅᓕᕆᓂᖅᓂᒃᑐᓂᒃ ᐊᑐᓕᓂᑲᒪᔨᖅᑎᑦ ᐊᐅᓪᓚᕐᓂᖅᓂᒃᑐᓂᒃ ᐊᑐᓕᓂᑲᒪᔨᖅᑎᑦ ᐊᐅᓪᓚᕐᓂᖅᓂᒃᑐᓂᒃ ᐃᓕᓴ

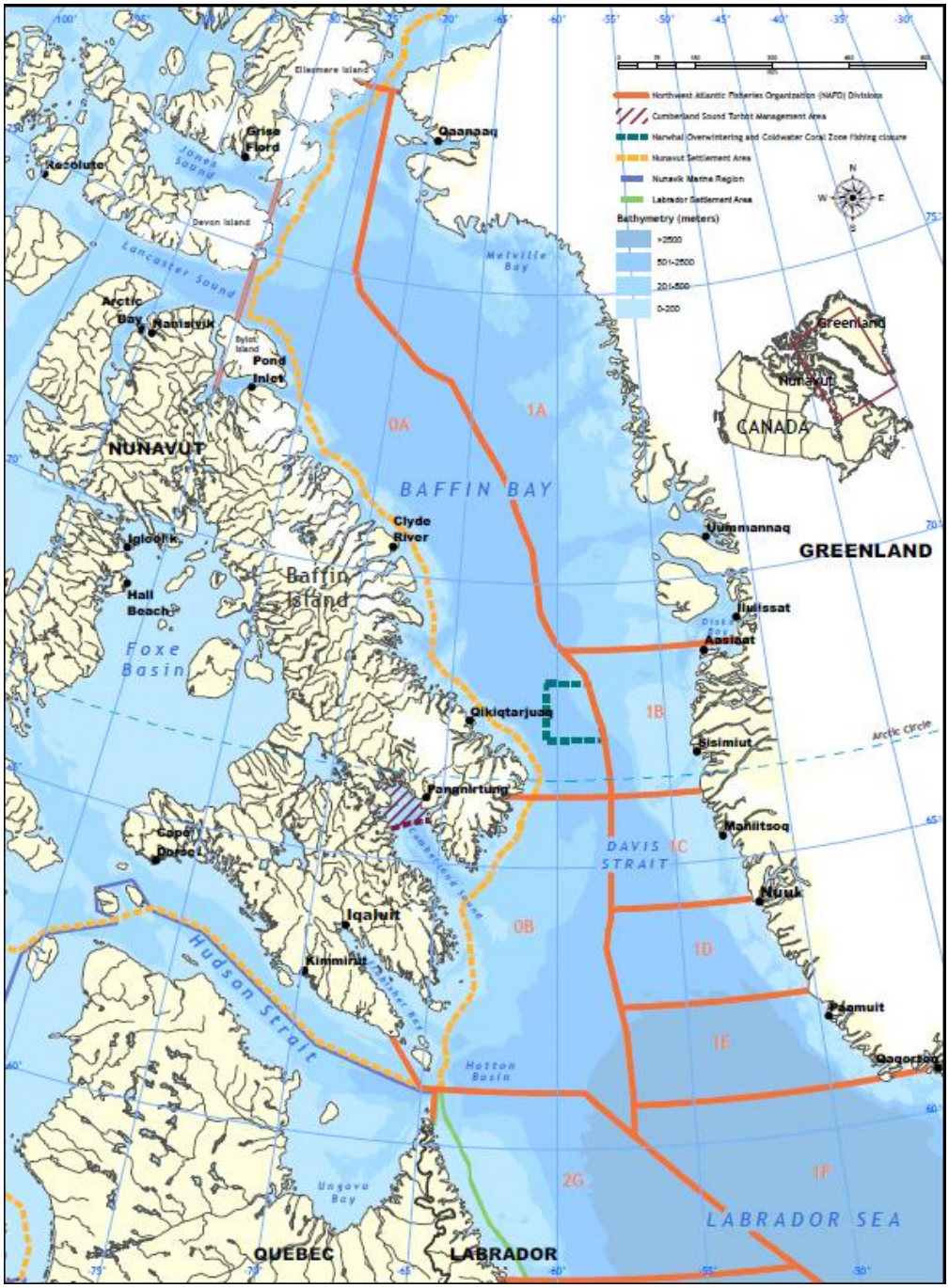
ᐊᐅᓪᓚᕐᓂᖅᑎᑦ <ᕀᓇᐅᑎᖅ, ᐊᓘᒐᕐᑕᐅᖅ ᓇᓕᔪᐅᓂᖓᓇᖅ ᐃᖃᓗᓕᕆᓇᖅ ᐃᓕᖅᑐᓚᓂᖏᓐᓂᒃ

ᑕᓚᐅᖏᒥ ᐅᐊᖏᑎᓴᑦ ᐅᒥᐊᖅᐸᐊᖅᑎᖃᑕᐅᓯᓪᓚᕐᓂᑦ ᑎᑯᑯᐊᑕᐅᓯᓪᓚᕐᓂᑦ ᐱᓕᕆᐊᖃᖅᑎᑐᑕᐅᓯᓪᓚᕐᓂᑦ ᓇᓗᓇᐃᔪᖏᑎᐊᓯᓚᕆᓗᑎᑦ ᐃᒪᐃᑕᐅᔭᓗᑎᑦ “ᑕᐅᑐᓗᑎᑦ, ᑎᑎᕋᖅᓗᑎᑦ ᐊᓚᒪᓗ ᐅᓂᒃᑲᐅᓯᐅᔭᖅᑕᕋᖅᓗᑎᑦ.” ᐱᑕᕆᐊᒃᓴᖏᑦ ᑐᑦᓴᓚᔪᑦ ᐅᐊᖏᑎᓗᑎᑦ ᐃᖅᑲᓚᓯᐊᕈᑎᓂᑦ ᖃᓄᐃᓕᑕᐅᕐᔪᓂᑦ, ᖃᐅᔨᓴᖅᑕᐅᓗᑎᑦ ᐅᑦᑐᕋᖅᑕᐅᓗᑎᓪᓗ ᐃᖅᑲᓚᓯᐊᕈᑕᐃᑦ, ᓄᐊᑕᐅᓗᑎᑦ ᐃᖅᑲᓇᑕᐅᕐᓂᖅᓂᑦ, ᑎᑎᕋᖅᓗᑎᑦ ᖃᐅᔨᓴᖅᑕᐅᓚᓴᖅᑐᓂᑦ, ᐅᐊᖏᑎᓗᑎᑦ ᓄᓇᕗᑦ ᐃᖅᑲᐃᐃᑦ, ᐊᓚᒪᓗ ᓇᓗᓇᐃᖅᑕᐅᓗᑎᑦ ᐅᑦᑯᓚᐃᓕᖅᑐᓂᑦ ᐊᓚᒪᓗ ᖃᐅᔨᓴᖅᑕᐅᓗᑎᑦ ᐱᓚᕆᐊᖃᖅᑐᓂᑦ.

ᑕᒪᐃᖅᑎ ᐅᒥᐊᖏᑦ ᐅᐊᖖᓇᖅᑐᓂᑦ ᖃᓇᑦ ᖃᓇᖖᓚᑕ ᐃᖅᑲᓗᓯᓚᖏᓪᓕᑦ ᑲᑐᔪᖅᑎᑦᓴᖅᑎᓚᖅᖃᖅ ᐊᐃᑐᑐᖅᓯᒪᓂᖅᑲᖖᓂᑦᑐᑦ 0 ᖃᑦᕈᑦᖖᓂᐊᖏᐃᓚᒥ ᑎᒍᒥᐊᓚᐊᓐᑦ ᐃᖅᑲᓗᑦᓚᔨᖅᑦ ᐊᓚᒪᓗ ᐃᒪᒥᐅᑕᓚᔨᖅᑦ ᖃᓇᑦᓗ. ᑦᖖᓇ ᖃᐅᔨᓴᐅᑎᐅᔨᖏᖅ ᖏᑦᒪᔭᐊᕐᑎᑕᐅᓯᓚᓴᑎᒍᑦ ᐊᑐᖅᑕᐅᓯᐃᖅᑐ. ᐅᐊᕐᑎᐅᔭᖅᑐᐅᔨᖏᖅ ᐃᒥᒪᓗ ᐊᑐᖅᑕᐅᓯᐃᖅᑐᓂᑦ ᐅᓇᓚᐃᓕᑕᐅᖏᑦᓗᓐᓂᑦ ᐅᐊᖏᑎᓄᐅᖏᑦᓗᓐᓂᑦ ᐅᒥᐊᖅᐸᐊᖅᑎᖃᑕᐅᓯᓪᓚᕐᓂᑦ ᐃᖅᑲᓚᓯᐊᕈᑎᓂᑦ ᐱᓕᕆᐊᖃᖅᑐᓂᑦ ᐃᖅᑲᓚᓯᐊᕈᑎᓂᑦ ᐊᓚᒪᓗ ᐃᖅᑲᓚᓯᐊᕈᑎᓂᑦ ᐃᖅᑲᓇᑕᐅᕐᓂᖅᓂᑦ ᐃᓕᓴᐅᓂᑦ ᐃᖅᑲᓚᓯᐊᕈᑎᓂᑦ ᐃᖅᑲᓇᑕᐅᕐᓂᖅᓂᑦ ᐃᖅᑲᓚᓯᐊᕈᑎ

ᐊᓗᓐᓯᐊᖅ 1. ᐅᐊᖏᓐᓇᒥ ᑲᓇᑕᑦ ᑲᓇᓐᓇᖏᑕ ᑕᕆᐅᖏᓚᓂᖅ ᐃᖅᑲᓗᓕᕆᔨᕐᓂ ᑲᑎᒪᔨᖅᑕᐅᓯᒪᔪᓐᓇᖅᑐᑦ ᐊᓚᒪᓗ ᐊᐃᕐᑐᖅᓯᒪᔪᑦ ᑐᑭᖏᓚᓂᒃᑯ ᖃᑦᑕᕐᓂᐊᓂᐊᓂᐊᖅᓯᓕᒫᑦ

ᖃᑦᑕᕐᓂᐊᓂᐊᖅᓂᖅ



ᑎᑎᖅᑐᒪᐅᕗᖅᓱᒪᓂᖅᓴ 2 - ᓇᓗᓇᐃᔨᐃᒍᑎᖏᑦ ᐅᓪᓗᒥᐅᕙᖅ ᐊᐅᓚᕆᒍᔪᑏᑦ ᐊᖏᖅᑭᓱᒪᓂᖏᑦ ᐊᕐᕌᒍᖅᓱᒪᓂᕆᔨᖅᓂᓂᓴᓂ 0 ᖃᓚᕋᓚᓐᓂᐊᖅᓂᖓᓪᓗᑎᑦ

ᐊᐅᓚᕆᓂᖓᓪᓗᑎᑦ ᖃᓄᐃᓚᓂᐅᕕᑐᐃᓚᓱᐊᕈᑎᑦ	ᓇᓗᓇᐃᔨᖅᓱᒪᓂᖏᑦ
ᑕᒪᐃᓐᓂ ᐃᖅᑲᓇᑕᐅᕙᓐᓇᖅᑐᑦ ᐊᒥᓱᓕᕆᓂᖏᑦ (TAC)	• ᒥᓂᑕᐅᕙᖅ ᓇᓗᓇᐃᖅᓱᓱᖅ ᑲᓇᑕᒥ ᖃᐅᔨᓂᓂ ᑕᒪᐃᓐᓂ ᐃᖅᑲᓇᑕᐅᕙᓐᓇᖅᑐᑦ ᐊᒥᓱᓕᕆᓂᕐᒪᑐᐃᓐᓇᖅᑐᑦᓂᓂ ᖃᓚᕋᓚᓐᓂᒃ
ᓯᐃᔨᓐᓱᑦ	• ᓯᐃᔨᓐᓱᖃᕆᐊᓕᑦ ᐃᖅᑲᓇᓱᐊᖅᓂᐊᖅᓗᑎᒍᑦ ᖃᐅᔨᓂᓂᒃ.
ᐅᒥᐊᖅᓯᐊᑦ	• ᓇᓗᓇᐃᖅᑕᐅᓱᒪᓗᒍᑦ ᐃᖅᑲᓇᓱᐊᖅᑭᓐᓇᐅᑎᓕᑕᑦ ᓯᐃᔨᓐᓱᒃᑐᒃ.
ᐅᒪᔨᐃᑦ, ᓱᑯᑎᐊᑯᓂᓐᓂᖏᑦ ᐊᕆᒪᓗ ᐃᖅᑲᓇᓱᐊᖅᑕᐅᔨᑦ	• ᐅᒪᔨᐃᑦ, ᐊᒥᓱᓂᖏᑦ ᐊᕆᒪᓗ ᓱᑯᑎᐊᑯᓂᓐᓂᖏᑦ ᐃᖅᑲᓇᓱᐊᖅᑕᐅᔨᓐᓇᖅᑐᑦ ᓇᓗᓇᐃᖅᑕᐅᓱᒪᕐᕉᑦ ᐊᖏᖅᑭᓱᒪᓂᖅᓴᓂᒃ ᓯᐃᔨᓐᓱᒃᑐᒃ ᐃᓕᓴᐅᑕᐅᓚᐅᖅᑐᑦ ᑕᒪᐃᓐᓂᒃ. • ᐃᓴᕐᐅᑎᖏᑦ ᖃᐅᔨᑎᐅᓕᖅᓇᒪᓐᓇᒃᑕ ᐱᓕᕆᖏᑦ ᐊᐸᔨᖏᑦᑕᑐᓂ ᓱᓇᒃᑯᑦᓴᓂ ᐃᓕᖅᑎᑕᐅᓱᒪᕐᕉᑦ ᐃᓕᓴᐅᑕᐅᓚᐅᖅᑐᑦ ᐊᕆᒪᓗ ᐃᓚᖏᐅᑕᐅᓚᐅᖅᑐᑦ ᐅᓇᓗᒥᒃ. • ᑯᑕᖏᑦ ᐊᖏᖅᑭᓕᕆᐊᖅᑕᐅᓴᐃᔨᑦ ᐊᑐᓕᖅᑎᑕᐅᓱᒪᑦ ᑕᒪᐃᓐᓇᒃ ᐅᐊᔨᑎᐊᑭᖅᓱᓇᖅ
ᐃᖅᑲᓇᓱᐊᖅᓇᖅ	• ᑭᓇᐅᔨᓕᕆᔨᖅᓂᐊᖅᑐᑦ ᓴᓂᖅᓭᖅᓚᐅᓂᖏᑦ ᑯᑕᓂᒃ ᐊᕆᒪᓗ ᐊᓴᐅᖏᖅᑎᑕᑐᒃ ᑯᑕᖅᑲᓂᑕᐅᕙᓐᓇᖅᑕᑦ, ᓴᓐᓄᐊᓕᓗ 1 - ᑎᓱᐸᓭ 31 (ᒪᓕᓕᑐᑦ ᓇᓗᓇᐃᖅᑕᐅᓱᒪᔨᑦ ᒪᑐᐅᓂᖅᑎᓐᓂᒃ ᐃᖅᑲᓇᓱᐊᖅᓇᖅᓯ). • ᐊᕐᕌᒍᖅᓱᒪᓂᖅᓴᓂᒃ 0B ᐊᖏᖅᑭᓱᒪᔨᑦ ᐃᖅᑲᓇᓱᐃᑦ ᐊᑭᑕᑐᖅᐅᔨᓄᑦ, ᓇᓗᓇᐃᖅᑕᐅᖅᑲᖅᑕᓂᐊᖅᑐᑦ ᖃᐅᔨᑎᐅᓴᖅᓯᒪᓂᖅᓴᖅᑎᑦ ᑯᑕᖅᑲᓂ ᐊᖏᖅᑭᓕᕆᓇᖅᑐᑦ.
ᑐᓴᖅᑎᑎᐊᑕᐅᓂᖅᓴᖅ ᒪᑐᓱᒪᓕᓂᓂᓚᖅᑐᖅ	• ᑐᓴᖅᑎᑎᐊᑕᐅᓂᖅᓴᖅ, ᖃᐅᔨᑎᐅᓴᖅᑎᑎᐊᑕᐅᓂᖅᓴᖅ, ᐅᕙᓪᓕᐊᓂᒃ ᐊᑐᓕᖅᑎᑕᐅᓱᒪᓂᖅᓴᖅ
ᐃᖅᑲᓇᓱᐊᖅᓂᖅᓴᖅ	• ᐃᒪᓐᓇᐃᑦ (ᓄᑖᖅᑲᐅᓱᒪᔨᑦ) • ᐊᓐᓴᐃᓚᑦ (ᐊᑭᖅᑭᖅᓱᒪᔨᑦ) • ᓄᓇᐊᑦ (ᐊᑭᖅᑭᖅᓱᒪᔨᑦ o ᓄᓇᐊᑦ ᓇᓗᓇᐃᓐᑦᑎᖅᑕᐅᓚᐅᖅᑐᑦ ᐊᑭᖅᑭᖅᓱᒪᓗᒍᑦ ᓄᓇᐊᑦ >ᑦᑕᓂᖅᓱᓕᓂᒃ ᐊᑐᓂ ᓄᓇᐊᖓᔪᑦ. • ᐃᖅᑲᓇᓱᐊᖅᑐᑦ ᐊᖓᑕᓂᕐᒪᐊᑕᐅᓂᖅᑐᑦ ᓇᓂᕆᐅᔨᖅᓂᖅᓴᖅᑐᑦ ᖃᐅᔨᑎᐅᓴᖅᑎᑎᐊᑕᐅᓚᐅᖅᑐᑦᓂᓂᖅ ᑕᒪᐃᓐᓇᒃᑯᑦᓴᓂᓂ • ᐱᓇᓱᐊᖅᓇᖅᓯᓂᒪᖅᑐᓂᒃ ᐱᓇᓱᐊᖅᑐᑦᓂᓚᖅᑐᒃ ᐊᓱᐅᔨᐅᓱᒪᔨᑦ ᓄᓇᐊᑦ ᐱᔨᐅᓇᓱᐊᖅᓂᐊᖅᑐᑦ. • ᐃᖅᑲᓇᓱᐊᖅᑎᑦ ᖃᑭᒪᑕᐅᔨᓚᐊᖅᑲᖅᓯᐅᑕᐅᓂᖅᓴᓂᐊᖅᑐᑦ ᐅᐊᑕᑎᑦ 72 ᑭᖓᔪᓕᓴᑦ ᐃᑲᖅᑕᓂᒃ ᐊᔪᓂᐅᑎᐅᓕᖅᑐᑦ

ᐊᐅᓪᓗᒍᓂᐊᕐᓂᖅ ᖃᓄᐃᓕᐅᕈᑎᐅᓯᐊᕈᑦ	ᓇᓗᓇᐃᔨᖅᑭᓯᓚᖓᓂᒃ
ᐃᖃᓗᓕᓯᐊᕐᓗᓂ ᑭᓴᓕᖅᑕᑎᖏᑦ	• ᐃᖃᓗᓕᓯᕿᓘᔨᐅᖏᑦᑐᖅ NSA ᐅᕐᓀᓕᓂᓐᓂᑦ ᓄᓇᓘᒃ ᐃᒪᖏᓐᓂ ᑭᓯᐊᓂ ᐊᖏᓕᕆᑎᐅᓕᐅᖅᑎᑦᓗᒍ ᐅᒪᒪᓕᖏᓐᓂᓄᑦ. • ᐃᖃᓗᓕᓯᐊᓚᐊᖃᖅᖏᑦᑐᑦ ᐊᖅᑲᐅᒪᔨᖅᓯᐅᑎᓂᒃ >19.9ᑦᑕ ᐃᒪᖅᑎᒃ <12 ᐅᖏᓕᓯᓂᓕᓂᑦ ᐃᒪᖅᒥ ᒪᐃᔨᑦᑐᑦ ᑲᓇᑕᑦ ᑲᓇᓐᖏᓐᓂ ᑕᓚᐅᖏᓐᓂᓂ. • ᐋᖅᑭᒃᓯᒪᔨᓄᑦ ᐃᖃᓗᓕᓯᓂᑎᓄᑦ ᒪᐃ 1-ᒥ ᐊᓖᒪᓗ ᑎᓯᖅᖃᐅᓂᒃ 31-ᒧᑦ, 20% ᑕᓚᐅᖅᒥ ᐅᐊᖅᑎᐅᖅᑲᓚᐊᓕᒃ. • ᐃᖃᓗᓕᓯᐊᖅᖃᐅᓕᐊᖅᖏᑦᑐᖅᓴᖅ ᐊᖃᕐᑐᖅᓯ,ᓂᖏᓂᒃ 0B ᓄᓗᐊᖅᓂᒃ ᐊᑐᖅᓗᑎᑦ ᓂᒋᐊᓂ 3°10’N ᐊᒃᑐᖅᐳ 1-ᒥ ᑎᑭᓪᓗᒍ ᑎᓯᖅᖃᐅᓂᒃ 31-ᒧᑦ. • ᐃᖃᓗᓕᓯᐊᓚᐊᖃᖅᖏᑦᑐᑦ ᐊᒃᓴᓇᔪᑦᓄᑦ ᐊᖃᕐᑐᖅᓯᒪᓂᖏᓂᒃ 0B ᓂᒋᐊᓂ 63°10’N ᐊᒃᑐᖅᐳ 1-ᒥ ᑎᑭᓪᓗᒍ ᑎᓯᖅᖃᐅᓂᒃ 31-ᒧᑦ ᑭᓯᐊᓂ ᐃᒪᖅᓂ ᐃᑎᓂᖅᑭᓂ >1372ᑦᑕ. • ᑎᓴᔪ ᐅᐟᓐ <<ᕐᑎᕆᖏᓐᓄ*, ᖃᑭᑭᖅᑕᓇᑦ ᐃᑭᕐᔾᔨᓂᓂ <<ᕐᑎᕆᖏᓐᓄ, ᐊᓖᒪᓗ Hᐊᕐᑕᓐ ᑕᓚᐅᖏᓐᓄ <<ᕐᑎᕆᖏᓐᓄ ᒪᑐᓯᒪᔨᖅ ᖃᓄᐃᓕᓚᐅᓂᓐᓂᐊᕐᓂᖏᓐᓄᑦᑐᓂ. • ᐊᖏᕐᑐᖅᓯᒪᓂᖏᓐᓄ 0A ᒪᑐᓯᒪᔨᖅ ᐅᒥᐊᕐᒥ ᐋᖅᑭᒃᓯᒪᔨᓄᑦ ᐃᖃᓗᓕᓯᐊᑭᑎᓂᓐᓂᓄᑦ ᓄᐃᓇᕐᓂᒃ 11 – ᑎᓯᓂᒃ 31-ᒧᑦ ᐊᖓᒪᓗ ᒪᑐᓯᒪᔨᖅ ᑕᒪᐃᓐᓂ ᐊᑐᓯᑦ ᐃᖃᓗᓕᓯᐊᑭᑎᓂᒃ ᑲᓐᓇᐅᐊᖓ 1 – ᒪᐃ 31 ᐊᖅᑲᒍᑕᒪᑦ. ᒪᑐᕆᐅᖏᓐᓄ ᐅᓪᓗᖅ ᐅᓐᓄᑕᐅᓕᒍᓐᓇᖅᑐᖅ ᒪᓕᕐᑐᑎᑦ ᓯᑯᑦ ᖃᓄᐃᓐᓂᕐᓂᑦ.
ᐱᕆᐅᓇᓯᖏᑦᑐᑦ ᐃᖃᓗᓄᕐᔪᑦ/ᐱᕆᐅᖃᓕᓂᑦ ᐊᖓᒪᓗ ᐃᑎᑕᐅᓂᖏᓐᓂ	• ᐃᖅᖃᖅᑎᐅᑕᐃᑦ ᐃᖃᓗᐃᑦ ᐱᓯᒪᔪᐅᕆᓚᐊᓪᓚᑦ (ᑭᓯᐊᓂ ᐅᖃᖅᓯᒪᒍᑎᑦ ᐊᓯᐊᔪᑦ ᖃᓄᐃᓚᓕᐅᑕᐅᕆᓚᐊᓪᓚᑦ ᓕᐃᓴᓐᓯᓴᖅᑐᓂ ᐃᓗᐊᓂ). • ᐊᓯᖏᑦᑐᑦ ᓇᓕᐊᑐᐃᓇᐃᑦ ᐃᖃᓗᐃᑦ ᐃᖅᖃᖅᑎᐅᑕᐅᖅᖏᑦᑐᑦ ᓴᖅᑯᑕᐅᕆᓚᐊᓪᓚᑦ ᐊᓖᑦᓗ, ᐅᒪᔪᑎᑦ, ᐋᓐᓂᖅᓱᖏᖅᑕᒪᓐᓇᖅᑐᑦ ᓴᓴᑯᑕᐅᕆᓚᐊᓪᓚᑦ. • ᐊᐅᓂ ᐱᕆᐅᓇᓯᖏᑦᑐᑦ ᐃᖃᓗᐃᑦ ᐱᕆᐅᔨᑦ ᐊᐅᓂ ᐃᖃᓕᓕᐊᖅᐊᓄᑦ ᐅᖏᓕᕈᐅᕆᓚᐊᖅᖏᑦᑐᑦ ᐅᖃᖅᖃᓕᒪᔨᑦ ᓴᓴᓐᓂᕐᓂᑦ ᐅᖃᖅᓯᒪᓂᖅᓱᓂᑦ ᖃᓄᐃᓕᓕᐅᑕᐅᔨᑦ.
ᑲᒪᒋᐅᐅᓂᖏᑦ ᐅᒪᔨᐃᑦ ᑎᑎᖅᑕᐅᓯᒪᔨᑦ ᐅᓪᓚᑦ ᐃᖃᓗᓕᓯᐊᕐᓗᓂ ᖃᓄᐃᓕᓕᐅᑕᐅᔨᑦ	• ᐃᒥᐊᓂ ᖃᓄᐃᓕᓕᐅᑕᐅᔨᑦ ᐃᖃᓗᓕᓯᐊᕐᓗᓂ ᐊᖓᒪᓗ ᐃᓕᓴᐅᔪᖅᑐᑦ ᐃᓕᓴᐅᔪᖅᑐᑦ ᐃᖃᓗᓕᓯᐊᕐᓗᓂ ᐊᖓᒪᓗ

ᐊᐅᓪᓚᑎᓂᖅ ᖃᓄᐃᓕᐅᑕᐅᓇᓱᐊᕐᓂᖅ	ᓇᓗᓇᐃᔨᖏᓪᓗᒥᓂᖅ
ᐃᖃᓗᓐᓯᓱᐊᖏᖅᑎ ᐅᓂᒃᑲᒥᐊᓕᕆᐊᓕᓚᖅᑐᑦ	• ᐊᐅᑦᓭᓪᓕᐅᖅᓂᓂᒥᓂᒃ ᐅᓂᒃᑲᖅ (ᑐᖓᓯᓱᐊᑎᓪᓗᐃ) ᑕᓚᐅᖅᒥ ᖃᑲᐅᔨᓴᖏᖅᑎ ᐊᐃᕐᕿᓚᔨᐅᓂᐊᖏᖅᑐᖅ. • ᖃᑲᐅᑕᒧᑦ ᑕᓚᐅᖅᒥ ᐅᓂᒃᑲᓭᐊᑦ (ᖃᑲᐅᑕᒧᖓᓱᐅᑦ). • ᑎᑎᕆᕐᐳᒃ ᐱᔨᓴᖏᑕᐅᕆᓭᓪᓗᑦᑐᓂ ᖃᑲᐅᑕᒧᑦ ᐊᓚᒪᓗ ᑐᓂᔨᐅᑐᓂ ᐃᖃᓗᓐᓯᓚᕐᓵᖅᑯᓂᓐᓂᒃ ᐊᓚᒪᓗ ᐃᒪᖅᒥᐅᑕᓚᕐᓵᖅᑯᓂᓐᓂᒃ ᐊᐅᑐᓂ ᐊᐅᑦᓭᓪᓕᐅᔨᓴᖏᖅᑎᓂᓐᓂᑦ. • ᓇᓗᓇᐃᔨᖏᓪᓗᒥᓂᖅᓯᓱᐊᕐᓂᑦ ᓯᓇᖅᑎᑦᓂᑦ ᓇᑎᑎᖅᑕᐅᔨᓴᖏᖅᑎᑦ. • ᐃᓯᓕᑎᓂᑐᐃᒍ ᐊᐅᑦᓭᓪᓗᒋᑦᑎᓂᒃ ᐅᓂᒃᑲᖅ (ᑎᐹᑦᑐᑦ) ᑐᓭᖅᑦᓂᒪᒥ ᐅᐊᖅᑎᓴᓂ ᑲᒪᑦᐸᓂᖏᓐᓄᑦ ᑲᑎᒪᔨᖏᑦ. • ᐊᓯᐅᓴᐅᓯᒪᔪᑦ ᐃᖃᓗᓐᓯᓱᐊᑎᓪᓗᒋᑦᑎᓂᒃ ᓇᑎᒪᐊᓕᕐᓂᒃ. • ᐱᔨᒥᓂᐅᔪᑦ ᐃᖃᓗᓐᓯᓱᐊᑎᓪᓗᒋᑦᑎᓂᒃ ᓇᑎᒪᐊᓕᕐᓂᒃ • >ᐃᓴᓂᑦ ᑕᑯᐅᐅᑎᓪᓗᑦᑐᓂ ᓇᑎᒪᐊᓕᕐᓂᒃ.
ᐅᒥᐊᖏᖅᑲ ᐅᐊᖏᑎᔨᔪᑎᑦ (VMS)	• ᐊᖓᕆᖅᑕᐅᓯᒪᔪᒥᒃ ᐊᓚᒪᓗ ᐊᑐᕐᕼᓘᒥᒃ ᐅᒥᐊᖏᖅᑲ ᐅᐊᖏᑎᔨᔪᑎᑎᖅᑲᐊᕐᓂᒃ. • ᐃᓗᐊᓂ NsA, ᐅᒥᐊᖏᔨᐊᑦ ᒪᖏᐹᓂᑐᒃ ᐅᒥᐊᖏᖅᑲ ᐅᐊᖏᑎᔨᔪᑎ ᖃᐅᔨᓴᖅᑲᐃᔨᔪᑎᑎᖅᑲᐊᕐᓇᑦ ᖃᓄᐃᒧᐊᓕᕆᑎᐅᑕᓯᒪᔪᓄᑦ ᐅᒥᐊᖏᖅᑎᒃᓯᓗᑎᑦ ᐃᓚᑎᑲᐊᑦ ᖃᓄᐃᒧᐊᓕᕆᑎᐅᑕᓯᒪᔪᑎᒍᑦ ᐊᐅᓪᓚᑎᓂᖅᓯᓱᐊᕐᓂᑦ
ᑕᓚᐅᖅᒥ ᐅᐊᖏᑎᓴᑦ	• 100% ᑕᓚᐅᖅᒥ ᐅᐊᖏᑎᓴᑕᖅᑲᐊᕐᓇᒃ ᐊᐃᔪᑕᖅᓯᒪᔪᓂᖅᓱᓂᒃ 0A ᑕᒪᒃᑭᓐᓄᒃ ᓄᑲᖅᓴᐅᔨᑦ ᐋᖅᑭᒃᓯᒪᔪᑦ ᐃᖃᓗᓐᓯᓱᐊᕐᓯᑦᑐᓂ ᐊᑐᖅᑕᐅᑎᓪᓗᑦᑐᓂ. • 100% ᑕᓚᐅᖅᒥ ᐅᐊᖏᑎᓴᑕᖅᑲᐊᕐᓇᒃ ᓄᑲᖅᓴᐅᔨᖏᑦ ᐃᖃᓗᓐᓯᓱᐊᕐᑎᓄᓐᓄᑦ ᐊᐃᔪᑕᖅᓯᒪᔪᓂᖅᓱᓂᒃ -B ᐊᖏᖃᒍᓪᓚᑎᖅᒥ ᐊᓚᒪᓗ ᐋᖅᑭᒃᓯᒪᔪᑦ ᐃᖃᓗᓐᓯᓱᐊᕐᓯᑦᑐᑎᖅᑐᓂ ᓴᓐᓄᐅᐊᓚ 1–ᒥ ᐊᓚᒪᓗ ᐃᐳᕐᑦ 30-ᒧᑦ. • ᑕᐃᒪᐃᑎᐊᖏᖅᑲᖅᑎᓪᓗᑦᑐᓂ, ᐊᐅᓪᓚᑎᓂᕐᓴᐅᔪᖅ ᐊᐅᑦᓭᓪᓕᐅᔨᓴᖏᖅᑲᖅᓯᒪᔪᓂᖅᓱᓂᒃ ᐃᖃᓗᓐᓯᓱᐊᕐᓇᐊᖏᖅᑐᖅᓱᓂᒃ ᑕᓚᐅᖅᒥ ᐅᐊᖏᑎᓴᐅᔨᔪᖅ ᐳᓯᐊᓂ ᐅᒥᐊᖏᖅᑎᒃᓯᕐᕆᓂᒃ.
ᐃᖃᓗᓐᓯᑕᒥᓂᕐ ᓄᓇᖅᑕᐅᑎᓪᓗᑦᑐᓂ ᐊᐅᓪᓚᔨᔪᓂᖅ	• ᓂᐅᔨᐅᓪᓗᑎᓂ ᑲᓇᑕᒥ ᓂᐅᔨᐅᔨᔪᓄᓇᖅᑐᖅᓱᓂᒃ ᐳᓯᐊᓂ ᑐᓭᑕᐊᖏᖅᑐᖅᓱᓂᒃ ᐅᐊᖏᑎᓴᒪᓕᒃ ᑕᑯᓄᓇᖅᑕᐅᑎᓪᓗᑦᑐᓂ. • ᓂᐅᔨᐅᓪᓗᑎᓂ ᑲᒋᓪᓕᒃ ᓄᓇᖅᑕᐅᑎᓂᒃ ᓂᐅᔨᐅᔨᔪᓄᓇᖅᑐᖅᓱᓂᒃ ᐳᓯᐊᓂ ᑐᓭᑕᐊᖏᖅᑐᖅᓱᓂᒃ ᐱᔪᓄᓇᖅᑐᑦᑕᐅᔨᔪᖅᑐᑦ ᑲᒪᒋᔭᖅᑐᓂ ᐊᐅᓪᓚᑎᓴᐅᔪᖅ ᐊᐅᑦᓭᓪᓕᐅᔨᓴᖏᖅᑎᓂᒃ ᐃᖃᓗᓐᓯᓱᐊᕐᓇᒃ ᐊᐅᓪᓚᑎᓂᖅᓯᓱᐊᕐᓂᑦ ᑲᒪᒋᔭᖅᑐᓂ ᑲᑎᒪᔨᖏᑦ ᐃᖃᓗᓐᓯᓱᐊᕐᓇᒃ ᐊᐅᓪᓚᑎᓂᖅᓯᓱᐊᕐᓂᑦ ᑲᒪᒋᔭᖅᑐᓂ

2019 ᐊᑯᑭᑐᑐ ᖃᓕᕋᓕᖏᑦIFMP ᑐᑭᓯᐊᕐᒋᑦ ᐊᓱᓴᖅᑐᓄᑦ

ᐃᓕᓐᓄ	ᐊᖅᑭᑦᐊᖅᓯᒪᔪᑦ ᑎᑎᕋᖅᓯᒪᔪᑦ/ᓇᓗᓇᐃᔨᖅᓯᒪᓂᖏᑦ ᐊᓱᓴᖅᑐᓄᑦ	ᐱᔪᔨᑎᖏᑦ ᐊᓱᓴᕋᓂᖏᓐᓄᓪᓄᑦ
1.3 ᐱᓯᓕᓕᐊᓂᑐᖃᖏᑦ ᐃᖃᓗᓕᔨᓂᕐᒧᑦ, 1.3.1 ᐊᓴᑐᖅᓯᒪᓂᖓ 0B	2017-ᒥᑦ, ᐊᐃᑦᑐᖅᓯᒪᓂᖓ 0B TAC ᖃᐅᓯᓚᐊᖅᑕᐅᓗᓂ 7,575t ᐅᑯᓄᖓ 2017 ᐊᓕᒪ 2018 ᐃᖃᓗᓕᓴᓴᓇᐅᔨᓂᑦ. ᑭᖑᓂᐊᒍᑦ TAC ᖃᐅᓯᓚᐊᖅᑎᓕᒥᑦ 0B ᐊᑐᓯᖅᑎᑕᐅᓴᓕᕐᓂ ᒥᓵᑕᒧᑦ ᒪᓂᕐᑕᒧᑦ ᐊᓕᒪ ᓇᓂᔨᐅᔨᓴᓇᖅᑐᑦ ᑲᕕᔪᐅᓴᖏᑦ ᑐᑭᓯᐊᕐᒋᑦᐃᖓᓂ ᐅᕝᕙ: http://www.dfo-mpo.gc.ca/decisions/index-eng.htm.	ᒪᓇᓕᕋᑦᓱ ᐊᓕᒪ ᓴᓄᓂᑕᐅᔨᓇᓂᑦ ᐃᓴᒪᑕᐅᓱᑦᑐᓂᑦ ᐊᑐᐃᓇᐅᓂᐊᖅᑐᑦ ᖃᕋᓴᖅᑎᓂᖓ ᐱᑕᖃᓂᖅᑐᑦ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐅᓪᓗᒥᒧᑦ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᑕᓪᓕ ᐃᓕᓐᓄ ᐊᑎᕐᓯᒪᒃᑯᒪᒃ.
1.3 ᐱᓯᓕᓕᐊᓂᑐᖃᖏᑦ ᐃᖃᓗᓕᔨᓂᕐᒧᑦ, 1.3.1 ᐊᓴᑐᖅᓯᒪᓂᖓ 0B	ᑯᑕᖓ ᐊᓕᒪ ᐃᖃᓗᓕᑕᐅᔨᓇᓂᑦ ᑐᓂᓯᔪᓐᓇᖅᑐᑦ ᖃᓄᐅᑎᓕᐅᔨᓴᖅᑐᑦ ᑐᖃᓕᓴᕐᓂᖓ ᐅᓯᒃᑐᒥᓂᖓ: http://www.dfo-mpo.gc.ca/stats/commercial/yrlist-eng.htm.	ᑲᑎᕐᑎᓂᒃ ᐊᓕᒪ ᐃᖃᓗᓕᑕᐅᔨᓇᓂᑦ ᑐᓂᓯᔪᓐᓇᖅᑐᑦ ᐊᑐᐃᓇᐅᓂᐊᖅᑐᑦ ᐃᓕᓴᕐᓂᕆᔨᐅᑎᓂᐊᖅᑐᑦ. ᑕᐃᒪᐃᓕᑎᓂᐊᕐᓗᑦ ᐅᓯᒃᑐᒥᓂᖓ ᐊᑐᐃᓇᐅᓂᐊᖅᑐᑦ ᐃᓕᓴᕐᓂᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐊᓕᒪ ᑎᑎᕋᖅᑕᐅᓯᒪᔪᑦ ᐊᓂᖅᑕᐅᓗᓂ IFMP-ᒥᑦ. ᖃᓕᕋᓴᕐᓂᖓ ᑕᐅᔨᓇᓂᑦ ᐊᑐᐃᓇᐅᓂᐊᖅᑐᑦ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐊᓕᒪ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐃᓕᓴᕐᓂᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐊᓕᒪ ᐱᓕᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐃᓕᓴᕐᓂᕆᔨᐅᑎᓂᐊᖅᑐᑦ ᐱᓕ

4.3.2 ᑲᓇᑕᐅᕐᓂ ᐃᒪᖏᓕᓂ ᐊᖏᒪᓂ ᓄᓇᖅᓯᓕᕐᓂ ᓴᓐᒥᕆᐅᓂᖏᓕᓂ ᑐᓴᖏᓂ	ᐅᑦᑐᐱᓚᒥ 2017, ᑲᓇᑕᐅᕐᑦ ᑲᓐᓇᐃᖅᓯᓪᓚᐅᖅᑐᑦ ᓯᒪᓕᓪᓕᐊᑦ ᑎᑭᑎᓯᓚᒥᑦ ᓴᓐᒥᕆᐅᓂᖅᑐᑦ 5% ᑲᓇᑕᐅᕐᑦ ᐃᒪᖏᓕᓂ ᐊᖏᒪᓂ ᓄᓇᖅᓯᓚᒥᑦ. ᓴᓭᒪᑐᖅᓯᑦ ᑲᓭᖏᖅᓯᒪᐃᓕᓇᖅᑐᑦ ᓴᓐᒥᕆᐅᓂᖅᑐᑦ 10%−ᖏᓕᓂ 2020−ᖃᓪᓗᖅᑲ. 2020 ᑐᑦᖅᑕᐅᓯᖅ ᑕᒪᑭᓂᓂ ᑲᑦᖅᑕᖏᑐ (ᑲᓇᑕᒥ ᐅᒪᔪᐃᑦ ᐊᓯᐲᖏᓕᑐᑦ ᑐᑦᖅᑐᒪ 1) ᐊᒪᒪ ᓯᓚᔪᐊᒥᑦ ᑐᑦᖅᑐᒪ ᑕᔾᔪᖅᑐᑦᑐᑦ ᑲᑎᒪᓂᖅᑐᑦ ᐊᓯᐲᖏᓕᑐᑦᑎᑎᓂᕐᓂᑦ Aichi ᑐᑦᖅᑕᖅ 11 ᐊᒪᒪ ᓯᓚᔪᐊᒥᑦ ᑲᑐᐲᖅᑎᑦᑕ ᑲᑎᓂᖅ 2030 ᐱᓕᓇᐊᖏᐃᐃ ᒪᑭᒪᑎᑦᖅᐅᔪᓂᓕ ᐱᓵᓕᓚᑎᑎᓂᕐᓂᑦ ᐊᑕᓂ ᑐᑦᖅᑐᒪ 14. 2017 ᐊᒪᒪ 2020 ᑐᑦᖅᑕᐅᓂᑦ ᐊᖏᖅᑕᐅᓯᒪᓚᒥᑦ ᐊᑐᖅᑕᐅᕐᑯᐅᓂᑦ ᑲᓇᑕᒥ ᐃᒪᖏᓕᓂ ᓄᓇᖅᓯᓚᒥᑦ ᑐᑦᖅᑕᖏᑐᓂᓂ. ᑐᑭᓯᒥᐊᑭᑎᑎᖅᓂᐅᓂᑦ ᓴᓐᓇᐃᕐᖅᓯᒪᓂᖏᓕᑦ ᐊᒪᒪ ᑐᑭᒪᑎᑦᖅᑐᑦ ᑲᑎᒪᓂᖅᑐᑦ ᐃᒪᖏᓕᓂ ᓄᓇᖅᓯᓚᒥᑦ ᑐᑦᖅᑕᖏᑐᓂᓂ ᐊᑐᐃᓕᓇᖅᑐᑦ ᐅᓪᓚᒥᑦ ᐊᑐᖅᑕᐅᓯᖅᑐᑦ ᐊᑐᐃᓕᓇᖅᑐᑦ ᐅᕐᓂᓂ http://www.dfo-mpo.gc.ca/oceans/conservation/index-eng.html. ᑎᑭᑖᓂᐊᖅᓯᒪᔭ ᑕᑖᓇ ᑐᑦᖅᑕᐅᓂᖅ, ᑲᓇᑕ ᓴᖅᑭᑎᑎᓂᕐᑦ ᐃᒪᐃᑦ ᓴᓐᒥᕆᐅᓂᖅᑐᑦ (MPAs) ᐊᒪᒪ “ᐊᓯᖏᑦᓗ ᐊᑦᑐᖅᑕᐅᓂᕐᑦ ᐱᐅᖅᓯᐊᖅᑕᐅᓂᖅᑐᑦᓴᓐᓂᖅᑐᑦ (“ᐊᓯᖏᑦ ᐊᑐᖅᑕᐅᓂᕐᑦ”), ᐅᖅᖃᑎᖅᖃᖅᑐᑎᑦ ᐱᓕᓇᖅᑎᑦᑐᓂᓕ, ᑲᖅᒪᐅᖅᑎᑦᑐᓂᓕ ᑲᑐᔾᔨᖃᑎᖅᑎᓂᓕ, ᐊᒪᒪ ᐊᓯᖏᑦ ᐱᐊᓕᓂᖅᑐᑦ ᐃᓇᓪᓚᒥᑦ ᑎᑎᐅᓂᕐᑦ. ᓇᓗᓇᐃᔨᖅᓯᒪᓂᖏᓕᑦ ᑕᑦᑯᐊ ᓴᓇᑦᑭᑎᑦ, ᐱᖅᓯᐅᑎᓗᒃ ᓇᓗᓇᐃᔨᖅᓯᒪᓂᖏᓕᑦ ᐊᑐᖅᑕᓪᓕᐊᑦᑐᓂᓕ ᐃᓇᖅᑐᓂᓕᖅᑐᑦ ᐃᖅᑲᓇᐅᑉ ᐊᑐᖅᑕᐅᓯᖅᑐᑦ ᐊᑐᐃᓕᓇᖅᑐᑦ ᐊᓯᖏᑦᓴᓐᓂᐊᖅᑐᑦ ᐃᖅᑲᓇᐅᑉ ᐊᑐᖅᑕᐅᓯᖅᑐᑦ ᐱᐊᓕᓂᖅᑐᑦ ᐊᑐᐃᓕᓇᖅᑐᑦ ᐅᕐᓂᓂ http://www.dfo-mpo.gc.ca/oceans/mpa-zpm-aoi-si-eng.html. ᐃᓗᐊᓂ NAFO ᓇᒋᔪᓂᖓᓕ 0 Hatton Basin, ᖃᑭᖅᑕᒍᔪᓕᐊᑦ ᐊᑯᑭᑦᑐᐃᔪᓗᓗ ᓄᓇᖅᓯᒪᑕ ᐃᑭᕋᓕᖓᓕ, ᐊᒪᒪ Disko Fan ᓴᓐᒥᕆᐅᓂᖅᑐᑦ ᒪᑐᐃᑦᑐᓂᓕ ᐃᖅᑲᓇᐅᑉᓴᓐᓂᖅᑐᑦ ᐊᑐᖅ ᐃᖅᑲᓇᐅᑉᓴᓐᓂᖅᑐᑦ ᒪᓕᓚᒥᑦ. ᑕᔪᑯᐊ ᓴᓐᒥᕆᐅᓂᖅᑐᑦ ᐃᑲᔪᖅᑭᑎᖅᑎᓂᕐᑦᑕᑦ ᑲᑎᒪᓂᖅᑐᑦ ᐃᖅᑲᓇᐅᑉᓴᓐᓂᖅᑐᑦ ᐃᖅᑲᓇᐅᑉᓴᓐᓂᖅᑐᑦ ᑐᑦᖅᑕᖏᑐᓂᓂ. ᓇᓗᓇᐃᔨᖅᓯᒪᓂᖏᓕᑦ ᑕᔪᑯᓇᓂ ᓴᓐᒥᕆᐅᓂᖅᑐᑦ ᐊᒪᒪ ᓴᓐᒥᕆᐅᓂᖅᑐᑦ ᐱᓚᒥᐊᓚᐅᖅᑕᖏᑐᓂ ᐊᑐᐃᓕᓇᖅᑐᑦ ᐅᕐᓂᓂ http://www.dfo-mpo.gc.ca/oceans/oeabcm-amcepz/refuges/index-eng.html	ᓄᑦᖅ ᐃᓕᓐᓄ ᐃᓕᓴᐅᕐᓂᖅ ᑕᑯᓴᓐᓄᑦᖅᑲ ᐃᓕᓴᐅᕐᓂᖅᑐᑦ ᐊᑯᑭᑦᑐᐃᔪᓗᓗ ᓄᓇᖅᓯᒪᑕ ᐃᑭᕋᓕᖓᓕ, ᐊᒪᒪ Disko Fan ᓴᓐᒥᕆᐅᓂᖅᑐᑦ.
4.4.6. ᖃᒪᐃᓇᖅᑕᐅᓯᒪᓚᒥᑦ, ᐊᓯᐅᓴᖅᐅᓯᒪᓚᒥᑦ ᐅᕐᖅᑕᓗᓂᕐᑦ ᐃᑎᑕᐅᓯᒪᓚᒥᑦ ᐃᖅᑲᓇᐅᑉᐸᑎᓂᕐᑦ	ᖃᒪᐃᓇᖅᑕᐅᓯᒪᓚᒥᑦ, ᐊᓯᐅᓴᖅᐅᓯᒪᓚᒥᑦ ᐅᕐᖅᑕᓗᓂᕐᑦ ᐃᑎᑕᐅᓯᒪᓚᒥᑦ ᐃᖅᑲᓇᐅᑉᐸᑎᓂᕐᑦ (ALDFG, ᐅᕐᖅᑕᓗᓂᕐᑦ “ᐃᑭᐊᖅᓯᐅᓯᒪᓚᒥᑦ ᐃᖅᑲᓇᐅᑉᐸᑎᓂᕐᑦ”) ᐃᒪᖏᑐ ᓴᓂᐅᓴᖅ ᐊᒪᒪ ᑐᑦᔪᑎᑐᓂᓴᖅᑐᑐᓂ, ᐊᒪᒪ ᑎᑭᔾᔪᑎᑐᓂᓴᖅᑐᑐᓂ ᐱᓕᓴᐅᓂᖅᑐᑦ ᐃᒪᖏᑐ ᐃᒪᖏᑐᓂ ᐃᒪᖏᑐᑎᑦᑕᐃᑦ ᐅᓪᓚᒥᑦ ᐊᒪᒪ ᐊᓯᖏᑦᓂ ᐃᒪᖏᑐᑎᑦᑕᐃᑦ ᐅᓪᓚᒥᑦ, ᐊᒪᒪ ᑎᑭᑭᓐᓇᐃᖅᑐᓂᓴᖅᑐᑐᓂ ᓴᓂᐅᓴᖅ ᑭᓴᐃᖅᑐᓂᓴᖅᑐᑐᓂ ᑲᓇᑕᒥ ᐃᒪᖏᑐᓂ ᐃᒪᖏᑐᑎᑦᑕᐃᑦ ᐅᓪᓚᒥᑦ G7 ᐊᖏᒪᓂᖅᑲᓪᓚᓂᖅᑐᑦ, ᑲᓇᑕ ᐱᖅᑲᓇᐅᑉᓪᓚᒪᓂᕐᓗᖅ ᓴᓪᓕᐅᐊᖅᑐ ᑎᑎᖅᑐᒪᐃᑦ ᖃᓄᐃᓕᖓᓂᓴᖅᑐᑐᓂ ᑕᖅᑭᐅᓂ, ᐃᒪᖏᑦᓂ, ᐊᒪᒪ ᓴᖏᓂᖅᖃᑎᖅᑎᓂᕐᑦᑐᑦ ᑲᑎᒪᓂᖅᑐᑦ ᐃᓚᖓᑕᐅᓂᖅᑐᑦ ᐃᓇᖅᑐᓂᓕᖅᑐᑦ ᐃᒪᖏᑐᓂ ᓴᓂᐅᓴᖅ. 2019-ᒥᑦ ᐊᑐᖅᑕᐅᑎᓂᕐᑦᑐᑦᑕᓂᕐᑦᑐᓂ ᑐᓴᖅᑎᑎᓂᕐᑦᑐᓂ ᐱᑕᖃᓚᐊᖃᖅᑎᑎᑕᐅᓯᒪᓚᒥᑦ ᑲᓪᓚᑕᐅᓯᒪᓚᒥᑦ ᐃᖅᑲᓇᐅᑉᐸᑎᓂᕐᑦ ᐃᓚᖓᑕᐅᓯᒪᓚᒥᑦ ᐊᓯᐅᓴᖅᐅᓯᒪᓚᒥᑦ ᐃᒪᖏᑐᓂ ᐃᒪᖏᑐᑎᑦᑕᐃᑦ ᐅᓪᓚᒥᑦ ᐃᒪᖏᑐᓂ ᐃᒪᖏᑐᑎᑦᑕᐃᑦ ᐅᓪᓚᒥᑦ ALDFG, ᑲᓇᑕ ᓴᖅᑎᑎᓂᕐᑦᑐᓂᓕᖅᑐᑦ ᐃᒪᖏᑐᓂ ᐃᒪᖏᑐᑎᑦᑕᐃᑦ ᐅᓪᓚᒥ

GHL IFMP ᑐᓴᖅᑎᑦᑎᓂᖅ

ᐊᑦᑭᒍ	ᑎᒍᒥᐊᖅᑎᐅᖅᑕᐅᔪᑦ	ᐊᐅᓪᓚᕐᔪᑎᖓ
2016-17	EAGSAC	ᑐᓂᔨᐅᓂᖏᑦ ᑐᓴᕐᓯᒍᑦ ᐃᖅᑭᒋᐊᖅᑕᐅᔪᓄᑦ ᐊᓘᒪ ᑎᑎᕆᕆᐊᓕᖓᖅᑕᐅᓕᒫᑦ IFMP ᐅᓐᓄᐊᑦ 18-19, 2017 EAGSAC ᑲᑎᒪᓂᖓ
2017-18	EAGSAC	ᑐᓂᔨᐅᓂᖏᑦ ᑐᓴᕐᓯᒍᑦ ᐃᖅᑭᒋᐊᖅᑕᐅᔪᓄᑦ ᐊᓘᒪ ᑎᑎᕆᕆᐊᓕᖓᖅᑕᐅᓕᒫᑦ IFMP ᐅᓐᓄᐊᑦ 23-24, 2017 EAGSAC ᑲᑎᒪᓂᖓ
2018-19	EAGSAC	ᑐᓂᔨᐅᓂᖏᑦ ᑐᓴᕐᓯᒍᑦ ᐃᖅᑭᒋᐊᖅᑕᐅᔪᓄᑦ ᐊᓘᒪ ᑎᑎᕆᕆᐊᓕᖓᖅᑕᐅᓕᒫᑦ ᐃᓐᓇᐊᑦ 19-20, 2019 EAGSAC ᑲᑎᒪᓂᖓ. ᑐᓂᐅᖅᑲᖅᑕᐅᓂᖏᑦ ᑭᖑᓪᓗᖅᓯᒪᑦ ᑎᑎᕆᕆᐊᓕᖓᖅᑕᐅᓕᒫᑦ ᐅᑯᓄᖓEAGSAC ᐃᓚᒋᔭᐅᔪᓄᑦ ᐅᖃᐅᓕᕆᖃᖅᑯᓂᐅᔪᓄᑦ (ᐃᒡᓗ 26, 2019). ᐅᖃᐅᓕᕆᔨᐃᑦ ᐱᔭᐅᔪᑦ ᐅᐊᓂ/ᓕᓐᓂᐊᓂ ᐃᐳᑐ 5, 2019.

ᐃᓕᓐᓇᐅᔪᒥᑦ EAGSAC ᐃᓕᒧᔨᐅᔪᑦ

ᐱᔪᒪᓂᖅᓱ ᖃᓄᐃᑦᑐᓂᖅᓱ	ᐃᓕᒧᔨᐅᔪᖅ ᑲᑐᔾᔨᖃᑎᒋᓴᔪᖅ / ᐱᔪᒪᓂᖃᖅᑐᖅ ᑎᒥᐅᔪᖅ
ᐊᐅᓪᓚᑦᑎᑲᑎᒋᑐᑦ ᑲᑐᔾᔨᖃᑎᒋᓪᑦ	ᓄᓇᕗᒃ ᐃᒪᖃᐅᑕᓂᑦ ᑕᐅᒪᔪᓂᑦ ᑲᑎᒪᔨᖓᑦ (NMRWB)
	ᓄᓇᑲᓗᒥ ᐃᓄᐃᑦ ᐅᒪᔪᓂᑦ ᑎᑎᕋᖕᖏᓐᓄ
	ᓄᓇᑲᓗᒥ ᐅᒪᔪᓪᓚᔨᖕᐊᐳᑦ ᑲᑎᒪᔨᖓᑦ (NWMB)
ᐱᔪᓇᐅᑎᓂᑦ ᑎᒍᒥᐊᖅᑕᑦ	ᓄᓇᑲᑦ ᑐᓐᖏᕕᒃ ᑎᒥᖓᑦ
	ᒪᑭᖕᕕᒃ ᑯᐊᐳᓴᓴᓐ
ᐃᒪᖅᓯᓪᓗᒥ ᓄᓇᓕᑦ	ᖁᑭᖅᑕᑐᓗᒥ ᐅᒪᔪᓪᓚᔨᐳᑯᑦ ᑲᑎᒪᔨᖓᑦ
ᑭᓇᐅᔭᓕᐅᕋᓱᓂᖏᓪᒧᑦ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑲᑐᔾᔨᖃᑎᒋᑦ	ᐊᑦᓯᓴᑎᒃ ᐃᖅᑲᒥᐅᑕᓂᑦ ᐃᖅᑲᓐᓂᑦ ᑲᑎᒪᔨᑦ
	ᐅᑭᐅᑕᖅᑐᒥᑦ ᑲᑐᔾᔨᖃᑎᒋᑦ
	ᓄᓇᑲᓗᒥ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ
ᑭᓇᐅᔭᓕᐅᕋᓱᓂᖏᓪᒧᑦ ᐃᖅᑲᓗᑎᒃᓯᒪᓕᒃᑐᑦ/ᐃᖅᑲᓕᑕᓂᑕᓂᐅᔪᑦ	ᐊᓂᔪᓐᓇ ᑕᐃᓕ
	ᐅᑭᐅᑕᖅᑐᒥᑦ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑲᑐᔾᔨᖃᑎᒋᑦ
	ᖁᑭᖅᑕᑦᓗᒥ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ
	ᐳᕕᐃᓐᓇᓕᒃ ᒪᔪᐸᑎᒃ
	ᑭᖕᕕᒃᑐᑦᑎᓐᓄᑦ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑭᖕᕕᒃᑐᑦᑎᓐᓄᑦ
	ᐅᑭᐅᑕᖅᑐᒥᑦ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑲᑐᔾᔨᖃᑎᒋᑦ
	ᐊᓂᔪᓐᓇ ᑕᐃᓕᒧᔨᐅᔪᑦ
	ᐅᑭᐅᑕᖅᑐᒥᑦ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ
	ᐃᖅᑲᓐᓂᑦ, ᓂᖅᑭᑦ ᐊᓯᒪ ᐃᖅᑲᓂᕆᔨᖅᑕᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ
	ᕿᐊᖅᓯ ᔪᖅᐃᓴ ᑭᖕᓴᖅᑲᓐᓂᖅᑐᑦ ᐃᖅᑲᓕᑕᓂᖏᓪᒧᑦ
	ᐊᐳᑎᑎᒥ ᑲᒪᑦᑎᐊᖅᑐᒥ ᖃᐅᓕᑕᓂᓂᐅᔪᑦ
	ᐃᖅᑲᐃᐃᑦ, ᓂᖅᑭᑦ ᐊᓯᒪ ᐃᖅᑲᓂᕆᔨᖅᑕᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ
	ᕿᑎᔾᔨᐊᓐ ᑕᐅᖅᓱᖃᑎᒋᓪᓂᖅ
	ᐊᐳᑎᒥ ᑲᒪᑦᑎᐊᖅᑐᒥ ᖃᐅᓕᑕᓂᓂᐅᔪᑦ
	ᐃᓄᐃᑦ ᐊᓯᒪ ᐃᖅᑲᓐᓂᑦ, ᓂᖅᑭᑦ ᐊᓯᒪ ᐃᖅᑲᓂᕆᔨᖅᑕᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ
	ᐃᓕᓴᕐᔪᐊᓘᓐ ᐃᓕᓴᕐᔪᐊᓘᓐ ᐃᓕᓴᕐᔪᐊᓘᓐ
	ᐃᓄᐃᑦ ᐊᓯᒪ ᐃᖅᑲᓐᓂᑦ, ᓂᖅᑭᑦ ᐊᓯᒪ ᐃᖅᑲᓂᕆᔨᖅᑕᑦ ᑲᑐᔾᔨᖃᑎᒋᓴᖓᑦ

ᐱᓕᕆᔨᒃᑲᓇᑕᒥᓂᖅ ᑲᑎᒪᔨᑦᑎᐊᓂᖅ	ᐊᑐᐊᓂᖅ
	ᐱᓕᕆᔨᒃᑲᓇᑕᒥᓂᖅ ᐱᖓᓱᒃᑐᒃᑯᑦ ᐱᓕᕆᔨᒃᑲᓇᑕᒥᓂᖅ
	ᐃᓚᖓᔾᔪᑎᒃᑲᓇᑕᒥᓂᖅ

ᑐᓂᓴᐅᔨᓪᓕ ᐅᑭᖏᑐ

ᓄᓇᑭᒥᒃ ᐅᒪᔨᓕᕆᔨᕐᕕᐊᖅᑭᑦ ᑲᑎᒪᔨᖏᑦ ᐊᒻᒪ ᓄᓇᕗᒻᒥ ᐃᒪᖏᓂᑦ ᐅᒪᔨᓄᑦ ᑲᑎᒪᔨᖏᑦ

ᐅᔭᖏᑐ

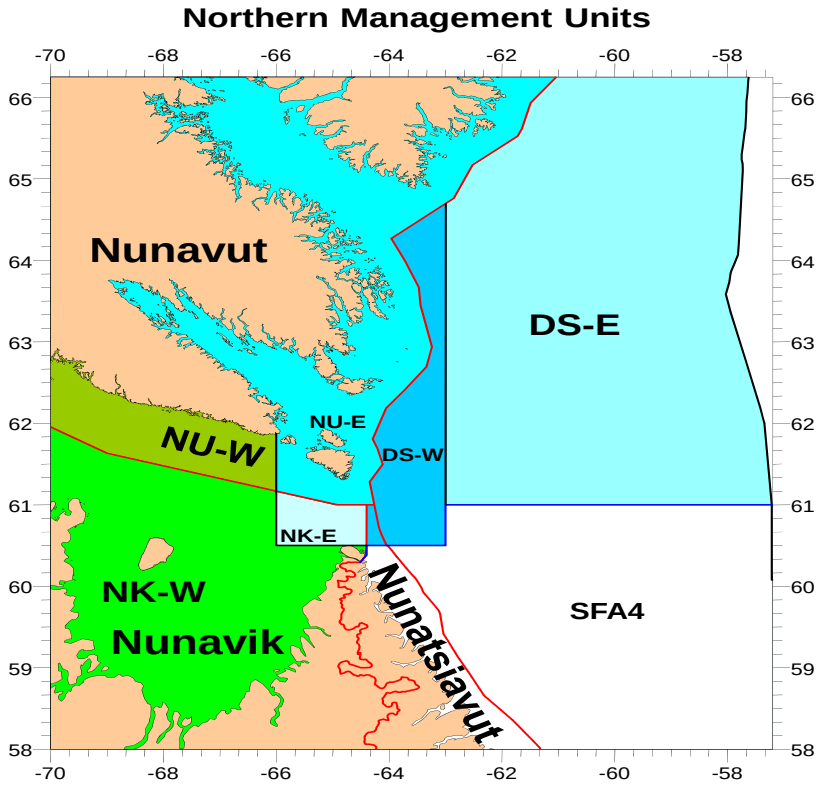
ᑐᑭᓯᐊᕈᓐᓇᑦ: ᐃᓱᒪᓕᐅᓂᖓᔪᑦ: X ᐊᑐᓕᖁᓴᐅᔨᖅ: X

ᐱᓕᕆᖏᑐ: 1) ᐊᐅᓪᓚᕆᓂᖓᔪᑦ ᐊᑐᖅᑕᐅᔨᑦ ᓄᓇᕗᒻᒥᑦ (NU) ᐊᒻᒪ ᓄᓇᕗᒃ (NK) ᐃᓕᓴᐅᖅᑲᖅᑕᐅᕕᒃᓴᖏᑦ NU/NK ᑲᓇᐅᖏᑐ (NU/NKE) ᐊᒻᒪ ᖃᑭᖅᑕᒋᔭᑦ ᐊᑯᑭᑐᐃᓐᓂ ᓄᓇᖏᓂ ᐃᑭᕐᓴᖏᑐ ᐱᖏᓐᓄᖏᑐ (DSW); 2) ᐋᖅᑕᐅᓂᖏᑦ ᐱᓵᐅᒋᐊᖏᑦᑐᑦ ᐱᓵᐅᔨᑦ ᓇᓗᓇᐃᖅᑕᐅᓂᖏᓐᓄᑦ ᐊᖅᓴᖁᓕᒥᒥ NU/NKE

ᓄᓇᖏᐊᖅ:

ᑐᖏᔨᖅᑐᑦ – ᑲᓇᐅᖏᑐᓂ ᖃᐅᔨᓴᖏᐃᔨᖅ ᑭᓕᓂᖏ

ᑐᖏᔨᑭᔨᑦᑐᑦ – ᐱᖏᓐᓄᖏᑐᓂ ᖃᐅᔨᓴᖏᐃᔨᖅ ᑭᓕᓂᖏ



ᓇᓗᓇᐃᔭᖅᓱᒪᓂᖓᓂ

ᒪᖁᑭᐃᖏᔨ ᑭᖏᔨᐸᐃᑦ ᐅᒪᔨᐃᑦ (*Pandalus montagui* ᐊᒻᒪ *Pandalus borealis*) ᓴᖅᑭᕋᖅᑕᖅᑐᑦ ᐅᑭᐅᑕᖅᑐᒥᑦ ᑭᖏᔨᐸᓂᐊᓂᕆᒥᑦ ᐃᖅᑲᓇᑕᓂᓂᕆᒥᑦ ᖃᑭᖅᑕᒋᔭᑦ ᐊᑯᑭᑐᐃᓐᓇ ᓄᓇᖏᓂ ᐃᑭᕐᓴᖏᓐᓄᑦ ᐊᒻᒪ ᑲᓇᐅᖏᑐᓂ ᖃᑭᖅᑕᒋᔭᑦ ᔨᐸᓐᓄᑦ ᓄᓇᕗᒻᒥ (NSA) ᐊᒻᒪ ᓄᓇᕗᒃ ᐃᒪᖏᑐ (NMR). ᑲᑎᑐᑎᑦ ᑭᖏᔨᐸᑦᑕᐅᔨᓐᓇᖅᑐᑦ (TAC) ᐊᑐᓂ ᑭᖏᔨᐸᓄᑦ ᒪᓕᒋᐛᒃ ᐊᔨᐱᓐᖏᑦᑐᒃ ᖃᐅᔨᓴᓂᖓᔪᑦ ᑭᓕᓂᖏᑦ (ᑲᓇᐅᖏᑐ ᐊᒻᒪ ᐱᖏᓐᓄᖏᑐ), ᐊᒻᒪ ᑐᓂᐅᖅᑲᖅᑕᐅᓕᖅᑕᕐᓂᑎᑦ ᐊᐅᓪᓚᑕᐅᔨᓄᑦ ᒪᓕᑐᑎᑦ ᓇᓗᓇᐃᖅᑲᓇᔨᓂᑦ ᑲᑎᑎᖅᑎᓐᓂᖓᔪᑦ ᐊᖅᑭᑕᐅᓯᒪᔨᓂᑦ.

ᐅᓇ ᑐᓴᕐᕋᐅᑦ ᑲᑎᒪᓕᓄᑦ ᐱᓱᕈᑎᒃᑯᑐᖅ ᐊᐅᓚᕐᕋᑎᓂᒧᑦ ᑭᑦᓂᖓᓂᑦ ᐊᐃᕐᑎᐊᓂᑦ ᓄᓇᕗᑦ / ᓄᓇᕃᖅ ᑲᓇᖏᓐᓂ (NU/NKE), ᐊᓪᒪ ᖃᑭᖅᑕᒍᔪᑦ ᐊᑯᑭᑐᑐᐃᓐᓗ ᓄᓇᖏᑕ ᐃᑭᕐᓴᖏᓐ (DSW) ᐊᐅᓚᑕᐅᓱᑦ ᑲᓇᖏᓐᓂᓂ ᖃᐅᓕᓴᕐᕕᐅᓱᑦ ᑭᑦᓂᖓᓂ (EAZ), ᐊᐅᓚᓂᕆᑯᑐᖅ ᓇᓗᓇᐃᓴᐃᓱᖅ ᓄᓇᕗᒻᒥᐊᕐᑐᕆᔨᓱᓂᑦ ᐃᒪᓕᕆᓂ ᓄᓇᕗᑦ ᐊᓪᒪ ᓄᓇᕃᖅ ᐊᓪᒪ ᐃᒪᐃᓐᓇᕐᒥᑦ. ᓄᓇᕗᑦ ᐊᓪᒪ ᓄᓇᕃᖅ ᐃᓕᓴᕐᕕᐅᕕᐅᓱᒪᓱᑦ ᐊᑐᓂ ᒐᕆᔨᓇᓂ ᐊᐅᓚᑕᐅᓱᓂᑦ, ᓇᓗᓇᐃᖅᑕᐅᓱᒪᓱᑦ ᐊᕐᔪᒃᓴᕐᑐᕆᔨᑦ ᐃᓱᒪᓕᐅᕈᑎᓂᑦ / ᐊᑐᓕᕆᔨᐅᓱᑦ ᐱᓕᕆᓚᐊᓂᕐᓂᑦ ᐅᒪᒪᓕᓚᓴᓂᑦ ᑲᑎᒪᓕᐅᕗᑎᓂᑦ ᐱᓕᕆᓚᐊᓂᕐᓂᑦ ᐅᒪᒪᓕᐅᕗᑎᓂᑦ DSW. ᖃᑭᖅᑕᒍᔪᑦ ᐊᔨᑭᑐᑐᐃᓐᓗ ᓄᓇᕗᑦ ᐃᓕᓴᕐᕋᐅᑦ ᖃᐅᔨᓴᖅᑎᑦᓯᓂᖅ ᐅᖓᓂ ᑐᓴᕐᕋᐅᒪᒥᑦ.

ᐊᑐᓂ ᐊᐅᓚᑕᐅᓱᑦ ᑐᓂᓯᐅᓯᒪᓱᑦ ᓇᒪᒥᓂᖅ ᑭᖑᓪᓕᔾᔪᑕᖅᑕᖏᓐᓂᓂ ᑯᑕᖏᓂᑦ ᐊᒪᓚᑦ, ᐃᓕᓴᕐᕕᐅᑦᑎᓂᑦ, ᒪᑕᓪᓂᑦ ᖃᓄᐃᓕᓴᕐᕕᐅᑦᑎᓂᓪᓕᐊᓂ ᓇᓗᓇᐃᖅᑕᐅᓯᒪᓱᑦ ᑕᒪᓐᓇᑦᑎᓂᓪᓕᐊᓂ ᓴᓇᔨᓂᖓᓂ. ᓴᓇᔨᓂᖓᓂ ᑎᒍᒥᐊᖅᑐᑦ ᑕᐃᒪᐃᓐᓇᖅᑐᑦ ᐃᓐᓇᐅᑉ ᐊᐅᓚᑕᐅᓱ

ᑲᑎᒪᓕᑦ 2012-ᒥᑦ ᑎᑎᕃᑲᕆᔨᑦ ᐅᑦᔪᕃᕆᕐᔪᑦ ᐅᑦᑐᐱᓗ 2 ᐊᓪᒪ 12 ᔪᖅᑭᓚᓱᕆᔪᑎᒃ, ᐊᓪᒪ ᐊᖅᑭᕃᑕᐅᑦᔪᓂᒃ ᑎᓴᒪᐊᕃᑕᐅᓴᒪᐊᕃᕆᕃᕆᔪᑦ ᐊᓕᕃᑲᑎᑕᐅᔨᓂ ᔪᑭᒧᐊᕃᑎᑕᐅᔨᓂ ᐃᕃᑲᔪᓚᓱᓂᒧᑦ ᐃᓕᓴᕐᕋᕃᑕᐅᔭᓕᔨᕃᕆ ᐅᑭᐅᑕᕃᔪᑎ ᑭᖑᓪᓕᖅᑐᑦ ᐅᕐᒥᓂ DSW. ᐃᕃᑲᔪᑕᐅᑎᖅᑭᑦᔪᑦ ᐃᓕᓴᕐᕋᕃᑕᐅᓂᖅᑐᑦ ᐃᓂᖓᑦᓴᕐᕋᕃᑕᐅᓂᖅᑐᑦ ᐃᓕᓴᕐᕋᕃᑕᐅᔨᓂᖅᑐᑦ ᓴᖅᑭᑎᑦᓴᕐᕋᕃᑕᐅᔨᓂᖅᑐᑦ ᓴᖅᑭᑎᑦᔪᓂ 2013 ᑭᓚᖓᓂ ᐊᓴᔨᕃᑕᐅᓂᖓ, ᐅᓇᐅᑎᓪᓗᔪ SFA 3 (ᐃᓐᓇ ᐱᖓᓐᓇᖓ ᕃᑲᐅᔨᓴᕆᔭᖅ ᑭᓚᖓᓂ) ᑭᖑᓪᓕᐊᑦ ᐃᓚᐅᕃᑲᕃᑕᐅᓂᖅᑐᑦ ᐃᓴᒪᑎᕆᐅᓚᐅᕃᑲᔪᑦ ᐱᕆᐅᑎᖓᑎᑐᓂᑦ ᐱᕃᑲᓴᓪᓗᑎ ᐃᓕᓴᕐᕋᕃᑕᐅᓂᖅᑐᑦ ᐱᕆᐅᑎᖓᑎᑐᓂᑦ ᐱᖓᓱᐊᕃᑕᐅᓂᖅᑐᑦ ᓴᖅᑭᑎᑦᔪᓂ ᐃᓕᓴᕐᕋᕃᑕᐅᓂᖅᑐᑦ ᐊᓕᕃᑲᑎᑕᐅᔨᓂ ᐱᖓᓱᐊᕃᑕᐅᓂᖅᑐᑦ. ᐱᕆᐅᑎᖓᑎᑐᓂᑦ ᐱᕆᐅᔨᓂᑦ ᐅᓇᐅᑲᖅᑕᐅᓴᒪᔨᕃᑕᖅᑎᖓᑎᑐᓂᑦᔪᖅ ᐅᓪᓚᐊᓄᑐᐃᓐᓇᖅ ᐅᒪᔨᓄᑦ ᐅᐅᑕᐅᔪᓂ WAZ.

ᐃᓐᓇᑕᒪᐅ ᐃᓐᓇᐅᔨᓕᑦ ᐱᕆᐅᑎᖓᑎᑐᓂᑦ ᐱᕆᐅᕃᖅᑕᕃᑐᑦ ᐃᓕᓴᕐᕋᕃᑕᐅᓂᔪᓕᑦ ᐃᒪᒃᓄᓇᕆᑦ ᕿᑭᖅᑲᒍᓱᑦ ᐊᑯᑭᑕᑐᐃᓪᓗᔪ ᓄᓇᖕᓂᒃᓴᓂ, ᐊᓪᒪ ᑐᑭᒧᐊᕃᑎᑕᐅᓴᒪᐊᕃᕆᕃᕆᔪᖅ ᐃᓗᐊᓄᑦ ᐊᖏᕈᑕᐅᔨᓴᒪᐊᕃᕆᕃᕆᔪᑦ. ᐱᑕᕐᓇᐃᓕᐅᔪᖅᑕᐅᓂᖅᑐᑦ ᐊᓕᕃᑲᑎᑕᐅᔨᓂᖅᑐᑦ ᐃᓕᓴᕐᕋᕃᑕᐅ

ᖃᓄᐃᓕᓐᓂᖅ ᐃᖅᑲᓇᐃᔨᖅ
BAFFIN FISHERIES

Société Makivik
Makivik Corporation

ᖃᐅᔨᓴᓕᓕᒃ ᑯᐊᐃᓕᓐᓂᒃ
ᓇᓗᓀᓐᓂᒃ ᑭᓇᐅᔭᓕᕆᔨᒃ
QIKIQTAALUK CORPORATION
& Group of Companies

LΔ 14, 2019

ᑕᓂᐅᓪ ᓄᑦᓱᒃ
ᐃᒃᓯᓲᐅᑦᖏᑦ
ᓄᓇᕗᑦ ᐅᒪᔨᓕᓚᔨᕐᔪᐊᒃᑯᑦ ᑲᑎᒪᔨᖅᑎᑦ
ᑎᑎᕆᒃᑯᐃᘁᖏᑲ 1379
ᐃᖃᓗᐃᑦ, ᓄᓇᕗᑦ
X0A 0H0

ᐅᖃᖅᑕᐅᔪᓐ ᐅᓱᖅᑎᑎᕆᔨᒃ : ᓄᑦᓱᒃ ᐊᒻᒪ ᒪᓱᐃᓐᓇ:

ᐱᔨᔪᑎᖅᕐᖓ : ᐃᖃᓗᓯᓂᖅ ᓄᓇᕗᒻᒥ ᐊᒻᒪ ᓄᓇᘁᒻᒥ ᑐᓂᔨᔪᐊᓂᖅᓂᒃᓂᑦ ᑯᑦᓯᖏᓐᓂᑦ ᖃᐅᖅᑲᓐᓂᑦ ᐃᑎᕐᕕᖏᓐᓂᑦ (DSW) ᐊᒻᒪ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᑲᓇᖅᓯᓐᓂ (NU/NKE)

ᐅᑯᐊ ᑎᑎᖅᑲᐃᑦ ᐊᑎᖏᓐᓂᑦ ᓇᓗᓇᐃᖅᓯᕐᓂᑦ ᐃᑲᔪᖅᑐᐃᓂᖏᓐᓂᑦ ᓄᓇᕗᑎᒻᒪᓗ ᑭᙳᕐᓂᒃᑯᑦᓂᒃᑎᑦ ᑎᒥᐅᔨᖏᓐᓂᑦ ᑐᓴᕐᕕᑎᓕᔭᒃᑯᑦᓂᑦ ᐃᖅᑲᓇᙳᖅᖏᓐᓂᑦ ᓄᓇᕗᑎᒻᒪᓗ ᑐᓂᔨᔪᐊᓂᖅᓂᒃᑯᑦᓂᑦ ᐃᖅᑲᓇᙳᖅᖏᓐᓂᑦ ᓄᓇᕗᑎᒻᒪᓗ ᐅᓂᒃᓴᖅᑎᑦ ᐃᒪᓕᖅᖏᓐᓂᑦ ᐃᓕᓴᕐᓂᖅ ᐊᑐᐊᓂᖅᓂᑦ ᖃᐅᖅᑲᓐᓂᑦ ᐃᑭᕐᕿᖓᓂᖅ ᐱᖓᓱᖏᓐᓂᑦ ᐃᓕᓴᕐᓂᖅ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᑲᓇᖅᓯᓐᓂ.

ᑲᓇᖅᓯᓐᓂ ᖃᐅᔨᓴᖅᑕᐅᔪᓕᕆᓕ ᐊᐢᑎᖅᓯᓐᓂᑦ ᐅᓂᔪᓂᑦ ᐅᓪᓗᓈᑦᑎᐱᐅᔨᕆᓕᑦ ᖃᐅᔨᓴᓂᐸᑎᑎᕐᓂᒃᓂᑦ ᐊᒻᒪᓗ ᐊᐱᖅᑐᖅᑕᐅᔪᓕᖅᑐᑦ ᐊᔨᑎᖏᓐᓂᑕᐅᓂᑦ (ᖃᐅᖅᑲᓐᓂᑦ ᐃᑭᕐᓴᖅᓯᓐᓂᑦ, ᖃᐅᖅᑲᓐᓂᑦ ᐃᑭᕐᓴᖅᓯᓐᓂᑦ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᑲᓇᖅᓯᓐᓂᑦ) ᐊᐅᓪᓚᑎᓂᖏᓐᓂᑦ ᐱᔨᕆᐊᑎᒥᓗᑦ. ᓇᓗᓇᐃᖅᑯᓪᓚᐃᑦ ᐊᑯᓂᖏᓐᓂᑦ ᖃᐅᖅᑲᓐᓂᑦ ᐃᑭᕐᓴᖅᓯᓐᓂᑦ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᑲᓇᖅᓯᓐᓂᑦ ᓇᓗᓇᐃᖅᑯᓪᓚᐃᑦ ᑭᓪᓕᔪᖅᖏᓐᓂᑦ ᐃᒪᘁᓐᓇᒻᒪ ᐊᑯᓂᖏᓐᓂᑦ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᐊᖏᔪᑎᓂᖏᓐᓂᑦ ᑭᓪᓕᔪᖅᖏᓐᓂᑦ. ᑭᓯᐊᓂ ᓄᓇᕗᑎᑦᓂ ᐱᔨᕆᐊᑎᒥᓗᑦ ᐃᖃᓗᓯᓂᖅᓂᑦ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᓇᓗᓇᐃᖅᑯᓪᓚᐃᑦ ᐊᐱᖅᑐᖅᑕᐅᔪᓕᖅᑐᑦ ᐃᑭᕐᓴᖅᓯᓐᓂᑦ ᐊᒻᒪᓗ ᓄᓇᕗᑦ ᓄᓇᘁᐅᑦ ᐃᓕᓴᕐᓂᖅ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅᑦ ᓇᓗᓇᐃᖅᑯᓪᓚᐃᑦ ᐊᐱᖅᑐᖅᑕᐅᔪᓕᖅᑐᑦ ᐃᑭᕐᓴᖅᓯᓐᓂᑦ ᐊᒻᒪᓗ ᓄᓇᕗᑦ/ᓄᓇᘁᐅ

ᑕᒪᓐᓇ ᐊᓯᔾᔨᓂᖅ ᑐᑭᖅᑐᖅ ᓄᓇᕗᑦ ᐊᒻᒪ ᓄᓇᕗᒻᒥ ᐃᖅᓚᓱᓐᓂᑦᓯᒪᑦ ᐱᓕᕆᔨᖅᑕᕐᒪᑦ ᑲᑎᑦᑎᑦ ᑐᓂᔾᔪᓯᐊᓂᖅᑐᓂᖅ
ᐃᖅᓚᖅᑕᐅᓚᑦᖅᓴᑦ ᓇᓕᐊᖅᓂᑐᐃᓐᓇᖅ ᒥᑦᓂᓇᓕᖅᑐᑦ ᐊᐃᑲᐅᖅᑕᐅᓯᒪᔨᖅᓂᖅ. ᐊᑐᓂ ᑲᑦᔪᓂᐅᔾᔨᑦ ᐊᑕᐅᓯᕐᒥᑦ
ᑐᓂᔾᔪᓯᐊᖅᓕᕆᖅᓴᑦ, ᑐᓴᕐᓯᒪᓕᒫᑦ ᖃᖅᑎᑦᑎᓂᑦ ᒥᑦᓂᓇᓕᖅᑐᑦ ᐊᐅᓚᕐᓂᖅᑐᓂᖅ ᐊᐅᓚᓯᖅᑐᓂᖅ ᐊᑐᐃᓐᓇᖅᑐᑦ
ᐊᔪᕐᓇᐅᔾᔨᑦ ᐊᖏᔪᑎᓕᒫᑦ ᐱᔨᑦᑎᖅᑐᑦ ᐱᓕᕆᔨᖅᑐᑦ ᐊᑯᓐᓂᖅᑐᓂᖅ ᐊᐅᓚᓯᖅᑐᓂᖅ ᑕᒪᒋᐊᖅ ᒥᑦᓂᓇᓕᖅᑐᑦ
ᐊᔪᕐᓇᐅᔾᔨᓕᕆᓂᖅ ᐊᖏᔪᑎᓕᒫᑦ ᐱᔨᑦᑎᖅᑐᑦ ᐱᓕᕆᔨᖅᑐᑦ ᐊᑯᓐᓂᖅᑐᓂᖅ ᐊᐅᓚᓯᖅᑐᓂᖅ ᐊᒻᒪᓕᒐᓱᐊᓂᖅ.

ᓇᐃᓴᑎᓐᓄᐃ, ᓄᓇᕗᑦ ᐊᒻᒪ ᓄᓇᕗᒻᒥ ᑭᖑᓪᓕᓴᕐᓂᖅᑐᓂᖅ ᐱᓕᓚᐅᖅᑭᓐᓄᑦ ᐃᓕᑎᐅᔾᔨᑦ ᐃᑲᔪᖅᐳᐃᖅᑐᑦ
ᑲᑎᖅᖏᓕᖅᑎᑦᓯᐊᕐᓂᖅᑐᓴᖅᓯᓐᓄᑦ ᓄᓇᕗᑦ/ᓄᓇᕗᒃ ᑲᓇᖅᓴᕐᖏᓂᑦ ᐊᒻᒪ ᖃᑭᖅᑕᓐᓄᑦ ᐃᑭᕋᓴᖅᑎᑦ ᐱᖓᓱᖅᓴᖅᑎᑦ
ᑐᓂᔾᔪᓯᐊᖅᑎᑦᓯᐅᔾᔨᑦ ᐊᒻᒪ ᐅᑕᐅᓂᓚᖅᓴᖅ ᐅᑭᐅᑦᑕᖅᑐᒥ ᑭᑦᑕᕐᓂᖅᑐᓂᖅ ᐃᓚᖅᑭᑐᖅᑐᓂᖅ ᐊᐃᑲᐅᖅᑕᐅᓯᒪᓕᖅᓴᖅ
ᓄᓇᕗᑦ/ᓄᓇᕗᒃ ᑲᓇᖅᓴᖅᑎᑦ. ᑕᑯᐊᐊ ᐊᓯᔾᔨᖅᑕᐅᔾᔨᑦ ᐊᑲᖅᓯᓂᖅᖃᖅᑐᑐᐊᑦ ᐃᖅᓚᓱᓐᓂᑦᓯᒪᑦ ᐃᓕᓴᕐᓂᖅᑐᓂᖅ ᓄᓇᕗᑦ ᐊᒻᒪ
ᓄᓇᕗᒻᒥ, ᐱᔾᓪᓚᐸᓂᖅᑎᑕᐅᔪᓂ ᐃᖅᓚᓵᑕᐅᔾᔨᑦ ᐅᓂᓱᑲᐅᔾᔨᓂᖅᑐᓂᖅ ᐱᓕᕆᔨᖅᓴᖅᑎᑦ.

ᓯᐅᓂᐊᔪᑦ ᖃᓄᖅᓇᒥᐅᓴᒃ ᑲᑎᒪᔨᑦ ᖃᓄᖅᑑᑭᑎᖃᖅᓯᓂᖅᑐᓄ ᓄᓇᕗᑦ ᐊᒻᒪ ᓄᓇᕗᒻᒥ ᐃᖅᓚᓱᓂᖅᑎᑦ ᐊᑐᓕᖅᓯᒪᔨᖅᓂᖅ
ᓂᑦ ᒎᓪᓇ ᐅᖃᐅᓯᐅᔾᔨᖅ ᐱᔾᔨᐊᑎᒋᓪᓗᑐ. ᐃᓯᒪᒋᔭᓪᓗᑐ ᐊᐅᑐᖃᖅᑐᓂᖅᑕᐅᓯᖅᓯᒪᓂᖅ ᐊᑐᓕᖅᓯᒪᔨᖅᓂᖅᑐᓂᖅ
ᓯᕗᒃᐊᒥᒃ ᐅᕐᑕᐅᓪᓗᑐ ᐊᑐᓕᖅᓯᒪᔨᖅᓂᖅᑐᓂᖅ ᐅᖃᐅᓯᐅᔾᔨᑦ ᐱᖃᓐᓂᖅᑐᓂᖅ ᐊᑐᓕᖅᓯᒪᔨᖅᓂᖅᑐᓂᖅ ᐊᑐᐃᓐᓇᖅᑐᑦ
ᐃᓚᖅᑭᑐᖅᑐᓂᖅ ᐊᑐᐃᓐᓇᖅᑐᑦ ᐊᑐᓕᖅᓯᒪᔨᖅᓂᖅᑐᓂᖅ ᐊᑐᐃᓐᓇᖅᑐᑦ ᐊᑐᓕᖅᓯᒪᔨᖅᓂᖅᑐᓂᖅ ᐊᑐᐃᓐᓇᖅᑐᑦ
ᐊᑐᓕᖅᓯᒪᔨᖅᓂ

ርL°ቈ ⊲ᓯᑦᓂᓐ ጋᑭᖏᑐᖅ ᓄቈᕐᒥ ⊲ᑦL ᓄቈᐃᖏᒥ Δᖏᓗᓚᖏᓂᑦᒧᑦ ᐱᓕᓚᐃᖏᑎᑦ ᑲᑎᑎᑎᑦ ጋᓄᐅᓯᐊᖃᖏᓂᕐᒪᖏᑎᑦ Δᖏᓗᑦᑕᐅᒪᓕᖏᑐᑦ, ᓇᓕᐊᖏᓂᑐΔᑦᓇᖏ ᑭᑦᑯᓂ ⊲ᐅᓕᑎᓂᑦᒧᑦ ⊲ᐃᐱᑐᖏᑕᐅᓯᒪᓕᖏᖏᓂᑦ. ⊲ᑐᓂ ᑲᑦᕿᓂᐅᓕᑦ ⊲ᑕᐅᓯᒋᑦ ጋᓄᐅᓯᐊᖏᓕᓕᖏᑐᑦ, ጋᖏᓐᓚᐱᒪᔪᑎᑦ ᖃᑦᓯᐅᓂᓗᖏᑎᑦ ᑭᑦᑯᓂ ⊲ᐃᐱᑐᖏᑕᐅᓯᒪᓕᖏᖏᓂᑦ ጋᓄᐅᓯᐊᖏᓚᓂᑦ ርL°ᒋᑦ ⊲ᕿᑎᐅᓕᓂᑦ ⊲ᖏᒻᒋᑎᒪᔪᑎᑦ ⊲ᐅᑦᑭᖏᑎᑦ ᐱᓕᓗᓯᓯᑎᑦ ⊲ᐅᓕᑎᓂᑦᒧᑦ ᑲᑎᒪᓕᐅᓕᑦ ⊲ᑦL⊃ ᒥᓕᑦᑕᖏᑎᑦ.

ᓇΔᓴᑎᑦᓗᒍ, ᓄቈᕐᒥ ⊲ᑦL ᓄቈᐃᖏᒥ ᑭᖓᓴᕐᓚᓯᖏᓂᑦᑕᖏᑎᑦ ᐱᓕᓗᓚᖏᖏᓄᑦ Δᓕᒥᓕᐅᓕᑦ ΔᑲᐊᖏᑐΔᕐᑐᑦ ጋᑦᓯᕐᐅᑎᓕᓗᓚᒥ ᑲᑎᖏᓐᓚᖏᑎᓯᒪᓂᑦᒧᖓᓕᖏᓂᑦᒧᖓᓚᖅᑐᑦ ᓄቈᕐᑦ/ᓄቈᐃᑲ ᑲᓇᖏᓐᓇᕐᓯᖓᓂᑦ ⊲ᑦL ᖃᑭᖏᒪᓂᑦ Δᑭᕐᓴᖏᓐᓇᖏᑎᑦ ᐱᖓᓐᓇᖓᓂᑦ ⊲ᐅᓕᑎᓂᑦᒧᑦ ጋᓄᐅᓯᐊᖏᑎᓯᓕᓕᖏᑐᑦ ⊲ᑦL ᓅᑕᐅᓂᓚᖏᖓ ᐅᑭᑎᓯᖏᓂᑦ Δᑭᕐᓴᑕᐅᓕᑦ ᑲᑎᑐᑎᖏᓂᑦ ᐱᖓᓐᓇᖓᓂᑦᒧᖓᓚᖅᑐᑦ ᓄቈᕐᑦ/ᓄቈᕐᑲ ᑲᓇᖏᓐᓇᖓᓂᑦ. ርᑲᑯᐊ ⊲ᓯᓕᓯᖏᑕᐅᓕᑦ ⊲ᑲᑐᖏᓯᖏᓂᐊᖏᑐᑕᐊᑦ Δᑭᓗᓚᖏᓐᓇᑦᓂᑦ Δᓕᒥᓕᐅᓕᓂᑦ ᓄቈᕐᒥ ⊲ᑦL ᓄቈᐃᖏᒥ, ᐱᓕᐊᑭᓕᖏᑎᑕᐅᓗᓂ Δᑭᓗᑕᐅᓕᑦ ᐅᓂᑲᑲᓯᓚᓕᐅᓂᑕᖏᑎᑦ ᐱᓕᓗᑦᓯᓚᖓᓂᑦ.

ᓯᖃᓂᐊᒍᑦ ᖁᓕᓐᓇᒥᑦᓯᕐᑲᒃ ᑲᑎᒪᒪᓕᑦ ᖇᓄᑐᖅᑭᑎᖇᑦᓂᐊᑦᓂᖏᑦᓐᓄ ᓄቈᕐᒥ ᓄቈᐃᖏᒥ Δᖇᓗᓚᓯᖏᑎᖏᑎᑦ ᐊᖃᓕᖁᓯᒪᓕᕐᖏᓐ-ᓂᑦ ᑕᓐᓇ ᐅᖇᐅᓯᐅᓕᖅ ᐱᑦᓯᐊᑎᒥᓐᓗᒍ. Δᓯᒪᒥᓐᓗᒍ ᐊᑦᑐᖇᓂᖇᓗᓕᐊᕐᖏᓐᓂᑕᓕᖏᓐ ᖁᔨᓂ ጋᑦᓯᕐᐅᑎᓕᓗᓚᓕᐅᓕᑦᖇᖅ, ᐊᐱᓚᑐᒍ ᓯᑲᑲᒥᑦ ᓇᓕᐅᖁᔭᓐᓗᒍ ᐊᖃᓕᖁᓯᒪᓕᕐᒪᖏᑎᑦ ᐱᖓᓐᓇᖓᓂᑦ, ርᒪᑲᑯᐊ ⊲ᖃᓕᖁᓯᒪᓕᕐᒪᖏᑎᑦ ᐱᖓᓐᓇᖓᓂᑦ Δᑭᓗᑕᐅᓕᑦ ᐊᓕᖁᓯᒪᓕᕐᒪᖏᑎᑦ ᐱᖓᓐᓇᖓᓂᑦ 2019Γ Δᖇᓗᓚᓯᖏᓂᑦᒧᑦ ᐊᑦᕿᒍᑎᓴᓕᖓᓂᑦ ᑲᑲᑯᓂ ⊲ᕿᑎᐅᓕᑦ.

Δᓕᐦᐧᐃᑎᑦ;

ᑦᐊᓂ Δᐅᑦ	ᐅᓪᓚᕐᖓ ⊲ᑭᓯᒃ	ᓴᓪᓕ ᐅᐊᑦ ᐊᖓᓚᐃᑎᖓ
Δᑲᓯᕈᐅᑕᖓᑦ	Δᑲᓯᕈᐅᑕᖓᑦ	ᐊᖓᓚᐃᖃᑲᖓᑦ
ᖃᑭᖃᑦᓴᓐᒥ Δᖃᓗᓕᓂᓵᑦ	ᖃᑭᖃᑦᓴᒃ ᑯᐊᐅᓚᓴᑦ	ᒪᑭᑲᒃ ᑯᐊᐅᓚᓴᑦ

ᐊᑲᑦᓂᓐᓂᑦ ጋᓂᓯᐊᐅᓕᑦ

ᑦᓇᑕᐦ ᐅΔᓕᑭᐦᓴᓐ, ᒥᓂᑦᑕᖓᑦ Δᒪᑦᒥᐅᑕᓕᓚᑲᑯᑦ, ᑲᓇᑕᐅᑦ ᒪᓕᒪᓂᓐᓂ
ᓗᑎ ᑐᒪᓕ, Δᑲᓯᕈᐅᑕᖓᑦ, ᐅᑭᐅᑦᑕᖏᑐᑦ Δᖃᑦᒥᐅᑕᓕᓚᓂᑦᒥ ᑲᑐᑦᓯᖃᑎᖓᓂᑦᓂᑦ
ᐱᐢᑕ ᖁᓕᓴᖃ, Δᑲᓯᕈᐅᑕᖓᑦ ᓴᖃᓂᑐᑦ Δᖃᓗᓕᓂᓵᓂᑦ ᑲᑐᑦᓯᖃᑎᖓᓂᑦᓂᑦ
ᓴᓱᐅᓯ ᓴᐅᓕᑐᓴᖃᑦ, Δᑲᓯᕈᐅᑕᖓᑦ, ᓴᖃᓂᑐᑦ Δᖃᓗᓕᓂᓵᓂᑦ

ርL°ል ⊲ᓂᐅᓂᖅ ጋᑭᖃᓄᖅ ᓄልᕐᒥ ᐋᒪᓕᕐᓂᐊᕐᓂᒃ ᐱᓕᕆᔨᒃᑯᑦ ᑲᑎᑦᑐᑦ ጋᓗᐊᓂᕆᔨᖅᑎᓂᒃᑯᑦ ᐃᖃᓗᐃᑦᑐᕐᓂᖅᓯᓄᑦ, ⊲ጋᓂ ᑲᒡᓂᐅᔪᓂᒃ ⊲ᑕᐅᓂᖅᓯᓂᒃ ጋᓗᐊᓂᕆᔨᖃᑦᑕᕐᓂᖅᓯᓄᑦ, ጋᓐᓇᕆᒪᓕᕐᒥᑦ ᖃᓄᐃᑦᑐᓂᒃᑯᑦ ርᐊᓪᓚᒃ ᐊᑦᑐᕐᓂᖅᓯᓄᑦ ᐊᑦᑐᕐᓂᖅᓯᓄᑦ ᐱᓕᕆᔨᒃᑯᑦ ᐊᑦᑐᕐᓂᖅᓯᓄᑦ ᐊᑦᑐᕐᓂᖅᓯᓄᑦ ᐊᑦᑐᕐᓂᖅᓯᓄ

ᖃᘐᑎᐅᐅᓪᓗᓂ ᑐᓂᐅᐅᔪᑦ ᐅᐊᖏᓐ

ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᓇᔨᔪᐊᕐᑭᔪᑦ ᖃᑎᒪᔨᖏᑦ ᐊᖏᒪ ᓄᓇᕕᑦ ᐃᒪᓐᓕᓂᑦ ᐅᒪᔪᓄ ᖃᑎᒪᔨᖏᑦ

ᑐᑭᓯᐊᕐᑐᑦ: X **ᐃᓱᒪᓕᐅᓂᖅᓯᒪᑦ:**

ᐱᔪᔾᑎᓐᓄ: ᓄᓇᕗᑦ ᐊᖏᒪ ᓄᓇᕕᖅᐃᖅᑲᓕᓚᓂᖅᓯᒪᑦ ᐃᓕᓴᐅᓱᒪᔪᓂᑦ ᑎᑎᕋᖅᑕᐅᓱᒪᔪᓂᑦ ᖃᓄᐊᕐᓴᐅᑦ ᓯᐅᒪᔪᓂᑦ ᐃᓕᓴᐅᓱᒪᔪᓂᑦ ᖃᐅᔨᓴᕐᓂᑦ ᐃᓚᖏᑕᐅᓂᑦ ᐃᒪᓕᕆᔨᓂᑦ ᐃᓕᓴᐅᓱᒪᔪᓂᑦ ᐃᒪᓐᓕᒫᓂᑦ ᐃᒪᓕᕆᔨᓂᑦ ᐅᒪᔪᓂᑦ ᐃᓕᓴᐅᓱᒪᔪᓂᑦ ᐃᒪᓕᕆᔨᓂᑦ ᐃᓕᓴᐅᓱᒪᔪᓂᑦ ᐃᓕᓴᐅᓱᒪᔪᓂᑦ ᐃᓕᓴᐅᓱᒪᔪ

NOAA ᐊᐳᐢᓚᐊᖅᑎᓴᕆᔨᓪᓕ ᓇᓂᕆᐅᔨᓪᓕ ᑐᓴᖏᓂᖅᓂᐊᖅᑐᑦ ᖃᓄᖅ ᓇᓗᓇᐃᖅᑕᐅᔪᒪᓂᖅᓴᓂᓕ (ᓄᓇᓗ.”ᐱᖃᓪᓚᕆᐅᖅᑐᑦ ᐅᕝᕙᓚᓴᓂᓕ ”ᓱᓕᑕᓄᖅᑐᓄᓕ”).

NOAA ᐅᖃᐅᖅᑎᐊᖅᓯᒪᔨᓪᓕ ᐱᕕᓴᖅᑕᖃᓂᐊᖅᑐᖅ ᐃᓕᓴᐅᓗ ᐃᖃᓗᓚᓂᓂᖅᑎᑦ ᓴᖅᑭᖅᑎᑕᐅᓕᐅᖅᑎᓴᓇᓕᑦ LOFF ᐊᖏᕿᔭᒃᓗ 2020. ᑎᑭᓕᐅᖅᑎᓴᓇᒍ Ċᓴᓇ ᐃᓕᒍᑕᕆᐊᓴᐅᔨᓪᓕ ᐃᓱᓕᕝᕕᖓ, DFO ᐃᖅᑲᓇᐃᖏᖅᑎᖃᐃᓴᓇᐊᖅᑐᑦ NMRWB ᐊᒃᒪ NWMB-ᔨᓴᓕᐊᒃᒪ ᐊᐅᓕᑎᖃᑕᐅᔨᓪᓕ ᑲᑐᐃᖃᑎᑎᑦ ᐊᑐᓕᖏᓕᓂᑎ ᐃᖃᓗᓚᓂᓂᖅᑐᖅᓂᑦ ᐱᖃᓪᓚᕆᐅᖅᑐᖅᓂᑦ ᐱᐅᓂᐸᖅᓯᒪᔨᓪᓕ ᑎᑎᐸᖅᓯᒪᔨᓪᓕ 1.

ᑎᑎᕋᖅᓯᒪᔨᓪᓕ 1. DFO (ᖃᑎᐊᓄᓪᓕ ᐅᑭᐅᑕᖅᑐᒍᓪᓗ) ᐃᖃᓗᓚᓂᓂᐅᔨᖅ ᓯᓚᓇᐅᔨᒥᑦ ᓱᓚᒋᓄᖅᑎᑎᕐᓂᑐᑦ (ᐅᕝᕙᓚᓴᓂᓕ ᑕᐃᒪᐃᓴᓂᓄᖅᑭᒪᔨᓪᓕ) ᐃᖃᓗᓴᓂᓕ ᐅᕝᕙᓚᓴᓂᓕ ᐃᖃᓗᓚᒥᓂᓂᓕ ᒥᐊᓚᑲᓇᓂᓕ ᓂᐅᕋᖅᑕᐅᔨᓪᓕ.

ᐃᖃᓗᓚᓂᓂᖅ	ᐅᒪᔨᐃᓕ	ᐃᖃᓗᓴᓇᓱᐊᕐᓵᑦ ᖃᓇᐃᓕᑐᓂᓕᖅᑐᑦ	ᑎᑎᕋᖅᑕᐅᓂᖓ
ᐃᖃᓗᑦᑐᑎᐊᖅ	ᐃᖃᓗᓇ	ᐊᖅᑭᑦᓱᒪᔨᓪᓕ; ᓄᓗᐊᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ
ᐸᓂᓂᖅᑐᖅ ᐃᑭᖓ	ᐃᖃᓗᓇ	ᐊᖅᑭᑦᓱᒪᔨᓪᓕ; ᓄᓗᐊᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ
NU ᐃᖃᓗᓕᓂᓂᖅ	ᐃᖃᓗᓇ	ᐊᖅᑭᑦᓱᒪᔨᓪᓕ; ᓄᓗᐊᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ
ᐸᓂᓂᖅᑐᐸ ᐃᑭᖓᓂ ᖃᓕᑎᓕᑦ ᐊᐅᓕᑕᐅᔨᓪᓕ (CSTMA)	ᐊᑯᑭᑐᑦ ᖃᓕᑎᓕᖏᑦ	ᐊᖅᑭᑦᓱᒪᔨᓪᓕ; ᐊᑦᑐᐃᓚᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ
NAFO ᐃᓂᖓ 0, 100 Cᓕᔪ **	ᐊᑯᑭᑐᑦ ᖃᓕᑎᓕᖏᑦ	ᐊᖅᑭᑦᓱᒪᔨᓪᓕ; ᐃᖃᓗᓴᓇᓱᐊᕐᑭᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ***
NAFO ᐃᓂᖓ 0¥	ᐊᑯᑭᑐᑦ ᖃᓕᑎᓕᖏᑦ	ᐊᖅᑭᑦᓱᒪᔨᓪᓕ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ
NAFO ᐃᓂᖓ 0¥	ᐊᑯᑭᑐᑦ ᖃᓕᑎᓕᖏᑦ	ᑲᓕᑦᑕᐅᔨᓪᓕ ᐃᖃᓗᓴᓇᓱᐊᕐᑭᑎᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ
ᑭᖓᔪᖅᐸᓂᑦ ᐊᐅᓕᑦᑕᐅᔨᓪᓕ (SMU) 0, 1, EAZ,WAZ¥	ᑭᖓᔪᐸᐃᑦ	ᑲᓕᑦᑐᑎ ᐃᖃᓗᓴᓇᓱᐊᕐᑭᑎᑦ	ᓱᓕᑕᓄᖅᑎᑦᑎᓂᖅ

** ᖃᐅᔨᓴᓂᔪᑦ ᐃᖃᓗᓕᓂᓂᖅ, ᓇᓗᓴᐃᖅᑕᐅᔨᓄᑦᓇᖅᑐᖅ ’ᐱᖃᓪᓚᕆᐅᖅᑐᖅᓂᑦᑐᖅ’ NOAA ᖃᑎᒥᖅᑕᓇᐅᖅᑎᓕᓪᓕᑦ

¥ ᓄᓇᖅᓴᐊᒋᑦᑐᖅ ᐊᐳᐢᓚᐊᖅ 1 (a) ᐊᒃᒪ (b) ᐊᑕᓂ.

ᐊᐳᐢᓚᐊᖅ1 (a). Groundfish ᐊᒃᒪ ᑭᖓᔪᐸᑦᓂᓄᑦ ᐊᐅᓕᑕᐅᔨᓂᖓᑦ ᐊᑦᓭᓴᑎᐢ ᑲᓇᑕᒥ

ᐊᐳᐢᓚᐊᖅ 1 (b). ᐅᑭᐅᑕᖅᑐᒥᑦ ᑭᖓᔪᐸᐃᓂᐊᕐᕕᒃᑦ ᑭᓪᓕᖓᑦ:
- ᑲᓇᓂᓇᒥᑦ ᖃᐅᔨᓴᓂᕕᐅᔨᓪᓕ ᑭᓪᓕᖓᑦᑎᑦ “EAZ” (blue)
- ᐱᖓᓂᓇᒪ ᖃᐅᔨᓴᓂᕕᐅᔨᓪᓕ ᑭᓪᓕᖓᑦᑎᑦ “WAZ” (green)

ᑐᓴᖅᑎᑐᓂᖅ:

DFO ᖃᑎᐊᓄᑦ ᐅᑭᐅᑕᖅᑐᒍᓪᓗ ᖃᐅᔨᖅᑲᐃᓕᔪᑦ ᑎᒍᒥᐊᖅᑎᐅᖃᖅᑕᐅᔪᓂᑦ ᑭᖓᔪᑦᓴᓂᑦ, ᐃᒪᐃᓐᓇᖅᑎᑦ ᐊᑯᑭᑐᑐᑦ ᖃᓕᕋᓕᖅᑭᑦ ᐊᒻᒪ ᐃᖅᓇᑦᑐᑎᐊᒥ ᐃᖅᓇᐃᑦ ᐃᓇᓕᓱᓂᐅᔪᓂᑦ ᐃᓄᑦᑎᒍᑦ ᑲᑎᒪᓂᖅᖅᑐᑎᑦ. ᐊᓚᒪᓇᑦᑕᐅᖅ, DFO ᓴᖅᑭᑎᑦᑎᓕᒪᔪᑦ ᒪᑦᑐᓐᓂᓄᑦ ᒪᐸᐱᓴᓂᓄ ᓇᓇᓀᐃᖅᓯᒪᔪᓂᑦ ᐊᑐᓂ ᐃᖅᓇᓕᓱᓂᕐᒪᑦ ᑎᑎᕋᖅᑕᐅᓕᒪᔪᑦ ᑎᑎᕋᖅᓯᒪᔪᓂᑦ 1.ᑕᐅᑯᐊ ᓇᓇᓀᐃᕐᖅᓯᒪᓂᖏᑦ ᑐᓂᐅᖅᖅᑕᐅᓕᐅᖅᑐᑦ ᒪᐃ 8, 2019 ᐊᐃᑦᑐᖅᓯᒪᔪᓂᑦ ᑎᒍᒥᐊᖅᑎᐅᖃᖅᑕᐅᔪᓂᑦ ᖃᑎᖃᕐᓴᐅᓕᓇᑎ ᐊᓚᒪᒪ ᑐᓴᕐᐃᖅᖅᑎᑦᑐᑎᑦ ᑐᓂᔾᔨᐅᓕᐅᖅᑎᓲᓇᒍ ᑭᖓᓕᓕᖅᕿᒥᑦ ᑐᓂᔾᔨᐅᔪᖅ. ᑭᖓᓕᓕᖅᕿᒥᑦ ᐅᖅᐅᓯᓴᐃᑦ ᐱᑕᖅᓀᐊᖅᑎᑕᐅᓕᐅᖅᑐᑦ ᒪᐃ 21, 2019 ᐱᖅᓯᐅᓴᕐᓗᑎ ᐱᓕᕓᓕᐊᓂᖓᓐᓄᓕᑦ ᐅᓂᒃᑲᓂᑦ ᑲᓇᑕ ᑐᓂᓂᐊᖅᑕᖓ NOAA-ᑯᓘᓕᑦ ᔪᓕᐃ 31, 2019. ᐱᓴᓱᖅᓯᒪᔪᓂᑦ ᑎᑎᕋᖅᑕᐅᓯᒪᔪᓂᑦ ᑐᓴᕐᑎᖅᑎᓂᓕᒻᒧᖓᓂᑦ ᐱᖅᓯᐅᑎᓯᒪᔪᖅ ᖃᐅᔨᓴᐃᖅᑲᐃᐅᕙᓗᒍ (TAB 2).

ᐊᑐᓕᕐᔪᐅᔪᖅ:

ᑕᒃᑯᐊ NWMB ᐊᒻᒪ NMRWB ᖃᐅᔨᒃᑲᐃᓗᑎ DFO ᖃᐊᓄᑦ ᐅᑭᐅᑕᖅᑐᒍᒧᓪᓗ ᖃᓄᐃᓕᑐᑐᐃᓐᓇᓂᖅ ᐃᓯᒪᓗᑕᐅᔪᓂᑦ ᐱᔭᔪᑎᖃᖅᑐᑦ ᓇᓪᓕᐊ ᐃᖃᓄᖅᑕᐅᓂᐅᔪᔪᑦ ᐱᖃᓯᒪᔾᔨᐅᓗᑎ ᖃᓄᐃᓕᑐᐃᓐᓇᓂᖅ ᐱᖅᓯᐅᑎᓯᒪᔪᖅ LOFF.

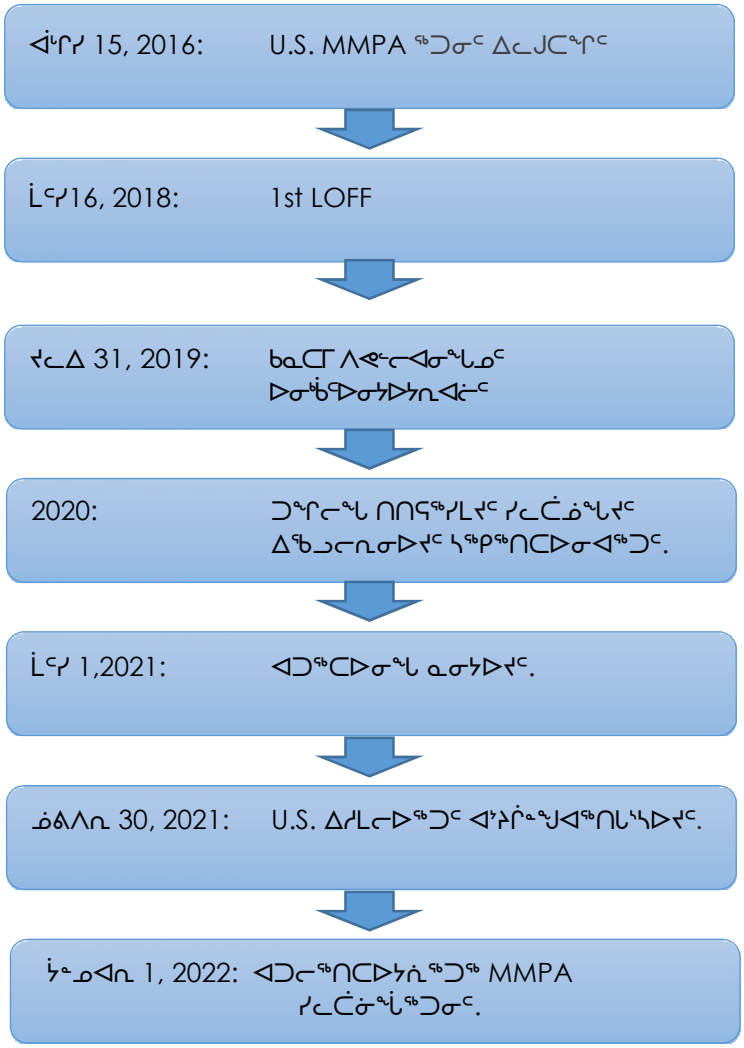
ᑎᑎᕋᖅᑐᖅ:

ᐊᐃᒥ ᓇᓐᓂ ᐊᒻᒪ ᑲᐃᑦᓚᓐ ᐸᕐᑎᐅᓪ, DFO, ᓄᓇᒋᓱᖅᑐᓂ ᐊᐅᓚᑦᑎᓂᖅ, ᖃᑎᐊᓄᑦ ᐅᑭᐅᑕᖅᑐᒍᒧᓪᓗ.
(204) 983-0599 / (204) 984-2338

ᐅᓪᓗᖓ:

LΔ 8, 2019

TAB 1. ᖃᖏᑦᓯᓂᖅᑕᑦ ᐊᑐᓕᖅᑎᑕᐅᓗᓂ ᑎᑭᑎᑦᑎᓂᕐᒧᑦ ᓱᓕᑦᔅᓂᕐᐃᓕᒍᑦᑕᖅᑕᑦ ᒥᐊᓕᑦᑎᓯᓇᑦ. MMPA.



TAB 2: ᑎᑎᕆᖅᓯᒪᔪᓄ ᐊᐃᓕᓐᑐᖅᓯᒪᔪᓂᖅ ᑎᒍᒥᐊᖅᑎᐅᖅᑕᐅᔪᓄ ᑐᓴᖅᑎᓂᑎᔪᓄ ᑐᓴᕐᐃᖅᑭᒪᓕᒐ DFO ᖃᑎᐊᓂᒃ ᐅᑭᐅᑕᖅᑐᒍᓕᒃᓗ LOFF ᑐᓂᔨᐅᔪᓄ

LOFF ᐃᓕᒍᑕᖓ	**ᑐᓴᖅᑎᑕᐅᔪᖅ ᑲᑐᔨᖃᑎᑎᒋᔭᖏᔪᖅ(ᑦ)**	**ᐊᑐᖅᑕᐅᔪᖅ ᑐᓴᖅᑎᑦᑎᓂᔪᑦ**
ᐃᖃᓗᓪᑐᑦᑎᐊᖅ ᐃᖃᓗᐃᓕᑦ	ᖃᑎᖅᑎᐅᑦ ᓂᖅᑭᑦ ᑎᒥᖓ (KF) ᐃᖃᓗᓪᑐᑦᑎᐊᖅ HTO ᐃᖃᓗᓪᑐᑦᑎᐊᒥ ᓄᓇᓕᒻᒥ ᐃᓚᓐᑐᖃᐃᑦ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᕆᔨᕐᓴᐅᕐᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᑦ ᑐᓴᖓᕕᐅᖅ ᑎᒥᖓᑦ ᓄᓇᕕᐅᖅ ᐃᒪᖅᒥᑦ ᐅᒪᔪᓕᕆᓚᓂ ᑲᑎᒪᔨᑦ ᒪᑭᕕᐅᖅ ᓄᓇᕗᑦ ᓴᖅᓪᓕᐅᕐᓂᑦ	KF ᐃᖃᓗᓐᓂᔪᓚᖅᑕᐅᔪᖅᑎᑦᓯᒪᓂᑦ ᑲᑎᒪᓂᖅ (ᐊᓚᓄᐊᓕᖅ 2019) ᑐᓵᔪᖅᑐᑦ ᐅᖃᐅᔾᔨᓂᑦ ᑐᓂᔨᐅᓪᓚᕆᐊᖅᑐᑦ (LΔ 2019)
ᐸᓚᓂᖅᑐᐸ ᐃᑭᖓᓂ ᐃᖃᓗᐃᓕᑦ	ᐸᓚᓂᖅᑐᖅ ᐅᒪᔪᓕᐊᑎᑲᐃᓕᓂᑦ ᑲᑎᒪᔨᑦ: ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᕆᔨᕐᓴᐅᕐᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᒻᒥ ᐃᓄᐃᑦ ᐅᒪᔪᓂᑦ ᑎᑎᕆᖃᖅᓴᖓ ᓄᓇᕗᑦ ᑐᓴᖓᕕᐅᖅ ᑎᒥᖓᑦ ᓄᓇᕕᐅᖅ ᐃᒪᖅᒥᑦ ᐅᒪᔪᓕᕆᓚᓂ ᑲᑎᒪᔨᑦ ᒪᑭᕕᐅᖅ ᓄᓇᕗᑦ ᓴᖅᓪᓕᐅᕐᓂᑦ	ᑐᓵᔪᖅᑐᑦ ᐅᖃᐅᔾᔨᓂᑦ ᑐᓂᔨᐅᓪᓚᕆᐊᖅᑐᑦ (LΔ 2019)
ᓄᓇᕗᑦ ᐃᖃᓗᓕᓂᒃ ᐃᖃᓗᑕᓗᓂᖅ (ᐊᒥᐊᐅᓂᖓ)	ᓄᓇᕘᓂᓂ ᐅᒪᔪᓕᐊᑎᑲᐃᓕᓂᑦ ᑲᑎᒪᔨᖏᑦ: ᑲᖏᖅᑐᑐᐃᓚᐅᖅ, ᖃᐅᔨᓪᓗᐃᑦᑐᖅ, ᒥᑦᑎᒪᒋᘁᓂᒃ, ᖃᑭᖏᑦᑕᕆᔨᐊᖅ, ᐃᐸᐃᐊᖅᓴᓂ, ᐃᓗᐊᓕᓂᒃ, ᓴᓂᖏᓂᒃ, ᐊᐅᔨᓪᓗᐃᑦᑐᖅ, ᑭᓇᐅᔭᓂᒃ, ᐃᖃᓗᐃᓕᑦ, ᑭᒍᒥᑦᑕ, ᓴᓂᖀᓇᐊᖅ, ᐅᖅᔨᖅᑐᖅ, ᑕᓚᕆᐊᖅ, ᑯᐃᕆᓂᒃ, ᕆᔨᓚᖅᑐᖅ, ᖃᓄᖅᑐᖅ ᖃᑎᖅᑎᐅᑦ ᐅᒪᔪᓕᓕᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᕆᔨᕐᓴᐅᕐᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᑦ ᑐᓴᖓᕕᐅᖅ ᑎᒥᖓᑦ ᓄᓇᕕᐅᖅ ᐃᒪᖅᒥᑦ ᐅᒪᔪᓂᓚᖅᑕᐅᔪᖅᑎᑦᓯᒪᓂᑦ ᑲᑎᒪᔨᑦ ᒪᑭᕕᐅᖅ ᓄᓇᕗᑦ ᓴᖅᓪᓕᐅᕐᓂᑦ	ᑐᓵᔪᖅᑐᑦ ᐅᖃᐅᔾᔨᓂᑦ ᑐᓂᔨᐅᓪᓚᕆᐊᖅᑐᑦ (LΔ 2019)
ᐸᓚᓂᖅᑐᐸ ᐃᖃᓗᐃᓕᑦ ᖃᓄᕐᓗᓂᑦ ᐊᐅᓪᒥᑕᐅᔪᓄ (CSTMA) ᐊᑯᖃᑐᑦ ᖃᓄᕐᓗᓂᖓᑦ	ᐸᓚᓂᖅᑐᖅ ᐅᒪᔪᓕᐊᑎᑲᐃᓕᓂᑦ ᑲᑎᒪᔨᖏᑦ ᐸᓚᓂᖅᑐᐸ ᐃᖃᓗᐃᓕᑦ ᐃᖃᓗᑕᓗᓂᖅ ᑎᒥᖓᑦ/ᐃᖃᓗᓕᓂᑕᖅᑲᐃᓕᑦ ᐸᓚᓂᖅᑐᑐᒥᑦ ᑎᒥᖓ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᕆᔨᕐᓴᐅᕐᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᒻᒥ ᐃᓄᐃᑦ ᐅᒪᔪᓂᑦ ᑎᑎᕆᖃᖅᓴᖓ ᓄᓇᕗᑦ ᑐᓴᖓᕕᐅᖅ ᑎᒥᖓᑦ ᓄᓇᕕᐅᖅ ᐃᒪᖅᒥᑦ ᐅᒪᔪᓕᕆᓚᓂ ᑲᑎᒪᔨᑦ ᒪᑭᕕᐅᖅ ᓄᓇᕗᑦ ᓴᖅᓪᓕᐅᕐᓂᑦ	NFA – DFO ᑲᑎᒪᓂᖅ (ᑎᓯᐱᓗ 2018) ᑐᓵᔪᖅᑐᑦ ᐅᖃᐅᔾᔨᓂᑦ ᑐᓂᔨᐅᓪᓚᕆᐊᖅᑐᑦ (LΔ 2019)
NAFO ᓇᒃᑐᓂᖓ 0, 100 ᑕᓵᕐ	ᓄᓇᕘᓂᓂ ᐅᒪᔪᓕᐊᑎᑲᐃᓕᓂᑦ ᑲᑎᒪᔨᖏᑦ: ᐃᐸᐃᐊᖅᓴᓂ, ᒥᑦᑎᒪᒋᘁᓂᒃ, ᑲᖏᖅᑐᑐᐃᓚᐅᖅ, ᖃᑭᖏᑦᑕᕆᔨᐊᖅ, ᖃᐅᔨᓪᓗᐃᑦᑐᖅ, ᐊᐅᔨᓪᓗᐃᑦᑐᖅ, ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᕆᔨᕐᓴᐅᕐᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᒻᒥ ᐃᓄᐃᑦ ᐅᒪᔪᓂᑦ ᑎᑎᕆᖃᖅᓴᖓ ᓄᓇᕗᑦ ᑐᓴᖓᕕᐅᖅ ᑎᒥᖓᑦ ᓄᓇᕕᐅᖅ ᐃᒪᖅᒥᑦ ᐅᒪᔪᓕᕆᓚᓂ Board ᒪᑭᕕᐅᖅ ᓄᓇᕗᑦ ᓴᖅᓪᓕᐅᕐᓂᑦ	ᑐᓵᔪᖅᑐᑦ ᐅᖃᐅᔾᔨᓂᑦ ᑐᓂᔨᐅᓪᓚᕆᐊᖅᑐᑦ (LΔ 2019)
NAFO ᓇᒃᑐᓂᖓ 0 ᐊᑯᖃᑐᑦ ᖃᓄᕐᓗᓂᖓᑦ (ᐊᑦᖁᕐᓯᒪᔪᓂ ᐊᐱᒪ	ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᕆᔨᕐᓴᐅᕐᓂᑦ ᑲᑎᒪᔨᖏᑦ ᓄᓇᕗᒻᒥ ᐃᖃᓗᐃᓕᓂᕐᒧᑦ ᑲᑐᔨᖃᑎᑎᒋᔭᖏᔪᖅ (NFA)	NFA – DFO ᑲᑎᒪᓂᖅ (ᑎᓯᐱᓗ 2018)

ᐃᓕᓴᕐᓂᐅᔾᔪᑎᑦ ᐃᓄᒃᑎᓯᔪᐊᕈᑎᒃ)	ᑲᓇᓕᖅᓯᓂᓄᑦ ᐅᑭᐅᑕᖅᑐᑎ ᐃᓅᒃᑎᐅᑕᓂᓄᑦ ᐃᓄᒃᑐᓕᓂᓄᑦ ᑎᒍᑕᐊᖅᑎᐅᖅᑕᐅᔪᑦ ᐅᖅᐅᓵᕆᐊᖅᑏᑦ ᑲᑎᒪᔨᕐᓯᒪᓕᖅᓴᓕᓕᓂᓄᑦ (EAGSAC) ᓄᓇᒃᑦ ᑐᓕᖏᓐᓄᑦ ᑎᑕᖏᑦ ᓄᓇᓐᓄ ᐃᒪᖓᑦ ᐅᒪᔨᖏᓕᓄᑦ ᑲᑎᒪᔨᑦ Board ᒪᑭᓐᓄ ᓄᓇᒃᑦ ᓴᖅᒪᓄᑦ	EAGSAC ᑎᒍᑕᐊᖅᑎᐅᖅᑕᐅᔪᑦ ᐅᖅᐅᓵᕆᐊᖅᑏᑦ ᑲᑎᒪᓕᖏᓐᓄ (ᐃᖃᓗᒃᑐᐊᓗ 2019) ᑐᓴᕐᓴᖅᑐᑦ ᐅᖅᐅᓯᓴᓕᓂᑦ ᑐᓂᔨᐅᓪᓗᐊᑕᐅᖅᑎᓖᑦᑕᑦ (ᒪᐃ 2019)
SMU 0, 1, EAZ, WAZ ᑭᖑᓪᓕᐃᑦ	ᓄᓇᒃᓗᒋ ᐅᒪᔨᓕᓗᕆᔨᖃᐅᓄᑦ ᑲᑎᒪᔨᖏᑦ ᐅᑭᐅᑕᖅᑐᑎ ᑭᓪᓕᓕᓗᓂᓄᑦ ᐅᖅᐅᓵᕆᐊᖅᑏᑦ ᑲᑎᒪᔨᕐᓯᒪᓕᖅᓴᓕᓕᓂᓄᑦ (NSAC) ᓄᓇᒃᑦ ᑐᓕᖏᓐᓄᑦ ᑎᑕᖏᑦ ᓄᓇᓐᓄ ᐃᒪᖓᑦ ᐅᒪᔨᖏᓕᓄᑦ ᑲᑎᒪᔨᑦ ᒪᑭᓐᓄ ᓄᓇᒃᑦ ᓴᖅᒪᓄᑦ	NSAC ᑲᑎᒪᓂᖅ (ᐃᓕᓴᐅ 2019) ᑐᓴᕐᓴᖅᑐᑦ ᐅᖅᐅᓯᓴᓕᓂᑦ ᑐᓂᔨᐅᓪᓗᐊᑕᐅᖅᑎᓖᑦᑕᑦ (ᒪᐃ 2019)

ᑐᓂᔨᐅᔪᖅ
ᓄᓇᕗᑦ ᐅᒪᔪᓕᕆᔨᖁᔭᐅᔪᑦ ᑲᑎᒪᔨᖏᑦᑐᓄᑦ
ᒪᐃ 2019

ᐃᓴᐹᑎᖓᑦ
ᑐᓴᕆᓴᖅ: X **ᐃᓱᒪᓕᐅᕆᐊᓕᒃ**:

ᐱᓴᐹᑎᓕᒃ: ᐊᖅᑭᑫᓱᐃᓂᖅ ᓄᓇᕗᑦ ᐃᖅᑲᓕᕆᓂᖏᓐᓇᑦ ᒪᓕᑎᖓᓂ

ᖁᓇᐃᔾᑐᓂᖏᓐᓄ:

ᑕᒪᐅᓇ ᓄᓇᕗᑦ ᐊᖏᑕᑎᖏᓐᓄ ᓴᖅᑭᓕᐅᖅᓯᒪᔨᓚᐊᖅ ᐊᑐᓕᖅᑐᓂᓗ 1993–ᒥ, ᓄᓇᕗᑦ ᐃᖅᑲᓕᕆᔨᐅᑦᑎᑦ ᓯᓕ ᐊᐅᓕᑕᐅᓴᒪᒃᑕ ᐊᐱᒋᐅᖏᓐᓄᑦᑐᓄᓇ ᒪᓕᑎᖓᓂ ᐃᓚᐊᓂ *ᓄᓇᕗᑎᐊᑦ ᐃᖅᑲᓕᕆᓂᖓᓕᒫᑦ ᒪᓕᑎᖏᑐᓂ, ᐊᑦᓭᓴᑎᒃ (Atlantic) ᐃᖅᑲᓕᕆᓂᖓᓕᒫᖅᖢᑦ ᒪᓕᑎᖏᑐᓂ ᐊᖏᓪᓗ ᐃᒪᓕᕐᕆᐅᑦᓄᑦ ᒪᓕᑎᖏᑐᓂ*

ᑕᕝᕙᓂ ᐃᓚᐊᓂ, ᓄᑖᑎᒥᑦ ᓄᓇᕗᑦ ᐃᖅᑲᓕᕆᓂᖓᓕᒫᑦ ᒪᓕᑎᐃᑦ (NFR) ᐅᓪᓚᒥᒫᑦ ᐊᖅᑭᒃᑭᓇᓱᖅᑐᐊᖅᑐᑦ ᒪᓕᑎᖓᓂ ᑐᓴᕐᕆᐅᑎᖅᑕᐅᓄᓄᑦ ᐃᖅᑲᓇᓂᖓᓕᒫᑦ ᓄᓇᕗᓪᓕᒥ ᐊᑦᑐᐊᓴᖏᓐᓂᓪᓗ ᐃᒪᓐᓄᑦ, ᐱᓚᐊᖅᑐᒥᑦ ᐃᓕᓴᕆᐊᓴᖏᓐᓂᓪᓗ ᐱᓕᕆᔨᐅᑎᖅᑕᐅᖁᖅᑐᖅ (HTO) ᐅᒪᔨᓕᕆᔭᐅᑦᓪᓗ (RWO), ᐊᖏᓪᓗ ᑲᒪᕐᕆᐊᓕᖓᓂ ᓄᓇᕗᑦ ᐅᒪᔨᓕᕆᔨᖁᔭᐅᔪᑦ ᑲᑎᒪᔨᖏᑐᓄᑦ ᑕᒪᐃᓐᓇᖅᑐᓂ ᐊᐅᓕᕐᑎᓂᖓᓕᒫᑦ ᐊᐅᓕᕐᑎᓂᖓᓕᒫᖅᖢᑦ ᓄᓇᕗᓪᓕᒥ, ᑕᐃᒪᒃ ᓇᓗᓇᐃᔾᔨᖅᑕᐅᓕᖏᓐᓂᒃ ᓄᓇᕗᑦ ᐊᖅᑭᑫᓱᐃᓂᖅᓴᓂ.

ᑲᑎᒪᔨᓪᓗᑎᑦ ᓴᖅᑭᖅᑕᐅᓕᖏᓐᓂᒃ ᐋᓕᐃ 9–ᒥ, 2018 – ᐊᖅᑦᔪᒍᑦ 25–ᖃᓕᖅᑎᓪᓗᑎᑦ ᑕᐃᒪᓐᓇᑐᓂ ᐊᑎᓕᐅᖅᑕᐅᓕᐅᖅᑐᓕᖏᓐᓂᒃ ᓄᓇᕗᑦ ᐊᖅᑭᑫᓱᐃᓂᖅ – ᐅᖅᑲᖅᑐᓕᖏᓐᓂᒃ ᐱᓕᕆᔨᖁᔭᐅᑎᓪᓗ ᓄᑖᒋᖅᑎᑎᒍᓐᓇᖅᓯᒪᓕ ᐊᖅᑭᑫᓱᐃᓚᑎᑦ ᐃᖅᑲᓕᕆᓂᖓᓕᒫᑦ ᒪᓕᑎᔭᖅᑐᑦ ᓄᓇᓕᓪᓕᒥ.

ᓱᑎᐸᐱᓚᒥ 2018, ᒪᑭᕕᒃ ᑯᐊᐳᓴᓵᖅᑯᑦ ᐅᖅᑲᖅᑕᐅᓕᕆᔨᐅᑦᑎᓕᖏᓐᓂᒃ ᑭᓱᖅᑕᐃᓕᓴᕐᓚᑎᑦ ᓄᓇᕗᓪᓕᒥ ᐃᓄᐃᑦ ᐱᓕᒪᔪᖅᑐᓂᒃ ᑕᐃᒃᑲᑦ ᑕᐃᒪᐃᓕᓴᕐᓚᐊᖅᑎᑕᐅᑭᐊᖅᑲᓂᖅᓯᒪᓂᒃ ᐃᓂᖅᑲᓇᐊᖅᑲᓂᖅᓯᒪᓂᒃᑐᓂᓗ ᓄᓇᕗᑦ ᐊᖅᑭᑫᓱᐃᓂᖅᓴᓄᑦ ᐃᓚᐊᑯᓗᒍᖅᑐᑦ ᓄᓇᕗᓪᓕᒥ ᐃᓄᐃᑦ ᓄᓇᕗᑎᑐᓂᖅᓴᓂ. ᑕᒪᓐᓇ ᓄᓇᕗᑦ ᐃᖅᑲᓕᕆᓂᖓᓕᒫᑦ ᒪᓕᑎᖓᓂ <ᖅᓇᕈᑎᓂᒃ ᐊᖅᑭᑫᓯᐅᑦᑎᓕᖏᓐᓂᒃ ᐃᖅᑲᓇᓂᖓᓕᒫᑦ ᒪᓕᑎᖓᓂ ᐊᒪᓪᓗ ᐃᓚᐊᕈᑦᓂᑎᑦ ᐃᖅᑲᓕᕆᓂᖓᓕᒫᖅᖢᑦ ᑲᓇᑕᒥᓂᖅᓯᒪᓂᒃ (DFO), ᓄᓇᕗᑦ ᑐᓴᓚᐃᑦᓂᖅ (NTI), ᓄᓇᕗᑦ ᐅᒪᔨᓕᕆᔨᖁᔭᐅᔪᑦ ᑲᑎᒪᔨᖏᑐᓂᒃ (NWMB), ᓄᓇᕗᑦ ᓴᖇᒪᑦᑎᓂᖅ (GN), ᐊᖏᓪᓗ ᒪᑭᕕᒃ ᑯᐊᐳᓴᓵᖅᑎᓂᖅ.

ᐊᕐᑐᐸᐱᓚᒥ 2018 ᐱᓕᕆᔨᖁᔭᐅᔪᑦ ᐊᖅᑭᑫᓱᐃᓂᖅᓴᓄᑦ ᐅᖅᓯᖏᓐᓇᐊᖅᑕᐅᔪᒪᓕᒃᑐᓂᓗ ᑐᓴᕐᕆᐅᑎᖅᑕᐅᓕᖏᓐᓂᒃ ᒪᓕᑎᕐᕆᐅᖅᑕᕐᓂᓕᓂᖅ. ᑐᓴᕐᕆᐅᑎ ᒪᓕᑎᕐᕆᐅᕕᒪᓕᒃᑐᖅ ᐃᓐᓇ ᐱᓕᐅᖅᖢᑦᑐᓂᓗᓂᖅ ᐃᖅᑲᓇᓂᖅ (ᐱᖅᓯᐅᑎᒃᓴᑦᑐᓂ ᐃᓚᖏᑕᐅᑕᐃᑦ) ᐊᒪᓪᓗ ᓄᓇᕗᑦ ᐃᓚᐊᓂ (ᐱᖅᓯᐅᑎᒃᓴᑦᑐᓂ ᑕᐃᒪᐃᓕᓴᕐᓚᐊᖅ ᐊᑐᓕᖅᑐᓂᓗ ᐃᖅᑲᓇᓂᖓᓕᒫᑦ ᓄᓇᕗᓪᓕᒥ). ᐱᖅᓯᐅᑦᔨᔪᒋᖏᓪᓕᒫᖅ ᓄᓇᕗᑦ ᐃᓚᖏᑕᐅᓕᓂᕐᓚᐅᖅᑲᖅᑐᓂᒃ, ᑲᖏᑐᖅᓯᐊᓗᒃᑐ ᐃᓚᐊᓂ ᐊᒪᓪᓗ ᓯᕿᓂᖏᓐᓇᐅᑦ (Hudson Bay ᐊᒪᓪᓗ James Bay). ᑕᒪᓐᓇ ᐱᒃᓀᐊᑐᒪᓕᒃ ᐊᑯᓂᑲᓕᕆᔨᐅᑦᑎᓕᖏᓐᓂᒃ ᐱᓕᕆᔨᖁᔭᐅᔪᑦ ᐊᒪᓪᓗ ᑲᑎᒪᔨᖏᑐᓄᑦ ᐊᒪᓪᓗ ᒪᓕᑎᖓᑎᑎᐅᑦᑎᓕᖏᓐᓂᒃ ᐊᒪᓪᓗ ᒪᓕᑎᖓᓂ ᑭᓕᓕᖅᑎᓂ ᓴᑎᓕᕐᓂᖅᑐᐅᑦᑎᓕᖏᓐᓂᒃ ᐊᒪᓪᓗ ᐊᓯᔾᔨᓀᐊᖅᑕᐅᖁᖅᑐᓂᒃ.

ᑐᖃᓚᑭᓂᖅ ᓱᑎᐸᐱᓚᒥ 2019 ᐊᖅᑭᑫᓱᐃᓂᖅᓴᓄᑦ ᑲᖅᑭᒃᓴᓂᓯᒋᐊᖅᑕᐅᓂᖅ ᑲᓇᑕᒥ ᓴᓱᒃᑯᓴᓂᖅ (Canada Gazette), ᓯᓂᓕᖅᑕᓐᓇᔭᓗᒍ I. ᐱᓕᕆᔨᖁᔭᐅᔪᓕᒃᑐ ᖁᑎᐅᓴᒪᓚᒃᑕ ᑕᒪᒃᑯᐊ ᐱᓚᕿᓚᓂᖅᑐᐊᖅᑕᐅᓕᖏᓐᓂᒃ ᐊᓗᓚᐃᖅᑲᓇᑕᐅᓚᐅᖅᑎᓕᖏᓐᓂᒃᑐᓂᓗ, ᐱᓕᕆᔨᖁᔭᐅᑦ ᑐᓴᕐᕆᐅᑎ ᐅᖅᑲᖅᑕᐅᓕᕆᔨᐅᒪᔭᖅᓱᓂᒃ ᑕᐃᒪᓐᓇᑐᓂ ᐅᒪᔨᓕᕆᔨᖁᔭᐅᔪᑦ ᓄᓇᕗᓪᓕᒥ 9, 2018 ᑲᑎᒪᔨᓪᓗᑎᑦ ᑎᑎᕋᖅᑲᓇᓕᖅᑎᓂ.

ᑲᑎᒪᔨᖏᑐᓂᖅ:

ᐱᓕᕆᔨᖁᔭᐅᔪᓕᒃᑐᑦ ᑲᑎᒪᔨᖏᑐᓪᓗᑎᓪᓗ ᑕᒪᑐᒥᓱᓂ ᐱᑎᐊᖅᑲᓇᓕᕐᓗᓂᓚᐊᖅᑎᑦ ᐊᒪᓪᓗ ᑲᔪᓯᖏᓐᓄᐊᖅᑐᐊᖅᑐᑦ 2019–ᒥ. ᑭᐹᐸᐱᓚᒥ 27-28, 2018, ᐃᓕᓴᕐᓂᐊᑕᕐᓂᑦ ᐃᖅᑲᓇᓂᖅ ᐊᐹᔅᓯᖏᓐᓄᐊᖅᑐᐊᖅᑐᑦ ᐅᒪᔨᓕᕆᔭᐅᑦᑎᓂᖅ ᓄᓇᕗᓪᓕᒥ

ᓄᓇᕗᒻᒥᓪᓗ ᐅᖃᐅᓯᐅᓂᐊᖅᑐᑎᓂ ᓄᓇᕗᒻᒥ ᐃᖃᓗᓕᕆᓂᖏᓪᓚ ᒪᓕᒐᓕᐅᖃᖅᓂᖏᓂ ᐊᒻᒪᓗ ᑲᑎᓪᓗᑎᓂ ᐃᓱᒪᑎᓐᖓᓂᒃᓴᓂ ᐅᖃᐅᓯᖃᖅᓯᓗᑎᓂ ᐊᕐᕌᖅᑐᒪᔪᓂ ᐅᒪᔪᓕᕆᓂᖅ.

ᐱᓕᕆᖃᑎᒋᓂ ᑲᒪᓕᐅᖅᑐᓂ ᑕᒪᑐᒥᖓ ᓄᓇᓗᓂᓂ >ᓕᕆᓚᐊᖃᖅᓂᖏᓂ 19-ᓄᓂ ᓄᓇᕗᒻᒥ ᓄᓇᓕᓂᓂ ᐃᓐᓇᐅᓗᒥ ᐃᓕᓴᐅᓗ, 2019. ᐱᓕᕆᖃᑎᖃᓕᐅᖏᓪᓕᐊᓂ ᑲᑎᒪᔨᑎᓪᓗᑎᓪᓗ ᐊᓯᖏᑐᓂ ᐊᓐᑐᖅᑕᐅᓂᐊᖅᑐᓂᓂ ᑎᒍᒥᐊᖅᑎᓂ ᑲᔪᓯᑐᐃᓐᓇᖅᑐᓂᒃᓴ ᐅᐱᕐᓴᖅᑎᒥᓂ ᐊᐅᖅᓯᓂ 2019.

ᑐᓴᐅᒪᖅᑭᑕᐅᑎᓕᕆᔨᒪᓕᒃ ᐸᖅᓇᕈᑎᓂᓂᐊᖅᑭᖏᓂᓂ ᐊᖅᑭᑦᑭᓯᒪᔪᓂ ᐊᒻᒪᓗ ᐊᐅᓕᕈᑎᑕᐅᓪᓗᑎᓂ ᑕᒪᑦᑯᓇᖏᑦᓴ ᐱᓕᕆᖃᑎᑕᐅᔪᓂᖅ. ᑐᓴᐅᒪᖅᑭᑕᐅᑎᓕᕆᓂᖓᓪᓗᑎᓂ ᐸᖅᓇᑎᑐᓂ ᐊᖅᑭᑦᑭᓯᐅᒪᔭᓕᒃᑐᑐᓂ ᑐᓴᐅᒪᖅᑭᑕᐅᑎᓕᕆᓂᖓᓪᓗᑎᓂ ᐸᖅᓇᐅᑎᓂᓂ ᐊᒻᒪᓗ ᐱᓕᕆᖃᑎᖃᖅᑐᓂ/ᑲᑎᒪᔨᑎᑕᑎᓂᓂ ᓄᓇᕗᒻᒥ ᐊᓐᑐᖅᑕᐅᓂᐊᖅᑐᓂᓂ ᓄᓇᕗᒻᒥ ᐃᖃᓗᓕᕆᓂᖏᓪᓗᑎᓂ ᒪᓕᒐᓕᐅᖅᑐᓂ.

ᑲᑎᒪᔨᑎᑕᐅᑎᓕᕆᔨᒪᓕᒃ ᖃᓄᖅᑭᑐᖅ ᖃᐅᔨᒪᔭᕐᓂᖅ ᐊᖅᑭᑦᑭᓯᐅᒪᔭᕐᓂᖅ: http://dfo-mpo.gc.ca/fm-gp/peches-fisheries/comm/consultation/nunavut-eng.htm

ᐱᖓᓱᐃᓕᓂ ᐊᖅᑭᑦᑭᓯᐅᓕᕆᔨᒪᔪᓐᓇᖅᑐᓂ ᖃᐅᔨᓴᖅᐅᓪᓚᕆᔨᖅᑕᒃᑯᓂ ᐱᓕᕆᖃᑎᒋᔭᖅᓂ <ᖓᓇᑎᓄᓂ ᓱᓕᑎᓕᕆᔨᖅᑕᐅᓂ 2019-ᒥ. ᑕᒃᑯᐊ ᐅᓂᒃᑲᓂ ᐊᖅᑭᑦᑭᓯᐅᓕᕆᔨᓱᓕᓂᓗ ᐱᖁᔭᓕᖅᑐᓂᓗ ᐱᓕᕆᖃᑎᒋᔭᖅᑕᒃᑯᓂ. ᑎᑎᕋᖅᑭᔨᓂ ᐊᖅᑭᑦᑭᓯᐊᓚᐊᓪᓕᓂ ᑎᓕᓯᔨᓯᑕᐅᓇᓂᒃᑲᖅᑐᓂ, ᑕᒪᓐᓇᐅᑎᓂᒃᑲᓂᖓᓪᓗᑎᓂ 2019.

ᐊᐃᑕᓵᑯᐅᔨᓴᓂ:

ᖃᒃᖓᑎᒃᑯᑎᑦᓂᖓᓪᓗᑎᓂᒃ ᓇᓗᓇᐃᖅᑭᓴᑎᑎᐊᓐᖓᓕᒫᖅᑐᐃᓐᓇᖅᑐᐃᓗᓚᒪᓕᒃ. ᑕᒃᑯᐊ ᓄᓇᕗᒻᒥᒃ ᐃᓕᓴᒥᐅᑕᓕᕆᓚᐅᖅᑯᓂ ᐃᓴᖅᓱᓕᒧᐊᓗᐃᑦ (ᓄᓴᓗ ᓇᓗᓇᐃᖅᑕᐅᓄᑭᓴᓂ ᓄᓇᕗᓂ ᐊᓐᑎᕐᓂᖓᓪᓗᑎᓂ ᐊᒻᒪᓗ ᐃᓴᖅᓱᓕᕆᓂᖓᓪᓗᑎᓂ ᐱᓕᕆᔨᖅᓯᓂᐅᑦ) ᐃᓕᓴᐅᓕᒪᓕᒐᓕ ᐱᓱᑕᐊᑐᓂᐊᓕᓗᒥ, ᖃᒃᖓᑎᒃᑯᑎᑦᓂᖓᓪᓗᑎᓂ ᐊᓄᔨᓴᖅᑕᐅᓱᑕᐊᖃᖅᑐᓂᐊᖅᑐᓂᓕᒫᒪᓕᒃ.

ᑕᒃᑯᐊ ᑐᓴᓂᕐᑐᖅᓄᓪᓗ ᒪᓕᒐᓕᐅᖅᑕᐅᓂᐊᖅᑐᓂᖓᓂ ᐱᐅᓂᖅᓴᓗᐃᖅᑭᑐᖅᑐᓂᐊᖅᑐᓂᓂ, ᐱᓱᑕᐊᑐᓂᐊᓕᓗᒥ, ᐊᒻᒪᓗ ᓂᕈᐊᑕᐊᖃᖅᑐᓂᖓᓪᓗᑎᓂᓄᑭᓴᓂᓕᒫᒪᓕᒃ.

ᑕᒪᒃᑯᐊ ᖃᒃᖓᑎᒃᑯᑎᑦᓂᖓᓪᓗᑎᓂ ᓇᓗᓇᐃᕐᓴᖅᓱᒪᑎᐊᓐᖓᓕᒫᖅᑐᓂᐊᖅᑐᓂᓂ, ᑲᔪᓯᑐᐃᓐᓇᖅᑐᓂᓂ ᐱᓕᕆᖃᑎᖃᖅᑐᓂᐊᖅᑐᓂᓂ ᐱᓕᕆᓚᒋᓐᓇᖅᑐᓂᓂ ᐱᓕᕆᖃᑎᖃᖅᑐᓂᐊᖅᑐᓂᓂ ᑲᑎᒪᔨᑎᑕᐅᓂᐊᖅᑐᓂᓂ ᐱᓕᕆᖃᑎᖃᖅᑐᓂᐊᖅᑐᓂᓂ ᐊᓯᖏᑐ

ᑭᓪᓗᒍ ᐅᖃᐅᓯᐅᓂᐊᖅᑐᑦ ᑐᓴᕐᕌᒍᑦ
ᓄᓇᕗᒻᒥ ᐃᖃᓗᓪᓚᕆᓂᖏᓪᓗᒍ ᒪᓕᒡᕕᐅᕙᓚᐅᖅᑕᒡᓕᒃ

Scale 1:5,000,000

Lambert Conformal Conic Projection; Standard parallels 60°N and 75°N

Kilometres
0 100 200 300 400 500
0 100 200 300
Nautical Miles

ᐊᓚᓄᐊᔪᖅᓯᒪᓂᖅᑎᑦ

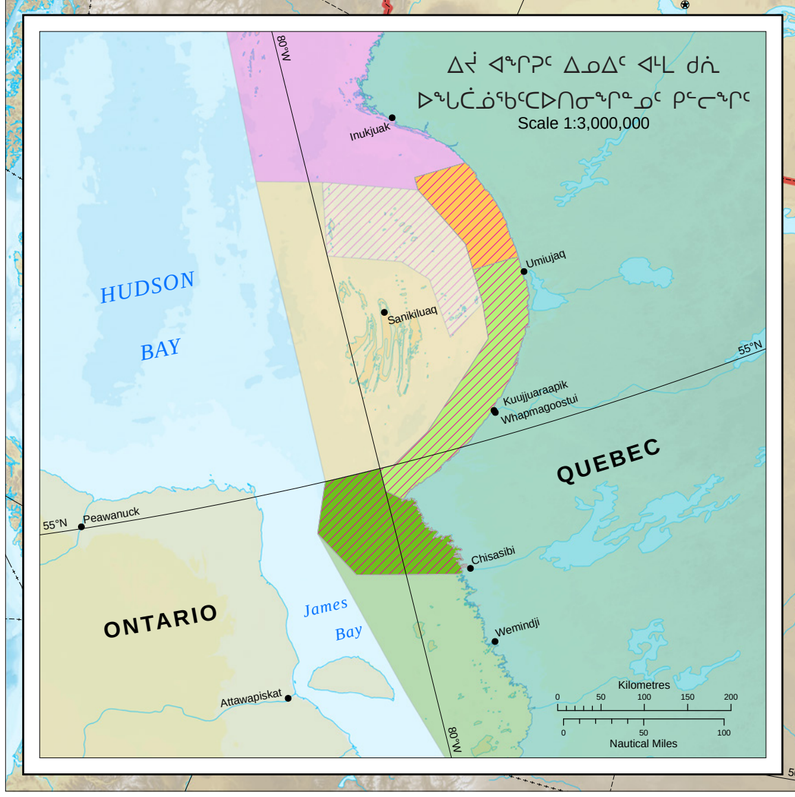
ᑭᓪᓗᒍᑦ
- ᓯᓚᕐᔪᐊᑦ ᓄᓇᓕᕐᕋᓄᑦ ᑭᓪᓗᒍ
- ᓯᓚᑎᕐᒪᓂ ᑭᓪᓗᒍ ᓄᓇᕗᒻᒥ ᐃᒪᓐᒪ
- EEZ/ZEE ᓯᓚᑎᕐᒪᓂ ᑭᓪᓗᒍ ᑐᕌᕐᒪᑐᐊᓂᖏᓪᓗᒍ ᑭᓪᓗᒍᓕᕐᓂᑦ
- ᓯᓚᕐᔪᐊᑦ ᓄᓇᓕᕐᕋᓄᑦ ᐃᒪᓐᒪᓄᑦ ᑭᓪᓗᒍ
- ᑲᑎᒪᔨᖅ ᐊᖏᔪᓕᕐᓂᖏᓪᓗᒍ ᐊᖏᔪᕐᓇᖅᑐᓂᒃ ᑭᓪᓗᒍ
- ᐊᑦᓯᓐᓂᒃ ᐃᖃᓗᓪᓚᕆᓂᖏᓪᓗᒍ ᒪᓕᒡᕕᐅᕙᓚᐅᖅᑕᒡᓕᒃ
- ᓄᓇᕗᒻᒥ
- ᓄᓇᓕᕐᕋᓄᑦ ᐃᖃᓗᓪᓚᕆᓂᖏᓪᓗᒍ ᒪᓕᒡᕕᐅᕙᓚᐅᖅᑕᒡᓕᒃ
- ᐃᒪᓐᒪ ᓴᓕᒥᓂᖅᑐᓂᒃ

ᓄᓇᕗᑦᑎᓂᑦ ᐊᖏᔪᓂᒃ
- ᓄᓇᕗᒻᒥ
- ᐃᓄᐃᓇᓗᐃᑦ ᑭᖑᓪᓕᖅᖢᓂ/ᐱᖓᓱᓇᓗᓂ ᐊᖏᔪᓂᒃ ᑲᑎᒪᔨᖅ ᐅᑭᐅᖅᑕᒡᓕᒃ ᓄᓇᕗᑦᑎᒧᑦ (1984)
- ᐊᖏᔪᓂᒃ (2008)
- ᓴᓪᔅᔨᐊᒥ ᓄᓇᕗᑦᑎᑦ ᐊᖏᔪᓂᒃ (2005)
- ᐃᓕ ᐃᒪᓐᒪᓂ ᓄᓇᕗᑦᑎᑦ ᐊᖏᔪᓂᒃ – ძᕆ (2012)
- ᐅᐊᓂᓪᓚᑦ ᑲᓇᓇᓐᒪᓂ ძᓪᐃᒥ ᐊᖏᔪᓂᒃ (NQA) (1978)
- ᓄᓇᕗᑦ/ᓄᓇᕗᓂᒃ ᐊᖏᔪᕐᓇᖅᑐᓂᒃ ᐅᖃᓪᓚᓱᖁᓂᐅᓂᖏᓐᓂᓐᓄᓪᓂᓂᒃ
- ᓄᓇᕗᑦ/ᓴᓪᔅᔨᐊᒃ ᐊᖏᔪᕐᓇᖅᑐᓂᒃ ᐅᖃᓪᓚᓱᖁᓂᐅᓂᖏᓐᓂᓐᓄᓪᓂᓂᒃ
- ᓄᓇᕗᑦ/ᐃᓕ ᐊᖏᔪᕐᓇᖅᑐᓂᒃ ᐅᖃᓪᓚᓱᖁᓂᐅᓂᖏᓐᓂᓐᓄᓪᓂᓂᒃ

ᓄᓇᕗᑦ/ᐃᓕ ᐊᖏᔪᕐᓇᖅᑐᓂᒃ ᐅᖃᓪᓚᓱᖁᓂᐅᓂᖏᓐᓂᓐᓄᓪᓂᓂᒃ ᑭᓪᓗᒍᑦ
- ᐃᓄᓐᓄᓪᓂ ᑭᓪᓗᒍ – ᓄᓇᕗᑦ ᓄᓇᕗᑦᑎᓂᒃ ᐊᖏᔪᓂᒃ
- ძᕆᓂᒃ ᑭᓪᓗᒍ – ᐃᓕ ᐃᒪᓐᒪᓂᒃ ᓄᓇᕗᑦᑎᓂᒃ ᐊᖏᔪᓂᒃ
- ᖃᐅᑎᖅᑕ ᐃᓄᐃᑦ/ძᕆ ᑭᓪᓗᒍᓄᓪᓂᓂᒃ

Bathymetry data from the GEBCO 2014 Grid, version 20150318. Limits and boundary data from Natural Resources Canada, Statistics Canada and Fisheries and Oceans Canada. Conservation area data from the Canadian Council on Ecological Areas. Treaty boundary data from Crown-Indigenous Relations and Northern Affairs Canada.

ᐃᓕ ᐊᖏᔪᓂᒃ ᐃᓄᐃᓇᓗᐃᑦ ᑭᖑᓪᓕᖅᖢᓂ/ᐱᖓᓱᓇᓗᓂ ᐅᑭᐅᖅᑕᒡᓕᒃ ᓄᓇᕗᑦᑎᒧᑦ (1984)


ᐃᓕ ᐊᖏᔪᓂᒃ ᐃᓄᐃᓇᓗᐃᑦ ᑭᖑᓪᓕᖅᖢᓂ/ᐱᖓᓱᓇᓗᓂ ᐅᑭᐅᖅᑕᒡᓕᒃ ᓄᓇᕗᑦᑎᒧᑦ (1984)
ᐅᐃᓕᖅᑐᖅ ᐃᓄᐃᑦ ᐊᖏᔪᓂᒃ ᐃᓄᐃᓇᓗᐃᑦ ᑭᖑᓪᓕᖅᖢᓂ/ᐱᖓᓱᓇᓗᓂ ᐅᑭᐅᖅᑕᒡᓕᒃ ᓄᓇᕗᑦᑎᒧᑦ (1984)

ᑐᓂᖅᑎᐅᓴᐃᑦ ᐅᑯᓄᖅᖃᓪᓚᑦ
ᓄᓇᕗᒻᒥ ᐅᒪᔨᓕᕆᔨᖅᑕᐅᑯᓂᖅ ᑲᑎᒪᔨᖏᑐᓄᑦ
ᐊᒻᒪ 2019-ᒥ

ᐅᑯᓄᖅᑲ:

ᖃᐅᔨᑎᑎᑭᖅᑕᐅᓂᖓᓕᒃ: ᐃᓕᓴᕆᕙᓗᒃᑐᑦ: X

ᐱᔨᔪᑎᖃᖅᑐᑦ: ᐃᖃᓗᐃᓕᓴᕆᔨᕐᓗᑎᑦ ᑕᕆᐅᕋᒥᐅᑕᓕᕆᓚᐅᑦᓴ ᑲᓇᑕ ᐊᒻᒪᓴ ᒥᕐᓯᐃᖅᓯᕐᓚᕆᓚᐅᑎᑦ ᑲᓇᑕ - ᐱᑭᐊᖅᑎᑐᒪᓕᕐᓂᑦ ᑐᓴᐃᔨᐃᑐᑐᒥ ᑕᕆᐅᕋᒥ ᓴᐅᓂᖅᑕᐅᓯᒪᓕᐲᖅᑐᓄᓪᓕᐊᓂᖓ

ᖃᓄᖅ ᐱᓕᕐᓂᐊᓕᐅᕋᒪᓐᓇᑦ:

ᐃᖃᓗᐃᓕᓴᕆᔨᐊᑎᑦ ᑕᕆᐅᕋᒥᐅᑕᓕᕆᓚᐅᑎᑦᓗᓂ ᐊᒻᒪᓗᓂ ᒥᕐᓯᐃᖅᓯᕐᓚᕆᓚᐅᑎᑦ ᓯᑲᒪᔪᒥᐊᖅᑎᑦᑎᔨᒃᑯᑦ ᑕᕆᐅᕋᒥ ᓴᐅᓂᖅᑕᐅᔨᕐᓄᓇᕗᑦᓴᖅ ᐃᓕᓐᓂᐊᕐᓴᖅᑐᑦ ᐅᑭᐅᖅᑕᖅᑐᑦ ᐃᑎᖅᑎᓂᐊᕐᓴᖅᑐᑦ ᑕᐅᔭᕐᓂᐊᕐᓴᖅᑐᑦ “ᑐᕐᐃᔨᐃᑐᑐᒥᖅᑐᑦ”. ᓴᐅᓂᖅᑕᐅᓄᑉᓴᖅ ᑕᒪᓐᓇ ᑐᕐᐃᔨᐃᑐᑐᒥᖅ ᐱᔨᔪᑕᐅᓪᓚᑎᑦ ᐱᒻᒪᓕᒐᔪᓂᖅᑕᑦ ᐊᕐᓇᑎᖅᑕᑦᓚᑦ ᐊᕐᓇᑎᒥᐅᑕᖏᑦᓚᑦ ᑕᐅᕐᓂ, ᐊᒻᒪᓗᓂ ᐃᓕᑎᔪᐅᕐᓇᖅᑐᓂᖓᓕᒃ ᐃᓄᐃᓂ ᐱᓕᕐᓂᐊᖅᑐᑦᓗᓂ ᐊᕐᓇᑎᖅᑎᑎᖅᑐᑦ, ᐃᖃᓗᐃᓕᓴᕆᔨᐊᑎᑦ ᐱᑭᐊᖅᑎᑎᖅᑐᑦ ᐃᓕᑎᔪᐅᕐᓇᖅᑐᓂᖓᓕᒃ ᐃᓄᐃᓂ ᐱᓕᕐᓂᐊᖅᑐᑦ ᐊᕐᓇᑎ

ᐊᖏᑭᓗᑎᒃ ᐃᓄᐃᑦ ᐊᐅᑐᖅᑕᐅᓯᒪᓂᖅᓯᓄᑦ ᐃᑲᔪᓯᐊᖅᑕᕐᓗᑎᐅᓴᖅᓯᖅᓂᕐᓂᖅ ᑕᑦᓯᑭᑎᐅᑦ ᐃᒪᖅᓱᓄᑦ ᓴᓪᓂᖅᓯᐊᖅᑕᐅᓯᒪᓕᕆᓴᒍᑦ, ᑕᒪᐃᖅᓄᑦ ᐃᓕᑎᕐᔨᓄᑦ ᐊᓯᖅᑭᓄᑦᓯ, ᐱᓕᕆᐊᖓᔪᖅᑕᕐᓂᖅᑐᖅᑐᓄᖅ ᐃᓕᑎᕐᔨᓯᑭᑎᑎᒃ ᓄᓇᕗᑦ ᑲᖏᒪᓂᕐᒥ, ᓴᓪᓂᖅᓯᐊᖅᑕᐅᓂᐊᖅᒪᑕ ᑐᖏᐃᒪᐃᑦᑐᖅᑐᖅ, ᐊᓪᒪᓗ ᐃᓯᒪᓴᖅᓯᐅᑦᑕᐅᓂᓗᑎᒃ ᓴᓪᓂᖅᓯᐊᑦᑕᐅᑲᐃᓇᖅᓯᐊᖅᑐᑦ ᐋᖏᑭᕐᓂᖅᑕᐅᖅᓱᒪᓕᐅᖅᓯᓗᑎᒃ ᑕᓇᐅᖅᓄᑦ ᓯᓚᑐᓂᑦ: ᑕᒪᑐᒪᓂᑦ ᐅᑭᓐᓄᖅᓯᐅᑦᑕᐅᓂᐊᖅᒪᑕ ᐃᓕᒪᓴᖅᓯᐅᑦᑕᐅᓂᐊᖅᒪᑕ ᐃᓕᓴᕐᓂᖅᑐᖅᑐᓄᖅ ᐃᓕᒪᓴᖅᓯᐅᑦᑕᐅᓂᐊᖅᒪᑕ ᐊᐅᑎᖅᓱᓄᖅ ᑎᑎᕋᖅᑕᐅᓂᐊᖅᒪᑕ ᐊᐅᑎᖅᓱᓄᖅ ᑐᓴᕐᓂᖅᑎᑦᑎᓂᖅᓯᓄᑦ

ᑕᔪᑎᕐᓂᐊᓪᓚᖅ 1. ᓄᓇᖓᐊᖅ (ᐃᓚᖓᐊᖅ) ᑐᙵᐃᓴᐃᑐᑐᒃ ᐱᑎᐊᖅᑎᑕᐅᔨᓚᓪᓕ ᑕᓇᐅᖏᑦ ᐃᓚᖏᓂ ᓴᓐᓂᒋᐊᖅᓂᕐᒧᑦ ᒪᓕᑦᑐᓂᖅ

ᑐᑭᓯᐊᖏᓕᕐᓯᒪᓕᒧᑦ

ᐅᓪᓗᒥᒧᑦ ᑎᑭᓪᓗᐊᔪ, ᐃᖏᓪᓗᓕᓚᕐᖏᐊᖅᔪᑦ ᐊᓚᒪᓗ ᒥᖏᖓᓴᓯᓚᓕᓚᕐᖅᔪᑦ ᖁᐅᕐᓚᒫᓂᕐᖅᑐᑦᑐᓂᑦ ᑕᒪᖅᔪᓄᖕᓂ ᐊᓚᒪᓗ ᑐᓴᖏᖅᑎᑕᐅᔨᓪᓗᑎᒃ, ᐊᒥᓯᓂᑦ ᐃᓕᓴᖅᑎᑕᐅᔨᓚᖏᖅᑐᑦ ᐱᙵᓪᓕᐊᔨᓂᑦ ᐱᓕᕐᓚᐊᖏᖅᑐᑦ ᑕᓇᐅᖏᒥᒃ ᓴᓐᓂᒋᐊᕈᑕᐅᐅᓇᓄᑦᓯᓂᒋᐊᖏᖅᑐᑦ ᑕᐅᓂᒋᐊᖅᑐᑦ ᐱᙵᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᑲᑎᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᐅᓇᓱᖏᓪᓕᐊᓂᖅᑐᑦ ᐊᐅᑐᖅᑐᑦᑐᓂᑦ ᐱᙵᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᑲᑎᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᐃᓕᓴᖅᑎᑕᐅᔨᓚᖏᖅᑐᑦ ᐱᙵᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᐃᓕᓴᖅᑎᑕᐅᔨᓚᖏᖅᑐᑦ ᑕᓇᐅᖏᑦ ᓴᓐᓂᒋᐊᖅᑐᑦ ᐱᙵᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᐃᓕᓴᖅᑎᑕᐅᔨᓚᖏᖅᑐᑦ ᐱᙵᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᐃᓕᓴᖅᑎᑕᐅᔨᓚᖏᖅᑐᑦ ᑕᓇᐅᖏᑦ ᓴᓐᓂᒋᐊᖅᑐᑦ ᐱᙵᖅᓴᕐᓇᓄᑦᓯᓂᒋᐊᖏᓪᓕᒡᓕ ᐃᓕᓴᖅᑎᑕᐅᔨᓚᖏᖅᑐᑦ ᑕᓇᐅᖏᑦ ᓴᓐᓂᒋᐊ

ᑕᐃᒪᓕ ᒎᓪᓇ ᑕᓇᐅᓴᒧᑦ ᐱᖁᓴᕐᕕᖏᓐ ᐊᕐᕕᑐᓂᖅᑐᑎᒍᖏᓐ ᓴᖅᑭᕐᓇᖅᓯᒪᔪᑦ ᑕᓇᐅᑦᔪᑦ ᓇᒥᑐᐃᓪᓇᖅ ᑕᓇᐅᖓᓂᐅᔪᒥᑦ, ᑲᓇᑕᐅᐸ ᓴᖅᒪᑐᖅᑭᑦᑎᓂ ᐊᖏᖅᑭᓕᐅᑎᒃᓯᑦ ᐱᑕᖅᓀᐊᖅᓂᕐᕕᑕ ᐱᓕᕐᓂᐊᖅᑐᓂᑦ ᐃᓕᓴᕐᓂᐊᕐᓂᑦᑐᑦ ᐃᓕᓴᖅᑭᑦᑐᓂᑦ ᐃᓕᓴᕐᓂᐊᖅᑐᑦ ᐃᓕᓴᕐᓂᐊᖅ

ᑐᑭᓯᓂᐊᑭᑕᐅᓂᖅᓗᑦ ᐃᒪᐅᑎᑦᑎᓂᖅᓗᑦᓗᓪ ᓇᐃᔪᖅᑕᐅᓯᒪᔾᔪᑦ ᐅᓂᒃᑲᓕᐊᑦ

ᐊᑦᑕᑦᓴᐃᑦ ᑕᑦᓚᑭᑎᐅᕐᓂ ᐃᒪᖏᓄᑦ ᑲᓇᒪᓕᒪᒥ ᓴᓗᐅᓂᐊᖅᑕᐅᓯᒪᔾᔪᑦ ᐊᓂᔪᖅᑐᓄᖅ

ᐅᑭᐊᒃᓴᓄᑦ 2018-ᒥ, ᖃᑭᖅᑕᓂ ᐃᓄᐃᑦ ᐅᖃᐅᓯᖃᓚᐅᖅᓴᑦ ᐱᖃᓯᐅᕐᓴᖅᑐᓂᖅᑐᑦ ᑐᓴᕐᓯᑎᓄᑦ ᐃᓄᐃᑦ ᐊᐅᑐᖅᑕᐅᓯᒪᓂᖏᓐᓄᓂ ᐃᑲᔪᓯᐊᖅᑕᐸᑭᑎᒃᓴᖏᓐᓄᓂ ᐊᖏᑭᑎᓄᓂ ᑕᐅᖅᓂ ᑕᑦᓗᑭᑎᐅᓴ ᐃᒪᖏᓂᒃ ᑲᓇᒪᓕᒪᒥᒻᒪ ᓴᓗᐅᓂᐊᖅᑕᐅᓯᒪᓕᒃᓴᐅᓂᖅᑐᑦ (ᑕᑦᓗᑭᑎᐅᓴ ᓴᓗᐅᓂᐊᖅᑕᐅᓯᒪᓕᖏᓐᓄᓂ). ᑲᓇᑕᒥ ᒥᖅᓱᐃᖅᓯᓴᖅᑎᑕᓚᐅᓵᑲᑯᑦ, ᑕᐃᒃᑯᐊᓗᓪ ᐃᓴᑲᓗᑕᓚᐅᓵᖅᑎᐊᕐᑲᑯᑦ, ᐊᐅᑦᓭᓯᑭᑎᑕᓚᐅᓵᑲᑯᑦᓗᓪ ᑲᓇᒪᓕᒥ ᐃᓭᒪᔾᔪᑎᒃᑕᐅᐅᓂᖅ ᐊᖏᑭᖅᑎᑎᑦᑕᐅᓂᖅᑐᑦ ᐃᓄᐃᑦ ᐊᐅᖅᑕᐅᓯᒪᓂᖏᓐᓄᓂ ᐃᑲᔪᓯᐊᖅᑕᐸᑭᑎᒃᓴᖏᓐᓄᓂ. ᑭᐅᔾᔨᓭᓗᑎᒃ ᖃᑭᖅᑕᓂ ᐃᓄᐃᑦ ᐱᑎᐊᖅᑎᑕᐅᓂᖅᓯᐅᕐᓂᖅᓴᖅᑭᑦ, ᑭᓱᓕᖅᑐᐃᓄᑦ ᒥᔅᓴᐃᖅᓯᓴᖅᑎᑕᓚᐅᓵᑲᑯᑦᓗᓪ, ᖃᑭᖅᑕᐅᓂᖅᑐᑦ ᐃᓄᐃᑦ ᐊᖏᑭᖅᑎᑎᑦᑕᐅᓂᖅᑐᑦ ᐊᖏᑭᑎᓄᓂ ᐃᓄᐃᑦ ᐊᐅᖅᑕᐅᓯᒪᓂᖏᓐᓄᓂ ᐃᑲᔪᓯᐊᖅᑕᐸᑭᑎᒃᓴᖏᓐᓄᓂ. ᑭᐅᔾᔨᓭᓗᑎᒃ ᐊᖏᑭᖅᑎᑎᑦᑕᐅᓂᖅᑐᑦ ᐃᓄᐃᑦ ᐊᐅᖅᑕᐅᓯᒪᓂᖏᓐᓄᓂ ᐃᑲᔪᓯᐊᖅᑕᐸᑭᑎᒃᓴᖏᓐᓄᓂ. ᐅᑭᐊᒃᓴᓄᑦ ᐊᖏᑭᖅᑎᑎᑦᑕᐅᓂᖅᑐᑦ ᐃᓄᐃᑦ ᐊᖏᑭᖅᑎᑎᑦᑕᐅᓂᖅᑐᑦ ᐃᓄᐃᑦ ᐊᐅᖅᑕᐅᓯᒪᓂᖏᓐᓄᓂ ᐃᑲᔪᓯᐊᖅᑕᐸᑭᑎᒃᓴᖏᓐᓄᓂ

ᑲᑐᔪᖅᑎᑦᑎᓂᑦ (ᑭᐊᐃᓕᖅᑯᑦ) ᓯᕗᓕᐅᑎᑎᕐᓯᒪᓕᐅᖅᓯᒪᑕ ᑕᐃᔨᓯᒪᓕ ᑕᐅᑐᒐᖅᑭᒪᓂᕐᓗᓂᕐᓗᓂᑦ ᐊᓯᖏᓐᓂᑦ ᐱᓕᕆᐊᕆᐊᕆᓕᐅᖅᑕᖏᓐᓂᑦ ᖃᑭᖅᑕᓂ ᐊᕐᓴᑐᖅᓯᒪᓂᖅᑭᑦ. ᐱᑎᐊᖅᓯᑎᑦ ᑕᐃᓐ ᓇᕗᐊᓗ 2019-ᒥ, ᐃᓕᐅᑎᑎᒧᑦᑕᐊᓕᕆᓕᐅᖅᑐᑦ ᐊᒥᓱᓂᖏᓐᓂᑦ ᓄᓇᓕᖏᓐ, ᐃᓕᖅᑲᓕᐅᖅᑐᑦᓂᑦ ᓄᓇᓕᖏᓐ ᓄᓇᑦᕋᐅᓯᒪᓂᖏᓐᓂᑦ. ᑕᒪᓐᓇ ᐋᖅᑭᑎᐊᖅᑕᐅᓂᖓ ᑕᐃᒪᐃᔨᐅᓕᐅᖅᑐᖅ ᒪᓕᕆᓯᑎᑦ ᖃᐅᔨᓴᖅᑕᐅᓯᒪᓕᐅᖅᑐᓂᑦ ᐃᓐᓇᐃᔨᖅᑕᐅᓯᒪᓂᖏᓐᓂᑦ ᐊᖏᑭᑐᑎᓄᑦ ᑐᖓᔪᓐᓇᖅᑐᑦ ᐱᔪᓂᑦ ᑕᓪᓕᐊᑐᑦ ᐃᒪᓐᖏ ᕿᓂᖅᑕᐅᓯᒪᓂᖅᑐᑦ ᐊᖏᑭᑐᑎᓄᑦ ᓴᓇᔨᓴᖅᑕᐅᓯᒪᓂᖏᓐᓂᑦ ᐊᐅᓚᑦᑐᖅ ᑲᑎᒪᔨᖅᑕᐅᓯᒪᓂᖏᓐᓂᑦ ᐊᖏᑭᑐᑎᓄᑦ ᓴᓇᔨᓴᖅᑕᐅᓯᒪᓂᖏᓐᓂᑦ ᐊᐅᓚᑦᑐᖅ ᐊᐅᖅᑲᓇᑕᖅᑭᑦᑐᖅᑎᑦ.

ᐃᓕᖅᑎᑕᕐᓯᒪᓂᖏᓐᓄᑦ ᐅᓪᓗᒥᒧᑦ

ᐃᕐᖢᓗᓕᓂᕐᔪᐊᖅᑯᑦ ᐃᓕᐅᑎᑎᕐᖢᑎᑕᖅᓯᒪᒧᑦ ᖢᓇᑕᒥᐅᑕᓂᖅ, ᐃᓕᑎᕐᐅᕐᓴᕆᓚᐊᕐᖢᖅᑐᖅᓂᖅ, (ᑕᒪᖅᑯᐊᓂ ᓯᕐᓗ ᓄᓇᖢᖅᑐᑐᕐᖢᐃᓕᑦ ᖢᑐᔪᖅᑎᑎᕐᖢᖅᑐᖅᓂᖅ, ᓄᓇᐅᑦ ᑲᙵᒪᖏᓐᓂᖅᑐᖅᓂᖅ, ᖢᖅᓂᑎᕐᐅᔨᓂᖅᓗ ᓄᓇᓕᓂᖅ), ᐊᓯᖏᓐᓂᖅᓗ ᖢᓇᑕᐅᑦ ᑲᙵᒪᖏᓐᓂᖅᓂ ᐊᑐᓂ ᐃᕐᖢᖅᑎᓇᖏᓐᓂᖅ, ᐃᓗᓇᕈᓐᓇᕐᓂᐊᕐᒪᑕ ᑕᒪᑎᒥᖅ ᖢᔨᓯᔨᑦ ᐊᖏᑕᐅᑕᐅᓯᒪᔨᓗ ᐱᓕᕆᓚᐅᕐᓂᖏᓐᓂᑦ ᑕᐅᙵᓂ ᑕᒪᓂ 12-ᖓᕐᖢᐅᔨᓂ ᑕᕿᐅᐅᔨᓂᖅ, ᖢᖅᐅᔪᑎᑕᐅᓗᑎᓗᓄᑦ ᑕᒪᐃᓐᓂᖅ ᕿᓂᐅᒥᐅᑐᐃᓐᓇᖅᑐᓇᖅᑎᓇᓐᓂᕐᓗᓂᑦ ᐊᑯᓂᐅᓂᐊᖅᑐᒍᑦᓗ ᓴᓇᔨᓴᖅᑕᐅᓯᒪᔪᐊᑎᖅᓴᖅᓂᑦ ᑕᓪᙰᐊᐊᑕᑐᑎᒥ. ᒪᐃ 2017-ᒥ, ᓄᓇᓕᖏᓐ ᑕᓚᐅᖅᓂᖅ ᓴᓇᓂᐊᖅᓂᖅᓗᓂᑦ ᑐᖅᑲᓚᔪᐅᔨᓂᖅ ᐅᖅᖢᐅᔪᐅᕐᓴᖅᑎᑦ ᖢᑎᒪᔪᕿᓕᑦ ᐃᓕᖅᑲᖅᓯᑎᑦ ᐊᖏᓪᓗᐊᑦᑲᓚᐅᐊᖅᑐᖅᓂᖅ ᑭᖑᓪᓕᖅᑐᐃᓐᓇᖅᑐᖅᓂᖅ ᓯᕐᒪᒃᑯᓂᖅ ᐊᐅᑎᓂᓚᔪᓐᓇᖅᑐᖅᓂᖅ ᓯᓚᑎᓪᓗᑎᒃ ᓯᓚᑎᓪᓗᑎᒃᑯᑎᑐᑎᓂᖅᓂᖅ ᐊᐅᑎᓂᓚᔪᓐᓇᖅᑐᖅᓂᖅ ᓯᓚᑎᓪᓗᑎᒃᑯᑎᑐᑎᓂᖅᓂᖅ ᓯᓚᑎᓪᓗᑎᒃᑯᑎᑐᑎᓂᖅᓂᖅ ᓯᓚᑎᓪᓗᑎᒃᑯᑎᑐᑎᓂᖅᓂᖅ (ECCC), ᒥᖅᑭᐃᓐᓇᕐᓂᐊᕐᓚᓚᔪᐅᔪᓂᖅ, PCA, ᖢᓇᑕᐅᓴᖅᑕᐅᓯᒪᓂᖅᓂᖅᑐᑦ ᐃᓕᖅᑲᖅᓯᑎᑦ ᐅᑭᐅᖅᑕᖅᑕᒥᓗ ᐱᓕᕆᔨᐅᖅᑕᐅᓯᒪᓂᖅᓂᖅᑐᖅᓂᖅ (CIRNA), ᐊᐅᓚᓂᕐᓗᓚᔪᐅᔪᓂᖅᓂᖅ ᖢᓇᑕᒥ, ᐃᓕᖅᑲᓚᔪᐅᔨᓂᖅᓂᖅ, ᓄᓇᕗᑦ ᓯᓚᒪᐃᑦ ᐊᖏᓪᓗᓚᔪᐅᔪᓂᖅᓂᖅ ᐊᓖᒪᓗ ᑐᓴᖏᓚᐅᐊᖅᑐᓂᖅᓂᖅ ᓴᕿᑭᑎᓚᐅᖅᑐᖅᓂᑦ ᑕᑭᐅᐅᖅᑎᐅᓂᐊᖅᑐᖅᓂᖅ ᑕᓚᐅᖅᑎᑦ ᓴᓇᓂᐊᖅᑕᐅᓯᒪᓂᖅᓂᖅᑐᖅᓂᖅ ᐊᓖᒪᓗ ᐃᓕᖅᑲᓚᔪᐅᔪᓂᖅᓂᖅᑐᖅᓂᖅ ᐊᖏᓪᓗᓚᔪᐅᔪᓂᖅᓂᖅᑐᖅᓂᖅ ᐃᓕᖅᑲᖅᓯᑎᑦ ᑐᑭᓯᐅᑦᓯᐊᓂᖅᓂᖅᑐᖅᓂᖅ ᐊᓖᒪᓗ ᑐᓴᖏᓚᐅᐊᖅᑐᖅᓂᖅ ᐊᖏᓪᓗᓚᔪᐅᔪᓂᖅᓂᖅᑐᖅᓂᖅ ᐊᓖᒪᓗ ᐃᓕᖅᑲᖅᓯᑎᑦ

ᑲᑐᔨᖅᑎᑦᓴᖅᓯᒪᔨᖅᑐᑦ ᐃᑲᔪᖅᓯᐃᓕᐅᕆᖅᑐᑦ ᓴᓪᓂᐊᕆᖅᑕᐅᖅᓯᐅᔨᓂᖅᑳ ᑕᐅᖅᓂ ᑭᓯᐊᓂ ᐃᓯᒪᓕᑎᖅᓴᓂᕿᐃᓕᐅᕆᖅᓂᕐᒧᑦ ᑲᓇᕐᑕ ᓴᖅᑭᑦᑐᓂᖅᑎᑦᓴᓂᕿᐃᓕᓴᐅᓂᕐᓂᒪᖅᓯᒪᓕᖅᓱᒻᒪᑦ.

ᔪᒃ 2018-ᒥ, ᐱᑎᐊᖅᑎᑎᕕᒪᓂᖓᑦ ᐱᓕᓕᐊᑎᑎᑐᓗᑎᒃ ᐱᐅᓯᕐᖅᑲᓗᑎᒃ ᑕᓚᐅᖅᑭᒃ ᓴᓪᓂᐊᕆᖅᑕᐅᖅᓯᐅᔨᓂᖅᓴ ᑐᓴᐅᒪᔨᑦᑐᖅᑭᒃ (ᐅᕝᕓᑲᓂᓐᓂᑦ ᑕᐃᔨᐅᕓᖅᑐᒃ ᖃᐅᑲᓇᓂᑦ “ᖃᐅᓯᑎᖅᑐᒥ ᐅᑭᐅᖅᑕᖅᑐᒥ ᐃᑎᐊᕕᔪᓂᖅᓯᓂᖅᓴ ᑕᓚᐅᖅ”) ᑐᓂᔨᐅᓕᐅᖅᑐᑦ ᓄᓇᕗᒻᒥ ᓯᕗᓕᐅᖅᑎ ᑲᑎᒪᔨᖃᕐᓂᑦ. ᑭᖑᓂᐊᒍᑦ, ᑎᑎᕿᖏᐅᓕᐅᖅᑐᑦ ᑕᒪᕿᑎᒃ ᑲᓇᓐᓇᕿᐅᑕᐃᓕᓐᓂ ᐱᖓᓐᓇᕿᐅᑕᐃᓕᓐᓂ (ᐅᐊᓕᓂᕿᐅᐅ) ᐃᓄᐃᑦ ᑭᓱᖅᑐᐃᔨᖓᑦ ᐊᒻᒪᓗ ᐊᑐᓂ ᐊᖏᔪᖅᑐᓕᒫᓂᖓᓂᓂᑦ ᓴᓪᓂᐊᕆᖅᑕᐅᓯᒪᓕᖅᓱᒻᒪᑦ ᐅᓂᒃᑳᖅᑕᐅᓕᐅᖅᑐᑦ ᑐᓴᐅᒪᔨᑦᑐᖅᑭᒃ ᓴᓪᓂᐊᕆᖅᑕᐅᓯᒪᓕᖅᓱᒻᒪᑦ ᑐᓂᔨᐅᓕᐅᖅᑐᑦ. ᑎᑎᖅᑲᐃᑕᐅᖅᓴ ᑐᓴᖅᑕᐅᓕᐅᖅᑐᑦ ᒥᓂᑕᐅᑦ ᑐᖓᓕᓕᖓᒪᓂᑦ ᐃᖅᑲᓗᓕᓚᔨᕋᐊᒃᑯᓂᖅᓱ, ᐊᖏᓚᔨᒃᖅᓯᒪᓂᖓᓂᓂᑦ ᑲᑎᒪᔨᖃᕐᓂᑦ ᐅᓂᒃᑳᖅᑕᐅᓕᐅᖅᑐᑦ ᐊᐅᓚᑦᑎᑦ ᐊᖏᔪᖅᑐᓕᒫᓂᖓᓂᓂᑦ ᐊᐅᓚᑦᑎᑦ ᐅᓪᓗᓕᓕᒫᓂᖓᓂᓂᑦ ᑲᑎᒪᔨᖃᕐᓂᑦ ᐊᐅᓚᑦᑎᑦ ᐅᓪᓗᓕᓕᒫᓂᖓᓂᓂᑦ ᐊᐅᓚᑦᑎ

ᑐᑭᓯᐊᖅᑎᑕᐅᓪᓗᒍ ᐊᓚᒪᓂ ᑐᑭᓯᐅᒪᔭᕆᑎᐊᖅᓂᓂᑐᖅ ᑕᒪᖅᑯᐊ), ᐱᔨᑦᑕᐅᓂᖏᓪᓂ ᓴᓐᓂᖅᑎᑕᐅᓯᒪᑐᐃᓐᓇᐃᓐᓇᑭᒪᓂᖏᓪᓂ, ᐊᓚᒪᓂ ᐱᔨᖅᓇᐃᓕᓯᒪᔨᑦᑕᐅᓂᐊᖅᑐᓂᖅ ᐊᑐᓵᔪᕆᐅᖅᓇᖅᑐᓂᖅ ᓄᓇᒥᐅᑕᖏᓐᓂᖅ/ᐃᒪᖏᒥᐅᑕᖏᓐᓂᖅ;

- ᐅᑭᐅᖅᑕᖅᑐᒥ ᐅᓗᓕᐊᓇᖅᑐᖅᓴᐊᖅᖏᓂᓂᒪᔪᑦ ᑐᑭᓯᓂᐊᖅᑎᐅᒪᑦ ᐅᖅᑲᒪᐅᖅᑐᑦ ᐃᑲᔨᖅᓯᐃᑕᑎᐊᖅᓂᖏᓪᓂᖅ ᓴᓐᓂᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅ ᑐᓴᐅᒪᔨᐃᑕᖅᑐᑦ;
- ᓄᓇᕗᒻᒥ ᐃᖅᑲᓕᑕᓚᓴᑦ ᑲᑐᔭᖅᑎᑎᒋᖏᑦ ᑐᓴᖅᑎᕆᑎᓕᐅᖅᑐᑦ ᐃᑲᔨᖅᓯᐃᓂᖏᓪᓂ ᐊᓴᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅ ᑐᓴᐅᒪᔨᐃᑕᖅᑐᑦ ᐊᓴᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅ ᑐᓴᐅᒪᔨᐃᑕᖅᑐᑦ ᐊᓴᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅ ᑐᑭᓯᐅᓇ. ᐅᖃᓕᐅᖅᑐᑦ ᓴᓐᓂᖅᑎᑕᐅᓂᖅᑕ ᐊᑐᑦᓯᓯᒪᔭᖏᓂᖅ ᐃᖃᓗᓕᓯᓂᐊᕐᓂᖅᑕᖏᓐᓂᖅ ᐱᓕᕆᓚᐅᕐᓂᖅᑕᖏᓐᓂᖅ;
- ᐅᑭᐅᖅᑕᖅᑐᑦ ᑕᕆᐅᑕᕆᓚᓂᖏᓂ ᐅᖃᖅᐅᓯᐊᓂ ᑐᓴᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅ ᐃᑲᔨᖅᓯᐃᓂᖏᓪᓂᖅ ᓴᓐᓂᖅᑎᑕᐅᑲᐃᓐᓇᐅᑎᓴᕐᓂᖅᓯᓂᖅ ᑕᐃᓚᖅᑐᑦ;
- ᓄᓇᖃᕐᐊᖅᑎᒥ ᐅᒪᔨᑕᕆᓂᖅᓯᒪᑦ ᑭᓇᐅᔭᓴᓇᓯᐊᕐᓂᖅᑐᑦ WWF-ᓂᑦ ᓴᓐᓂᖅᑎᑕᐅᑲᐃᓐᓇᐅᓯᒪᔨᑦ ᑕᐃᒪᖅᑐᑦ ᑕᒪᑐᒥᖅ ᓯᓂᖃᐅᓇᖅᑐᑎᑦᑐᓂᖅᑐᓂᖅ ᐊᒥᓯᓂ ᐊᖏᕐᓯᓂ ᐊᖏᕐᓴᓂᑦ. ᐅᒪᔨᑕᕆᓂᖅᓯᒪᑦ ᑭᓇᐅᔭᓴᓇᓂᓯᕕᓐᓂᑦ ᐅᖃᖅᐅᓯᐊᓂ ᑐᑦᓯᓂᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅᑐᖅ ᐃᑲᔨᖅᓯᐃᓂᖏᓪᓂᖅ ᐊᑐᖅᑕᐅᑲᐃᓐᓇᐅᓂᐊᖅᑐᓂᖅ ᐊᑐᖅᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅᑐᖅᑐᑦ ᐊᑐᖅᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅᑐᖅᑐᑦ ᑕᒪᑐᒥᖅ ᐊᓴᖅᑎᑕᐅᑲᐃᓐᓇᐅᓐᓇᖅᓵᓂᖅᑐᖅᑐᑦ ᑕᒪᑐᒥᖅ ᐊᑐᖅ

ᐊᖇᑎᓕᓚᐅᒃᑯᑦ ᐊᖇᑎᒥᐅᑕᓕᓚᐅᓵᓗ ᐱᓕᓚᕐᓚᓂᓂᔪᓐᓇᖅ. ᓄᕕᐱᓚ 2018-ᒥ ᑕᒪᐅᑯᐊ ᖃᐅᔨᒪᔨᐸᑎᐅᒃᑲᐃᑦ ᑐᔨᖅᑕᐅᓕᐅᔾᔪᒥᒃ ᐅᒥᐊᖏᔨᐊᓕᓚᐅᓵᐅᕐᓄᑐᑦ, ᐊᖅᑯᑎᑎᑭᐅᓵᓗᑎᒃ ᑲᓇᖃᒥ ᑕᓚᐅᖏᓪᒧᑦ ᐅᖏᐅᑭᔨᐸᖅᑎᐅᓂᔪᒥᑦ ᑲᑎᒪᔨᑦ ᑎᑎᖅᑲᖅᑕᖏᓐᓂᕐᒥᑦ, ᐃᓕᖅᑎᒪᑕ ᕼᐊᓂᓇᓕᐅᖅᑐᓂᑐᓂᖅ ᑐᑐᖅᑐᓕᐅᖅᑐᓂᑐᓂᖅ.

ᑎᓯᐱᓚ 2018, DFO-ᑯᑦ ᑲᑎᒪᓪᓕᓚᓕᖅᓯᑎᒃ ᓴᖅᑭᑎᑎᓕᐅᖅᑐᑦ ᐱᑎᐊᖅᑐᑎᑕᑐᒪᓴᖏᓂᓂᒃ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᓚᓵᖅᓴᐊᖅᑯᑦ ᑲᑎᒪᔨᖏᓂᓂᒃ, ᐅᖅᑲᐅᓴᐅᑎᐊᕐᓕᓚᑐᒃ. ᑕᐃᖅᑯᐊ DFO-ᑯᑦ ᑎᑎᖅᑲᖅᑕᓕᐅᖅᑐᑦ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓕᓚᓵᖅᓴᐊᖅᑯᑦ ᑲᑎᒪᔨᖏᓂᓂᒃ 2, 2019-ᒥ ᐅᖅᑲᐅᓯᖅᑲᖅᑐᓂᒃ ᐃᓕᐅᖅᑲᓂᔭᐊᓂᖓᓂᒃᑐᓂᒃ ᐅᖅᑲᐅᔨᑎᐊᕐᓇᖅᓯᓕᐅᖓᓂᖓᓂᒃ. ᑭᖑᓂᐊᔪᑦ ᐊᑐᓯᑯᔨᑕᐊᓚᖏᓂᖓᓂᒃ ᖃᓄᖅᑎᑎᑐᒥᒃ ᐅᒪᔪᓕᓚᓵᖅᓴᐊᖅᑯᑦ, <ᖃᓇᐅᑕᐅᓕᒪᐅᖅᑐᑦ ᐃᓕᐅᑎᖅᑎᕈᐊᖅᑲᓇᓂᒃᑐᓂᒃ ᐃᓕᐅᔨᒪᒍᓂᒃ ᓄᓇᕗᒻᒥ ᐊᐅᔨᐃᐃᑐᓗ, ᐊᐅᔨᐃᐃᑐᑐᒪ ᑐᑭᓯᐅᑦᑐᓂᒃ ᐃᓕᓴᐅᒪᓂᖅᑐᑦ ᐊᑐᖅᑕᐅᔨᓇᖅᑐᓂᒃ ᑐᑭᓯᐅᑦᑐᓂᒃ ᑐᓴᐃᔨᐃᑐᑐᒥ. ᖃᑭᖅᑐᒥ ᐃᓕᐅᖅᑲᖅᑕᖏᓐᓂᕐᒥᑦ ᐃᓕᐅᔨᒪᒍᓂᒃ ᐊᐅᔨᐃᐃᑐᑐᒪ ᑲᑎᒪᔨᖏᓂᓂᒃ ᐊᐅᔨᐃᐃᑐᑐᒪ ᐃᓕᐅᔨᒪᒍᓂᒃ ᐊᐅᔨᐃᐃᑐᑐᒪ ᑐᑭᓯᐅᑦᑐᓂᒃ. ᑕᒪᑎᒃ ᐊᖏᔪᓕᓚᓵᖅᓴᐊᖅᑯᑦ ᐃᓕᐅᔨᒪᒍᓂᒃ ᐊᐅᔨᐃᐃᑐᑐᒪ ᐃᓕᐅᔨᒪᒍᓂᒃ ᐊᐅᔨᐃᐃᑐᑐᒪ ᐃᓕᐅᔨᒪᒍᓂᒃ ᐊᐅᔨᐃᐃᑐᑐᒪ ᑐᑭᓯᐅᑦᑐᓂᒃ. ᑕᒪᑎᒃ ᐊᖏᔪᓕᓚᓵᖅᓴᐊᖅᑯᑦ

ᖃᐅᓕᓂᖅᑎᓂᒃ ᐅᓂᒃᑲᖏᐅᓂᖅᑎᓂᒃ. ᑲᑎᒪᔨᖓᓕᑦ ᐃᓕᑕᖅᓯᓕᐅᖅᑐᑦ ᖃᓄᐃᓕᓕᐅᖅᑎᐅᓂᖏᓐᓂᒃ ᐱᖓᓕᓕᐊᓂᖓᓚᐅᓕᑦ ᑕᒪᐃᓄ ᐊᑐᓕᖃᐅᓕᓕᐅᖅᓯᑎ ᑲᑎᒪᓴᖅᑕᐅᓕᓕᐅᖅᓯᑎ ᐊᑐᓕᖃᐅᓕᓕᐅᖅᓯᑎ ᐃᓕᓕᐅᑎᑦᑎᐊᖅᑐᑦ ᐃᓕᓕᐅᖅᑐᑦᑎᐊᖅᑐᑦ ᐃᓕᓴᐅᓂᖏᓐᓂᒃ ᐃᓕᓕᐅᖅᑐᑦᑎᐊᖅᑐᑦ

ᐅᖅᑐᑎᑎᕆᖅ ᓴᓪᓕᒥᕆᖅᖃᓂᓗᒍᑦ ᑐᓴᐃᔨᐃᓕᑦᑐᖅ

(ᖃᓄᑦᑎᖅᑐᑦ ᐅᑭᐅᖅᑕᖅᑐᐊᑦ ᐃᑎᖅᓴᖓᑕᐅᑉ)

P. Coupe

P. Coupel

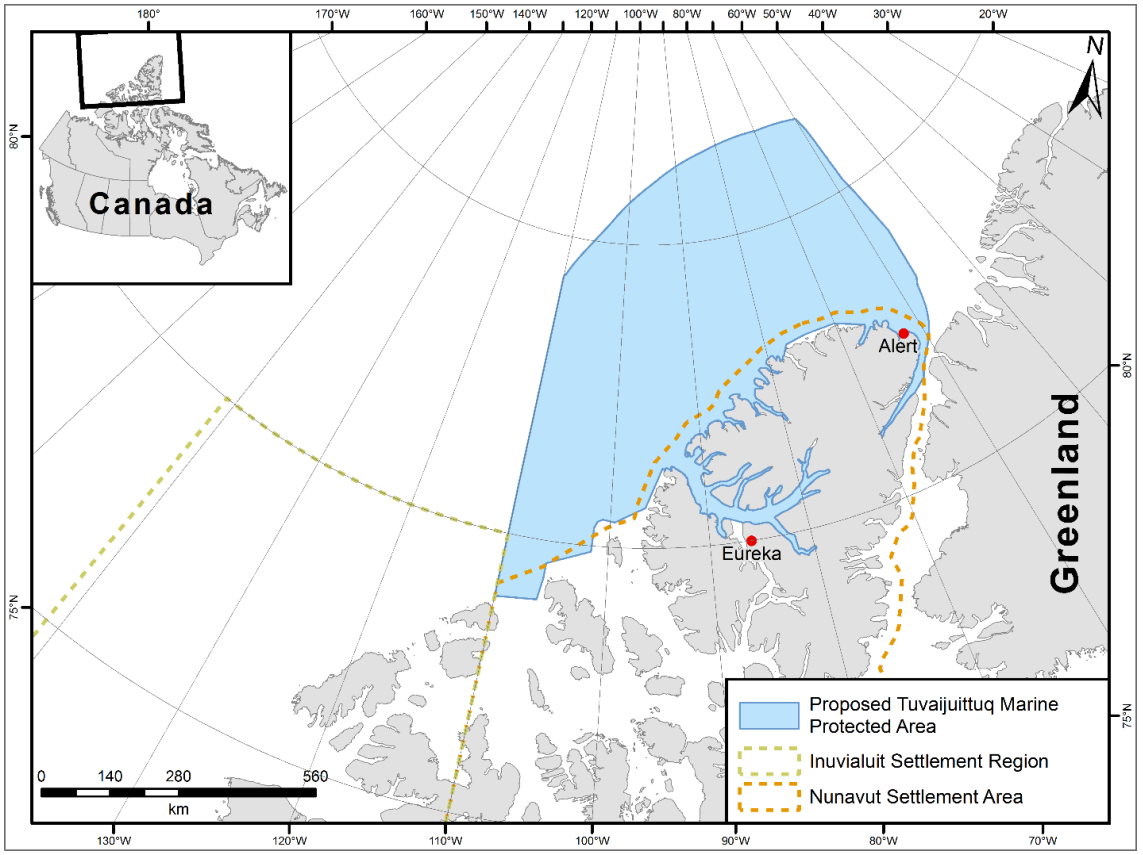
ᐅᒃᑐᑎᑭᖅ ᓴᖅᑭᓇᓂᔪᓂᒍᑦ ᓴᐃᒥᖃᐅᓂᒃᑐᒃ ᐃᓂᐅᒃ ᐃᓗᐊᔪᑦ ᑐᓐᓇᐃᑎᒍᖅ ᑎᑭᓐᓄᑦ ᐊᖏᖅᓯᓗᓂ ᐊᖏᖅᓯᓗᓂ

- ᑐᓐᓇᐃᑎᒍᖅ
  - ᑐᒃᓯᕆᐅᑦ ᐊᖏᑭᖅᑕᐅᖅᓯᓪᓗᒍ
  - ᖃᒪᓂᖅ ᐱᓐᓪᓗᐊᓕᐅᖅᓂᖏᓐᓇ
  - ᐊᓐᑎᖏᖓ ᐊᓐᑎᑎᒋᐊᓪᓗ ᐱᒻᒪᕆᐅᓂᖏᓕᕐᓂᑦ
  - ᐃᓄᐃᑦ ᐃᓅᔨᐸᑎᒃᓴᐃᑦ ᐃᓯᒪᑎᓪᓗᑎᑦ
  - ᓴᓐᓂᖅᓯᐊᖅᑕᐅᓂᖅᓗᑦ ᐊᑐᖅᑕᐅᖏᓐᓄᑦᑎᐊᓂᖅᑎᓯᒪᓕᖅᑎᑦ
  - ᑐᑭᓯᓂᐊᕐᑎᑦ
  - ᑭᖑᓂᐊᔪᑦ ᐱᔨᒃᓴᐃᑦ

Canada

ᑐᒃᓯᕿᐅᑦ ᐊᖏᑭᖅᑕᐅᕐᓂᑦ ᑐᓗᒍ

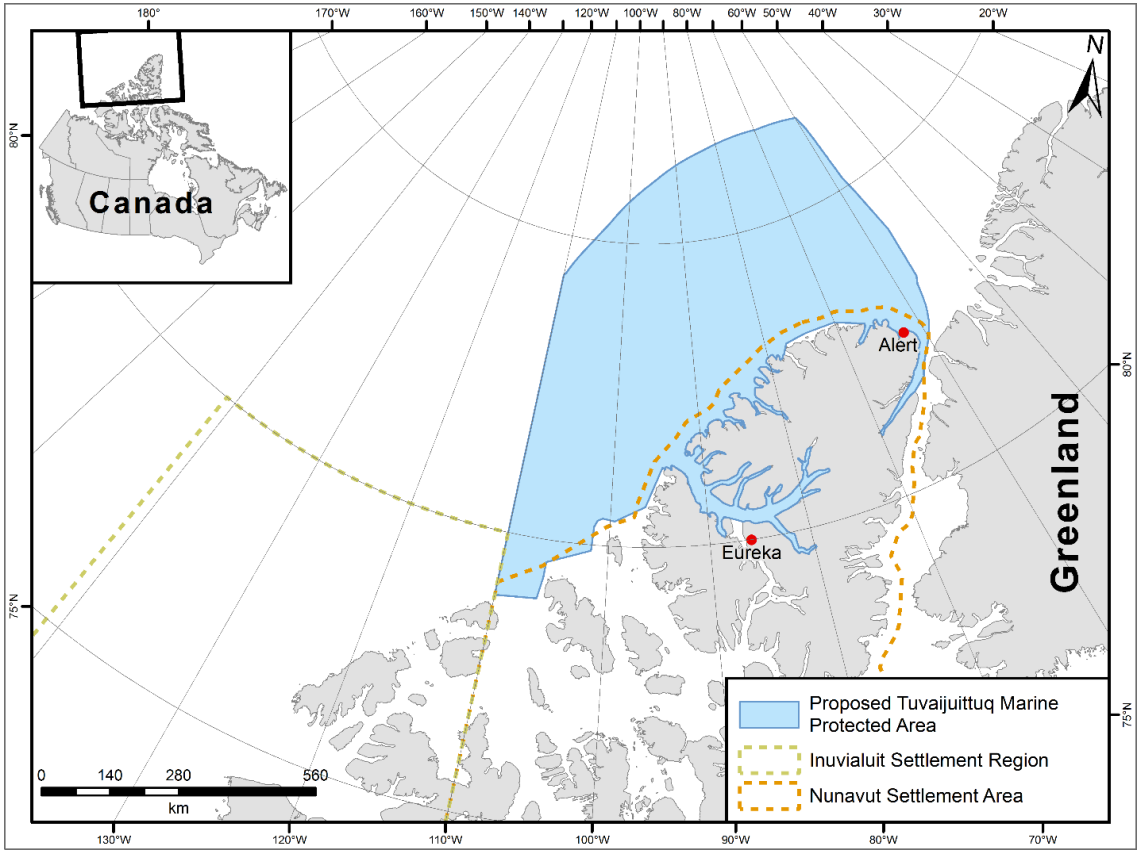
ᐃᕐᒃᑐᑦᓯᒪᔨᕐᓴᐊᓂᑦ ᐊᖏᑭᖅᑕᐅᕐᓂᔨᐅᑦ ᓄᓇᕗᒻᒥ ᐅᒪᓴᑦᓯᒪᔨᕐᓴᐊᒃᓂᑦ ᒃᑎᒪᔨᖏᓪᓗᓂ ᓴᓴᐃᓴᖏᓪᓗᒃ ᑕᕆᐅᖏᓚᓂᒃ ᑐᓐᓇᐅᑉᐳᑦ, ᑎᒃᓂᐊᖅᑕᐅᑐᓄᓂ ᑕᕆᐅᕐᒥ ᓴᓴᐃᓴᖅᑕᐅᓵᓴᐅᑎᒃᓴᖏᓐᓄᑦ ᒥᑕᒧᑦ ᒪᓕᑐᓐᓂᕐᓂᒃ ᐊᑐᕐᓄᓂ ᑎᓕᔪᐅᑎᓂᓂ, ᐱᕈᐊᖅᑎᑕᐅᓵᒪᓵᓂᒃ ᑕᕆᐅᕐᒪᑦ ᒪᓕᑐᕐᓂᒃ ᐋᖅᑭᒋᐊᕈᑎᓂᒃ (ᒪᓕᑐᒃᓴᖅ ᒃᓭᑕᒥ C-55-Γ). ᒪᓕᒃᑕᐅᑎᑐᓗ ᓄᓇᕗᒻᒥ ᓄᓭᒋᕈᑎᓄᑦ ᐊᖏᒋᕈᑎᑦ ᑎᑎᕐᓯᒪᓂᖅᓯᒋᐊᓂ ᑕᒧᓂ 9.3.2



Canada

ᑐᓴᐃᓗᐃᓕᑐᖅ “ᐃᒪᓐᓄ ᑐᓴᖅᑲᐃᓂᓇᐅᔭᖅᑐᖅ ᓗᑯᓐᓄ ᐊᐅᖅᑕᓐᓕᒫᒪᒪᓕᓕᓕᒃ”

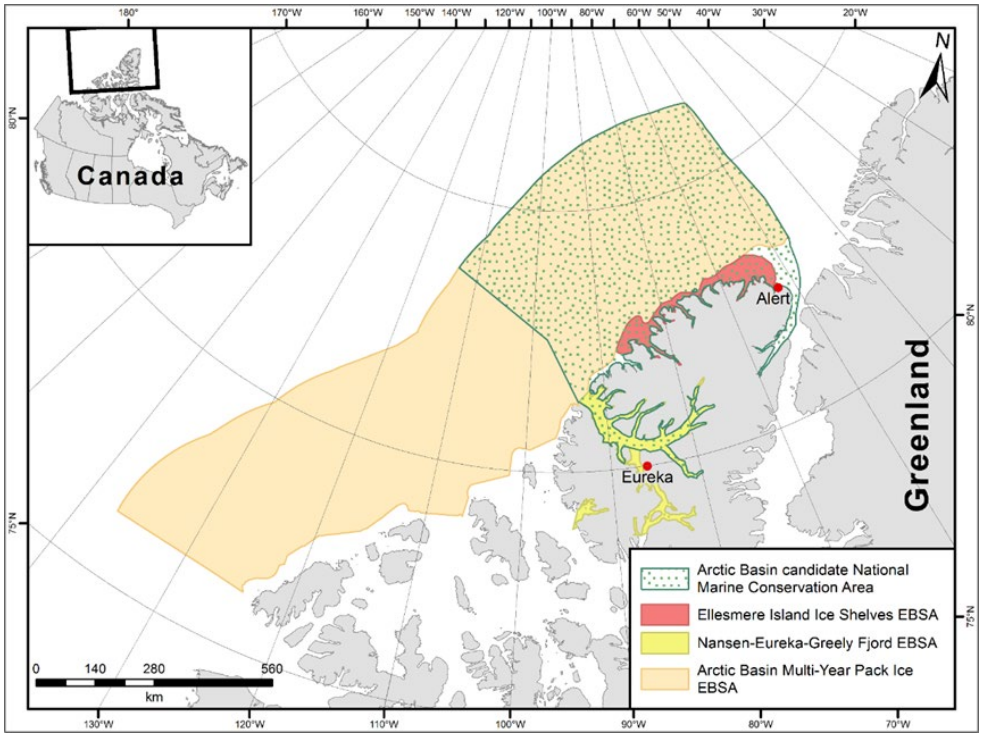
- ᐃᓂᒋᓪᓚᖅᓴᖅ ᐱᓐᓄᓂᓰᔪᑦ ᐃᐅᑦᓱᒥᐅᓪ ᖃᑭᖅᑕᐅᔪᓐᓇᑦ ᐊᒻᒪᓗ ᐅᓇᓕᑕᐅᓪ ᐅᑭᐅᖅᑕᖅᑐᓐᓄᑦ ᖃᑭᖅᑕᒪᓴᓐᕈᑦ
- ᐱᑕᖅᓯᖅ ᐃᑕᖅᓂᒌᕐᔪᓂᐅ ᐳᖅᑕᓂᖅ ᓯᑯᑐᖅᓂᑲ ᐅᑭᐅᖅᑕᖅᑐᓪᓚᒥᑦ
- ᒎᓂᓭ ᐃᓂᐅᔪᖅ ᓂᓚᐅᕈᕐᖅᖃᐃᐅᔪᖅ ᐊᐅᕐᖅᑯᑦ ᔪᑯᖅᐃᓂᓭᓕᐊᑕ ᐊᑯᓂᐅᒌᕐᔪᒃ ᔪᑯᖅᖅᑕᑦᓗᓂ



Canada

ᓯᓇᐅᖏᒪᑦ ᑐᖓᐃᓯᐃᑦᑐᖅ?

- ᐃᓕᑭᕐᖏᑕ ᐃᓂᐅᓴᑦ ᑕᕆᐅᖏᑕᒍᑦ ᓇᓗᓇᐃᖅᑕᐅᓯᒪᔪᖅ ᐃᒪᒻᒥᐅᑕᓕᕆᔨᓴᓇᓂᑦ ᐅᒪᔪᖏᑎᒍᑦ ᓂᕆᔪᑎᖅᑭᑦᑕᓗ ᐃᓱᐊᓂᖅᓴᖅᑕᐃᑦᑐᖅ ᐃᓂᑭᓂᖅ (EBSAs)
- ᑲᓇᑕᐅᑦ ᒥᖏᒍᐃᖅᓯᖓᓕᕆᔨᖏᑦ ᓇᓗᓇᐃᕆᓴᖅᑲᖅᓯᒪᔪᑦ ᐃᓕᑭᕐᑕ ᑐᖓᐃᓯᐃᑦᑐᖓᑦ ᓴᓐᒪᖅᑕᐅᓯᒪᓂᐊᑕ ᑲᓇᑕᒥ ᑕᕆᐅᑲᓂᑦ ᓴᓐᒥᖅᑕᐅᓯᒪᔪᑦ ᐃᓂᑭᓄᓂᑦ (NMCA)
- ᓚᖓᒪᐅᖏᑦᑐᑦ ᑎᒥᖏᓂᖏᑦ ᐃᓕᓴᓂᐊᖅᖁᓴᐃᕆᐅᓂᖅᓯᖏᓐᓄᑦ ᓗ ᖃᐅᐅᓴᖅᑲᖅᓯᒪᔪᑦ ‘ᑭᖑᑦᓕᐅᓂᖏᓴᓂᒃ ᓯᑎᑐᖅᑲᖅᑲᖅᖁᐅᓂᖅ’ ᐃᓕᑕᕆᓴᐅᓯᒪᓕᑐᓂ ᑕᒪᓴᓐ ᓴᓐᓴᓂᐊᖅᑕᐅᓯᒪᓕᕆᖅᖃᖅᓂᖅ



Canada

ᓯᖏᒡᑕᕐᓂᐊᓕᖅᑐᖅ ᐱᒻᒪᓗᐅᕐᓂᖅ?

- ᐊᖏᖅᔪᒍᓴᖅᓂᒃ ᓯᑯᑐᖁᓕᐃᓪᑕ ᐱᑕᖏᒃᐅᐃᖅᓗᖅ>ᑦ ᑕᐅᕐᓂᓂ
  - ᐱᒻᒪᓗᐅᔨᓂᒃ ᐅᒪᔨᖏᒃᓯᖅᖅ ᓯᑯᒥᓯᖏᔨᓂᒃ
    - ᓯᖏᓐᓄ ᐅᑲᐃᑦ, ᐊᓂᓂᐃᑦ, ᐊᑦᓯᑦ, ᐊᐃᐃᓴᑦ
  - ᐊᑐᑎᓪᓕᑦ ᓴᓗᐢᓂᐊᑭᑕᐅᓯᑭᑎᑦ ᓯᑯᐠ ᐊᑕᓪᓂᓯᖏᔨᓂᒃ
    - ᑐᖏᑲᐊᐅᔨᑦ ᓂᐊᓕᓐᓂᓐᓄ ᐊᐃᓴᑐᖏᓯᓐᖏᓕᓂᓐᓄ ᐅᖏᑭᓴᖅᑎᑎᓐᓂᐅᔨᓂᒃ ᐊᓯᖏᓐᓂᒃ ᐅᒥᑐᐃᓐᓕᖅ ᐊᓂᖏᓐᓂᒃ ᐅᒪᐅᖅᑕᖅᑐᒥ.
- ᓂᓚᐅᑎᔾᐅᔨᑦ ᐱᑕᖏᒃᓚᐊᓕᓴᓚᓗᓚᐊᔭᓕᓐᓂᐊᖅᑐᑦ ᓯᑯᒦᓯᖏᔨᓂᒃ ᐅᒪᔨᓂᒃ ᐊᔨᔪᓚᓴᐅᔨᓐᓕᖏᓐᓂᖅ ᓯᑯᐠᓄ ᐊᑕᓂᓯᖏᔨᓂᒃ ᐅᒪᓐᖏᒃᑐᑦ, ᓯᕐᓂᓂᒃᓴᒥ ᑕᓚᐅᖏᒥ ᓯᑯᐃᔾᖅᕿᓕᓕᐊᔭᐊᓕᖅᕿᑦ

Canada

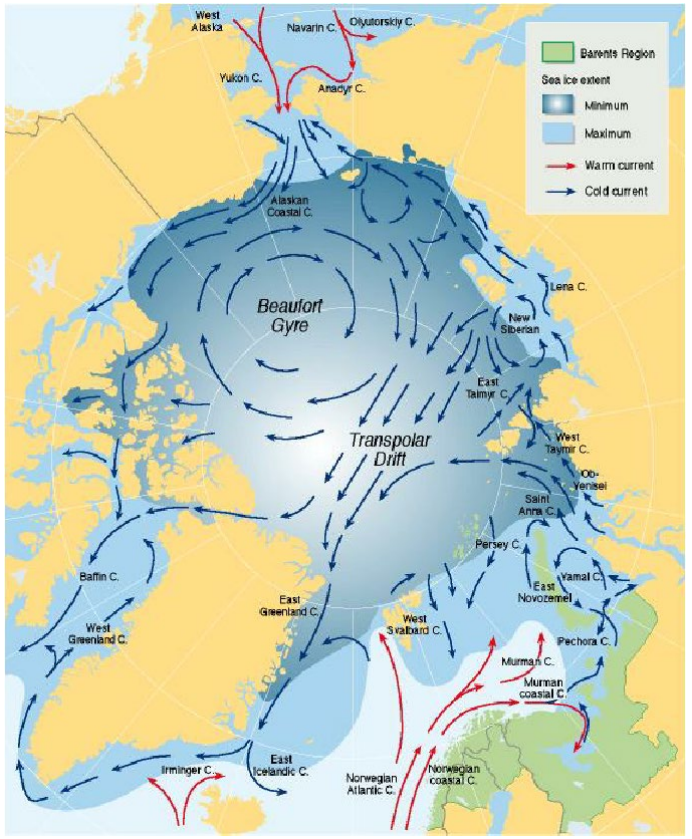
ᓯᖏᒃ ᑐᓂᐃᔨᐃᓕᑐᖅ ᐱᒪᒪᓗᐅᓂ??

ᐊᓱᓕᖅᖃᕐᕋᖅ ᐅᒪᔨᑦ ᐊᐅᓪᓚᔨᓯᖏᓐᓄᑦ

- ᒋᐦᓇ ᐸᐅᘐᐊᑦ ᑲᐃᓂᓪᓕᐅᒋ, ᐅᑭᐅᖅᑕᖅᑐᑦ ᑕᓗᐅᖏᓕᑕ ᐃᖏᕐᕋᓂᓂᖏᓐᓄᑦ ᐃᓪᓕᖅᑕᖅᑐᖅ ᓄᓇᕗᐊᓪᓕᒻᒥᑦ ᑲᑎᖅᓯᖏᓐᓇᕋᖅ ᐃᒪᕐᑎᐊᓂᐅᓂᖏᓐᓄᑕ ᑲᓇᑕᐅᐸ ᓄᓇᕋᑯᒋᓂᖅ ᐊᓪᓚᓗ ᖃᓯᐊᖅᑯᑦ ᐅᑭᐅᖅᑕᖅᑐᖅᑐᐊᑕ ᑯᖏᑦᓂᖅ

- ᑕᒪᐦᓇ ᐃᒧᑐᐸᑦ ᐃᖏᕐᕋᓂᖏᓐᓄ ᐱᓕᕆᐅᓗᐊᖅᑐᖅ ᐊᒥᓱᓄᑦ ᓱᑯᑐᖅᑭᖅᑦᓪᓗᑦ ᐅᑭᐅᖅᑕᖅᑐᒥ, ᑕᐅᘐᓇ ᐃᖏᕐᕋᓭᖏᓐᓄᓕᓄᑦ ᐅᑭᐅᖅᑕᖅᑐᖅᑐᐸ ᖃᓇᑕᐅᖅᑕᖅᑐᕐᓂᖏᓐᓄᓕᓄᑦ

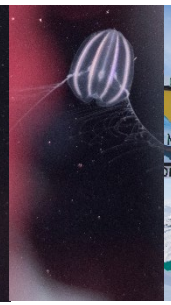
- ᐊᓭᑎᓐᓄ ᐊᓭᑎᒥᐅᑕᖏᑦᓯᓪᓗ ᑕᒧᓂ ᖃᐅᔨᓪᓚᕐᐅᑎᓂᖏᓐᓄᑦ ᖃᐅᔨᓴᖅᑕᐅᑎᓂᐊᕆᐊᖅᑐᖅᑐᑦ ᑐᑭᓯᐅᖅᑕᐅᓗᑎᒃ ᖃᓄᐃᑦᑐᓂᖏᓐᓄᓕᓄᑦ ᐊᓭᑎᒥᐅᑕᒪᓴᖏᓐᓄᑦ



Canada

ᑲᔪᓯᔨᒃ ᖃᐅᔨᓴᖅᑕᐅᓂᖏᓐᓄᑦ

- ᐊᓴᐲᖏᓐᓂᑐᑎᓂᒃ ᐅᑭᐅᖅᑕᖅᑐᒥ ᐱᓕᕆᐊᖓᔪᓂᖏᓐᓂᒃ (MAP) ᐱᒋᐊᓕᐅᖅᑐᑦ 2018-ᒥ; ᑲᔪᓯᓂᐊᖅᑐᑦ ᑕᒪᐅᖏᓐᓄᑦ 2022-ᒧᑦ
    - ᐊᑐᖅᑕᐅᓂᖏᓐᓂᒃ ROVs/AUVs ᑐᐃᖅᓴᐅᔨᔪᒋᑦᓗ ᐊᑐᑭᓐᓇᖅᑐᑦ ᔪᑾᔦᑎ ᐊᑕᓂ ᐊᓪᒪᓗ ᐱᑕᖅᖃᓂᖏᓐᓄᓂᒃ ᖃᐅᔨᒪᔪᓐᓇᕐᓂᖅ ᖃᑎᖅᑐᓴᖅᓯᐅᖓᑐᓂᒃ ᐅᒪᔪᓂᒃ ᐃᖅᑲᐅᓴᐃᓗᑦᓗ ᐊᓴᐲᖏᓐᓄᑦ ᐱᑕᖅᓯᐅᓂᖏᓐᓂᒃ
    - ᔪᑾᑾᔪᑦ ᐃᑾᑕᖅᐃᐅᓗᐊᓂᒃ, ᑲᑎᖅᔪᐃᐃᐅᓂᐊᖅᒪᑕ ᐱᑕᖅᑲᓪᖅᕿᓪᓕᐊᓂᖏᓐᓄᓂᒃ ᐊᐋᑎᖏᓐᓄᓪᓗ ᔪᑾᖏᓐᓄᑦ ᖃᐅᔨᓴᖅᓴᖅᑲᐃᑦ ᐊᒧᖅᑲᐃᐃᐅᓗᐊᓂᒃ ᑕᐅᕐᓂ ᑲᑎᖅᔪᐃᓂᖏᓐᓄᓂ ᖃᑌᐅᕐᑕᓂᒃ ᖃᐅᔨᓴᕆᐅᕙᓪᓕᐊᓂᐊᖅᑐᖅᑐᓂᒃ
    - ᖃᐅᓴᑕᔪᑾᑦ ᖃᐅᔨᓴᖅᑕᐅᓗᐊᓂᒃ ᑕᐅᐋᓴᑦ Archer Fiord ᑲᓇᖅᔪᓴᓂᒃ



ᐃᓄᐃᑦ-ᐃᓄᒃᓴᐋᑎᒃᑲᓇᖅᑐᖅᑕᖅᓯᒪᔪᑦ ᐃᓗᒪᑎᑲᐅᓂᖅᓯᒪᔪᑦ

- ᐊᓐᓇᐅᔾᔨᖅ ᐊᒥᓗᖅᓯᒪᔨᖅᑐᑦ ᐱᓕᕆᔨᑦ ᑕᐅᖅᓂ ᐅᖅᓱᓗᖅᓯᒪᓂᓄᑦ ᑎᑭᑕᐅᓴᕐᓂᖅᓯᒪᓂᓕᒃᑐᓗ (12-ᖃᓴᖅᑲᐅᔨᓂ)
  - ᑭᓇᐅᔪᓕᐅᕐᕿᐅᖅᓯᒪᔪᖅ ᐃᕐᑭᐊᖃᐃᔭᐅᑕᐅᓂᕐᒧᑦᓗ ᑲᔪᓚᐊᖅᑐᕐᖄᓂᕐᒧᑦ ᐃᕐᖢᓗᓕᕐᕿᐅᕿᖅᓯᒪᔪᖅ ᑕᓚᐅᒥᐅᑕᓂᖅᓗ ᐱᓕᕆᔨᕐᖄᖅᓯᒪᔪᖅ
  - ᐱᑎᐊᑭᒋᓂᒃ ᖃᒥᕐᑭᓇᖅᑕᐅᓂᖅᓱᓄ ᓇᓗᓇᐃᓚᓕᐅᖅᑐᖅ ᐱᑕᕐᖏᓚᔨᓗᐊᖅᓴᑎᑕ (ᐱᑕᕐᖏᓂᓂᑭᓂᖅ) ᓂᕐᑭᖏᓕᒥᓂᑦ ᐊᖅᐳᓇᓕᖏᑕᓂᒃ
  - ᐅᓴᕐᖄᓂᐊᕐᕿᐅᖅᓯᒪᔪᖅ ᐅᖅᓯᐊᓗᖅᓯᐅᕐᕿᐅᖅᓯᒪᔪᖅ ᓴᓴᒋᒃᖏᐅᕐᕿᐅᖅᓯᒪᔪᖅ ᐱᔨᓂᐅᑎᒋᓐᓇᓂᑦ
  - ᐅᒥᐊᔪᔨᐊᑦ ᐊᖅᑯᐃᓴᖅᔨᓇᖅᑐᑦ (ᒪᓕᒃᔨᔭᒍ ᖃᒥᕐᑭᓴᐅᓂᐊ AIS ᓇᐞᓴᕿᐅᑎᕐᑎᖅᑐᒃᓂᒃ) ᐊᑯᓂᓂᐊᒍᑦ ᐃᓴᔪ 2017 ᐊᓚᒪᓗ ᓇᕆᐱᓇ 2018
  - ᐊᒥᓗᖅᓯᒪᔪᑦ >ᓕᕐᓗᑦ ᐃᕐᑭᐊᖃᐃᔭᕐᓇᐊᖅᑐᖅᓴᖅᑲᓐᓗ ᑕᐅᕐᖓᖅ
- ᐅᑭᐅᖅᑕᖅᑐᒥ ᓗᓕᐅᔪ ᐊᓗᔪᖅᓴᓐᓕᐊᓂᖅᓯᒪᓂᓄᑦ ᑕᓇᐅᕐᒥ ᓗᑯᐃᔭᖅᓴᓐᓕᐊᑐᐃᓐᓇᖅᑐᖅ (ᓗᑯᑐᑲᓗᖅᓂᒃ ᐊᕐᕌᒍᒪᕐᖅᓂᒃ).
  - ᐱᔪᓕᐅᔨᕿᖅᑐᑦ ᐊᑯᓂᐅᓂᖅᖄᖅ ᐅᒥᐊᕐᔪᐊᑕᐅᖅᑲᑕᕿᖅᓂᖅ ᐅᒥᐊᔨᐅᑯᕐᓇᐅᔨᕿᖅᓗᓂᕐᒧᑦᓗ
  - ᐱᖃᓕᓕᕈᐅᔨᕿᖅᑐᑦ ᒪᑯᓄᖅ ᐅᓴᕐᓂᓂᐊᕐᓇᐅᔨᕿᕐᓂᖅ, ᐅᖅᓯᓗᖅᖄᓕᐅᓂᖅ, ᓴᖅᑭᑎᒋᖅᓯᐅᕐᓇᐅᓂᖅ, ᑭᓇᐅᔪᓕᐅᕐᓂᕐᒧᑦ ᐃᖅᑲᓗᓯᐅᕐᓂᖅ, ᖃᐅᐅᔪᖅᓇᐅᓂᖅ, >ᓕᕐᖏᓕᓚᓂᕐᒧᑦᓗ

Canada

ᓴᓗᐅᓂᐊᖅᑕᐅᓂᖓᓪᓚᑦ ᑐᖅᐃᓯᐃᑦᑐᖅ

- ᐃᓱᒪᒋᑦᓗᒋᑦ...
  - ᐊᖏᕐᕈᖓᓂᖅᓂᑦ ᐱᓕᒪᓐᓇᐅᓂᖓᓂᓗ ᓱᑯᑐᖅᐃᑦ ᑐᖅᐃᓯᐃᑦᑐᖅᑎᒥ
  - ᖃᓄᑕᒥᐅᔨᒥ ᐊᓱᖏᕈᖅᓯᓂᖓᓚᓂᖅ ᓱᑕᐅᑦ ᐊᓱᖏᖅᑭᑦᑕᐊᓂᖓᓚᓄᑦ
  - ᓱᖅᑲᔨᒥᖅ ᐱᕈᖅᑐᖅ ᐱᔨᒪᓂᖅᑲᓂᖓᑎᑕ ᓄᓇᕐᔨᐊᑕᒧᒪᒥᑦ ᐅᒥᐊᕐᔨᐊᖅᑯᑦ ᐅᓱᑲᖅᑕᐃᓂᒪᒥᖅ ᐅᑎᑲᖅᑕᖅᑐᖓᓚᓂᖅ (ᓴᖅᑲᖅᓯᓂᓗᒃᑐᖅᑐᒥᒃ ᓱᑕᐅᑦ ᐊᖏᕈᑎᓂᓪᓗᒍᑦᑐᖅᑐᒥᒃ ᐅᑭᐅᖅᑕᕐᓂᖅᓯᒪᓂᖅ ᐊᖅᑐᐃᓂᖓᓚᓄᑦ ᐃᓂᐊᓄᑦ); ᐊᓕᓗ
  - ᐊᐊᑎᖓᓚᓂᓂ ᐊᐊᑎᒥᐅᑕᖓᓚᓂᓂᓗ ᐱᑕᖅᑕᕐᓂᖓᓂᑦ (ᓴᖅᑲᓱᓯᑦ, ᓄᑎᖅᑲᓱᓯᑦ ᐊᓂᖅᓂᖅᖅᓴᖓᑕᑦᑕᓚᐊᖅᑯᓯᐊᓪᓗ).
- ᓴᓗᑎᓴᐅᓯᒪᔪᑎᖅ ᐃᒪᓂᓚᐃᑕᖓᓕᕆᓂᖅᓂᐊᖅᑐᑦ...
  - ᓄᓗᓄᓚᐃᓴᐃᓂᖅ ᐱᓕᒪᓐᓇᐅᓂᖓᓚᓂᓂᓂᖅ ᐃᑕᒋᖅᓯᕈᑕᐅᓗᑎᑎᓗ ᐱᑕᓇᖅᑎᑎᕈᐊᖅᖅᓂᖅᓗᑦ ᖃᐅᔨᖅᓂᖅᓚᓂᓂᖅ ᑕᐅᔨᓂᓄ
  - ᖃᐅᔨᐃᑕᖓᓕᕆᓂᖓ ᑐᓴᐅᒪᓴᐅᓂᖅᓴᐃᐅᓗᓂᓂ ᖃᐅᔨᒪᓴᐅᔨᓂᕈᑕᐊᓗᓂᓗ ᑲᓄᑕᕐᑕᐅᑕᓄᑦ ᐃᓄᖅᓴᓐᓂᖅᓗᑦ ᐃᓄᓕᕆᓂᓗ ᐃᑕᒋᓴᐅᓴᓇᖅᑎᖅᖅᓂᖅᓚᓂᓂᖅ ᐱᑕᖅᓴᕐᓇᒋᓴᖓᓄᑦ ᐃᓄᒋᓴᐅᒪ ᐊᖏᖅᓴᒪᓴᖓ ᑕᐅᑐᓗᑦᑐᓗ.

Canada

ᓴᓗᐅᓂᖅᑕᐅᓂᖓ ᑐᙵᐃᓱᐃᑦᑐᑦ

- ᐅᒃᑐᑎᑭᒃᑯᑦ ᐊᑐᖅᑕᐅᓂᖓ ᓯᕗᓪᓕᖅᐸᒥᒃ ᐊᑯᐅᓂᖅᓯᑎᒥᒃ ᓴᓗᒥᓪᓚᕐᓇᑎᓂᒃ ᓴᖅᑭᑦᓴᖅᑐᖅ ᑕᓚᐅᐸ ᐱᖅᑯᓴᕐᐊᖓᓚᑦ ᑕᓚᐅᓕᒥ ᓴᓗᓕᖅᑕ ᐃᓂᖓᓚᒃ ᐊᖅᑯᑎᖅᓴᓗᑎᒃ ᒥᓂᐅᓚᕐᓇᑎᒥᒃ (ᒪᓕᒃᑲᖅᑐᖅ ᐊᖏᑭᖅᑕᐅᓂᖓᓄᑦ ᑯᐃᖏᒪᑦ ᐱᖅᑯᓴᕐᐊᒃᓴᖅ C-55).
- ᐊᑯᓂᖅᓯᑎᒃᑯᑦ ᓴᓗᒥᑲᐅᓂᖓᓚᑦ ᒋᒃᑯᐊ ᖁᑭᖅᑕᓂᒃᑯᑦ, ᓄᓇᕐᐸᐅᑦ ᓚᕿᒪᖏᑦ ᐊᒪᒪᓗ ᒃᓇᑕᐅᐸ ᓚᕿᒪᑐᖅᖏᑦ ᐱᓕᓗᑎᒥᒃ ᒃᑐᓵᖅᑎᑭᒥᓂᒃ ᐃᓕᐅᖅᑕᐅᕙᕐᓇᖅᖅᑐᓂᓪᓗ ᐱᒋᐊᓚᖅᑭᓗᑎᒥᒃ ᐱᔪᒪᓇᖅᑎᑦᓯᐊᑕ ᖁᑭᖅᑭᑎᖅᑯᓗᑎᒥᒃ ᒃᑐᖅᑭᑦᓴᖅᑎᒥᒃ ᐃᓕᐅᖅᑕᐅᕙᕐᓇᖅᖅᑐᓂᓪᓗ ᖁᑭᖅᑭᑎᖅᑯᓗᑎᒥᒃ ᐊᑯᓂᖅᓴᓗᑎᒥᒃ ᐃᓕᐅᖅᑕᐅᕙᕐᓇᖅᖅᑐᓂᓪᓗ ᐃᓕᐅᖅᑕᐅᓂᖓᓄᑦ ᐃᓕᐅᖅᑕᐅᓂᖓᓄᑦ ᐊᑯᓂᖅᓴᓗᑎᒥᒃ ᐊᑯᓂᖅᓴᓗᑎᒥᒃ ᐊᑯᓂᖅᓴᓗᑎᒥᒃ

ᑕᓇᐅᕐᒧᑦ ᒪᓕᒐᒪᐃᑦ ᒥᓂᑕᐅᑦ ᑎᓕᓴᕐᕕᑎᖅᑎᑦ

- ᒥᓂᑕᒧᑦ ᑎᓕᐅᓇᓱᕈᑦ MPA ᐊᑐᖅᑕᐅᓗᓂᓂ ᑕᓇᐅᑦ ᐱᖁᔪᓱᐊᖅᓯᓂᒃ* ᐃᒪᐃᒃᑯᓱᖅᑐᖅ:
  - ᖁᓕᐊᖅᑎᑕᐅᓗᑎᒃ ᐱᓕᓇᓂᓇᔪᐅᓴᑦ ᐊᒃᑐᐃᐅᓂᖅᖽᐊᖅᑎᑦ ᐃᓂᓯᓂᒃ ᑎᑭᓪᓗᐃ ᑕᑦᓕᒪᑦ ᐊᖁᔅᑐᐃᑦ, ᐃᓗᐊᒍᑦ ᐱᓕᓇᓄᑦ ᑕᒪᐃᐅᓂᒃ ᓴᖅᑭᓴᒪᔨᐊᓄᑦ ᐃᓂᑎᒍᑦ ᑕᓓᓇᓗᑕᑦ 12 ᑕᖅᑭᖅ ᓴᐅᓂᐊᑕ ᑎᒃᓓᐊᖅᑕᐅᓂᓇᔪᑕ (ᐊᔨᖏᑎᑎᑕᐅᔨᓪᓗ ᒃᔨᓴᔨᑦ) ᐊᔨᖅᑎᑕᐅᑦᓴᖏᑦᑕᑦ ᒃᔨᓴᑎᑕᐅᖅᑎᑭᖏᓪᓗᑎᒃ
  - ᐃᓗᐊᒍᑦ ᑕᑦᓕᒪᑦ ᐊᖁᔅᑐᐃᑦ, ᐊᓴᖏᑦ ᐅᖅᓯᓴᖅᑎᓗᒍᑦ ᓴᓪᒥᑎᓇᓱᐊᖅᑐᓴᑦ ᑭᖑᓂᖅᓱᓕᔅᓴᖅᑐᑦ ᒥᓂᑕᐅᑦ ᑎᑎᕋᖅᑕᐅᖅᑐᓂ. ᑕᐃᒪᓕ ᑕᐅᕐᖏᖅᓯᒪᔪᖅ ᓴᓪᒥᑎᓇᓱᐊᑎᑦ ᓴᖅᑭᑕᐅᖅᑭᒃᑯᑎᒃ, ᑕᓐᓇ ᒥᓂᔅᑐᑦ ᑎᓕᐅᓇᓱᕈᑦ MPA ᐋᖅᑕᐅᓂᐊᖅᑐᖅ
  - ᓇᓕᓕᐊᑐᐃᓇᖅ ᑕᒪᐃᓐᓂᓗ ᓴᓪᒥᑎᓇᓱᐊᖅᑐᓴᑦ ᒪᓕᒋᐊᖅᑎᑦᑐᑦ ᓄᓇᕗᑦ ᓄᓇᕆᑕᒋᓄᑦ ᐊᖏᕋᒋᐊᓂᖅ
  - ᒃᔨᓴᑎᑕᐅᓕᖅᑐᒃ ᐱᓕᓇᓂᑦ ᐃᓕᓴᖅᑐᑦ ᐅᔨᓂᖅ:
    - ᖁᑲᐅᓴᓴᐃᓕᒋᑦ
    - ᐅᓇᕈᖅᑐᑲᓴᑯᑦ ᐱᓕᓇᓂᖏᑦ
  - ᒥᓂᑕᒧᑦ ᑎᓕᐅᓇᓱᕈᑦ ᐊᔨᖏᑎᑎᕐᓂᓗᑐᓂᓂ:
    - ᑐᐊᐅᖁᓇᖅᑐᑦ, ᑭᒃᑯᓕᒪᓄᑦ ᐅᑐᓚᐊᓇᔨᐃᐃᓕᒋᑎᑦ, ᒃᓇᕈᒥᒃ ᓴᓪᒥᑎᓇᓱᒋᑦ, ᒃᓇᕈᒥᒃ ᐅᑐᓚᐊᓇᕆᒋᑦ ᐅᖏᓐᓄ ᐱᖁᔪᓂᒃ ᐊᑐᖅᑎᑦᑎᓂᖅ
    - ᖁᑲᐅᓴᓴᐃᔨᑦ ᑐᑭᓯᔪᓕᒋᓴᓇᔨᒥᓂᒃ
    - ᐃᓕᓴᖅᑦ ᐱᓕᓇᓂᐅᓯᑦ ᒃᓚᑭᔨᑦ ᒃᓇᕈᐅᑦ ᐊᓯᖅᑭᓄᑦ, ᑎᒥᐅᔨᒧᑦ, ᐅᒥᐊᔨᔪᐊᓕᒧᑦ ᐅᖏᓐᓄ ᓄᓇᓕᔨᔪᐊᖓᔪᖅᑎᑭᓄᑦ

*ᑐᖓᓚᐊᖅᑯᑦ ᒪᓕᓚᓕᐊᖓᔪᖓᓂᒃ ᒪᓕᓚᒃᖅᑯ ᒃᓇᕈᒥ C-55-ᒥᒃ.

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IIBA

- ᐅᖃᐅᓯᖅᑕ 9.4.1 ᓄᓇᕗᑦ ᓄᓇᕆᔨᑎᖏᓐᓄᓇᓂᑦ ᐊᖏᔪᑎᐊᓂ ᑎᓪᓗᐅᓚᖃᖅᖃᑐᖅ ᓚᖓᒪᓂᑯᑦ ᑎᒥᖃᑎᖏᓪᓗᑦ ᐱᓪᓚᐃᖃᖅᖃᑐᑦ ᓴᓪᒥᑕᔭᓂᑦ ᐃᓂᐅᔨᓂᒃ ᐊᓪᒪᓗ ᑎᒃᑯᐊᖅᑕᐅᓯᒪᔨᑦ ᐃᓄᐃᑦ ᑎᒥᖏᓪᓗ ᐊᑦᑐᖃᑎᖃᖃᖅᑐᑦ ᐃᓄᖏᓂᒃ ᐊᑦᑕᐃᓂᒪᓗᑦ ᐱᖓᓪᓕᐅᑎᓄᑦ ᐊᖏᔪᑎᐊᓂᒃ (IIBA)
- ᖃᓄᑕᒥ ᐊᓪᒪᓗ ᖁᑭᖅᑕᓂ ᐃᓄᐃᑦ ᖃᑐᐸᖃᑎᖃᑎᒋᖏᓪᓕᑦ ᐊᑦᑐᖃᑎᖃᑎᒋᑐᖅ ᑕᓪᓗᑭᑎᐅᓕ ᐃᓚᖏᑕ ᖃᓄᑕᒥ ᕆᓚᐅᖅᓯᒪᔨᑦ ᐊᓪᒪᓗ ᖃᑐᐸᖃᑎᖃᑎᒋᑐᖅ (TINMCA)
- ᐊᖏᕐᖃᑎᒋᒃᓴᓪᒪᓕᐅᑦ ᑭᓐᓇ ᐊᑦᐱᑕᑕᐅᔨᖅᖃᐊᖏᑕᑭ ᖁᑎᕐᑐᐃᓚᕆᓕᐅᖅᖃᐊᓱᑕᓂᒃ ᓴᖅᑭᑦᑐᓐᓇᖅᑐᓂᒃ ᕆᓚᐅᒃᑯᑦ ᓴᓪᒪᐅᑦᑕ ᐃᓕᓐᓂ ᑐᓐᓇᐃᔨᐃᑦᑐᖅ ᐃᓕᓴᕐᕕᐊᓂᒃ ᐊᑦᑐᖃᑎᒋᒃᑐᑦ ᓴᓪᒪᓕᐅᑦ TINMCA ᑎᑭᓪᓗᒍ ᔪᑦᓂ 2019

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ᑐᑭᓯᓂᐊᑭᑎᕆᔨᖅᓂᖅ ᑐᖅᐃᓕᐃᑦᑐᒥᒃ ᐱᓕᓂᖅ

- ᐃᓕᐅᑎᑕᐅᓂᖏᑦ ᓄᓇᓕᑦ ᐃᓕᑭᐳᐅᓗᐊᑕᕆᓂᐊᖅᑐᑦᓗ 2018-ᒥ ᐃᓱᒪᑭᐳᐅᓂᖏᓐᓂᒃ ᐊᑐᖅᑕᐅᑲᐃᓐᓇᑌᐊᕆᔨᖅᑐᑦ ᓴᓗᓐᓂᐊᓕᒻᒥᑦ ᑐᖅᐃᓕᐃᑦᑐᕐᒥᒃ
- ᐃᓕᐅᑎᑕᐅᓕᕐᒥᔾᔨᑦ ᐊᖏᓄᓇᓱᑦᑕᑦ ᑲᑐᔪᕆᑎᑦᑕᖏᓐᓂᒃᓴ ᓄᓇᓕᓕᓗ ᖃᐅᔨᐃᑦᑐᕐᒥ, ᐊᐅᔨᐃᑦᑐᕐᒥᒃ, ᐃᒃᓚᐊᖅᔪᖓᒥ ᕿᑐᐊᓇ 2019-ᒥ
- ᑐᑭᓯᐅᒪᖃᑎᒌᒍᑎ ᐊᖏᕐᖃᑎᒋᐊᕐᔪᒃ (MoU) ᐊᑎᓕᐅᓇᓴᐅᓕᐅᖅᑐᖅ ᑲᓇᑕᐅᕐ ᓴᖅᒪᑐᖅᑎᓂᒃ, ᓄᓇᕗᕐ ᓴᖅᒪᖓᓂᒃ ᐊᓪᒪᓗ ᖁᑭᖅᑕᓂ ᐃᓄᐃᑦ ᑲᑐᔪᖃᑎᒌᖓᓂᒃ ᐄᓴ 2019 ᐃᓗᐊᓂ ᓴᖅᑭᑦᓗᑎᒃ ᐊᐅᓪᓚᐱᑎᒃᓴᓂᒃ ᖁᒥᖅᑭᐳᐅᓂᖓᓪᓕᓂᑦ ᐱᔪᒪᓇᖅᑐᐊ ᑲᔪᓯᑎᐊᑭᒥᓇᑭᑦᓗ ᓴᓪᒥᒻᒪᓂᒃ ᑐᖅᐃᓕᐃᑦᑐᖅ
- ᑐᖅᐃᓕᐃᑦᑐᖅ ᑐᑭᒧᐊᓇᔨᑦ ᑲᑎᒪᔨᖅᑕᖏᑦ (ᐃᓕᓴᕐ 2019) ᓴᖅᑭᒥᐅᓕᐅᖅᑐᖅ ᐊᒋᔪᑦ ᐊᖏᕐᖃᑎᒌᒍᑎᐊᒥ MoU ᑐᑭᒧᐊᓇᓗᐅᑦ ᖁᒥᖅᑭᐃᓕᐊᑎᒌᓂᑦ ᐊᐅᓚᓂᓇᑲᓂᒃ ᐱᖅᓯᐅᑎᓕᓗᒃᑑᒃ:
  - ᑲᒪᒥᖅᑲᓂᓂᖅ ᖃᑲᑲᓂᓂᐊᔪᑦ ᑐᑭᒧᐊᓇᑲᑎᒌᓂᒃ ᑲᑎᒪᐊᖅᓯᒪᓂᒃ ᑲᑐᔪᖃᑎᒌᑲᓂᑦ ᐊᐅᓚᕆᑎᖃᑎᒌᒐᕐᓂᒃ ᐃᓚᐅᓂᖃᖅᑐᑐᒃᓗᑦ;
  - ᐅᖃᐅᔨᔨᖃᓯᐃᓗᑎᒃ ᐊᑐᖅᑕᐅᔾᔪᓇᖅᓂᖏᓚᒋ ᐊᑯᓂᓂᐊᔪᑦ (ᓇᐃᕐᑐᑯᓂ) ᓴᓗᒥᖅᐳᐅᓂᖏᓐᓄᑦ; ᐊᒻᒪᓗ
  - ᐊᑐᕆᔭᕆᒥᓂᒃ ᓴᖅᑭᘅᓗᑎᒃ ᓴᖅᑭᖅᓴᖅᑭᕈᒪᓂᕐᒧᑦ ᐊᕐᕌᒍᑕᓴᑯᓂᑦ ᓴᓗᒥᑌᓴᐅᔾᔨᒋᓂᑦ ᑐᙵᓚᘅᓴᖏᓂᒃ

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ᑭᖏᑦᑕᖅᐳᓪᓗᑦ ᐱᒃᖃᓗᒃᑯᑦ ᐊᑯᓂᐅᔪᑦ ᓴᓱᒥᐅᑎᓄᑦ

- ᑕᓚᐳᖅᓯᑦ ᒪᑕᑌᐃᑦ ᐋᖅᑭᒋᐊᖅᑕᐳᕐᓗᑎᒃ
- ᒪᑕᒋᐊᑦᑕ ᓄᓇᕗᒻᒥ ᐊᖏᑭᑭᑎᓂᒃ
  - ᒪᑕᖏᒪᖏᑎᑕ ᖃᐅᔨᕐᐳᓗᑎᒃ ᒪᑕᒋᐊᑦᖏᓂᒃ ᓄᓇᕗᒻᒥ ᐸᖅᓇᐃᔨᑦ ᑲᒥᓯᓇᑦᑯ ᖃᓄᐃᐊᑦᖏᑎᖅᖏᖏᓂᒃ
  - ᐊᕐᑎᑎᑕᓚᔨᑯᑦ ᑲᑎᒪᔨᖏᑦ ᖃᒥᖅᑭᐊᖅᓗᑎᒃ ᒪᑕᒃᓯᒪᖏᒪᖏᑎᑕ
  - ᓄᓇᕗᒻᒥ ᐅᒪᔪᑦᑕᓚᔨᖅᔪᐊᒃᑯᓐᓄᑦ ᐊᖏᑭᑭᑕᐳᓂᖏᑦ
- ᐱᓴᓱᖅᑕᐳᓂᖏᑦ ᐃᓄᐃᑦ ᐊᒃᑐᖅᑕᐳᓯᒪᓂᖏᓐᓄᑦ ᐃᑲᔪᓯᐊᒃᑲᐃᑦ ᐃᑦᑕᑕᓚᖅᐳᓂᖏᓐᓄᓇᑦ/ᑐᕐᐃᔪᐃᑦᑐᖅᑭᑦ
- Canada Gazette I

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ᐊᐱᕐᓂᑎᖅᑕᕐᒃ?

ᐱᑭᐊᖅᑎᑕᐅᔾᔨᒪᔨᑦ ᓴᖅᑭᑎᑎᔨᒪᓂᖅ ᓴᐳᔅᓂᐊᖅᑕᐅᓯᒪᓕᕆᓂᖅ: ᑐᓴᐃᔨᐃᔅᑐᖅᑎ

ᐃᖃᓗᓕᓕᒫᔨᔨᐊᕆᔭᑦ ᐊᖏᑭᕈᑎᕆᓴᖅᓯᐅᖅᑐᑦ ᐅᒪᔨᓕᓕᒫᔨᔨᐊᕆᔭᓄᓪᓗᓂ ᓴᖅᑭᑎᑕᐅᔨᓄᖅᑕᐅᓂᖏᓐᓂᖅ ᓴᐳᔅᓂᐊᖅᑕᐅᓯᒪᓕᕆᓂᖓᓱᓂ ᑐᓴᐃᔨᐃᑐᑉ ᐊᑐᖅᑐᑎᓗ ᑎᓕᐅᖅᑕᐅᓯᐊᖅᑕᐅᓂᖅᓴᖅᑎᑎᔨᓄᖅᑕᐅᓂᖅᑐᖅ ᒥᓂᑕᖏᓐᓄᑦ ᑎᑎᕋᐅᓕᖅᑐᑦ ᐊᑐᖅᑐᑎᓗ ᑕᓚᐅᖅᓯᑦ ᒪᓕᒐᓴᑦ ᐋᖅᑭᑎᐊᖅᑕᐅᔨᓗᑎᒃ (ᑲᓇᑕᒥ ᒪᓕᒐᓱᐃᔨᐸᒃ ᐱᔨᓄᐊᕆᔨᕐᓂᔨᐊᖅᑐᒥᒃ C-55), ᒪᓕᓴᕆᔨᓗᑎᒃ ᑕᐃᖅᑯᓗ ᓄᓇᕗᒻᒥ ᓄᓇᕗᑦᑭᑎᑎᓄᑦ ᐊᖏᑭᕈᑎᖏᑐᓂᒃ ᐊᑐᖅᑐᑎᓗ ᑕᓚᐅᖅᓯᑦ ᒪᓕᒐᓴᑦ ᐋᖅᑭᑎᐊᖅᑕᐅᓂᖅᓴᖅᑎᑎᔨᓄᖅᑕᐅᓂᖅᑐᖅ (ᑲᓇᑕᒥ ᒪᓕᒐᓱᐃᔨᐸᒃ ᐱᔨᓄᐊᕆᔨᕐᓂᔨᐊᖅᑐᒥᒃ C-55), ᒪᓕᓴᕆᔨᓗᑎᒃ ᑕᐃᖅᑯᓗ ᓄᓇᕗᑦᑭᑎᑎᓄᑦ ᐊᖏᑭᕈᑎᖏᑐᓂᒃ

ᐱᓕᕆᐊᓂᑦ ᑐᑭᖏᖅᓯᖅ ᓇᓪᓕᐊᑐᐃᓐᓇᖅ ᐱᓕᕆᓂᐅᓵᓱᖅ ᓴᖅᑭᑕᐅᓵᒪᔪᑦ ᐃᓗᐊᒍᑦ ᐅᖅᑐᑎᑭᓴᐅᓵᔨᑦ ᓴᓐᒥᑦᓴᓇᖅᓴᖓᓂ MPA ᑕᖅᑭᑦ 12 ᐃᓗᐊᒍᑦ ᓯᓚᓂᐊᑕ ᑎᑦᑯᐊᖅᑕᐅᓂᓵᔨᑕ (ᐅᕐᓂᓗ ᐊᖏᕐᕋᖅ ᐱᓕᕆᔨᐅᓇᓵᓂᕐᓂᖅᑕᖅ) ᑲᔪᓯᓇᖅᑎᑕᐅᓴᖅᑐᑦ ᐃᓗᐊᓂ ᐊᑐᖅᑕᐅᔪᒪᔪᑦ ᓴᓐᒥᑦᓴᖅᖃᕐᒥᑦ MPA. ᖃᑲᑲᓂᐊᒍᑦ, ᖃᐅᔨᓴᖅᑎᑦ ᖃᐅᔨᓴᐃᓵᔪᓯᖏᓂᑦ, ᐅᓗᓕᐊᓇᐅᑎᑦ, ᓄᖅᑲᖅᑎᑦᑎᐊᑎᑦ ᐊᒪᓗᓂ ᑐᐊᐃᓴᖅᑐᑦᓯᔨᑦ ᐱᓕᕆᔨᑦ, ᐊᒪᓗᓂᑦᑕᐅᖅ ᑕᒪᑦᔨᐊ ᐱᓕᕆᓂᐅᔨᓯᑦ ᑲᓇᑕᐅᖅᑕᑐᑐᑦᑐᒍᑦ ᓄᓇᓵᔪᐊᖓᖅᑎᑦᓄᑦ, ᑎᑎᐅᔪᒍᑦ, ᐅᒥᐊᓵᔪᐊᖏᒍᑦ ᐅᕐᓂᓗ ᓄᓇᓵᔪᐊᖓᖅᑎᑦ, ᐊᔪᖅᑎᑎᑕᐅᓴᖅᑐᑦ ᐅᑐᑎᑭᓴᐅᓵᔨᑦ ᐊᖏᕐᕋᖅ ᐅᖃᓕᓴᐅᑎᓪᓗᑎᑦ ᐱᓕᕆᐅᓵᔨᑦ MPA. ᐃᓄᐃᑦ ᐊᖓᔪᓇᓕᖅᑐᑦᓯᒪᒍᑦ ᐃᑲᐱᑎᑦᑕᐊᓐᓂᖅᑐᑦᑐᑦ ᐃᓕᓴᐅᔪᖅᑐᑦ ᑐᓴᑎᑦᑕᐅᓴᖅᑐᑦ ᐊᖏᕐᕋᖅ ᐱᓕᕆᓂᐅᓵᔨᑦ ᐃᓕᓴᐅᔪᖅᑐᑦ ᐊᑐᖅᑕᐅᓴᖅᑐᑦ

ᑕᓚᐅᖏᒃᑐ (ᑕᒻᒥᖏᖅ 1). ᑕᒪᖁᑎᒃᖏ ᐊᖏᑭᓯᓂᖅᖏᖅ ᑕᐃᓂ ᐃᓂᓕᐅ ᐊᑭᑎᐊᖏᑭᓐᓄᑐᑦ ᒪᖁᑉᐸᒃ ᓄᓇᑕᑭᑎᑎᒍᑦ ᐊᖏᑭᑭᑎᖅᖁᐅᓂᖏᑭᓐᓄᑐᑦ, ᑕᐃᒃᑯᓄᖏᒃ ᐃᓄᙳᐊᓗᐃᑦ ᓄᓇᕋᑎᖏᑭᓐᓄᑐᑦ ᐊᒪᓚᓗ ᓄᓇᖅᓱᒪᒥᐅᑦ ᓄᓇᑕᖁᕕᓯᒪᔪᖏᑭᓐᓄᑐᑦ. ᑭᓯᐊᓂᓕᑕᐅᖅᑐ, ᑕᐃᑯᐊ ᐱᑎᐊᖅᑎᑕᐅᔨᒪᓂᖏᑦ ᖁᑯᓕᐅᑎᓯᒪᔪᕐᓇᖅᑐᑦ ᓄᓇᖅᓱᒪᒥ ᓄᓇᑕᖁᕕᓯᒪᓂᖏᑭᓐᓄᑐᑦ, ᑕᐃᒪᐃᓂᖏᓚᓄᓕᒪ ᓄᓇᖅᓱᒪᒥ ᓄᓇᑕᑭᑎᓂ ᐊᖏᑭᑭᑎᖅᑎᑕᐅᖅᑐᑦ ᐃᓗᓕᑎᖅᑐᖅᑐᑦ ᐱᔨᔪᑎᖅᑐᖅᑐᖅᑐᑦ ᓴᓂᑭᑎᖅᑎᑕᑭᑭᓐᓇᖅᑎᓯᒪᔪᖅᑐᑦ. ᑕᒪᓐᓇ ᑲᓇᖅᑐᖅᓯᐅᓂᒃᑯᑦ ᐃᓂᖅᑎᒪᑦ ᑕᐅᖅᓯ ᓪᓚᖅ Lincoln-ᑕᓚᐅᕐᓗᐊᓂᖅ. ᑲᓇᑕ ᑕᓚᐅᕐᔪᑎᖅᑎᓂᖅᑎᓗᓄᑦ ᑭᓱᓪᑎᖁᓕᐅᓂᖏᓐᓂᖅ ᐊᑭᕋᖅᑐᑐᑎᖅᒪᑦ ᓄᓇᓪᓂᖅ Dᐃᐊᓐᒪᑲᒥᖅ ᑕᒪᓐᓇ ᐱᓪᓗᒍ ᐊᒪᓚᓗ ᑐᐱᕆᔭᐅᓪᓚᐅᖅᓯᒪᖅ ᓄᓇᐱᓚᓐᓇᖅ 28, 2012-ᖓᒍᓪᓗᒍ ᐊᖏᕋᖅᑎᑦᒍᑎᖅᑲᑭᐊᑕᑦ ᓴᕋᑎᑕᐅᓪᓚᐅᖅᓯᒪᓂᖏᓐᓂᖅᑐᖅ ᓪᓚᖅᑐᖅ ᑕᓚᐅᖅᑎᑦ ᑭᓱᓪᖁᖅ ᐱᔪᓐᓇᖅᑎᑐᖅᑐᖅ ᑎᑭᐅᑎᓯᒪᓂᖏᓐᓂᖅᑐᖅ ᐊᒪᓚᓗ ᐊᖏᑭᑐᑎᓪᓚᕆᔭᖅᓯᒪᓂᖏᓐᓂᖅᑐᖅ ᑲᔪᓯᐊᖅ. ᐊᖏᑭᑐᑎᐅᓪᓚᐅᑭᓐᓇᑐᑎᖅ, ᑲᓇᑕᒥᓗ ᐊᑯᑭᑐᑦᑎᐅᑕᓗᓗ ᓄᓇᖅᓴᐊᖅᑎᐅᓄᑦ ᑭᓱᓪᑎᖁᓕᐅᑎᑕᐅᓪᓚᐅᖅᑐᖅᑐᖅ ᐃᓕᑕᑲᐅᓗᒍᑎᖅ, ᑕᒪᓐᓇ ᐃᓕᑎᓴᖅ ᑭᓱᓪᑎᖁᓕᐅᓂᖏᓐᓄᑦ ᐃᓂᖅᓴᐅᑎᖅᑐᖅᑐᖅᑐᑦ ᑲᓇᖅᑐᖅᓯᐅᓂᒃᑯᑦ ᐃᓕᑕᑲᐅᓗᒍᑎᖅᑐᖅᑐᖅᑐᑦ ᐱᔪᓐᓇᖅᑎᑐᖅᑐᖅᑐᑦ ᐊᑭᕋᖅᑐᑐᑎᖅᒪᑦ ᐃᓂᖅᑎᒪᑦ ᐊᑭᕋᖅᑐᑐᑎᖅᒪᑦ ᐱᔪᓐᓇᖅᑎᑐᖅᑐᖅᑐ

ᑲᑐᔨᖅᑎᑦᓴᓂᑦ, ᓴᐳᓂᓕᖅᑕᐅᓯᒪᔪᑯᔨᓂᒪᔭᒪᑦ ᐱᓪᒪᓚᐅᓂᖅᑲᔪᓄᑦ ᑕᒪᓂᓗ ᐱᔪᑎᑕᐅᓗᓂ ᓯᓚᐅᑉ ᐊᓯᔨᖅᑐᑦᓚᐊᓂᖓ.

ᐊᒥᓯᑎᓗᓂᑉᓂᒃ ᓯᑯᑐᖅᐅᓂᖏᑦ ᑕᒪᓂ ᐱᓪᒪᓚᐊᓗᓂᒥᒃ ᓚᐖᑎᓚᐅᐅᖓᒪᑦ ᓯᑯᓕᖏᑦᑕᓚᐊᖅᑲᖅᑐᓄᑦ ᐅᒪᔪᓄᑦ, ᓂᕈᑭᓴᓴᖅᑲᖅᓯᐅᓂᖓᓄᑦ, ᓴᓐᓂᒃᑯᖅᓯᐅᖅᓯᐅᓂᖓᓄᑦ, ᓂᓚᒃᑯᖅᓯᐅᖅᓯᐅᓂᖓᓄᑦ ᐊᓪᒪᓗ ᐊᒥᓯᑐᖅᑦᓕᓚᐊᖅᓯᐅᓂᖓᓄᑦ ᐊᓯᖏᓐᓂᑉᓗ ᐱᐅᓯᖅᑐᖅᓯᐅᓂᖓᓄᑦ. ᑕᒪᓐᓇ ᐱᑕᖅᑲᖅᓯᐅᓂᖓᑦᑐᖅᓴᖅ ᓯᑯᒪᓚᐊᑎᒍᖅ ᓯᑯᑦ ᐊᑦᒍᒋᑦ ᓚᐖᑎᖅᓯᑦ ᑐᖓᔪᖅᑲᖅᓯᐅᓂᖓᓄᑦ. ᑕᒪᓐᓇ ᐱᑕᖅᑲᖅᓯᐅᓂᖓᓂᑕᐅᖅ ᐊᓯᖏᓐᓂᓐᓂᒃ ᐊᓯᖏᓐᓂᓐᓂᒃ ᐊᓯᖏᓐᓂᓐᓂᒃ

ᐃᖃᓗᓕᕆᔨᖅᑕᐅᓂᒃ ᖃᐅᔨᓴᖃᑦᑕᕐᓂᖅᓯᒪᑦ ᐱᓕᕆᓚᐅᖅᓴᑦ ᑕᐃᔨᐅᔪᓂᒃ ᐊᒥᓱᓂᒃ ᐊᔨᔪᖏᓪᓗᑎᐅᒍᓱᓂᒃ ᐃᓕᓴᓂᐊᑭᑕᐅᓚᓚᐅᖅᓱᓂᒃ ᐅᑭᐅᖃᑕᖅᑐᑦ ᐱᓕᕆᓚᐅᖅᓱᓂᒃ, ᐃᓕᓴᖏᓪᓕ ᐱᔨᔪᑎᖃᖅᓯᓗᑎᒃ ᑐᓚᐅᔪᐃᓕᓴᐅᑎᓪᓗᒥ. ᑕᒪᓐᓇ ᖃᐅᔨᓴᖃᑕᐅᔪᒪᓂᖅᓴᒃ ᑐᖏᓐᓄᓚᐅᖅᓴᒃ ᖃᐅᔨᔭᐅᔪᒪᓂᖅᓯᖏᓐᓂᓂᒃ ᐊᒥᓱᐊᓗᓚᓄᓂᒃ ᐊᖏᕐᒪᓄᓂᒃ ᓱᓇᑐᐃᓗᓂᖃᖅᓯᖏᓐᓂᓂᒃ ᑲᓇᑕᒥ ᐱᓕᕆᐊᖓᓂᖅᓯᒪᓂᒃ ᐅᑭᐅᖃᑕᖅᑐᑦ. ᑕᒪᓐᓇ ᖃᐅᔨᓴᖃᑕᐅᔪᒪᓂᖅᓴᒃ ᑐᖏᓐᓄᓚᐅᖅᓴᒃ ᐊᒥᓱᐊᓗᓚᓄᓂᒃ ᐊᖏᕐᒪᓄᓂᒃ ᐊᓯᐅᓚᐅᖅᓯᒪᓂᒃ ᐱᓕᕆᓚᐅᖅᓱᓂᒃ ᐊᒥᓱᐊᓗᓚᓄᓂᒃ ᐊᓯᐅᓚᐅᖅᓯᒪᓂᒃ ᐊᒥᓱᐊᓗᓚᓄᓂᒃ ᐊᖏᕐᒪᓄ

ጋዖሥσ⊲ፖርᐅσ‷ᖁᓂᓄᑦ ᐅσᓄbᓪ⊲ᓄᑦ. ርᒪᖁ⊲ ᐅσᓄbᓪ⊲ᑦ ᐅᖃᐅሥbᓄ⊲ᕀσ⊲ᖃᕐᑐ ᒧᓐᓇᐅᕐᖅ ᐱርᖃᓪᖃᕐᑐσᓄ ⊲ᒪᒪᓄ ᓴᖅᑐᑐᓐᓇᕀσ⊲ᕀᒥᔪσᓄ ᐃᓄᖁᐊᑎᓄᖁᓐᓂᓄᑦ ᐃᓄᔨᐊᑎᓄᓴᐅᐅᓐᓂᕀᒧᑦ ᐱᓪᓚᕀᖓᐊᖅᑐᑐᓄᓄᑦ ርᐅᐅσ ⊲ᒥᔨᐊᖅᑐᑐᓄ. ᓄᓇᒥᐅርᑐᖃᓪᓚᔨᐅᐊᓄᑦ ᐅσᓄbᓪ⊲ᓪ⊲ᑦ ᐊᑐᖃᑕᐅᓪᒥᒥᑦ ርᐃᔭᐃᔪσᓄ “ᖁᔨᕎᑐᖃᑐ ርᓚᐅᖅᒥ ᐃᑎᔪσᖅᑎᓄᓄ (ᑐᐊᐃᔪᐃᑐᖃᖅ) ᐅᖃᓗ⊲ᔪᑦᓴᖅᖁᖃᖅᖁᐅᐊᓄᓇᖅᑐᓄᓄ”. ርᐃᓄᔪ⊲ ᐃᓄᖁᐊᑎᓄᖁᓐᓂᓄᑦ ᐃᓄᔨᐊᑎᓄᓴᐅᐅᓐᓂᖃᑐ ᐱᓪᓚᕀᖓᐊᖅᑐᑐᓄᓄᑦᓂᓪᑐ ᐅσᓄbᓪ⊲ᕀᒥ ⊲ᑐᖃᑕᐅᓪᓚᐅᐊᖅᑐᒥᑦ ⊲ᕀᖃᑕᐅᓪᓚᐅᐊᖅᑐᒥᑦ-ᐃᑲᔨᐊᑎᓄᓴᖁᓐᓇᖅᑐᓄᓪᓗᑦ ጋዖርᐅᖁርᐅσᓪᓕᒥᑦ ᐱᔨᐊᑎᖁᖃᑐσᓄ ᐱᒋᐊᖅᑎᖁᓄᓄ ᒪσᓪᑕᒧᑦ ᑎᓪᓚᒧᑦ ᑎᓪᓕᓪᑐᑐᓄ ᒪᖁᑎᖁσᓄ.

ᐃᓚᓐᐅᐊᖅ ⊲ᒥᔨᓄᑦᓪᓗ ርᓚᐅᖅᒥ ᐱᓪᓚᕆᐅᐊᓂᓐᓄᑦ ርᐅᐊσ, ᐃᓗ⊲ᖃᑐᒥᑲ ᐅᓇᔨᓐᓂᓚᓄᓪᓗ. ᑎᒪᓐᓇ ᐃᒪᖃᑲ ᐅᓇᔨᓐᒪᑦ ᓄᓇᓪᓄᑦ ⊲ᒪᒪᓗ ርᒪᐃᓐσᑎᒃᓴᓗ ᔨძᔨᒪᐃᓐᓇᔨᓐᒧᓗσ ⊲ᕀᖁᔨᓪᐃᖅ. ርᒪᑲᖃᖅᑐσᓄ ᔨძᐃᓐᓇᐅᐊᓂᓐᓄᖅ ርᐃᓂ ᐱᓗ⊲ᖃᖅᑐᖃᖅ ᐃᓐᑎᕀᖁᐅᔪᓐᓇᓐᓇᖅᑐᖃᑎᓕᕀσᓄᓄᓚᓄ. ᐅᑭᐅᖃᑎᖅᑐᓪᒥᒥ ᔨᓪ ⊲ᔨᔪᓐᖅᑐᓪᓚᕆᐅᐊᓂᓐᓄᑦᓂᓪᑐ ᐱᓪᓚᕆᐅᐊᓂᓐᓄᑦᓂᓪᑐ. ርᒪᐅᖅᑐᓪᓚᕆᐅᐊᓂᓐᓄᑦᓂᓪᑐ ⊲ᒪᒪᓗ ᐱᓗ⊲ᖃᑐᒥᑲ ᐃᓐᑎᕀᖁᐅᔪᓐᓇᓐᓇᖅᑐᖃᑎᓕᕀσᓄᓄᓚᓄ ᐱᓪᓚᕆᐅᐊᓂᓐᓄᑦᓂᓪᑐ

ᒪᓪᒃᑕᐅᓪᓗᑎᒃ ᐊᖏᕈᑎᑐᕆᓄᑦ ᑐᖓᔪᒃᕆᐊᑦ ᐊᑎᓪᐅᕝᕕᑕᐅᓪᓕᕝᕕᕝᕕᖅᑐᑦ Δ> 2019-ᒥ, ᒥᖅᐳᐊᕝᕕᓂᖅᓯᓪᓚᕐᓇᓕᒃᑯᑦ ᐊᓘᓚᓗ ᐃᖅᑲᓗᓪᓚᕐᔾᔨᐊᒃᑯᑦ (ᑭᓯᑐᐃᓪᓗᑎᒃ ᑲᓇᑕᒥᑦ), ᓄᓇᕗᑦ ᑲᕐᒪᖏᖅᓄᓄᑦ, ᐊᓘᓚᓗ ᖃᐅᖅᑕᓂ ᐃᓄᐃᑦ ᑲᑐᔾᔨᑲᑎᑦᑎᓂᖅᓄᑦ ᐊᑎᓪᐅᕝᕕᕝᕕᖅᑐᑦ ᑐᑭᓯᐅᑎᓂᖅ, ᐅᓂᒃᑳᓕᕐᓂᖅᓄᑦ ᐃᓕᓴᐅᑎᑦᑎᓂᖅᓄᑦ ᐊᐅᓚᑦᑎᓂᖅᓄᑦ ᐃᓕᓴᐅᑎᑦᑎᓂᖅᓄᑦ ᐊᖏᕈᑎᑐᕆᓄᑦ ᑲᑎᒪᔨᖅᑎᑦᑎᓂᖅᓄᑦ ᐃᓕᓴᐅᑎᑦᑎᓂᖅᓄᑦ ᐊᑎᓪᐅᕝᕕᕝᕕᖅᑐᑦ ᑲᑎᒪᔨᖅᑎᑦᑎᓂᖅᓄᑦ ᐊᑎᓪᐅᕝᕕᕝᕕᖅᑐᑦ ᑕᑯᓂᖅ ᓴᖅᑭᑎᒃᓴᓂᖅ ᐊᑐᓕᖅᑐᑦ ᐃᓕᓴᐅᑎᑦᑎᓂᖅᓄᑦ ᐊᑎᓪᐅᕝᕕᕝᕕᖅᑐᑦ ᐊᑎᓪᐅᕝᕕᕝᕕᖅᑐᑦ ᑕᑯᓂᖅ ᓴᖅᑭᑎᒃᓴᓂᖅ

ᐊᓱᒪᑎᐊᖅᑎᓂᕐᓂᖅ ᒪᓕᑦᓂᔭᑐᐃᑦ ᐱᔨᐅᑎᖅᑐᑦ ᒥᑭᔪᐅᐅᓂᖅᑲᓂᖅ ᐊᓱᒪᓗ ᐊᖏᕐᔪᐅᐅᓂᖅᑲᓂᖅ ᐅᒥᐊᕆᔨᐊᓂᖅ ᐊᐅᑐᖅᑕᐅᑦᑕᐃᓕᒪᔨᕆᐊᖅᑎᖅᑎᓂᕐᓂᓄᑦ ᓄᓇᑦ, ᑕᓚᐅᖅᒥᐅᑕᐃᑦ ᐅᒪᔨᖏᑦᑎᓂ ᐊᓄᖏᑦᓗ ᐅᒪᓐᒥᒃᑭᑦ ᐱᑭᖅᑭᐅᑦᓗ ᑕᒪᖅᑯᐃ ᐱᓕᕆᔨᑕᕐᕕᖏᑎᒃ ᓄᑦᓗ ᑕᒪᖅᑯᓄᖕᓇ ᐅᖅᓄᐊᓗᓄᕿᖅᓄᐅᖅᑐᓄᑦᓗ ᓴᓴᓕᖅᓄᓄᖅᑐᓄᑦᓗ ᐊᓱᒪᓗ ᖃᐅᔨᓴᖅᑎᐅᓚᐅᖅᑐᓗᓄᑦᓗ.

ᐃᓕᑎᒋᓄᓄᒪᓕᒃᑯ ᑭᖅᑯᐊ ᐱᓂᐊᖏᓂᖅᑐᑦ ᐱᓴᓱᖅᑕᐅᖏᑯᓂᖅ ᐃᓄᐃᑦ ᐊᖅᑕᖅᑕᐅᓄᖅᑐᓄᓄᑦ ᐃᓄᐊᓄᐊᕈᑕᑭᐅᑎᖅᑐᑦ ᐊᑎᑕᐅᑎᖅ ᑕᐅᑐᒃ ᐃᒪᓐᓇᓂᖅ ᓴᓪᓂᓗᑕᐅᑎᓄᑦᓗ ᐃᒪᓐᓇᓂᖅᑐᑦ ᑕᓪᓕᐊᓂᖅ 2019-ᒥ. ᑕᒪᓐᓇ ᑎᑎᕋᖅᑐᖅ ᐃᓄᓕᒪᓄᑦ ᐊᑐᐃᓐᓇᐅᑕᓂᐅᔪᖅᑐᖅ ᐊᑎᓕᐅᖅᑕᐅᑕᐅᔪᑐᓂᖅ ᑕᒪᐃᓐᓇ ᕿᓐ 2019-ᒥ. ᑕᒪᓐᓇ ᑎᑎᕋᖅᑐᖅ ᐃᓄᓕᒪᓄᑦ ᐊᑐᐃᓐᓇᐅᑕᓂᐅᔪᖅᑐᖅ ᐊᑎᓕᐅᖅᑕᐅᑕᐅᔪᑐᓂᖅ ᑕᒪᐃᓐᓇ ᕿᓐ 2019-ᒥ. ᐊᓱᕋᓐᓂᕐᓂᓕᕆᐅᖅᑐᓂ ᑕᐃᕋᓐᓂᑦᑐᖅ ᐊᐃᑉᐸᓂᖅ ᑕᐃᕋᓐᓂᑦᑐᖅ ᐊᑎᓕᐅᖅᑕᐅᑕᐅᔪᑐᓂᖅ ᑕᒪᐃᓐᓇ

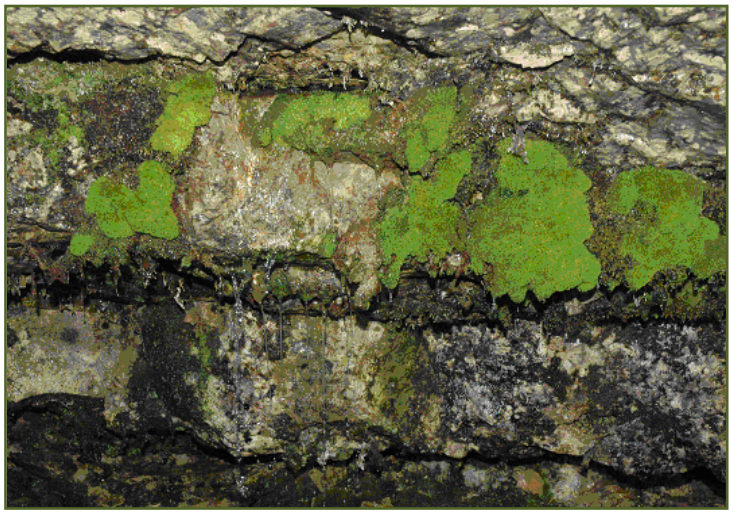
ᑲᓴᒋᒥ ᐃᓕᐅᕙᖅᓂᐊᖏᑦᑐᑦ ᒋᐃᑲᔪᓂᖅᓱ ᖁᑭᖏᒋᓂ ᐃᓄᐃᑦ ᑲᑐᔾᔨᖏᑎᑦᓯᓕᕐᓂᖅᓂᖅ ᐊᖅᑭᖏᑎᑦᔪᒋᐅᓚᓕᖅᓂᖅᓂ ᐃᓄᐃᑦ ᐊᑲᑐᖏᑎᒋᓕᖅᓂᖅᓂᓄᑦ ᐃᑲᔪᓕᐊᖏᒡᓗᑎᑦᓂᖅᓂᓄᑦ ᐊᖅᑭᖏᑎᒋᒪᖅᓯᓂ ᒋᒪᓕᖅᓯᐅᑎᑦᓂᖅᓂᓄᑦ ᒋᓪᓗᑐᐊᕐᓂ ᐃᒪᓐᓇᖅᑐᖅ. ᐃᓕᑭᐅᐅᔪᑦ ᐱᓕᕆᖏᑎᑦᓯᓕᕐᓂᖅᓂᓄᑦ ᐃᓄᐃᑦ ᐊᑲᑐᖏᑎᒋᓕᖅᓂᖅᓂᓄᑦ ᐃᑲᔪᓕᐊᖏᒡᓗᑎᑦᓂᖅᓂᖅᓄᑦ ᐊᖅᑭᖏᑎᓂᖅ ᒋᐅᖅᓯᓂ ᒋᓪᓗᑐᐊᕐᓂ ᐃᒪᓐᓇᖅᑐᖅ ᓴᓐᓂᖅᓂᓄᑦ ᒋᐃᓕᓂ ᔪᓐ 2019-ᒥ ᐊᖅᑭᖏᑎᓴᑲᓕᑎᓗᑎᒃ ᐃᓄᐃᑦ ᐊᑲᑐᖏᑎᒋᓕᖅᓂᖅᓂᓄᑦ ᐃᑲᔪᓕᐊᖏᒡᓗᑎᑦᓂᖅᓂᓄᑦ ᐊᖅᑭᖏᑎᓂᖅ ᓴᑲᓄᖅᑐᑐᑎᑦᓂᐅᕐᓂᖅᓂᓄᑦ ᒋᓐᓇᕐᒥ ᑲᑎᕐᓂᖅᓂᓄᑦ ᓴᓕᕐᓂᖅᓂᓄᑦ ᐅᑲᓂᐊᖅᑎᑦᓂᐅᕐᓂᖅᓂᓄᑦ ᑕᐅᕐᓂᖅᓂᓄᑦ ᐃᓕᓴᕐᓂᖅᓂᓄᑦ ᐅᑲᓂᐊᖅᑎᑦᓂᐅᕐᓂᖅᓂᓄᑦ ᑕᐅᕐᓂᖅᓂᓄᑦ ᐃᓄᐃᑦ ᐊᖅᑭᖏᑎᓂᖅ ᐃᑲᔪᓕᐊᖏᒡᓗᑎᑦᓂᖅᓂᖅᓂᓄᑦ ᐊᖅᑭᖏᑎᓂᖅ ᐃᓄᐃᑦ ᐊᖅᑭᖏᑎᒋᓕᖅᓂᖅᓂᓄᑦ ᐊᖅᑭᖏᑎᓂᖅ ᑕᒪᓐᓇᖅᑐ

Environment Canada Environnement Canada

ⲥⲟⲛⲃⲣⲁⲥ ⲟⲉⲫⲥⲣⲣ ⲃⲗⲓⲥⲟⲙⲁⲛⲁⲣⲁⲇⲣ ⲃⲘⲖⲁⲭⲣⲣ (NWMB)–ᵇᵈᵒₒᶜ ⲃⲥⲖⲭⲓ

ⲥⲫⲣⲣⲁⲫⲣⲘⲟⲫᶜ: **ⲁⲣⲓⲖⲣⲃⲫⲣⲘᶜ:** X

ⲁⲣⲓⲖⲟⲛⲛ: ⲥᵇⲣᵍᵇⲋⲥⲃⲣⲁᵍᵇ ⲣᵋⲓⲣⲣᵍᵇⲥᶦᵍᵇ ⲗⲣⲣⲘⲣⲁⲫⲣⲘⲖᶜ ᶜᵋⲫⲃⲣⲘ ⲃᵋⲁⲁⲫⲟⲟᶜ *(Mielichhoferia macrocarpa)* ⲃⲫⲥⲣⲣ, ⲃⲖⲁⲁᶜ ᵍᵈᵎⲁⲫ ᵍᵇⲋⲣⲣⲥⲋⲟᶜⲖⲣⲣⲏᵍᵇ (SARA) ⲥᵒᵒⲏⲙⲣⲣⲣⲋⲋ

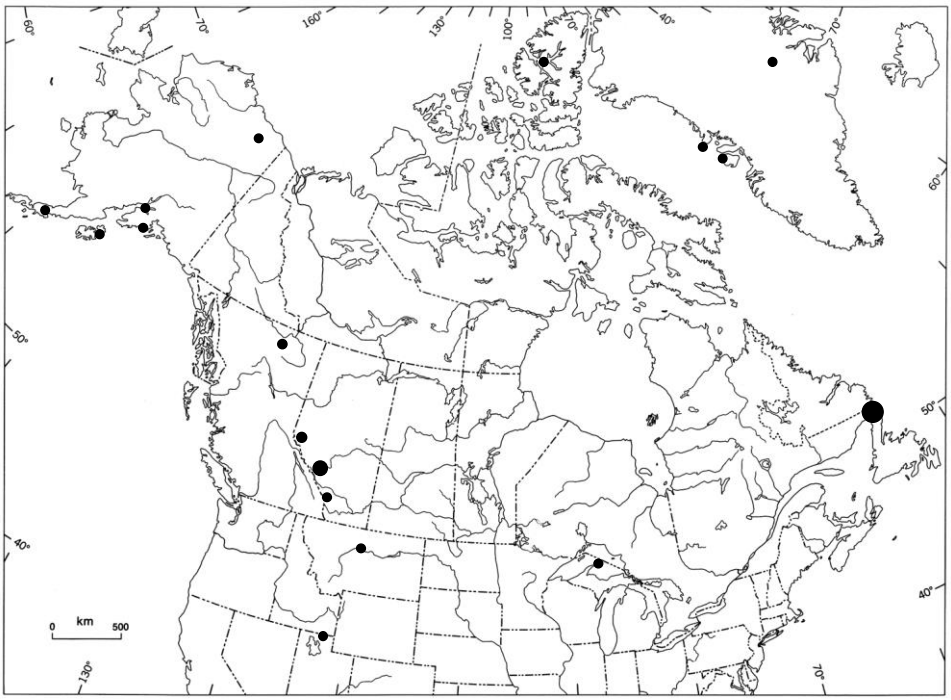


Ⲓᵎⲫⲃⲣⲁᵍᵇ ᵍⲃⲟⲁⲣⲣⲏⲟⲣⲣᶜ:

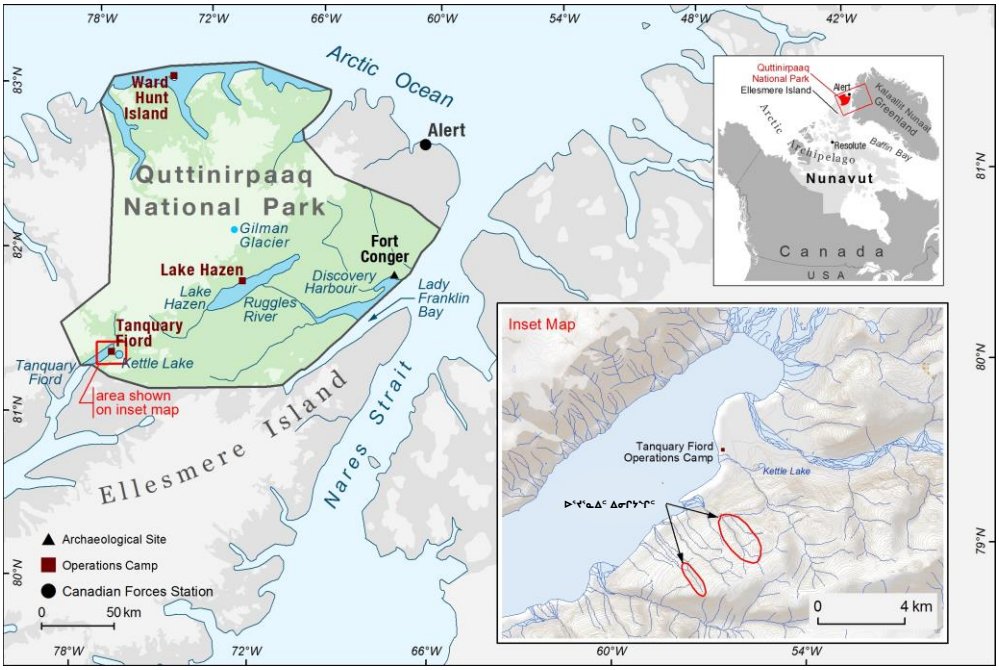
ⲃᵋⲁᵋⲫⲁⲁᶜ ⲘⲘⲘᵍᵇⲥⲣⲃⲣⲖⲁᵍᵇ ⲟᵋⲁⲣⲘⲃⲣᵍᵇⲘⲣᵎ ⲁⲣⲥⲣⲘⲫⲣⲋⲣᵎⲋⲟ ⲃⲣⲖⲁⲁᶜ ᵍᵈᵎⲁⲫ ᵍᵇⲋⲣⲣⲥⲋⲟᶜⲖⲣⲣⲏᵀ (ⲃⲁⲖⲋⲓ 1; 02/04/2011). ⲘⲘⲘᵍᵇⲥⲣⲃⲣⲖⲟⲣⲏ ⲖⲣᵇⲋⲖ, ⲁᵎⲫⲫⲟᵋⲖᶜ ᵋⲫⲖⲘ ⲁⲃⲁᵍᵇⲘⲣᵎⲋⲘᵇ ⲁᶦᵍᵖᵇᵎⲣⲃᵍᵇⲋᶜ, ⲁⲓⲖⲋ ⲁⲟⲣⲁᵎⲟⲟⲣᶜ ⲁⲓⲖⲋ ⲁⲟⲣᵎⲏⲣⲣⲟⲟᶜ ⲗⲫⲁᵇⲥⲣⲃⲣᵋⲃᵍᵇⲋᶜ ⲃᵋⲁᵋⲫⲁⲁᶜ ⲁᵎⲫᵋⲁᵈᵋⲋⲣᶜ ⲃⲫⲥⲣⲣ.

ⲟᵋⲁⲣⲘⲃⲣᵍᵇⲘⲣᵎ ⲚⲢᵖⲥⲋⲟ ⲁᵎⲁᵋⲟⲫⲥ ⲗⲋᵍᵇⲥⲣⲥᵍᵇⲋⲟᵇ ⲃⲣⲃⲁⲃⲣⲟⲋⲟᵇ ⲃⲣⲖᵋⲋⲟᵇ ⲃⲘᵒᵒⲓⲁᶜ ⲁⲓⲖⲋ ⲫⲁᵍᵇᵋⲏⲋⲟᶜ ᵍᵇᵍᵇⲃᵍᵇⲋⲟᵇ ⲁⲟⲣⲃⲁⲟᵇ ᵍᵇⲃⲣᵍᵇⲋⲟᵇ ᵍᵇⲇᵍᵇⲣⲣⲓⲁⲟᵇ ⲁᵋⲫⲋᵋⲁⲟᵇ ⲁⲖᵋⲦ ⲇⲫᵋⲫᵍᵇⲋᵍᵇⲋⲚᵇ ⲁⲓⲖⲋ ⲃⲢⲃᵇⲇᶜ ⲁⲖᵍᵇⲃⲢⲁᵍᵇⲥᵍᵇⲋⲚᵇ. ⲃᵋⲁᵋⲫⲁⲁᶜ ⲁⲟⲣⲁᵎⲫᵒⲟⲣᶜ 17–ᵒⲓⲁᶜ ᵍᵇⲃⲣⲖⲓⲃⲣⲁᶜ ⲃⲫⲥⲣⲣ ⲁⲓⲖⲋ ⲁⲟⲣᵎⲏⲣᶜ ⲫⲚⲣⲋⲁᵎᵍᵇⲋⲟᵇ, ⲁⲥⲣᵍᵇⲋⲟⲟ ⲁⲟⲣⲃⲁⲟᵇ ⲁⲃⲣⲋⲣⲥ, ⲣⲘᵎ ⲃⲥⲖⲘⲓⲣ, ⲟⲣⲃᵋᵋⲓⲟᵒ ⲁⲓⲖⲋ ⲓᵎⲣⲋⲁⲓ, ⲁⲓⲖⲋ ⲟⲫⲟⲣⲏ (ⲘⲘⲘᵍᵇⲣⲖⲁᵍᵇ 1). ⲁⲟⲣⲁᵎⲫⲟⲟⲣᶜ ⲤⲖⲁⲥ ⲁⲓⲟⲣⲟⲟⲣᶜ 960–ⲟᵇ ⲃⲘᵒᵒⲓⲁᵍᵇⲃⲣᵋⲋⲟᵇᵍᵇ ⲃⲫⲥⲣⲣ. ⲟⲫⲟⲣⲏⲣ, ᵍᵈᶜⲘᵒⲟᵍᵇᵋⲋⲣᶜ ⲣᶜᵎⲁⲁᵍᵇⲣᵍⲁᵒⲣⲋⲁⲓᵍᵇ ⲫⲟᵎⲃⲣᵍᵇᶜⲣᵍᵇⲋ, ⲗᵒⲓⲣᶜ ⲁⲟⲣⲁᵎⲫⲟⲟⲣᶜ ⲫⲟⲟⲁⲁⲁᵍᵇⲥⲣⲃⲣⲁⲟᵇ ⲃⲣⲣᵒⲓᵇ ⲟⲫⲁᵒⲓⲥ ⲁⲥᵒⲓⲥ ⲁⲟⲣⲟ (ⲘⲘⲘᵍᵇⲣⲖⲁᵍᵇ 2).

Canada



ᑎᑎᕃᓱᒪᔪᖅ 1. Ċᓵᓇ ᑎᑎᕃᓱᒪᔪᖅ ᑕᐻᕆᐅᑎᑎᓂᑎᔪᖅ ᐊᒥᐊᓕᑲᒥᒃ ᐅᖃᕐᓂᓵᑕᒃ. ᐱᖓᒍᓂ ᑎᑎᕃᓱᒪᔪᖅ ᐃᐅᐳᑐᒥ ᐃᓕᓴᕐᓂᖅ 6–ᓂ ᐃᓚᒋᐊᓵᓂᖏᓂ ᐃᓗᓕᖅᑐᑦᑐᓂ, ᐊᑕᐅᓱᖅ ᑎᑎᕃᓱᒪᔪᖅ >ᑎᑦ ᑲᓕᑦᐱᕆᒥ 1–ᒥ ᐃᓚᒋᐊᓵᓂᖏᓂ ᐃᓗᓕᖅᑐᑦᑐᓂ, ᓂᐅᕙᓵᓪ ᐊᑦᒪᓗ ᓪᓕᐊᑐᑦ 7–ᓂᑦ ᐃᓚᒋᐊᓵᓂᖏᓂ ᐃᓗᓕᖅᑐᑦᑐᓂ ᐊᑦᒪᓗ ᑎᑎᕃᓱᒪᔪᖅ ᓄᓇᕗᑦ ᖃᓂᒋᔭᑎᑦ ᐅᑭᐅᕗᑦᑕᖅᑐᑎᑦ ᐃᓕᖅᑐᑦᑐᓂ 3–ᓂᑦ ᐃᓚᒋᐊᓵᓂᖏᓂ.



ᑎᑎᕃᓱᒪᔪᖅ 2. Ċᓵᓇ ᑎᑎᕃᓱᒪᔪᖅ ᑕᐻᕆᐅᑎᑎᓂᑎᔪᖅ ᐅᖃᕐᓂᓵᑕᒃ ᐃᓂᖏᓐᓂᑦ ᖃᓂᒋᔭᓂᓂᖅᓴᖅ ᑎᓪᓗᐃᓂᓂᕆᔨᖅ ᑲᓇᑕᒥ.

ᐱᓕᓐᓂᐊᕐᓗᑎᒍᑦ ᓴᖅᓇᐅᑎ

ᐅᒪᓕᐊᑦ ᖃᓄᖅᓇ ᖃᐅᔨᓪᓗᓂᑦ ᒪᓕᕆᓂᖅ (*SARA*) ᐊᒋᓂ, ᑎᑎᕋᖅᓯᒪᔨᖅᖃᐅᑦ ᐊᑦᑕᕐᓇᖅᑐᐅᑦᑐᐅᒻᒥᑦ, ᐋᓐᓇᖅᓂᕐᒧᑦ ᐸᕐᓇᐅᑎᒻᒥᑦ ᐊᑦᑭᓯᓂᖅᑕᐅᔪᓚᐅᓕᖅᓂ. ᐋᓐᓇᖅᓂᕐᒧᑦ ᐸᕐᓇᐅᑎ ᒪᕐᒉᖅᓂᑦ ᑎᑎᕋᖅᑲᓂᑦ ᐸᕐᓇᐅᑎ: ᐋᓐᓇᖅᓂᕐᒧᑦ ᐸᕐᓇᒍᑎ ᐊᑦᒪᓗ ᐱᓕᕆᐊᑭᑐᒧᑦ ᐸᕐᓇᐅᑎ. ᑭᖑᓪᓕᖅᐸᖅ ᐋᓐᓇᖅᓂᕐᒧᑦ ᐸᕐᓇᒍᑎ ᐅᕐᔪᓇᐃᓂᑦ ᑲᓇᑕᒥᔪᖅᑭᑕᐅᔪᓂᖅ ᐄᓕᔪᐃᑦ ᖃᐅᔨᓪᓗᖅᑐᐅᓂᐅᑦ ᑲᒪᒋᔭᐅᓂᖅ ᑕᑕᑎᓐᖓᑦ ᑎᓪᓗᒍ 2017–ᒥ. ᐋᓐᓇᖅᓂᕐᒧᑦ ᐸᕐᓇᒍᑎ ᐅᖃᐅᔨᖃᕐᓗᐊᖅᓱᓂ ᐋᓐᓇᖅᓂᕐᒧᑦ ᑐᕆᒪᐅᔪᓂᑦ ᐊᒪᓗᓂ ᐱᓇᓱᑕᐅᔪᓂᑦ, ᐱᓕᕆᐊᑭᑐᒧᑦ ᐸᕐᓇᐅᑎ ᐅᖃᐅᔨᖃᕐᓗᐊᖅᓱᓂ ᐋᓐᓇᖅᓂᕐᒧᑦ ᐱᓕᕆᐊᑭᑐᒧᑦ ᓂᕐᔪᐊᑎᓂᖅᓱᓂ ᐅᓂᒃᑲᖅᓯᒪᔪᖅ ᐱᓕᕆᐊᑭᑐᒧᑦ ᐱᓕᕆᐊᑭᑐᒧᑦ ᐅᓂᒃᑲᖅᓯᒪᔪᖅ ᐊᑦᑕᕐᓇᖅᑐᐅᑦᑐᐅᒻᒥᑦ ᐱᓕᕆᐊᑭᑐᒧᑦ ᐅᓂᒃᑲᖅᓯᒪᔪᖅ ᐊᑦᑕᕐᓇᖅᑐᐅᑦᑐᐅᒻᒥᑦ ᐊᑦᑕᕐᓇᖅᑐᐅᑦᑐᐅᒻᒥᑦ ᐊᑐᓂ, ᐊᒪᓚᓗ ᐊᐅᓇᖅᓯᒪᔨᐸᑦ, ᓂᓪᔪᐊᑎᐅᖅᑎᑦ ᓴᖅᑭᑎᓴᓂᓪᓕᓚᑦ ᐃᓂᓕᑦᓯᒪᔪᓂᑦ ᐊᒪᓚᓗ ᖁᐅᓴᖅᑐᐅᑦᑐᐅᒻᒥᑦ ᐊᑦᑕᕐᓇᖅᑐᐅᑦᑐᐅᒻᒥᑦ ᐊᑦᑕᕐᓇᖅᑐᐅᑦᑐᐅᒻᒥᑦ ᐊᑐᓂ, ᐊᒪᓚᓗ ᐊᐅᓇᖅᓯᒪᔨᐸᑦ, ᓂᓪᔪᐊᑎᐅᖅᑎᑦ ᓴᖅᑭᑎᓴᓂᓪᓕᓚᑦ ᐃᓂᓕᑦᓯᒪᔪᓂᑦ ᐊᒪᓚᓗ ᖁᐅᓴ

ᒥᕐᓯᐃᓕᓴᕆᓚᓕᒃᑯᑦ ᑲᓇᑕᒥ ᐊᐃᕐᓭᓪ ᐊᓪᒪᓗ ᓂᐧᐤ 2016–ᒥ, ᒪᓪᓴᓗᐃ. ᑐᑭᓯᖏᑐᑕᐅᔪᖅ ᐱᓪᓭᕆᐊᕈᑎᒍᑦ ᕿᕐᓇᐅᑎ ᓴᖅᑭᑕᐅᓯᒪᔪᖅ ᐅᒪᔪᐃᐃᑦ ᖃᐅᔨᓴᖅᑐᑐᒍᓐᓇᐅᑦ ᑲᒪᕐᑎᐃᓐᓇᑦ ᑎᑭᐱᓚᓐ 22, 2017–ᒥᑦ, 60–ᓄᑦ ᐅᓪᓗᓄᑦ ᐃᓄᓪᒦᓄᑦ ᐅᖃᐅᓐᓇᕐᓂᒍᑦ ᐱᐃᓴᓚᒥᒃ, ᐃᓱᓪᓕᐅᖅᓯᓗᓐ ᓂᐲᐅᐊᓚᓐ 20, 2018–ᒥ. ᐅᖃᐅᓱᓴᐃᑦ ᐃᓱᒋᔭᐅᓕᐅᖅᑐᑦ ᐊᓪᒪᓗᓐ ᐅᓪᓗᒥ ᑲᑎᑭᓱᖏᓪᓕᖅ ᑭᖑᓪᓕᖅᐸᖅ ᐱᓕᕆᐊᕈᑎᒍᑦ ᕿᕐᓇᐅᑎ ᑎᑎᕋᖅᑕᐅᓪᓚᓂᓪ.

ᑭᖑᓪᓕᒃ ᐱᒃᓴᑦ:

ᓄᓇᕗᑦ ᐅᒪᔨᓕᓚᓴᕐᔪᐊᑦ ᑲᑎᒪᔨᖏᑦ (NWMB)–ᑲᑯᓂᓄᖑᐅᓂᐊᖅᑐᒍᑦ ᐊᐱᓚᒃᖅᑐᕐᓗᓕ ᐊᖏᕐᖅᑕᐅᕐᑯᓗᒍ ᑐᑭᓯᖏᑐᑕᐅᔪᖅ ᑭᖑᓪᓕᖅᐸᖅ ᐱᓕᕆᐊᕈᑎᒍᑦ ᕿᕐᓇᐅᑎ. ᐱᓕᕆᐊᕈᑎᒍᑦ ᕿᕐᓇᐅᑎ ᐋᖅᑭᑎᐊᖅᑕᐅᑲᑦ ᖃᐅᔨᓴᕐᑎᕆᐊᖃᓂᓪᓕᐊᖅᑐᖅ ᐊᓪᒪᓗᒍ ᓄᒋᑎᓪ 5 ᖃᖏᕐᕋᓚᒥᑲ ᐅᕝᕙᓗᓂᑦᓂᓕᒫᑦ ᑕᐃᒪᐃᓕᑕᕆᐊᖃᖅᐸᑦ.

ᐋᖅᑭᑲᓇᖅᑕᐅᔪᖅ ᐅᒧᓐᓄ:
ᑲᓇᑕᒥ ᓂᕐᔪᑎᑦ ᐱᓕᕆᔨᑎᒋᓂᓪ, ᓴᓗᓇᐃᕝ 14 ᐊᐃᕐᓭᓪ 2019

ᐅᒪᔪᐊᑦ ᖃᓄᖅᓱᓇᖅᑐᓕᕐᓂᖅᓯᓄᑦ ᒪᓕᒐᖅ (SARA) ᐊᖏᓪᓗ ᐃᓐᓇ

ᑐᓴᕐᖁᑎᐅᔪᖅ ᑭᖑᓪᓕᖅᕿᖅ ᐅᖃᕐᓇᐃᓄᑦ ᐱᓕᕆᐊᖃᑎᒍᑦ ᕿᖃᐅᑎ ᓇᐃᐸᓴᖅᓴᕐᓂᑦ ᖃᐅᔨᓴᖅ

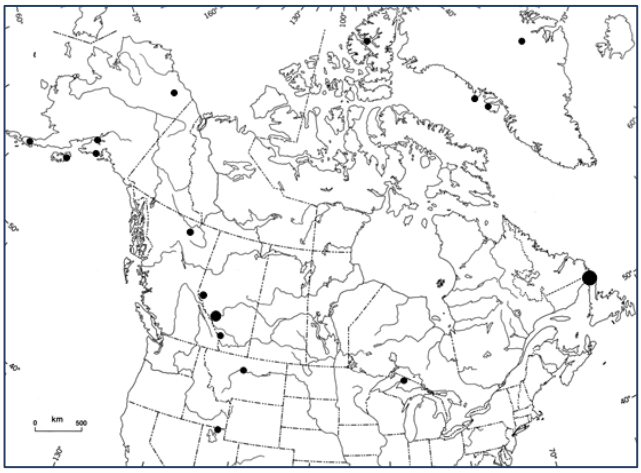
ᓇᐃᐸᓴᖅᓴᕐᓂᑦ

ᐅᓇ ᓇᐃᐸᓴᖅᓴᕐᔪᐊᑎᒥ ᐊᑐᐃᓐᓇᐅᑎᑕᐅᔪᖅ ᑐᓴᕐᖁᑎᐅᔪᖅ ᑭᖑᓪᓕᖅᕿᖅ ᐱᓕᕆᐊᖃᑎᒍᑦ ᕿᖃᐅᑎ ᐅᖃᕐᓇᐃᓄᑦ (*Mielichhoferia macrocarpa*) ᐃᓱᒪᖅ ᓂᕐᔪᑎᐅᔾᔨᖅᑎᑭᓱᓕᓂᒃ. ᓄᕆᐊᓗ 2003-ᒥ ᐅᖃᕐᓇᐃᓕᑦ ᖁᐅᓴᓴᖅᑕᐅᓕᐅᔾᔨᖅᑐᖅ ᑲᑎᒪᓴᑦ ᖁᓇᐃᓕᓐᓄᑦᓂᓐᓄᑦ ᓂᕐᔪᐊᑦ ᐊᑕᕐᓇᖅᑐᒃᓯᒪᓕᒃᓂᒃ ᑲᓇᕐᑕᑦ (COSEWIC)-ᒃᑯᓄᐅ ᐊᑕᕐᓇᖅᑐᒃᓯᑕᒥ. 2011-ᒥ, ᑲᑎᒪᓴᑦ ᖁᓇᐃᓕᓐᓄᑦᓂᓐᓄᑦ ᓂᕐᔪᐊᑦ ᐊᑕᕐᓇᖅᑐᒃᓯᒪᓕᒃᓂᒃ ᑲᓇᕐᑕᑦ (COSEWIC) ᖁᐅᓴᓴᓂᒃᓂᓕᓄᐅ ᖁᓇᐃᓕᑕᐅᑎᑕᐅᑎᒃ ᒪᑕᒃᓯᑎᒃ, ᐅᖃᕐᓇᐃᓕᑦ ᐃᓕᓴᐅᑕᐅᔾᔨᖅᑐᓕ ᐅᐃᒍᖓ 1-ᒧᑦ ᐅᓪᓗᕆᐊᑦ ᓂᕐᔪᐊᑦᓯᑕᒃᓯᓕᓂᑦ ᒪᓕᒐᖅ (SARA)-ᓄᑦ, ᐊᖏᓪᓗ ᐊᓂᓐᓇᓂᓕᒫᑦ ᕿᖃᐅᔨᑎᑎᑦ ᐊᑲᑎᐸᓴᕐᔪᐊᖅᑐᓂᒃ ᐊᖏᓪᓗ ᓴᖅᑭᑕᐅᑦᓯᓄᑦ ᑎᓴᒪᓪᓗ 2017-ᒥ. ᑐᓴᕐᖁᑎᐅᔪᖅ ᑭᖑᓪᓕᖅᕿᖅ ᐱᓕᕆᐊᖃᑎᒍᑦ ᕿᖃᐅᑎ ᐊᑲᓴᖅᑕᓪᓕᖅᑐᖅ ᓂᕐᔪᐊᑦ ᐊᑕᕐᓇᐃᓕᑦ ᐊᖏᓪᓗ ᐱᓕᕆᐊᖃᑎᒍᑦ ᕿᖃᐅᑎ ᐃᓐᓇᖅᑐᑦᑎᐊᓕᒫᑦ ᓂᕐᔪᐊᑦ ᐃᓚᖓᑕᒥ, ᐅᓂᒃ ᑲᓪᓚᐱᓚᒋᑦ, ᓂᐅᕐᓯᓐᓂᖅ ᐊᖏᓪᓗ ᐊᓐᓄᓕᒫᑦ. ᑕᒪᐃᑎᒍᑐᐊᓕᓇᖅᑎᑦᓯᐊ, ᒡᓘᓐᓇ ᓇᐃᐸᓴᖅᓴᕐᔪᐊᖅᑐᓂᒃ ᐱᓕᕆᐊᖃᑎᒍᑦ ᕿᖃᐅᑎ ᐅᓂᒃᑲᖅᑕᐅᔾᔨᖅᑐᓂᒃ ᐃᓐᓇᖅᑐᑦᑎᐊᓕᒫᑦ ᐊᖏᓪᓗ ᐃᓐᓇᖅᑐᑦᑎᐊᓕᒫᑦ ᐊᑐᐃᓐᓇᐅᑎ

- ᑐᓴᒪᓚᕐᔪᓐᓇᖅᑕᖏᓐᓂ ᐅᓇᖏᓐᓂᑎᓂᓐᓇᓄ ᐊᓐᓇᐅᒪᓂᕐᒧ ᐊᐅᑐᖅᑕᐅᓂᕐᓂᖅᑐᖅᑐᑦ ᐊᑦᑕᕐᓇᖅᑐᒥᑦ, ᓄᑦᓗ ᒪᑦᔨᓇᓴᓂᕐᒥᑦ, ᐅᓇᑦᓂᓚᖏᓐᓂ ᐊᖏᐅᓐᓂᒪᓗᐊᕐᓂᖏᓐᓂᓐᓄᓇ, ᐱᓐᓱᐊᕐᓂᕐᒧᑦ ᖃᓄᐃᓇᓕᐅᕐᓂᖅᑐᓂᖏᓐᓂ ᐊᓚᒪᓗ ᐃᓇᐅᔨᑦᑕᖅᑕᐅᔪᓐᓇᖏᓐᓂᑦᑐᑦ ᖃᓄᐃᓇᓕᐅᕐᓂᖅᓂᒃ ᐃᓗᓇᓕᓂᕐᓂᓐᓄᓇ.

ᐃᓕᓐ ᐅᖃᕐᓇᐃᓕᑦ ᐃᓂᖏᐅᑦᓯ?

- ᐃᓂᖏᐅᕐᓂᖅᑐᑦᑕᖅᑐᑦ ᓂᕆᔨᖅ-ᐅᐊᓕᓂᐅᓂᐅᔪᑦ ᓴᖑᓚᓂᓕᒫᖅ.
- 17–ᖃᓚᑦ ᖃᓄᐅᓕᒫᖅᓯᐅᑦ ᐃᓇᓂᐊᖑᓂᖏᓐᓂᑦ ᑲᓇᑕᒥ:
  - 1–ᖃᓚᕐᓇᖅ >ᑎᑭ ᑲᑎᒪᓕᕆᒥ
  - 6–ᖃᓚᑦ ᐃᐅᔾᔨᑕᒥ
  - 7–ᖃᓚᑦ ᓂᐅᕐᐳᓪ ᐊᓚᒪᓗ ᓪᓚᑐᐊᒥ
  - 3–ᖃᓚᑦ ᓇᓇᕗᓕᒥ



ᑎᑎᕋᕐᓚᕐᔪᖅ 1. ᐅᖃᕐᓇᐃᓕᑦ ᐃᓂᓕᓴᖏᓐᓂ ᐊᓚᒪᓗ ᓴᖅᑭᖅᑕᕐᓂᖅᓂᖏᓐᓂ ᓂᕆᔨᖅ-ᐅᐊᓕᓂᐅᓂᐅᔪᑦ ᓴᖑᓚᓂᓕᒫᖅ. ᐱᓐᓱᓕᑦ ᑎᑎᕋᕐᓚᕐᔪᖅ ᐃᐅᔾᔨᑕᒥ 6–ᓂᐅ ᐃᓇᓂᐊᖑᓂᖏᓐᓂᑦ ᐃᓗᓇᓕᓂᕐᓂᖅᑐᖅᑐᑦ, ᐊᑦᑕᐅᓂᕐᓂᖅᑐᖅᑐᑦ ᐃᐅᔾᔨᑕᒥ ᑎᑎᕋᕐᓚᕐᔪᖅ >ᑎᑭ ᑲᑎᒪᓕᕆᒥ 1–ᒥ ᐃᓇᓂᐊᖑᓂᖏᓐᓂᑦ ᐃᓗᓇᓕᓂᕐᓂᖅᑐᖅᑐᑦ, ᓂᐅᕐᐳᓪ ᐊᓚᒪᓗ ᓪᓚᑐᐊᒥ ᑎᑎᕋᕐᓚᕐᔪᖅ 7–ᓂᐅ ᐃᓇᓂᐊᖑᓂᖏᓐᓂᑦ ᐃᓗᓇᓕᓂᕐᓂᖅᑐᖅᑐᑦ, ᐊᓚᒪᓗ ᑎᑎᕋᕐᓚᕐᔪᖅ ᓇᓇᕗᓕᒥ ᖃᓄᕐᓂᐅᑐᑦᑕᖅᑐᒥ ᐅᓄᐅᖅᑕᕐᓂᖅᑐᖅᑐᑦ ᐃᓇᕐᓂᖅᑐᖅᑐᑦ 3–ᓂᐅ ᐃᓇᓂᐊᖑᓂᖏᓐᓂᑦ.

ᐊᐅᓚᓂᓕᒧᑦ ᐱᓕᕆᔨᐊᕈᑎᑦ

ᑲᓇᑕᓪᓚᒥᑦ ᐊᐅᓚᓂᓕᒧᑦ ᐸᑐᓚᔪᓂᑦᑕ ᖁᑦᓚᔪᑦ, ᐊᐃᔭᑐᖅᓯᒪᔪᓕᐊᑦ ᐊᐅᓚᓂᓕᒧᑦ ᑎᑎᕋᖅᑲᐃᑦ ᐅᖃᕐᓇᐃᓕᐃᓄᑦ ᐋᖅᑭᒃᖢᖅᑐᖅᑐᑦ ᐃᐅᔾᔨᑕᒥ ᐊᐃᓚᖅ ᓂᐅᕐᐳᓪ ᐊᓚᒪᓗ ᐃᓇᐅᔨᑦᑕᖅᑕᐅᔪᓐᓇᖅᑐᖅᑐᑦ. Ċᖅᑯᐊ ᑎᑎᕋᖅᑲᐃᑦ ᐊᐃᒐᓕᓂᕐᓂᖅᑐᖅᑐᑦ ᐊᐃᐊᑐᖅᑕᐅᔪᓐᓇᖅᑐᖅᑐᑦ-ᑐᑭᖓᓚᔪᓄᑦ ᐃᓂᓕᓴᖏᓐᓄᓄᑦ ᐊᓚᒪᓗ ᐊᔪᖅᑕᕐᓂᑎᓂ ᐋᖅᑭᒃᖢᖅᑐᖅᑐᑦᓂ, ᐋᑦᑕᖅᑕᐅᔪᑎᓂᕐᒥᑦ, ᐊᓚᒪᓗ ᐊᐅᓚᓂᓕᒧᑦ ᐱᓕᕆᔨᐊᑦᑎᓐᓄᑦ. ᑲᓇᑕᓪᓚᒥᑦ ᐊᐅᓚᓂᓕᒧᑦ ᐸᑐᓚᔪᓂᑦ ᐊᐃᒐᓕᓂᕐᓂᖅᑐᖅᑐᑦ ᐃᓇᓂᐊᖑᓂᖏᓐᓂᓐᓄᓇ ᐊᓚᒪᓗ ᐃᓂᓕᓴᖏᓐᓂᑦ ᐱᓇᓱᐊᕐᓂᖏᓐᓄᓇ ᐅᖃᕐᓇᐃᓕᑦ ᑲᓇᑕᓪᓚᒥᑦ ᐅᖃᕐᓇᐃᓕᐃᓄᑦ ᐅᖃᖅᓯᐅᑦ ᐃᓕᓐ ᐅᖃᕐᓇᐃᓕ

ᖃᓄᐃᓕᖅᖢᖅ ᓯᑭᐸᓚᑎᑎᑦᑕᐅᓴᖅᓄᑦ:

- ᐊᐃᒃᐅᑐᖅᓯᒪᓴᕐᓂᖅ ᐊᒻᒪᓗ ᓇᓇᖅᑲᖅᑎᑎᐅᓴᕐᓂᖅ ᐱᓕᕐᓂᐊᖅᑐᑐᒃᓴᓂᐅᖅ ᐱᓴᖅᑭᓯᓚᐅᖅᑎᑦᑎᓂᖅ ᐱᓕᕐᓂᐊᖅᑐᑐᒃᓴᓂᐅᖅ ᐱᐅᓯᒪᓴᕐᓂᖅ ᐱᓕᕐᓂᐊᖅᑐᑐᒃᓴᓂᐅ

ᐊᒻᒪᓗ ᐃᓄᑐᐃᓇᖅᑐᓂᖅ ᓇᔨᖏᑕᐅᔪᑦ ᐱᓕᕆᐅᔪᑦ ᓴᐳᑎᓇᓕᒫᑦ ᑲᒪᒋᔭᐅᓂᖅ ᑲᓇᑕᒥᒃ ᓄᓇᕗᑦᑐᐊᕐᓂᑦᑐᑦ, ᐊᐅᓚᕆᑎᓄᑦᓯᓗ ᐱᖓᓱᖏᑎᕆᑎᓂᖅ ᐱᐅᖏᓪᓗᐊᖅᓯᓂᖅ ᖃᓄᐃᓕᐅᕈᑕᐃᓂᖅ.

ᐃᓄᐅᓂᖅᓗᒍ-ᑭᓇᐅᔨᓕᕆᓂᖅᓗᒍ ᖃᐅᔨᓴᖏᖅᑕᐅᓂᖅᓯᓂᖅ

ᐃᓄᐅᓂᖅᓗᒍ-ᑭᓇᐅᔨᓕᕆᓂᖅᓗᒍ ᖃᐅᔨᓴᖏᖅᑕᐅᓂᖅᓯᓂᖅ ᐅᖃᐅᔨᓴᐅᔪᖅ ᐱᓕᕆᔨᐊᖑᑎᒍ ᐸᖅᑖᓐᓇᖅᑐᖅ ᐅᖃᐅᔨᓴᓕᖅ ᐅᑎᖅᑕᓐᓄᑦ ᐃᓄᐅᓂᖅᓗᒍ-ᑭᓇᐅᔨᓕᕆᓂᖅᓗᒍ ᐊᐅᖓᖅᓄᓇ ᐱᓕᕆᖏᖅᑎᑎᓂᖅᑎᒋᑦ ᐱᓕᕆᔨᐊᖑᑎᒍ ᐸᖅᑖᓐᓇᖅᑐᖅ ᐅᖃᐅᔨᓴᓕᖅ ᑕᔪᐃᐅᓪᓚᒥᒃ, ᐃᓄᐅᓂᖅᓗᒍ ᐊᒻᒪᓗ ᐊᐖᑎᓂᓕᕆᓂᖅᓗᒍ ᐃᒃᔦᑎᓂᖅ ᐃᒃᔦᑎᑐᓴᖅᑐᑦ ᐱᓕᕆᔨᐊᖑᑎᒍ ᐸᖅᑖᓐᓇᖅᑐᖅ ᑕᒪᓐᓇᐅᔪᖅ ᐱᖓᓱᖏᑎᕆᑎᓂᖅ ᐸᖅᑖᓐᓇᖅᑐᖅ ᐃᓕᓴᕐᓂᐊᕐᓂᖅ.

ᑎᑎᖅᑲᖅ ᐅᓂᒃᑲᖅᑐᖅ ᐅᓄᖅᑐᓂᒃ ᒪᓕᑕᐅᔭᕆᐊᒃᑯᓕᒍᑦ, ᐱᔨᑦᓕᕆᓂᖅᓯᓂᖅᑎᑦ ᐊᐅᓚᕆᑎᓄᑦᓯᓗ ᐃᓕᓴᕐᓂᐊᕐᓂᖅ ᐱᓕᕆᔨᐊᖑᑎᒍ ᐸᖅᑖᓐᓇᖅᑐᖅ

ᐃᓂᒋᔭᐅᓇᓄᑦ ᐊᑐᖅᑕᐅᓂᖓᑦ

ᐅᖃᕐᖢᐃᑦ ᐱᑕᖃᖅᑐᖅ ᐊᕕᑐᖅᓯᒪᔨᓂᒃ, ᑲᓇᑕᒥᒋᑦ ᐊᒻᒪᓗ ᐃᓄᖕᓄᑦ ᓇᖕᒥᓂᕆᔭᐅᔨᓂᒃ, ᐊᒻᒪᓗ ᐃᓂᐅᔨᑦ ᑕᒪᐃᖏᓂᕐᒪᑦ ᓴᓇᑎᐅᖃᖅᑐᑦ ᐃᓂᒋᔭᖏᑦᑐᑦ ᐊᑐᖅᑕᐅᓂᖓᑦ ᐊᖏᖅᑎᑦ ᐊᑐᐃᒃᑐᖅᑎᒋᑦ ᐊᓐᓇᐅᒪᔨᖅᑎᑦᑎᓐᓂᒃ. ᐊᖏᖅᑎᑦ ᐊᑐᐃᒃᑐᖅᑎᑦᑎᓐᓂᑦ ᐱᖓᓱᖅᑎᑦᑎᓐᓂᑦ ᐱᓪᓚᑕᐅᔨᓂᒃ ᐊᓐᓇᐅᒪᓂᖅᓯᒪᓂᖅ ᐱᓪᓚᑕᐅᔨᓂᒃ ᐊᓐᓇᐅᒪᓂᖅᓯᒪᓂᖅ ᖃᐅᔨᓴᖅᑐᐅᓂᖅ ᐃᓄᖕᓄᑦ ᐃᓄᐊᕐᓇᓐᓂᐊᓂ ᐃᓄᖕᓄᑦ ᐊᓐᓇᐅᒪᓂᖅᓯᒪᓂᖅ ᐱᓪᓚᑕᐅᔨᓂᒃ ᐊᓐᓇᐅᒪᓂᖅᓯᒪᓂᖅ ᐊᑐᖅᑕᐅᓂᖓᑦ ᖃᐅᔨᓴᖅᑐᐅᓂᖅ ᐱᓪᓚᑕᐅᔨᓂᒃ

Species at Risk Act
Action Plan Series

Action Plan for the Porsild's Bryum (*Mielichhoferia macrocarpa*) in Canada

Porsild's Bryum

2019

Government of Canada
Gouvernement du Canada

Canada

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For copies of the action plan, or for additional information on species at risk, including the Committee on the Status of Endangered Wildlife in Canada (COSEWIC) Status Reports, residence descriptions, recovery strategies, and other related recovery documents, please visit the Species at Risk (SAR) Public Registry[1].

Cover illustration: © René J. Belland 2005

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[1] www.canada.ca/en/environment-climate-change/services/species-risk-public-registry.html

Preface

The federal, provincial, and territorial government signatories under the Accord for the Protection of Species at Risk (1996)[2] agreed to establish complementary legislation and programs that provide for effective protection of species at risk throughout Canada. Under the *Species at Risk Act* (S.C. 2002, c.29) (SARA), the federal competent ministers are responsible for the preparation of action plans for species listed as Extirpated, Endangered, and Threatened for which recovery has been deemed feasible. They are also required to report on progress within five years after the publication of the final document on the SAR Public Registry.

Under SARA, one or more action plan(s) provides the detailed recovery planning that supports the strategic direction set out in the recovery strategy for the species. The plan outlines what needs to be done to achieve the population and distribution objectives (previously referred to as recovery goals and objectives) identified in the recovery strategy, including the measures to be taken to address the threats and monitor the recovery of the species, as well as the proposed measures to protect critical habitat that has been identified for the species. The action plan also includes an evaluation of the socio-economic costs of the action plan and the benefits to be derived from its implementation. The action plan is considered one in a series of documents that are linked and should be taken into consideration together. Those being the COSEWIC status report, the recovery strategy, and one or more action plans.

The Minister of Environment and Climate Change and Minister responsible for theParks Canada Agency is the competent minister under SARA for the Porsild's Bryum and has prepared this action plan to implement the recovery strategy, as per section 47 of SARA. To the extent possible, it has been prepared in cooperation with the provinces of British Columbia, Alberta, and Newfoundland and Labrador and Nunavut Territory, as per section 48(1) of SARA.

Success in the recovery of this species depends on the commitment and cooperation of many different constituencies that will be involved in implementing the directions and actions set out in this action plan and will not be achieved by Environment and Climate Change Canada and the Parks Canada Agency, or any other jurisdiction alone. All Canadians are invited to join in supporting and implementing this action plan for the benefit of the Porsild's Bryum and Canadian society as a whole.

Implementation of this action plan is subject to appropriations, priorities, and budgetary constraints of the participating jurisdictions and organizations.

The recovery strategy sets the strategic direction to arrest or reverse the decline of the species, including identification of critical habitat to the extent possible. It provides all

[2] www.canada.ca/en/environment-climate-change/services/species-risk-act-accord-funding.html#2

Canadians with information to help take action on species conservation. When critical habitat is identified, either in a recovery strategy or an action plan, SARA requires that critical habitat then be protected.

In the case of critical habitat identified for terrestrial species including migratory birds SARA requires that critical habitat identified in a federally protected area[3] be described in the *Canada Gazette* within 90 days after the recovery strategy or action plan that identified the critical habitat is included in the public registry. A prohibition against destruction of critical habitat under ss. 58(1) will apply 90 days after the description of the critical habitat is published in the *Canada Gazette*.

For critical habitat located on other federal lands, the competent minister must either make a statement on existing legal protection or make an order so that the prohibition against destruction of critical habitat applies.

If the critical habitat for a migratory bird is not within a federal protected area and is not on federal land, within the exclusive economic zone or on the continental shelf of Canada, the prohibition against destruction can only apply to those portions of the critical habitat that are habitat to which the *Migratory Birds Convention Act, 1994* applies as per SARA ss. 58(5.1) and ss. 58(5.2).

For any part of critical habitat located on non-federal lands, if the competent minister forms the opinion that any portion of critical habitat is not protected by provisions in or measures under SARA or other Acts of Parliament, or the laws of the province or territory, SARA requires that the Minister recommend that the Governor in Council make an order to prohibit destruction of critical habitat. The discretion to protect critical habitat on non-federal lands that is not otherwise protected rests with the Governor in Council.

[3] These federally protected areas are: a national park of Canada named and described in Schedule 1 to the *Canada National Parks Act*, The Rouge National Park established by the *Rouge National Urban Park Act*, a marine protected area under the *Oceans Act*, a migratory bird sanctuary under the *Migratory Birds Convention Act, 1994* or a national wildlife area under the *Canada Wildlife Act* see ss. 58(2) of SARA.

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Executive Summary

Porsild's Bryum (*Mielichhoferia macrocarpa*) is a small brilliant green moss, often associated with waterfalls and calcareous rock and known to occur in at least 17 populations throughout Canada. It was listed as Threatened on the *Species at Risk Act* (SARA) Schedule 1 in 2011.

This action plan complements the *Recovery Strategy for the Porsild's Bryum in Canada* (Environment and Climate Change Canada 2016) and will be implemented in British Columbia, Alberta, Nunavut and Newfoundland. The proposed recovery measures in this action plan address the objective set out in the recovery strategy for the entire population and distribution of Porsild's Bryum in Canada.

No additional critical habitat is identified in this action plan, but it is expected that as the Schedule of Studies is completed, additional critical habitat may be identified and presented in an updated recovery strategy or action plan(s). Critical habitat identified in the species' recovery strategy is located on non-federal land and a federal protected area and proposed measures to protect this critical habitat are presented in section 1.4 of this action plan.

The recovery measures included in this action plan are required to implement the recommended recovery approaches outlined in the recovery strategy. Recovery measures proposed for the Porsild's Bryum are related to five broad strategies: (1) inventory and monitoring, (2) research, (3) outreach / stewardship, (4) habitat management, and (5) reintroduction and/or population augmentation.

The socio-economic evaluation was completed and it was determined that the direct and indirect costs of implementing this action plan are anticipated to be low over the short term (2019-2023) and the long term (2023 onwards). The implementation will benefit other species, habitat and ecosystems.

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1. Recovery Actions

1.1 Context and Scope of the Action Plan

The taxonomic designation of Porsild's Bryum has changed over time. At the time of the 2003 COSEWIC assessment, it was considered to be in the genus Mielichhoferia (*Mielichhoferia macrocarpa* (Hooker) Bruch & Schimper ex Jaeger & Sauerbeck). It was then placed in the genus Bryum (*Bryum porsildii* (I Hagen) Cox & Hedderson) and more recently in the genus Haplodontium (*Haplodontium macrocarpum* (Hooker) Spence). The currently accepted name of the species is *Haplodontium macrocarpum*. These names are synonomous and all refer to the Porsild's Bryum.

Porsild's Bryum was assessed as Threatened by the Committee on the Status of Endangered Wildlife in Canada (COSEWIC) in 2003, then subsequently listed as such on Schedule 1 of the *Species at Risk Act* (SARA) in 2011 using the name *Mielichhoferia macrocarpa*. As such, this is the name used for the purposes of this action plan.

Porsild's Bryum is a small brilliant green moss associated with shaded calcareous cliffs or rock outcrops and continuous or intermittent seepage (COSEWIC 2003; Environment and Climate Change Canada 2016). The distribution extent has changed little since it was assessed by COSEWIC in 2003. It is known to occur in at least 17 populations in Canada: 1 in British Columbia, 6 in Alberta, 7 in Newfoundland and Labrador, and 3 in Nunavut (Environment and Climate Change Canada 2016). There is limited information available to determine reliable trends in the population though loss of individuals and colonies, and a decline in habitat quality, has been noted at some locations (COSEWIC 2003). Porsild's Bryum has slow regeneration, limited dispersal ability, and narrow substrate requirements that likely make recovering from threats such as drought, temperature extremes, recreational activities, or stochastic events difficult (COSEWIC 2003; Belland and Limestone Barrens Species at Risk Recovery Team 2006).

The recovery strategy identifies the following population and distribution objective for Porsild's Bryum:

> "To maintain or increase the number of colonies, and sub-populations for all known extant populations of Porsild's Bryum, while also maintaining or increasing the distribution of colonies and sub-populations within each population, and, where feasible, to reestablish the species to locations where it has been extirpated and previously known to exist."
> (Environment and Climate Change Canada 2016)

This action plan addresses all populations of Porsild's Bryum in Canada and should be considered along with the *Recovery Strategy for Porsild's Bryum in Canada* (Environment and Climate Change Canada 2016). The recovery strategy provides more details on the strategic direction and approaches for recovery of Porsild's Bryum, critical habitat information, and background information on the species and its threats.

Provincial recovery documents for Porsild's Bryum have been developed in Alberta (Alberta Porsild's Bryum Recovery Team 2010) and Newfoundland and Labrador (Belland and Limestone Barrens Species at Risk Recovery Team 2006). These documents summarize provincial-specific distribution and habitat patterns, threats, recovery initiatives, etc.

1.2 Measures to be Taken and Implementation Schedule

Table 1. Implementation Schedule

#	Recovery Measures	Priority[a]	Threats or objectives addressed	Timeline
Broad Strategy: Inventory and Monitoring				
1	Conduct field surveys to locate Porsild's Bryum populations and subpopulations, both within and adjacent to the species' known range and in other potential locations deemed to have suitable habitat, to determine the species complete population size and distribution. Alberta Porsild's Bryum Recovery Team (2010), Belland and Limestone Barrens Species at Risk Recovery Team (2006), and Environment and Climate Change Canada (2016) identify areas of particular interest.	High	Knowledge gaps	Ongoing to 2023
2	Survey all sites to determine baseline population sizes, and identify threats and their impacts.	High	Knowledge gaps; All threats	Ongoing to 2021
3	Develop and implement a long-term monitoring program which examines population sizes and dynamics, colony numbers, threats, habitat trends (e.g., air temperature, relative humidity, and water quality), and microclimate trends at selected sites throughout the species' range.	High	Knowledge gaps; All threats	By 2021, then regularly (frequency dependent on location)
Broad Strategy: Research				
4	Develop and implement a research plan to determine the detailed biological needs of the species (e.g., physiological tolerances to light and temperature, water chemistry and substrate requirements, and resilience to disturbance), and habitat conditions.	High	Knowledge gaps; All threats	Ongoing to 2021
5	Further identify limiting factors and natural threats not already presented in the *Recovery Strategy for the Porsild's Bryum in Canada* (Environment and Climate Change Canada 2016).	High	Knowledge gaps	By 2021

6	Create a habitat model to predict species presence at potential sites.	Low	Knowledge gaps	Ongoing to 2022
7	Develop minimum viable population estimates.	Low	Knowledge gaps	By 2028
Broad Strategy: Outreach / Stewardship				
8	Develop educational material (e.g., brochures, displays at interpretative centres, and signage within protected areas) and other outreach initiatives to increase public understanding of threats to the species and promote stewardship.	Medium	Recreational activities	Ongoing to 2021
9	Work with various levels of government, stakeholders, and the general public to identify solutions for minimizing known threats (e.g., preventing campfires at Whitehorse Creek, Alberta, or preventing recreational use of cliffs at Ribbon Creek, Alberta).	Medium	Recreational activities; Industrial activity	Ongoing to 2021
10	Where appropriate, collaborate with industrial partners to minimize the effects of industrial activities (e.g., road dust in Mountain Park, Alberta). Avenues for collaboration include (but are not limited to) regular meetings / discussions, the development of beneficial management practices, and reviewing work procedures.	Medium	Industrial activity	Ongoing to 2021
11	Encourage the involvement of the public and industrial stakeholders in implementation efforts, including monitoring (where feasible) (e.g., through *Adopt-a-Plant Alberta* program).	Medium	Recreational activities; Industrial activity; Knowledge gaps	By 2021
Broad Strategy: Habitat Management				
12	Ensure critical habitat for extant populations on federal lands is legally protected.	High	Recreational activities; Industrial activity	Completed)
13	Work with provinces and landowners to secure effective protection of critical habitat for extant populations on non-federal lands.	High	Recreational activities; Industrial activity	By 2020
14	Install and maintain fencing, signage, etc. in strategic locations to conserve subpopulations vulnerable to recreational activities	Medium	Recreational activities	Ongoing to 2021, then as required

	(e.g., Ribbon Creek Lower and Upper, Whitehorse Creek 2, and Whitehorse Creek Boulder), if deemed necessary for population survival and recovery.			
15	When feasible, restore habitat at damaged locations. The necessity, extent and type of restoration will be site specific.	Medium	Recreational activities; Industrial activity; Stochastic events	As required
Broad Strategy: Reintroduction and Population Augmentation				
16	Develop reintroduction protocols.	Medium	All threats	By 2022
17	Determine the feasibility of reintroduction and population augmentation and identify priority sites for implementation.	Medium	All threats	By 2023
18	Re-introduce plants to restored habitat and/or implement population augmentation, if deemed feasible.	Medium	All threats	As required
19	Monitor effectiveness of reintroductions.	Medium	All threats	For at least 5-10 years post reintroduction

[a] "Priority" reflects the degree to which the measure contributes directly to the recovery of the species or is an essential precursor to a measure that contributes to the recovery of the species. High priority measures are considered those most likely to have an immediate and/or direct influence on attaining the population and distribution objectives for the species. Medium priority measures may have a less immediate or less direct influence on reaching the population and distribution objectives, but are still important for the recovery of the population. Low priority recovery measures will likely have an indirect or gradual influence on reaching the population and distribution objectives, but are considered important contributions to the knowledge base and/or public involvement and acceptance of the species.

1.3 Critical Habitat

1.3.1 Identification of the Species' Critical Habitat

Critical habitat of Porsild's Bryum was partially identified in section 7 and Appendix A of the federal recovery strategy (Environment and Climate Change Canada 2016). The recovery strategy also contains details about the identified critical habitat including its geospatial extent and biophysical attributes (section 7.1) (Environment and Climate Change Canada 2016). Please refer to that document for details.

Given the best available information, no additional critical habitat for Porsild's Bryum can be identified in this action plan. Critical habitat will be updated in an amended recovery strategy or additional action plan once the Schedule of Studies is completed. Refer to section 7.2 of the federal recovery strategy for a Schedule of Studies necessary to complete critical habitat identification (Environment and Climate Change Canada 2016).

1.3.2 Activities Likely to Result in Destruction of Critical Habitat

Examples of activities likely to result in the destruction of critical habitat may be found in section 7.3 of the federal recovery strategy (Environment and Climate Change Canada 2016).

1.4 Proposed Measures to Protect Critical Habitat

1.4.1 Proposed protection measures on Federal Lands

Critical habitat of Porsild's Bryum in Nunavut is identified within Quttinirpaaq National Park of Canada and has been legally protected by the process outlined in subsection 58(2) of SARA, as well as by the *Canada National Parks Act*. In addition, the Parks Canada Agency may use existing management tools to prevent destruction of critical habitat, such as posting notices, restricting access to the area, and educating visitors.

1.4.2 Proposed protection measures on Non-federal Lands

With regard to critical habitat on non-federal lands, Environment and Climate Change Canada will work with the Governments of British Columbia, Alberta, Nunavut, and Newfoundland and Labrador to report on critical habitat protection.

SARA requires that if, after consulting with the appropriate provincial or territorial minister, the Minister is of the opinion that the laws of the province or territory do not effectively protect the critical habitat of the species, then the Minister must make a recommendation to Governor in Council that an order be put in place to prevent the destruction of critical habitat.

The implementation of conservation measures is an important complementary strategy for preserving this species' critical habitat. Environment and Climate Change Canada

will work with the applicable provinces, non-governmental organizations, and individuals to facilitate the implementation of conservation measures.

2. Evaluation of Socio-Economic Costs and of Benefits

The *Species At Risk Act* requires that an action plan include an evaluation of the socio-economic costs of the action plan and the benefits to be derived from its implementation (SARA 49(1)(e), 2002). This evaluation addresses only the incremental socio-economic costs of implementing this action plan from a national perspective as well as the social and environmental benefits that would occur if the action plan were implemented in its entirety, recognizing that not all aspects of its implementation are under the jurisdiction of the federal government. It does not address cumulative costs of species recovery in general nor does it attempt a cost-benefit analysis. Its intent is to inform the public and to guide decision making on implementation of the action plan by partners.

The protection and recovery of species at risk can result in both benefits and costs. The Act recognizes that "*wildlife, in all its forms, has value in and of itself and is valued by Canadians for aesthetic, cultural, spiritual, recreational, educational, historical, economic, medical, ecological and scientific reasons*" (SARA 2002). Self-sustaining and healthy ecosystems with their various elements in place, including species at risk, contribute positively to the livelihoods and the quality of life of all Canadians. A review of the literature confirms that Canadians value the preservation and conservation of species in and of themselves. Actions taken to preserve a species, such as habitat protection and restoration, are also valued. In addition, the more an action contributes to the recovery of a species, the higher the value the public places on such actions (Loomis and White 1996; Fisheries and Oceans Canada 2008). Furthermore, the conservation of species at risk is an important component of the Government of Canada's commitment to conserving biological diversity under the *International Convention on Biological Diversity*. The Government of Canada has also made a commitment to protect and recover species at risk through the Accord for the Protection of Species at Risk. The specific costs and benefits associated with this action plan are described below.

2.1 Policy Baseline

The provinces of British Columbia, Alberta, and Newfoundland and Labrador, the Territory of Nunavut, and the federal government have access to many legislative, regulatory, and management tools for the conservation and stewardship of Porsild's Bryum (e.g., endangered species legislation, protected areas legislation, and environmental assessments). For example, Porsild's Bryum is listed as Endangered under Alberta's *Wildlife Act* and Threatened under Newfoundland and Labrador's *Endangered Species Act*. In addition, the populations in Nunavut occur within a national park and are subject to the *Canada National Parks Act* while the population in British Columbia occurs within a provincial park subject to British Columbia's *Park Act*.

Both Alberta and Newfoundland and Labrador have published recovery strategies for the species and recovery activities have been initiated in Alberta (AESRD 2013). For example, microclimatic sensors were installed at the Mountain Park population in Alberta (2011-12) to document key temperature and relative humidity conditions during the growing season. In addition, baseline site characteristics, including water and rock chemistry data, was obtained from several of the Alberta populations (AESRD 2013). A sign was erected at one site in Alberta near a popular campground by to inform campers about several plant species present, including Porsild's Bryum, and the importance of protecting them. Nationally, field surveys have recently (ca. 2015) been conducted at several of the populations of Porsild's Bryum in Canada in support of an updated COSEWIC status report.

Industrial policies and work procedures already in place may also contribute to the implementation of this action plan and thus the conservation of Porsild's Bryum. For example, dust levels along a haul road servicing a coal mine site adjacent to Mountain Park, Alberta, have been monitored to inform potential mitigative measures to reduce any impact to the species (Alberta Government 2014).

Additionally, many recovery measures can be carried out by federal or provincial species at risk funding programs, contributions by recovery biologists, or research by university partners.

2.2 Socio-economic Profile and Baseline

The measures outlined in this action plan relate primarily to inventory and monitoring, research, outreach and education and habitat management. Populations of Porsild's Bryum occur primarily within federal and provincial protected areas and parks. There are few communities or individuals that would be affected by the implementation of the measures identified in the action plan for Porsild's Bryum.

Within British Columbia, Porsild's Bryum is only known to occur within Muncho Lake Provincial Park and in Nunavut all populations are currently known from Quttinirpaaq National Park. Quttinirpaaq National Park is within the Nunavut Land Claims Agreement and is an area particularly important to Inuit from Grise Fiord and Resolute Bay.

However, the implementation of this action plan (i.e., inventory and monitoring, research, outreach and education and habitat management) is expected to have little to no effects on these communities.

Although Porsild's Bryum is found outside of federal or provincial protected areas in Newfoundland and Labrador, a non-governmental organization is actively involved with conservation and stewardship initiatives in the nearby limestone barrens. This recovery and conservation partnership has been ongoing in the area for several years in an effort to restore habitat and promote the long term protection and conservation of Porsild's Bryum and other species at risk in the area.

In Alberta, most of the recovery measures for the species will take place in various provincial protected areas with varying levels of protection. Potential affected stakeholders include transmission and telecommunication companies with dispositions on provincial lands and the mining industry. Porsild's Bryum is found within traditional territories of numerous First Nations in Alberta, but the implementation of the action plan (i.e., inventory and monitoring, research, outreach and education and habitat management) is expected to have little to no effects on these communities.

2.3 Socio-economic Costs of Implementing this Action Plan

Costs are those directly associated with the implementation of the recovery measures identified in the implementation schedule (Table 1), as well as those encountered as a result of that implementation. Only the incremental costs are considered and therefore do not include ongoing actions or initiatives discussed in section 2.1 (Policy Baseline). The direct and indirect costs of implementing the action plan are expected to be low (between $0 and $5 million) over the short term (2019-2023). Costs at the regional or provincial scale are expected to be minimal. Long-term (2023 onwards) costs are also expected to be minimal.

Social costs are the potential costs associated with implementing the action plan, which may have an impact on various stakeholders. Because there are a small number of known occurrences, the majority of occurrences are in protected areas, and there is lack of human-use associated with this species, the social costs anticipated from the implementation of this action plan are low.

2.4 Benefits of Implementing this Action Plan

2.4.1 Value of biodiversity to Canadians

It is anticipated that this action plan will contribute to the recovery of Porsild's Bryum and lead to the achievement of the population and distribution objective and the conservation and protection of habitat for the species.

Biodiversity is essential for healthy ecosystems, human health, prosperity, security, and wellbeing. Canadians derive many benefits from biodiversity including recreational,

aesthetic, educational, cultural benefits as well as ecological goods and services essential to human survival. Care for the environment is consistently ranked as one of Canada's top priorities in public opinion polls[4]. An opinion poll found that three quarters of Canadian respondents feel that preserving natural areas and the variety of native plant and animal life in Canada is important to them[5].

The total value of endangered species consists of non-consumptive use values (such as recreation, spiritual/cultural, research, and education), indirect use values (value of the ecological role of a species in an ecosystem), and non-use values (i.e., preserving the benefits of nature for future generations)[6]. Implementing the recovery measures of this action plan will have a positive impact on society. The direct value of this implementation, for the preservation or the enhancement of biodiversity, is not easily estimated.

2.4.2 Eco-tourism and cultural values

Eco-tourism is the fastest-growing area of the tourism industry (Mastny 2001; UNEP 2013). In 2004, this market grew three times faster than the industry as a whole and the World Tourism Organization estimates that global spending on eco-tourism is increasing by 20% a year, about six times the industry-wide rate of growth (TEEB 2008). Many of the Porsild's Bryum subpopulations are already located in or near parks (see Table 2 of the recovery strategy for details), but it is possible that education and stewardship activities may lead to a small increase in eco-tourism activity.

2.5 Distributional Impacts

Porsild's Bryum occurs on provincial, federal, and private properties, and the majority of sites are within protected areas. Thus, private landowners are not expected to absorb the direct incremental costs for the species' recovery. Any indirect incremental costs resulting from the implementation of recovery measures will be shared. Should additional populations of Porsild's Bryum be discovered on private land through activities identified in this action plan, the distributional impacts will be re-assessed.

[4] Canada's Fourth National Report to the United Nations Convention on Biological Diversity, 2010. Available online http://www.cbd.int/doc/world/ca/ca-nr-04-en.pdf Accessed December 3, 2010.
[5] Ipsos Reid Opinion Poll "Nine in Ten (87%) Canadians Say That When Connected to Nature They Feel Happier." Released January 7, 2011, www.ispsos.ca
[6] Non-use values include bequest value (satisfaction of knowing that future generations will have access to nature's benefits), altruist value (satisfaction of knowing that other people have access to nature's benefits) and existence value (satisfaction of knowing that a species or ecosystem exists).

3. Measuring Progress

The performance indicators presented in the associated recovery strategy provide a way to define and measure progress toward achieving the population and distribution objectives.

Reporting on implementation of the action plan (under section 55 of SARA) will be done by assessing progress towards implementing the broad strategies.

Reporting on the ecological and socio-economic impacts of the action plan (under s. 55 of SARA) will be done by assessing the results of monitoring the recovery of the species and its long-term viability, and by assessing the implementation of the action plan.

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Appendix A: Effects on the Environment and Other Species

A strategic environmental assessment (SEA) is conducted on all SARA recovery planning documents, in accordance with the *Cabinet Directive on the Environmental Assessment of Policy, Plan and Program Proposals*[7]. The purpose of a SEA is to incorporate environmental considerations into the development of public policies, plans, and program proposals to support environmentally sound decision-making and to evaluate whether the outcomes of a recovery planning document could affect any component of the environment or any of the Federal Sustainable Development Strategy's[8] (FSDS) goals and targets.

Recovery planning is intended to benefit species at risk and biodiversity in general. However, it is recognized that implementation of action plans may also inadvertently lead to environmental effects beyond the intended benefits. The planning process based on national guidelines directly incorporates consideration of all environmental effects, with a particular focus on possible impacts upon non-target species or habitats. The results of the SEA are incorporated directly into the action plan itself, but are also summarized below in this statement.

The measures set out in this document are expected to have no negative effects on other species. Many of the measures pertain to inventory / monitoring or research and therefore should not adversely impact other species. Other actions pertaining to outreach / stewardship and habitat management may create benefits for the surrounding habitat and ecosystems.

[7] www.canada.ca/en/environmental-assessment-agency/programs/strategic-environmental-assessment/cabinet-directive-environmental-assessment-policy-plan-program-proposals.html
[8] www.ec.gc.ca/dd-sd/default.asp?lang=En&n=CD30F295-1

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1. ᑕᒪᐃᓐᓄᖏᓪᓗ ᑲᓇᑕᓕᒫᒥ ᓇᓄᐃᑦ ᐊᐅᓪᑕᐅᓂᖏᓐᓄᑦ ᓯᖏᐅᑎ

2. ᐱᐅᖅᓯᐊᓂᖏᓪᓗᑦ ᑐᕋᖏᐃᖏᑦ ᐊᒻᒪᓗ ᖃᓄᐃᓂᕆᖓ ᐊᐅᓪᓚᑎᓂᖏᓪᓗᑦ ᓯᖏᐅᑎ ᐅᕐᓂᔪᓐᓄᑦ
ᐱᓴᐅᑲᓄᓂᓂᖏᓪᓗᑦ ᐊᑐᐊᒃᐃᑦ ᑲᓇᑕᐅᕐᓴ ᐊᕕᑲᑐᖅᓯᒪᕕᐊᓂ

3. ᐊᐅᓪᓚᑎᓂᖏᓪᓗᑦ ᐱᓯᓐᓇᐅᑎᑦ

4. ᒪᓂᒪᑐᖅᑯᓂᓂᑦ ᐃᓪᓕᐅᓴᕐᕆᐅᓯᑦ ᐃᓗᓕᓐᖏᑦ

5. ᑯᐊᐅᓱᓴᑯᑦ ᐊᒻᒪᓗ ᐊᐅᓪᓚᑎᓵᑦ ᐃᓪᓕᖅᖃᑐᑦ ᐊᐅᓪᓚᑎᓂᖏᓪᓗᑦ ᐱᓕᓇᖅᑎᑭᖓᓐᓂᑦ

6. ᓂᓇᐅᑭᕐᕆᐅᓯᑦ ᐱᕕᖅᖃᕐᓴᓴᓐᓄ

7. ᓇᓴᒥᓂᖅᓂᓪᓗᑦ ᑎᒍᒥᐊᖅᑏᑦ ᑲᑎᒪᑎᑦᑎᓂᖅ ᐊᒻᒪᓗ ᑲᑎᒪᓵᑦ ᓇᓴᒪᕐᓴᖅᓂᓪᓗᑦ ᖃᓄᐃᓕᓐᓄᓂᐅᕐᓴᓇᐊᓪᓕᑦ

8. ᖁᔪᓐᓇᒃᑐᒃ ᐊᒻᒪᓗ ᐊᐱᖅᑯᑎᑦ

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1. ᑕᒪᐃᓐᓄᖅᑐᖅ ᑲᓇᑕᓕᒫᒥ ᓇᓄᐃᑦ ᐊᐅᓚᑕᐅᓂᖅᑭᓐᓄᑦ ᓴᖅᓇᐅᑎ

- ᑖᓐᓇ ᑲᓇᑕᓕᒫᒥ ᓇᓄᐃᑦ ᐊᐅᓚᑕᐅᓂᖅᑭᓐᓄᑦ ᓴᖅᓇᐅᑎ ᐱᖃᓯᐅᕐᓴᓂᐊᖅᑯᖅ ᐱᖓᓱᓯᕐᖅᑐᓕᕐᓂᑦ:
  - ᐃᓚᖓ 1 : ᑲᕿᒪᑐᖃᓐᓂᑦ ᐃᓚᓕᐅᕐᔪᐅᕐᔪᑦ
  - ᐃᓚᖓ 2-7 : ᐊᓗᑐᖅᓯᒪᔪᓂᑦ ᐊᓪᒪᓗ ᐊᕆᑦᑐᖅᓯᒪᔪᓂᑦ ᑎᑎᖅᑲᑦ
    - ᐃᓄᕆᐊᓗᐃᑦ ᐊᕆᓴᖅᑐᓯᒪᕆᐊᓂ (ᓄᓇᑦᓯᐊᖅ/ᓴᓗᓇᐃᖏ) (ᐃᓚᖓ 2) ᐊᓪᒪᓗ ᐋᐦᑎᐅᓕᔨ (ᐃᓚᖓ 5) ᐱᓴᓰᖅᓯᒪᔨᓴ
    - ᓄᓇᕗᑦ (ᐃᓚᖓ 3) ᐱᓴᓰᓕᐊᑭᓪᓚᔪᖅ
    - ᐃᓕᓐᑐᕐ (ᐃᓚᖓ 4), ᑯᐸᐃᓴ -ᐃᐅᐳ ᐃᒪᕆᒪᒥ ᐊᕆᑦᑐᖅᓯᒪᔪᓂᑦ - ᓄᓇᕆᓴ ᐃᒪᕆᒪᒥ ᐊᕆᑦᑐᖅᓯᒪᔪᓂᑦ (ᐃᓚᖓ 6) ᐊᓪᒪᓗ ᓄᕐᓂᓪᓂ ᐊᓪᒪᓗ ᓂᕐᑐᐊᖏ (ᐃᓚᖓ 7) ᐃᖅᑲᓇᐃᓴᖅᑲᓪᓂᓕᐊᖅᑯᑦ

- ᑎᑎᕿᖅᑕᐅᑭᐊᓚᖅᓯᒪᔪᖅ ᓴᕐᒪᑐᖅᑲᑦᑯᓪᓂᑦ ᐃᓚᓪᐅᕐᔪᐅᕐᔪᑦ ᐱᓴᓰᖅᓯᒪᓪᓚᑯᑦ. ᑭᓯᐊᓂ, ECCC ᐱᓇᓯᓴᐅᓪᓚᑦ ᑭᖑᓪᓪᓚᖅᓯᐅᓂᑦ ᐱᖃᓯᐅᕐᔪᓴᐅᓪᓚᐅᑐᑦ ᐃᓚᓪᐅᕐᔪᐅᔪᓂᑦ ᑕᒪᐃᓐᓂᓪᓕᐸᓗᒃ (ᐅᕙᕐᓇᑲᓐᓂᑦ ᑕᒪᕿᑦᓪᓕᒃᑦ) ᐊᓗᑐᖅᓯᒪᔪᓇᑦ ᐊᓪᒪᓗ ᐊᕆᑦᑐᖅᓯᒪᔪᓇᑦ ᑎᑎᖅᑲᐃᑦ ᐱᓴᓰᖅᑕᐅᓚᐅᐅᑭᐅᓂᖅᑭᓐᓄᑦ.

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2. ᐱᐅᖅᓯᐊᓂᖏᓐᓇᑦ ᑐᖃᖏᕐᕕᒃᑐᑦ ᐊᒻᒪᓗ ᖃᓄᐃᓂᖏᑎᒍᑉ ᐊᐅᓪᓚᑎᓂᖏᓐᓇᑦ ᕿᖏᐅᑎᑎ ᐅᓴᓐᓄᓕᒃᓂᑦ ᐱᕆᐅᖅᖃᓂᖏᓐᓂᖏᓐᓇᑦ ᐊᑐᐊᒃᐃᑦ ᖃᓇᑕᐅᐸ ᐊᕕᖅᑐᖅᓯᒪᕕᐊᓂ

ᑭᓪᓗᖏᕕᒃᑐᒃ	ᒪᓕᒃᓕᐅᖏᓂᖅ	ᑐᖃᖅᑕᒃᑐᒃ	ᐊᖏᖃᒍ	ᖃᓄᐃᓕᓕᖏᓂᒃᑐᒃ
ᓂᐅᓴᓐᓯᓐ ᐊᒻᒪᓗ ᓯᓚᐃᓐᓇᖅᑐᖅ	*ᐅᒪᔪᐃᑦ ᐊᑦᑕᓇᖅ ᖃᑐᒡᑐᑦ ᐱᖏᓂᖅ*	ᐊᑦᑕᖏᓇᑐᐃᓐᓇᓗᐊᓯᓪᓗᑦ*	2002	ᐱᓴᓪᓕᐊᔪᑦ
ᒪᓂᑐᓴ	*ᐅᒪᔪᐃᑦ ᐊᑦᑕᓇᖅ ᖃᑐᒡᑐᑦ ᐱᖏᓂᖅ*	ᐊᑦᑐᖅᑕ	2008	ᐱᓴᓪᓕᐊᔪᑦ
ᐊᓐᑎᐅᓗᓚᑦ	*ᐅᒪᔪᐃᑦ ᐊᑦᑕᓇᖅ ᖃᑐᒡᑐᑦ ᐱᖏᓂᖅ, 2007*	ᐊᑦᑕᖏᓇᖅᓯᓴᓪᓕᐊᔪᑦ	2009	ᐱᓕᕆᖅᓯᒪᔪᖅ
ᓂᐸᐃᒃ	*Loi sur les Espèces Menacées ou Vulnérables*	ᐊᑦᑕᖏᓇᑐᐃᓐᓇᓗᐊᓯᓪᓗᑦ*	2009	ᐱᓴᓪᓕᐊᔪᑦ
ᓄᓇᕐᑕ	*ᓄᓇᕐᒥ ᐅᒪᔪᓕᓗᓂᖏᓐᓂᑦ ᐱᖏᓂᖅ*	ᑎᑎᕆᖅᓯᒪᔪᖅᓴᕐᑕᑐᖅ	-	ᐱᓴᓪᓕᐊᔪᑦ
ᓄᓇᑦᓯᐊᖅ	*ᐅᒪᔪᐃᑦ ᐊᑦᑕᓇᖅ ᖃᑐᒡᑐᑦ (ᓄᓇᑦᓯᐊᖅ) ᐱᖏᓂᖅ*	ᐃᓱᒪᓕᓕᒫᑦ	2014	ᐱᓕᕆᖅᓯᒪᔪᖅ (ISR ᑲᑐᔨᓪᓗᑎᑦ ᐊᐅᓪᓚᑎᓂᖏᓐᓂᖏᓐᓇᑦ ᕿᖏᐅᑎᑎ)
ᔨᑲᓪ	*ᔨᑲᓪ ᐅᒪᔪᓕᓗᓂᖏᓐᓂᑦ ᐱᖏᓂᖅ*	ᑎᑎᕆᖅᓯᒪᔪᖅᓴᕐᑕᑐᖅ	-	

3. ᐊᐅᓪᓗᓂᓂᕐᒧᑦ ᐱᔨᓴᐅᑎᑦ

- ᑕᓴᓇ ᓴᖏᒪᑐᖅᑯᓴᓂᑦ ᐃᓪᓚᐅᕐᓴᐅᔨᑦ ᐃᓚᓴᖅᓯᕐᖅ ᑕᒪᓴᓇ ᓇᓂᕐᓂᑦ ᐊᐅᓪᓗᓂᖅ ᑲᑐᕐᐅᓚᑯᑦ ᑲᒪᒋᔭᓂᐅᑯᑦ ᑲᑎᒪᔨᖅᓯᒪᑦ ᐊᕐᒪᑐᖅᓯᒪᔨᓂᑦ/ᐊᕐᕿᑦᑐᖅᓯᒪᔨᓂᑦ ᓴᖏᒪᓂᑦ ᐊᕐᒪᓗ ᐅᒪᔨᓚᓱᕐᔨᐊᑯᑦ ᑲᑎᒪᔨᖅᑭᓐᓄᑦ

- ᑕᒪᓴᓇ ᐱᓚᓱᐊᓱᓂᐊᖅᑐᒍ, ᑕᒪᓴᓇ ᓴᖏᒪᑐᖅᑯᓴᓂᑦ ᐃᓪᓚᐅᕐᓴᐅᔨᑦ ᐃᓕᖅᖅᑎᑦᑎᓄᖅ ᓴᖏᒪᑐᖅᑯᑦ ᐃᖅᑲᓇᐃᓐᓱᓱᓱᓚᐅᓚᖅᓯᓂᑦ ᐊᐅᓪᓗᓂᖅᓯᒪᔨᑦ ᑲᑎᒪᔨᖅᑎᓂᕐᒥᑦ ᒪᓪᓕᒐᒍ ᐊᑦᑐᒍ ᓴᖏᒪᑐᖅᑯᑦ ᑭᐃᓕᖅᑭᑐᖅᓯᒪᔨᓂᑦ ᐊᕐᒪᓗ ᐱᖓᓱᑐᒍᓐᓇᖅᑐᑦ ᐊᑦᑐᒍ ᐃᓚᖅᑭᑐᖅᓯᒪᔨᓂᑦ ᐊᑐᐊᖅᓯᒪᓂᖅ ᐊᑦᑐᓂᖅ ᐱᔨᓴᐅᑎᑦᑎᓂᕐᓂᑦ

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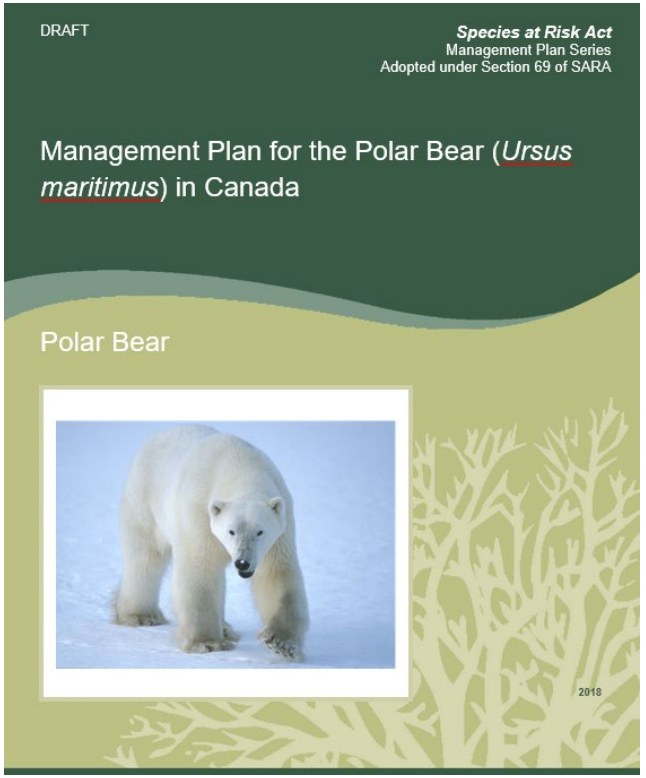
4. ᑲᓴᒪᑐᖏᑦᓂᓪᓂᑦ ᐃᓕᓕᐅᔪᕆᐅᔾᔪᑦ ᐃᓗᓕᓐᖏᑦ

ᓇᓗᓇᐃᖏᓪᓕᔾᔪᑦ ᐃᓗᓕᓐᖏᑦ

1. COSEWIC ᐅᒪᔾᐃᑦ ᖃᐅᔨᓇᓯᑕᐅᓂᖏᑦ
2. ᐅᒪᔾᐃᑦ ᖃᓄᐃᓕᓐᖑᓕᕐᓂᖏᑦ
3. ᐅᒪᔾᐃᑦ ᒥᔪᓴᓄᑦ ᑐᑭᓯᑎᐊᕈᑎ
4. ᐊᑦᑕᕐᓇᕈᑎᑦ
5. ᐊᐅᓪᓚᑦᑎᓂᕐᒧᑦ ᑐᖅᑖᑦ
6. ᐃᓗᓕᖃᐅᖅᑐᑦ ᐊᑐᐊᒃᔪᓴᐃᑦ ᐊᓪᒪᓗ ᐊᐅᓪᓚᑦᑎᓂᕐᒧᑦ ᖃᓄᐃᓕᓐᐅᕈᑎᔪᓴᑦ
7. ᖃᐅᔨᔪᕐᓂᕐᒧᑦ ᐱᓭᓕᓐᐊᓂᖓ
8. ᖃᐅᔪᕈᐊᕐᕆᒡᓴᑦ



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5. ᑯᐊᐳᓱᓴᕆᑯᑦ ᐊᓚᒐ ᐊᐅᓪᓕᑎᓵᑦ ᐃᓪᖏᖅᑐᑎᑦ ᐊᐅᓪᓕᐅᓂᖅᓴᖏᑦ ᐱᓪᓇᖅᑎᑎᓯᓐᓂᖅᓴᖏᑦ

- ᑲᑐᔨᖅᑎᑎᑦ ᐃᓪᒥᑎᐅᖅᑕᐅᔾᔪᑦ ᑕᐃᑲᓂ ᑲᓇᑕᒥ ᓇᓂᐃᑦ ᐊᐅᓪᓕᐅᓂᑎᔨᖏᓐᓄᑦ ᑲᑎᒪᔨᕆᓪᑦ (PBAC), ᐱᖅᓯᐅᑎᓪᓗᑎᑦ ᓄᓇᕗᑦ ᓴᖅᓪᒪᑦ, ᓄᓇᕗᒻᒥ ᐅᒪᔨᓪᓚᕆᐊᖅᑯᑦ ᑲᑎᒪᔨᖏᑦ ᐊᓚᒐ ᓄᓇᕗᑦ ᑐᙳᓴᕐᕄ, ᖃᐃᖅᑯᓯᐅᓪᓚᖅᑯᑦ ᐃᖅᑲᓇᐃᔨᖅᑎᖅᑐᓪᓗᑎᑦ ECCC ᐱᓕᕆᓪᓚᐊᑎᑦᑎᓂᖏᑦ ᑎᑎᕆᖅᑕᐅᕈᐊᐅᓱᖅᓯᓚᕐᓗᑎᑦ ᓴᓪᓕᑐᖅᑐᓯᓂᖅᓴᖏᑦ ᐃᓪᓚᕆᐅᔪᖅᓯᔪᑦ ᐃᖅᑲᓇᐃᔨᖅᑎᑎᓯᓐᓂᖅ ᐱᓪᓇᐃᑦ.

- ᑕᒪᐅᑯᐊ ᐱᖅᑭᖅᑕᐅᔾᔪᑦ ᑎᑎᖅᑲᓪᓚᐊᖓᔪᒪᔨᓂᖅ ᐊᐅᓕᓪᓚᖅᑎᑕᐅᓪᓚᐅᖅᑯᑦ ᖃᑎᖅᑕᕆᐅᔪᓐᓇᖅᑐᓪᓗᑎᑦ ᓇᓂᐃᑦ ᐊᐅᓪᓕᐅᓂᑎᔨᖏᓐᓄᑦ ᑲᑎᒪᔨᕆᓪᑦ (PBAC) ᑲᑎᒪᔨᖏᓐᓄᑦ ᐊᓚᒐ ᑲᓇᑕᒥ ᐅᒪᔨᐃᑦ ᐱᔨᑎᖅᑕᐅᓂᖏᓐᓄᑦ ᐊᐃᑦᑐᖅᑕᐅᔾᔪᑦ ᑕᐃᑲᓂ ᓄᐃᐱᓗ 2018

- ᓇᓂᐃᑦ ᐊᐅᓪᓕᐅᓂᑎᔨᖏᓐᓄᑦ ᑲᑎᒪᔨᕆᓪᑦ (PBAC) ᑲᑎᒪᔨᖏᑦ ᐊᓚᒐ ᑲᓇᑕᒥ ᐅᒪᔨᐃᑦ ᐱᔨᑎᖅᑕᐅᓂᖏᓐᓄᑦ ᑭᓪᓕᖃᖅᐃᖏᓐ ᐱᐃᓴᖃᖅᑎᑕᐅᑲᓂᓪᓂᐊᖅᑯᑦ ᐊᑐᐃᓇᖃᖅᓂᖅᑎᑦ ᐅᖃᐅᔾᔨᖅᑲᓂᑦ ᐃᓪᓚᕆᐅᔪᖅᓯᔪᑦ ᑕᐃᑲᓂ ᔨᓂ 2019

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6. ᓂᓚᐅᑎᑦᑐᐅᔪᑦ ᐱᕕᖏᕐᕕᓴᓴᖅᓯᒪᓕ:

ᓄᕕᐱᓚ 2018 - ᑲᓪᓄᐅᓚ 2019

- ᓇᓄᐃᑦ ᐊᐅᓪᓚᑎᐱᖓᓪᓄᑦ ᑲᑎᒪᔨᕐᓯᓪᓕᑦ (PBAC) ᑲᑎᒪᔨᖓᑦ **(ᐱᖏᓪᐅᑎᑦᓯᒍᑦ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓂᓚᔨᕐᔪᐊᖏᑦ)** ᐊᓴᒪᓗ ᑲᓇᑕᒥ ᐅᒪᔪᐃᐃᑦ ᐱᔨᑎᖏᑕᐅᓂᖓᓪᓄᑦ ᐊᕕᑦᑐᓴᓕᔪᓂᑦ ᐱᓪᓇᐅᖏᑦ ᑎᑎᕋᓚᐊᖏᓪᓕᕐᑕᑦᑕᐅᓕᔪᒥᑦ ᒪᓂᒪᑐᖏᓕᓂᑦ ᐃᓪᓗᓂᐅᓴᔪᑦ ᖃᒥᖅᑭᐅᑐᔪᓐᓇᖅᑐᓂᑦ ᐊᓴᒪᓗ ᐅᖏᐅᔨᓴᖏᓐᓇᐅᓗᑎᑦ (ᔨᓂᓪᓂᖏᓱᒥ ᐱᕕᖏᓴᓴᖏᑎᓂᑦ ᐅᖏᐅᔨᓴᖏᐃᑦ)

ᔪᓂ 2019

- ᕚᖓᓪᓂᖅᒥ ᑎᑎᖏᓪᓂᐊᖓᔪᖏᖅ ᑕᐃᑯᖓ ᒪᓂᒪᑐᖏᓕᓂᑦ ᐃᓪᓗᓂᐅᓴᔪᐅᔪᑦ ᐊᐅᓪᓚᖏᑎᑕᐅᓂᐊᖏᑯᑦ ᑕᑲᑯᖓ ᓇᓄᐃᑦ ᐊᐅᓪᓚᑎᐱᖓᓪᓄᑦ ᑲᑎᒪᔨᕐᓯᓪᓕᑦ (PBAC) **(ᐱᖏᓪᐅᑎᑦᓯᒍᑦ ᓄᓇᕗᒻᒥ ᐅᒪᔪᓂᓚᔨᕐᔪᐊᖏᑦ)** ᖃᒥᕐᑕᔪᐅᔪᓐᓇᖅᑐᓂᑦ (ᔨᓂᓪᓂᖏᓱᒥ ᐱᕕᖏᓴᓴᖏᑎᓂᑦ ᐅᖏᐅᔨᓴᖏᐃᑦ)

ᐅᑭᐊᓴᖅ 2019

- ECCC ᖃᒥᕐᑕᓂᒥᐊᖏᑦ ᐊᓴᒪᓗ ᓴᖏᑭᑎᑎᓚᑎᑦ ᐊᐃᕐᕐᓇᓂᑦ ᐅᖏᐅᔨᓴᖏᓐᓇᖅᓴᓴᖏᑎᑦ
- ECCC ᓂᓚᐅᕐᑐᑦ ᑕᑲᐞᓂᓕᒫᖏᓪ (ᐅᕐᓂᔪᓪᓂᑦ ᑕᒪᐃᓪᓂᓕᒫᑐᓗᒃ) ᐊᓱᔪᑕᖏᓕᕐᔪᓂᑦ ᐊᓴᒪᓗ ᐊᐱᑎᑕᖏᓕᕐᔪᓂᑦ ᕕᖓᑦᓴᖏᓕᓂᐊᓚᐊᖓᑦ

ᓄᖓᔪᐊᓂ 2019/ᐱᒋᐊᓕᕆᔨᓂᑦ 2020

- ᑲᑎᒪᑎᑦᑎᐸᓪᓂᓂᕐᒥᑦ ᐱᑕᖏᐸᓪᓂᓚᐊᖏᕐᒪᖓᓗᖏᓪ ᖏᐅᔨᓴᖏᑕᐅᓂᐊᓚᕐᓂᖅ ᐱᔪᓚᐊᖏᖏᐸᑕ ᑕᑲᐃᓂᓂᓚᐅᓪᓗᖏᓪ ᐃᓪᓗᓂᐅᓴᔪᐅᔪᑦ, ᐱᖏᓪᐅᑎᑦᓯᒍᑦ ᓇᓪᓕᐊᑐᐃᓪᓚᖏᓪ ᐱᔨᔪᑎᓪᓕ ᓇᒥᓂᖏᓴᓂᕐᒧᑦ ᑎᒍᒥᐊᖏᑎᐅᓂᖏᓪ ᑕᑯᓄᓪᓚᖏᑦ ᐅᒪᔪᓂᓚᔨᕐᔪᐊᖏᑦ ᑲᑎᒪᔨᖓᑦ ᐅᖏᐅᔨᓴᖏᐃᑦ.

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6. ᓂᓚᐅᒋᔭᐅᓪᓕᑦ ᐱᕆᖏᕐᕆᓴᖅ (ᑲᓪᓚᓪᖅ) :

ᐱᒋᐊᓕᕆᑦᑕᖏᓂᑦ-ᖃᑎᒌᓚᐊᓄᑦ 2020

- ECCC ᖃᒥᖅᑭᓂᐊᖅᑯᑦ ᐊᓱᒪᓗ ᓴᖅᑭᑎᑦᑎᓗᑎᑦ ᐅᖏᐅᓯᓴᐃᑦ ᐱᔭᐅᓯᒪᓪᓕᑦ ᑕᐃᒃᓂ ᑭᙳᔨᓕᖅᑭᒋ ᑲᑎᒪᓂᖏᓕᖅᑕ
- ᑲᓇᑕᒥ ᓇᓄᐃᑦ ᐊᐅᓪᒥᐅᓂᖅᓯᓄᑦ ᕿᖅᓇᐅᑎ ᓴᖅᑭᖅᑕᐅᓯᒪᔪᖅ “ᑐᑭᓯᖏᐅᑎᑎᔪᑦ” ᑎᑎᖅᑲᐅᑦᓗᔪ ᑕᐃᒃᓂ SAR ᑭᒃᑯᑐᐃᓇᐃᑦ ᐊᑎᓕᐅᖅᓱᖏᓂᑦ ᐅᑦᓗᓄᑦ 60 ᐅᖏᐅᓯᓴᖏᖅᓱᐃᐅᔪᓇᖅᓗᓂᓂ(ᑭᙳᔨᓕᖅᑭᒋ ᐱᘎᓴᖏᖅᓱᐃᕐᔪᖅ ᐅᖏᐅᓯᓴᖏᖅᓱᐃᐅᓂᙳᖓ)

ᖃᑎᒌᓚᐊᓂᑦ-ᓄᙳᒍᐊᓄᑦ 2020

- ᐱᑕᖏᓇᐊᖏᖅᑲᖅ ᐁᖅᑭᑭᐊᓇᐊᓕᓂᓂᑦ ᓴᖅᑭᖅᑕᐅᕆᖏᖅᓯᐊᓇᕐᓂ
- ᑲᓇᑕᒥ ᕿᖅᓇᐅᑎ ᑐᓂᓂᖅᑯᑕᐅᓕᐅᖅᑐᖅ ᐅᒪᔪᕆᓇᔪᖅᕆᐊᖅᑯᑦ ᑲᑎᒪᔪᙱᓚᓄᑦ (**ᐱᖏᕆᐅᑎᑦᓗᑎᑦ ᓄᓇᕐᒥᒃ ᐅᒪᔪᕆᓇᔪᖅᕆᐊᖅᑯᑦ ᑲᑎᒪᔪᙱᓚᓄᑦ**) ᐃᓱᒪᓕᐅᕐᑕᐅᖅᑯᑦᓗᔪ

ᓄᙳᒍᐊᓂ 2020/ᐱᒋᕿᓕᕆᑦᑕᖏᓂᑦ 2021

- ᐋᒪᒪᕆᔭᖅᕿᑦ, ᑲᓇᑕᒥ ᓇᓄᐃᑦ ᐊᐅᓪᒥᐅᓂᖅᓯᓄᑦ ᕿᖅᓇᐅᑎ ᓴᖅᑭᖅᑕᐅᓯᒪᓪᓗᓂ ᑭᙳᔨᓕᖅᑭᒋᓗᒃ ᑎᑎᖅᑲᐊᒍᔪᖅ ᑕᐃᒃᓂ SAR ᑭᒃᑯᑐᐃᓇᐃᑦ ᐊᑎᖅᑭᖅᑕᐅᓯᒪᓂᑦ

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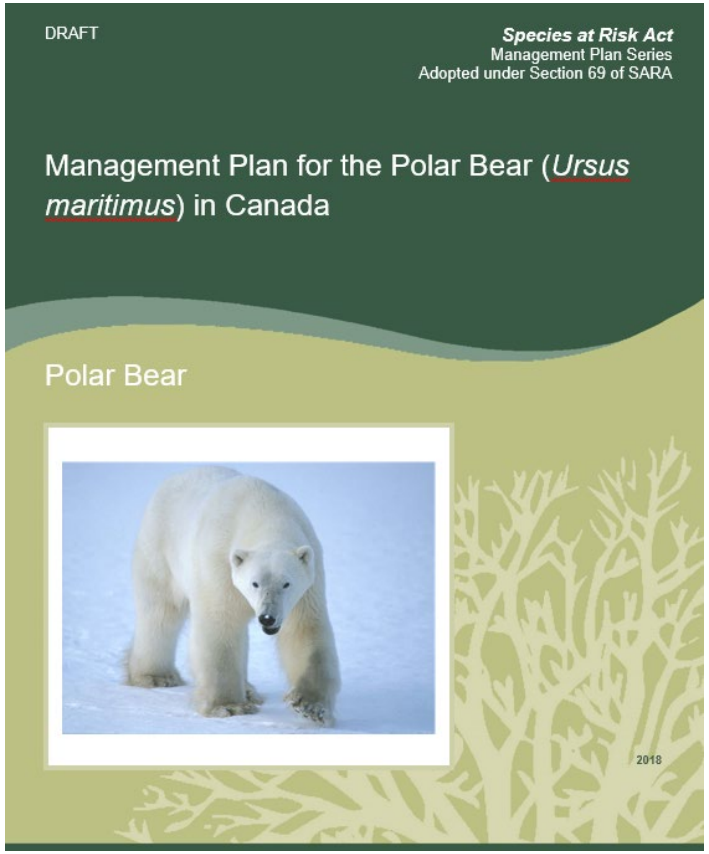
7. ᓇᓚᒥᓂᖅᓂᕐᓗᑦ ᑎᒍᒥᐊᖅᑏᑦ ᑲᑎᒪᓂᖅ ᐊᑐᓚᓗ ᑲᑎᒪᔨᑦ ᐃᓚᒋᑕᖅᓂᕐᓗᑦ ᖃᓄᐃᓕᓴᖅᑎᑦ

ᑲᑎᒪᔨᑦ/ᓄᓇᖅᑲᖅᓯᒪᔪᑦ ᑲᑐᔾᔨᖅᑎᑦᑎᑦ	ᓄᓇᕗᓂᔅᓴᖅᑐᒥᐅᑦ ᑲᑎᒪᑦᑎᓂᕐᓗᑦ ᖃᐅᔨᖅᑲᑕᐅᓯᒪᕗᖅ ᑲ?	ᐊᑐᕐᓂᐅᕆᐊᔪᑦ ᖃᓄᖅ ᑲᑎᒪᔨᓄᑦ ᒪᓕᑦᑕᐅᔭᕆᐊᖅᓯᒪᓂᖅ ᓇᓚᒋᑕᐅᓂᕐᓗᑦ		
		ᑲᑎᒪᑐᐃᓐᓇᖅ	ᑎᑎᖅᑲᑎᒍᑦ ᐃᓄᓕᒻᓂᑦ ᐃᓕᑦᑎᓂᕐᑎᓂᖅ	ᐃᓄᐹᓂᑦ ᐃᓄᓕᒻᓂᑦ ᐃᓕᑦᑎᓂᕐᑎᓂᖅ
ᐃᓄᓇᕕᒻᒧᐃᑦ ᐅᓪᓗᔪᓄᑦ ᑲᑎᒪᔨᑦ	ᐊᑦᑐ	✓		
ISR-ᐅᓪᓗᔪᑦᓯᔪᕕᐹᓂᑦ ᑲᑎᒪᔨᖅᑎᑦ ᑲᑎᒪᔨᕐᓇᓐᖓᑦ ᓂᒋᐅᑉ ᓯᕗᖅᑐᒃᖅᓯᒪᓂᑦ	ᐊᑦᑐ	✓		
ISR-ᐅᓪᓗᔪᑦᓯᔪᕕᐹᓂᑦ ᑲᑎᒪᔨᖅᑎᑦ ᑲᑎᒪᔨᕐᓇᓐᖓᑦ ᓄᓇᕗᑦᔪᐊᖅ				
ᓄᓇᕗᒻᒥ ᐅᓪᓗᔪᑦᓯᔪᕕᐹᓂᑦ ᑲᑎᒪᔨᖅᑎᑦ				
ᓄᓇᕕᒃ ᐃᓪᓕᕕᒻᒥ ᐊᕕᑦᑐᖅᓯᒪᔪᓂᑦ ᐅᓪᓗᔪᑦᓱᓂᖅᓂᕐᓗᑦ ᑲᑎᒪᔨᑦ	ᐊᑦᑐ		✓	
ᐊᖏᓚᓯᒃᓂᕐᓗᑦ, ᐃᖅᑲᓗᓯᒃᓂᕐᓗᑦ ᐊᑎᓚᓃ ᒥᑭᐊᖅᓂᐊᓂᕐᓗᑦ ᐊᐅᓪᓚᑦᑎᓂᕐᓗᑦ ᑲᑎᒪᔨᕐᓇᓐᖓᑦ (ᑯᐸᐃᒃ)	ᐊᑦᑐ	✓		
ᐃᐅᐳ ᐃᓚᒃᓕᒻᒥ ᐊᕕᑦᑐᖅᓯᒪᔪᓂᑦ ᐅᓪᓗᔪᑦᓱᓂᖅᓂᕐᓗᑦ ᑲᑎᒪᔨᑦ				
ᑐᖅᖢᓂᑦ WPCB/ᓄᓇᕐᓯᐊᐅᓂᑦ ᓴᕿᒪᖢᖅ	✓			✓

- ECCC ᐃᖅᑲᓇᐃᔨᖅᑎᖏᑦ ᐃᖅᑲᓇᐃᔨᖅᑎᖅᑐᖅᓯᐊᖅᑯᑦ ᓄᓇᕗᒻᒥ ᐅᓪᓗᔪᑦᓯᔪᕕᐹᓂᑦ ᑲᑎᒪᔨᖅᑎᑦ ᐃᖅᑲᓇᐃᔨᖅᑎᖅᑐᖅᓯᒪᓂᑦ ᖃᐅᔨᖅᑲᑕᐅᓯᒪᔪᑦ ᓇᓚᒋᑕᐅᓂᕐᓗᑦ ᐱᖅᓯᐅᑦᔨᔪᓂᖅᓯᒪᓂᖅᓂᕐᓗᑦ ᑲᑎᒪᔨᖅᑎᑦ ᐊᐅᓚᑦᑎᓂᕐᓗᑦ ᑲᑎᒪᔨᖅᑎᑦ ᐊᐅᓚᑦᑎᓂᕐᓗᑦ ᓄᓇᕗᒻᒥ ᐅᓪᓗᔪᑦᓯᔪᕕᐹᓂᑦ ᑲᑎᒪᔨᖅᑎᑦ ᐊᑐᐅᓪᓚᑦᑎᓂᕐᓗᑦ ᑲᑎᒪᔨᖅᑎᑦ ᑕᒪᓐᓇ

ᖃᐅᔨᓴᓗᒃ

- ᐊᐱᖅᑯᑎᓴᒃᑲᐃᓕᓐᓂᖅ?



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